

# Lepidoptera are Relevant Bioindicators of Passive Regeneration in Tropical Dry Forests

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**Abstract:** Most evaluations of passive regeneration/natural succession or restoration have dealt with tropical rain forest or temperate ecosystems. Very few studies have examined the regeneration of tropical dry forests (TDF), one of the most damaged ecosystem types in the world. Owing to their species diversity and abundance, insects have been widely used as bioindicators of restoration. Butterflies were among the most abundant and useful groups. We sampled four sites with different levels of anthropogenic disturbance in a Mexican TDF (Morelos State) and compared butterfly communities. A first goal was to examine whether adult butterflies were significant bioindicators owing to their specificity to restricted habitats. A second aim was to determine if differences exist in butterfly communities between some fields abandoned from 4–8, 8–15 and 15–30 years and a reference zone considered as primary forest. We found 40% to 50% of the species of butterflies were specifically related to a habitat and/or a level of anthropogenic disturbance. The time it takes for passive regeneration and recovery of the Mexican tropical dry forest is much higher than 25 years (our older zone), considering that almost none of the butterflies found in our conserved reference zone were present in our 25 year aged study zone.

**Keywords:** dry broadleaf forest; Mexico; butterflies; bioindicators; abandoned fields; Sierra de Huautla UNESCO Biosphere Reserve

## 1. Introduction

While they only represent 7.7% of the earth's land surface [1], tropical dry forests (TDFs) are home to exceptional biological diversity [2–5]. About 54% of all TDFs are located in the Americas. In Mexico, for example, TDFs (also called tropophilous forest [6], Figure A1) contain a wealth of unique biodiversity [7,8] and one of the greatest levels of endemism for plants (40%) when compared to the rainforest (5%) [9–11]. Yet conservation efforts remain directed toward rainforests [10,12]. The mean annual temperature of TDFs is higher than 17 °C, and the rainfall usually ranges from 250 to 2000 mm [13]. The main particularity of the TDF is the contrast between rainy season (May to November) which allow the development of luxurious vegetation, and dry season (November to May) when the vegetation becomes dry and less attractive [14]. This characteristic certainly leads to the lack of interest from conservationists to create reserves. Hence, only a small fraction (9%) of all TDFs remains intact [1,7,15–18] and few conservation efforts are made with the exception of Costa Rica (Guanacaste) and three main spots in Mexico (Chamela, Sierra Gorda and The Sierra de Huautla). As a result, TDFs are among the most threatened types of tropical forests [8,19].

The main cause of the decline of TDF cover is the conversion to agriculture or extensive cattle ranching [7,8,20,21] which generate degraded patches [7]. The resulting deforestation diminishes the ecological functions of the forests and adversely affects the associated biodiversity [19,22]. As a result of these risks and the negative perceptions of dry forests which made them unattractive [8,16], there is an urgent need to support conservation efforts [23–25]. There is, however, a lack of detailed quantitative studies and much of the data required for conservation planning are incomplete or absent [10,26,27].

Considering the loss of primary forest, some restoration programs have been set up to rehabilitate degraded areas by natural regeneration not directed by humans [28,29]. This method of passive conservation has proven to be successful with limited costs [30]. According to one study, the natural regeneration of a rainforest in Puerto Rico after pastures and coffee plantations were been abandoned, produced a secondary forest after 30 to 40 years [31]. This resulted in biodiversity similar to that of the island's mature undisturbed forest. However, it was not possible to accurately match the species composition with that of the primary forest, which means some ecosystem functions might not have been restored. Studies of passive regeneration of TDFs are comparatively scarce [32–34]. In particular, there is a lack of reliable indicators of the successional stages and ecological health of the systems during regeneration.

The dynamics of plant communities when they are subjected to environmental changes is slower than the response of their pollinators [35]. Many insects have already been used as bioindicators, including beetles, ants, grasshoppers, flies, hoverflies and butterflies [36–41]. Studying the changes in species richness of well-studied groups such as butterflies allows for characterization of the dynamics and disturbance levels of biodiversity in temperate and tropical areas [14,42–46]. Lepidoptera species are highly sensitive to climate change and environmental factors such as temperature or precipitation range [47]. They also have the advantage of being considered “flagship taxa” that capture the attention of environmental managers [48,49]. In tropical forests, species richness decreases with increasing habitat degradation [50]. Within a given genus, some species have a high mutual exclusion rate and are reduced to very specific ecological niches [41]. Around 1825 species of butterflies exist in Mexico, and approximately 500 of them have been recorded in Sierra de Huautla Biosphere Reserve (REBIOSH), a TDF protected area in Morelos, Mexico [43]. In March 1993, the Morelos State declared that most of the southern part of the state was subject to an ecological conservation plan (first as a state reserve then in 1999 with the Mexican biosphere status), mainly focused on intact or near-intact TDFs. Since 2006, this protected area is a UNESCO Biosphere Reserve (Figure A2). The history of soil management around the villages of the REBIOSH is well known. After the disappearance of extensive farms (*fincas*) following the Mexican revolution led by Emiliano Zapata (1910), former farm workers slowly restarted subsistence farming. The population gradually declined since the 1970's due to emigration, mainly to the USA [51]. One of the consequences was that many agricultural fields were abandoned, resulting in a spontaneous succession of plants and animals.

The aim of this work was to evaluate whether butterfly communities could be used as ecological markers of a dry tropical forest regeneration following abandonment of cultivation.

In order to answer our question, we sampled butterfly species in four different sites representing four different successional stages (i.e., sites differing in the number of years since cultivation was abandoned). We examined whether the dynamics of the butterfly community across successional stages denoted a return to the reference composition represented by the most preserved zone. The species richness and the species composition during the rainy and dry seasons were evaluated. The use and the limits of adult Lepidoptera as bioindicators for studying dry tropical forest dynamics was further discussed.

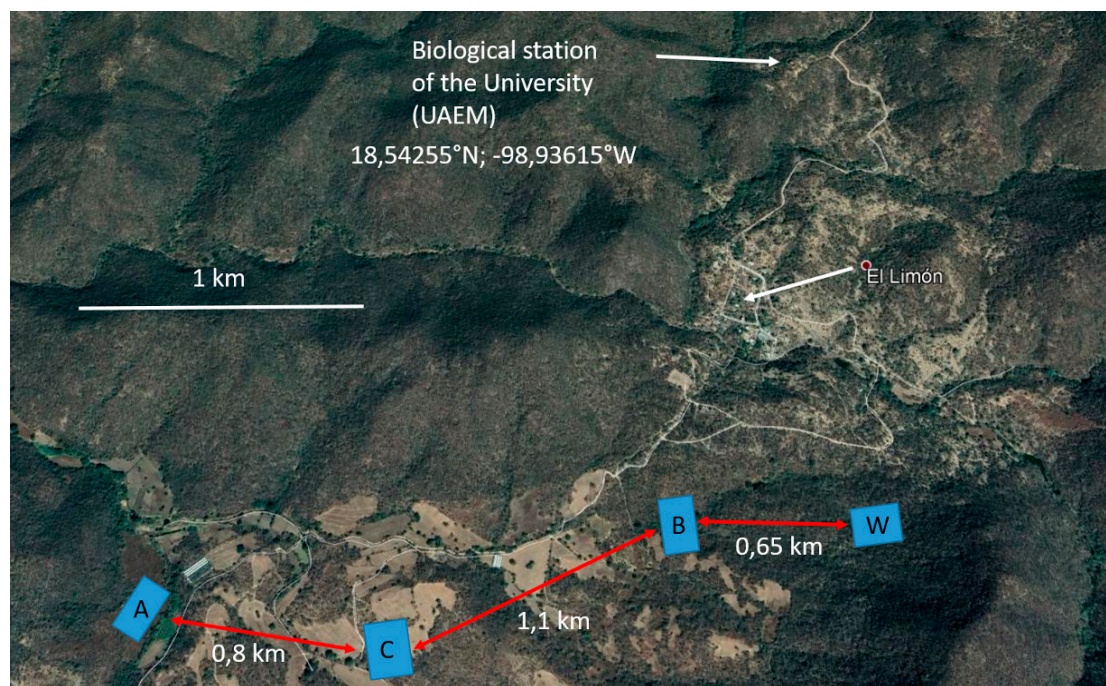
## 2. Materials and Methods

### 2.1. Study Sites

The four study sites were located in the REBIOSH, close to the village and the biological station of El Limon (Figure 1). All four sites were selected on the basis that they shared similar environmental characteristics: elevation, slope orientation and gradient, geology and pedology. Moreover, their proximity ensured that they were subject to the same precipitation patterns. Each site was characterized by a specific time period of non-exploitation for agriculture. Zone A has not been exploited for 5–6 years (knowing that our sampling was done in two consecutive years), Zone B for 11–14 years (with a small range of uncertainty) and Zone C for 25–26 years. The history of human exploitation of fields in this region is difficult to trace. This is why, in addition to the practical difficulty of collecting insects, it was not possible to perform replications on plots with equivalent orientation, altitude and humidity conditions between them.

The Wild Zone (W) was our reference of primary (undisturbed) TDF (Figure 1). Zone A is an open field with only herbaceous vegetation; Zone B is characterized by small shrubs; Zone C is similar to Zone B, but the percentage of tree cover is higher. The size of each sample plot (forest patch) is 2 ha. We did not sample close to the edges to avoid possible biases due to ecotone vegetation.

From early January until the end of May, the trees are leafless, and the herbaceous vegetation cover is low. From June to the end of December, the vegetation is luxurious (see [52] for pictures). Therefore, we separated our sampling into two sets, namely dry season and rainy season, based on the vegetation cover.



**Figure 1.** Localities of sampling (northeast of the Biosphere reserve of the Sierra de Huautla, Morelos State, Mexico (see Figure A2)). Reference: biological station of “El Limon” 18°32′33.12″ N; 98°56′10.22″ W (Source: INEGI/CNES 2020). Zone A has not been exploited for 5–6 years; Zone B for 11–15 years; Zone C for 25–26 years; and Zone Wild (W) the reference, representing primary dry tropical forest. Each sampling zone (in blue) is a 2 ha plot, approximately 180 × 110 m.

### 2.2. Butterfly Sampling

Our sampling covered two successive years in order to have a replicate for both the rainy and dry season. The butterflies were collected with a standard Lepidoptera hand net (handle 1.5 m long, net diameter 40 cm) for one hour, one day per month and per zone. To limit data collection bias, only

one of us (LL) performed all the sampling. Some species were collected a single time and were excluded from analysis (of the 119 species collected, 93 were included in the analyses). The two years of sampling were compiled after performing preliminary analyses and the average values used for further analysis.

### 2.3. Data Analyses

#### 2.3.1. Diversity Indices

We calculated Shannon's entropy and Simpson's evenness indices every month at each site to evaluate and compare spatial and temporal changes in diversity. The Chao1 index (calculated using EstimateS© 8.20 software) was used to estimate the species richness [53]. The  $\alpha$ -diversity was recorded at each site and compared to Chao1 estimates.

We estimated overall beta diversity ( $\beta_{cc}$ ) partitioned by its components, species richness differences ( $\beta_{rich}$ ) and species replacement ( $\beta - 3$ ) [54,55].  $\beta_{cc}$  represents the proportional effective species turnover. Interestingly, these components are additive (i.e.,  $\beta_{cc} = \beta_{rich} + \beta - 3$ ), as previously demonstrated [55] and reflect the breakdown of the constituents of  $\beta$ -diversity in the loss or gain and the replacement of species between communities.

#### 2.3.2. Species Distribution Patterns

Ecological data such as organism counts often vary and co-vary in a nonlinear fashion. Nonlinear modeling methods such as artificial neural networks (ANNs) were therefore deemed relevant to analyze shifts in species distributions among our study sites (A, B, C and W), in relation to humidity, drought or vegetation cover. Specifically, we used the self-organizing map algorithm (SOM; [56]), which combines ordination, clustering and gradient analysis.

The network consisted of 28 independent samples (4 sites  $\times$  7 months) in the rainy season, 20 in the dry season (4 sites  $\times$  5 months), characterized by 93 butterfly species (descriptors) for the rainy season and 42 species for the dry season. A vector that includes all descriptors represents each sample, and there are as many vectors as samples. The input layer is comprised of  $p$  nodes, or neurons. The output layer forms a rectangular two-dimensional map with neurons laid out over a hexagonal lattice. Each neuron of the output layer is also called a "cell" (visualized as a hexagon) and is linked to the neurons of the input layer by connections that have weights associated with them, forming a vector.

These weights represent the virtual values for each descriptor in every output neuron such that every cell in the output layer stores a virtual vector of connection weight. These virtual vectors characterize the coordinates of the centers from groups of similar input vectors, where similarity is measured in terms of Euclidean distance for all neurons. The aim of the SOM is to organize the distribution of sample vectors in a two-dimensional space, using their relationship to the virtual-vector distribution. Similar input vectors are allocated to the same virtual vector and it changes with the addition of new input vectors. The virtual vectors that are neighbors on the map (neighboring neurons) are expected to represent neighboring groups of sample vectors. Consequently, sample vectors that are dissimilar are likely to be distant from each other on the map. The SOM units (hexagons) were divided into clusters according to the weight vectors of the neurons, and clusters were justified according to the lowest Davis–Bouldin index, i.e., for a solution with low variance within clusters and high variance between clusters [57].

#### 2.3.3. Cladistic Analysis

A matrix was created based on presence/absence and relative abundances for every month and zone. Different patterns are visible, which allowed us to qualify the specificity of each species sampled. We coded the observed degree of openness of the living environment for each species as follows: "open," A; "shrubs," B and C; "trees," W. Humidity conditions were divided into the three following categories: xeric (\*), semi-xeric or moderately humid (\*\*) and humid (\*\*\*). We noted 0: absence; 1: one isolated individual; 2: two to seven individuals; 3: 8 to 15 individuals and 4: more

than 15 individuals. The data set was split in two: rain season and dry season, as mentioned above. Only the rain season is presented here (most of the dry season's butterflies are generalist species). A cladogram was generated using PAUP\* 4.0b10 software [58], with all characters organized (Wagner) in order to use the abundance data for each species. Starting trees were obtained by stepwise addition. All heuristic searches for optimal trees were carried out by TBR (tree bisection and reconnection) branch swapping.

We did not attempt to conduct a phylogenetic classification. Thus, we decided not to include any outgroup in our data set so that the tree was unrooted. As some multiple subequal trees were found, a majority rule (50%) consensus tree was calculated.

Considering the presence per sites, we checked if species from the same zone were represented together. Specialized species from one zone were expected to be grouped in the same clade, contrary to the habitat and/or distribution generalists (the term generalist in this work does not include host-plant generalists, rather it refers to the specificity for a type of habitat within an ecosystem).

In order to test our hypothesis, a comparison was made between the theoretical habitat of each species and the results obtained using the SOM approach. When congruence was found between the cladistics approach and SOM analysis, we annotated Y, and when congruence was not found, we annotated N. When a species was considered a generalist for one approach (theoretical or observed) and considered a specialist for the other, we noted 0. When its theoretical habitat tended to be like the SOM result but there was no 100% congruence, we termed it tendency. A bioindicator in a strict sense is a species found in one type of habitat and when the congruence is “Y”. The two sites (B and C) have been considered together.

### 3. Results

We found 119 species from five families (Papilionidae, Pieridae, Nymphalidae, Lycaenidae (including Riodininae) and Hesperidae), (Table A1). A total of 5122 individuals were collected and identified. Individuals belonging to *Codatractus melon* and *Codatractus sallye* (Hesperidae, Eudaminae) were pooled due to uncertain identification. Twenty-six rare species (less than 10 individuals in two years) were removed from the dataset prior to analysis to avoid biases due to the low number of individuals collected. We thus retained 93 species. The most common species was *Eurema daira* (556 individuals in total), followed by *Pyrisitia proterpia* and *Nathalis iole* (229 and 160 individuals, respectively). All three species belong to Pieridae, Coliadinae.

#### 3.1. Diversity Indices

In the rainy season, Shannon's entropy almost doubled compared to the dry season (dry season: 1.51, rainy season: 2.71) (Table 1). Shannon index values did not differ significantly among sites for both seasons. However, higher diversity was always found in August in the four zones. Simpson's evenness differed between the dry and rainy season (respectively average/season 0.66 and 0.88) but not within seasons for the four zones.

On average, the  $\alpha$ -diversity (28.5 versus 9.4) and Chao values (33.9 versus 10.4) were about three times higher during the rainy season than the dry season (Table 2). Zone A during the rainy season showed greater values for overall diversity. For both seasons, the shrubby zones (B and C) were the least diverse.

**Table 1.** Various diversity indices by months (first set is the dry season, second set is the rain season) and by sites (A, B, C and W). Highlighted in gray are the highest values per season for each site.

	Shannon (H)	Simpson (1-D)	H Mean	(1-D) Mean	Chao1	Div Alpha
Dry Season						
Zone A (open)			1.44	0.58	12.27	11.8
January	1.26	0.50				
February	0.86	0.35				
March	1.51	0.64				
April	2.17	0.78				
May	1.39	0.61				
Zone B (shrubs)			1.55	0.73	8	6.8
January	1.84	0.79				
February	1.52	0.74				
March	1.43	0.72				
April	1.58	0.74				
May	1.40	0.68				
Zone C (shrubs)			1.55	0.69	8.17	7.4
January	1.11	0.47				
February	1.63	0.75				
March	1.81	0.82				
April	1.70	0.78				
May	1.48	0.65				
Zone W (forest)			1.51	0.65	13	11.6
January	0.88	0.40				
February	1.13	0.55				
March	1.60	0.72				
April	1.70	0.73				
May	2.24	0.86				
Rainy Season						
Zone A (open)			2.88	0.89	44.53	34.57
June	2.87	0.93				
July	3.13	0.91				
August	3.36	0.94				
September	3.03	0.89				
October	2.98	0.90				
November	2.62	0.85				
December	2.19	0.83				
Zone B (shrubs)			2.64	0.88	28.89	25.86
June	2.47	0.87				
July	3.08	0.94				
August	3.19	0.94				
September	2.78	0.92				
October	2.78	0.89				
November	2.27	0.84				
December	1.70	0.74				
Zone C (shrubs)			2.49	0.83	30.20	25.43
June	1.70	0.70				
July	2.85	0.91				
August	3.13	0.93				
September	2.99	0.93				
October	2.80	0.91				
November	2.82	0.90				



December	1.15	0.52				
Zone W (forest)			2.83	0.91	31.90	28.14
June	2.52	0.90				
July	2.90	0.91				
August	3.21	0.94				
September	3.16	0.95				
October	3.04	0.92				
November	2.96	0.93				
December	2.01	0.80				

When comparing  $\beta$ -diversity values (overall  $\beta$ -diversity:  $\beta_{cc} = \beta_{rich} + \beta - 3$ , average values from Table 2, see also Material and Methods section), we observed that between seasons, the greater component (in bold) was the species richness differences ( $0.71 = 0.63 + 0.08$ ) while between sites (for dry  $0.62 = 0.23 + 0.39$  and rainy season  $0.45 = 0.08 + 0.37$ ) the greater component (in bold) was the species replacement.

**Table 2.**  $\beta$ -diversity values between sites (ZA to ZW: Zone A, B,C and W; see Material and Methods/Study sites for more details) and between seasons (see Material and Methods for calculations). Above diagonal calculated values [ $\beta_{cc} (\beta_{rich} + \beta - 3)$ ], below diagonal, “a,b,c” values (see Material and Methods); a is the number of species common to both sites, b is the number of species exclusive to the first site, and c is the number of species exclusive to the second site.

	ZAsec	ZArain	ZBdry	ZBrain	ZCdry	ZCrain	ZWdry	ZWrain
<b>ZAdry</b>		0.59 (0.56 + 0.03)	0.67 (0.24 + 0.43)		0.66 (0.41 + 0.25)		0.64 (0.13 + 0.51)	
<b>ZArain</b>	28;1;40			0.43 (0.05 + 0.38)		0.43 (0.14 + 0.29)		0.59 (0.10 + 0.49)
<b>ZBdry</b>	11;15;7			0.79 (0.67 + 0.12)	0.5 (0.14 + 0.36)		0.59 (0.17 + 0.42)	
<b>ZBrain</b>		47;20;16	14;4;49			0.36 (0.09 + 0.27)		0.43 (0.08 + 0.35)
<b>ZCdry</b>	11;17;4		11;7;4			0.78 (0.72 + 0.06)	0.69 (0.29 + 0.40)	
<b>ZCrain</b>		44;22;11		47;17;10	13;2;45			0.48 (0.03 + 0.45)
<b>ZWdry</b>	14;15;10		12;6;11		10;5;13			0.67 (0.57 + 0.10)
<b>ZWrain</b>		36;30;21		43;19;13		39;19;17	20;3;37	

### 3.2. Species Distribution Patterns

#### 3.2.1. Rainy Season

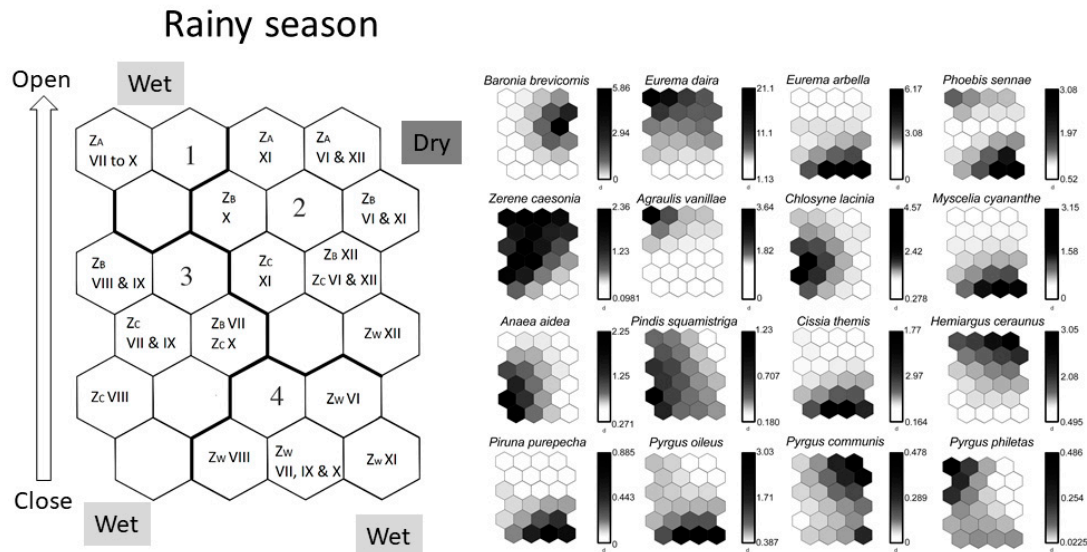
The sites/dates were classified into four subsets, forming distinct clusters (Figure 2). Cluster 1 corresponded to Zone A from July to October (humid and the most open locations), Cluster 4 represented Zone W from June to November (humid and closed areas), Cluster 2 was composed of Zone W in December, Zones A,B and C in June, November, and December and Zone B in October, illustrating rather dry periods in shrubby areas. Cluster 3 showed a mixture of Zone B from July to September and Zone C from July to October (that is, the peak of rainy season in shrubby areas).

The SOM map showed the strength of the association of each species with a type of habitat/date along two gradients. The vertical gradient on the SOM was related to openness (lack of cover) of the location and the horizontal gradient portrayed the humidity level of the location (Figure 2).

Figure 2 showed a selection of species representing most distribution patterns encountered in our sampling, including very common generalist species such as *Zerene cesonia* (Coliadinae) but rare in preserved areas, regular generalist species with a preference for closed but humid locations such as *Pindis squamistriga* (Satyrinae), species specialized in one type of habitat (humid closed forests) such as *Eurema arbella* (Coliadinae), or in contrast, *Agraulis vanillae* (Heliconiinae) for open habitat. A single species (among all collected during this study) was specialized on dry shrub areas (*Baronia*

*brevicornis*, Baroniinae). We noted that the three species of the genus *Pyrgus* (Pyrginae), all generalists, each show a clear tendency for having their own specific habitat.

The highest probability of occurrence for a species during the rainy season (21.1; Figure 2) was for *Eurema दौरा* (Coliadinae), which showed a clear preference for open/dry locations during the rainy season. Note that among the genus *Pyrgus*, the rate of occurrence of the two species *Pyrgus communis* (0.478) and *Pyrgus philetas* (0.486) was much lower than for the third species, *Pyrgus oileus* (3.03).



**Figure 2.** Probabilities of occurrence and distribution (SOM) of sampled Lepidoptera for rainy season. Left side: Kohonen (SOM) map for the four studied sites (2 years of sampling). Four clusters can be distinguished. Cluster 1: humid + open field; Cluster 3: humid + shrubs; Cluster 4: humid + conserved forest; Cluster 2: driest months of the rain season independent of type of habitat. Roman numerals are months, from I = January to XII = December). Zones A to W are noted Za to Zw. Right side: Gradient analysis of the abundance of Lepidoptera on the trained SOM represented by a shaded scale (dark = high abundance, light = low abundance). Each small map representing species can be compared to (or superimposed on) the map of Kohonen, thus showing the distribution patterns of the various taxa (in shades of gray) within each subarea of the SOM. Generalist species are indicated when SOM patterns are almost black. On the right of each map, the relative scale of occurrence probabilities is shown.

### 3.2.2. Dry Season

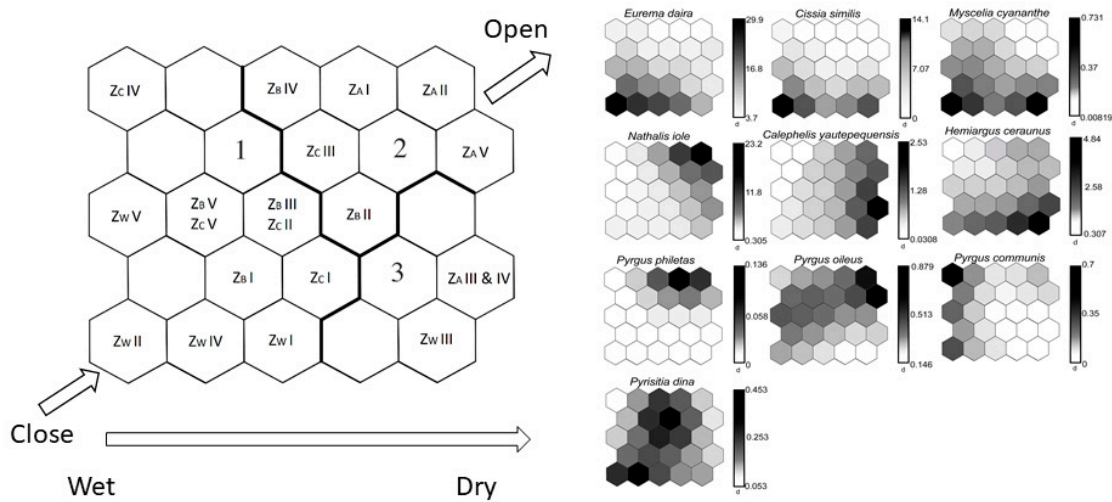
The sites/dates were classified into three clusters (Figure 3). These clusters were less defined than those of the rainy season. Cluster 2 was composed of the driest and open locations, mainly corresponding to Zone A. Clusters 1 and 3 were linked. Cluster 1 was composed of the most humid and closed locations of the zone W, while Cluster 3 consisted of the two driest/hottest locations/dates (Zone A, March and April, often above 40 °C).

A horizontal gradient of humidity was rather clear on the SOM. There was also a gradient of openness (from close to open areas) from lower-left to upper-right in the Kohonen map. The two Coliadinae, *Eurema दौरा* and *Nathalis iole*, dominated the selection of species presented here. Interestingly, the first species was, for the dry season, specialized in closed humid areas while it occurred in open field during the rainy season. The second species, *N. iole*, was more frequent in open areas during the dry season, as were some species such as *Calephelis yautepequensis*.

Considering the *Pyrgus* genus, *Pyrgus oileus* was shown to be a rather generalist species but the two other *Pyrgus* species were more specialized, especially *Pyrgus communis* which had a higher probability of occurrence during the dry season (dry: 0.7; rainy: 0.48). Some species showed a similar probability of occurrence independent of the season, e.g., *E. दौरा* (dry: 29.9; rainy: 21.1) and *Hemiargus*



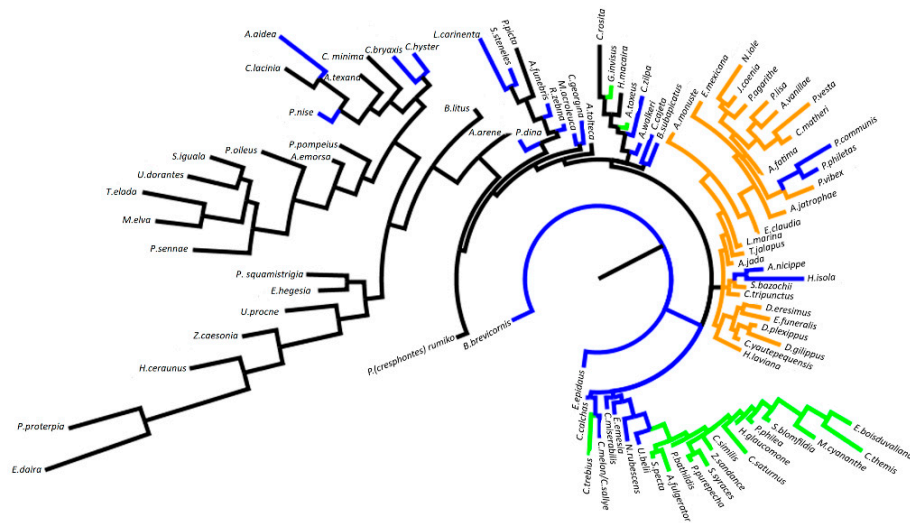
*ceraunus* (dry: 4.84; rainy: 3.66), while most species, such as *Myscelia cyananthe* (dry: 0.73; rain: 3.15), were more scarce during dry the season.



**Figure 3.** Probabilities of occurrence and distribution (SOM) of sampled Lepidoptera for dry season. Left side: Kohonen (SOM) map for the four studied sites (2 years of sampling; roman numerals are months, Zones A to W are noted Za to Zw). Horizontal gradient from left to right side denotes from humid to dry. Diagonal from left bottom corner to up right is from closed to open habitat. See Figure 2 for more information.

### 3.3. Cladistic Analysis

Generalist species (black lines, Figure 4) were all at the end of the spiral tree, leading us to assume that the most specialized species were placed in the center of the spiral. *Baronia brevicornis*, taking into account both flight periods and locations, was found to be the most specialized species of our sampling, while *Eurema daira*, a common species from USA to Argentina flying all year round, was our most generalist species. A first group of species corresponding to shrub specialists was at the center of the spiral (blue lines), followed by species specialized in primary forest (green lines). Open field specialist species were grouped (orange lines), followed by some less specialized shrub species and then by the generalist species (black lines). As mentioned above, most of these generalist species were also those with wider overall distribution in the Americas.



**Figure 4.** Cladistic tree based on presence/absence and relative abundances for every month and zone, and the degree of openness of the living environment for each species and humidity conditions. A pattern code was attributed to each species depending on the type of vegetation. Green is for species found only or preferably (more than 60%) in the forest (Zone W); blue is for species found only or preferably in zones covered by shrubs (Zones B and C); orange is for species found only or preferably in open areas (Zone A); black is for species which can be found in two or more of these categories with none exceeding 60% (generalist species). Three classes of characters—temporality, habitat specificity and abundance in each zone—were used to build this figure. We assumed that species placed on the center of the spiral are the most specialized (*Baronia brevicornis*) while those on the periphery (*Eurema daira* and *Pyrisitia proterpia*) are the most generalist species.

### 3.4. Comparison of the Two Analyses

For the rainy season (Table A2), 5.4% (5 out of 93) of the butterflies, were not congruent between theoretical cladistics and SOM analyses of occurrences. This percentage was 16.7% (7 species out of 42) for dry season (Table A3).

During the rainy season (Table A2), 46.24% of the species (43 out of 93) tended to occupy only one type of habitat and could therefore be considered relevant bioindicators. Zone A had 7 out of 13 specific species (54%), Zones B and C had 4 out of 15 (27%) and Zone W had 6 out of 15 (40%). For the dry season (Table A3), we found 26% species (11 out of 42) in just one type of habitat, and only a single bioindicator of Zone A (1 out of 9, 11%) and W (1 out of 2, 50%).

## 4. Discussion

The main aim of our study was to examine the potential of butterfly communities as bioindicators to assess different states of TDF regeneration. Specifically, we expected the community of Lepidoptera after 20–30 years without anthropogenic action to converge towards that of the control primary TDF (Zone W). Given that adults are usually easier to identify than caterpillars, students in biology and environmental managers of different sites could easily use butterflies to routinely evaluate restoration states. This approach has some limitations, especially when the number of insects is low. However, a study of North American butterflies, with a sampling limited to around 30 species, [59,60] showed that a general linear model allowed samples of less than 10% of the total number of species to explain 77–88% of the variation in the species richness at a local scale.

#### 4.1. Specificity of Lepidoptera Habitat

Some Lepidoptera species have been shown to possess high habitat specificity [35,54,61] and can be used as bioindicators [40]. Checking the congruence between the theoretical occurrence and the crude results of presence (SOM analysis), we obtained (regardless of the season) around 49% of Lepidoptera species specific for a type of habitat. In a very distinct context, but to illustrate the habitat specificity of imagoes, a former study performed in the Mediterranean eastern part of the Pyrenees (France) found that 36% of species were considered to be relevant [41].

During the rainy season, species with the highest level of specificity occurred in open fields (Zone A): *Eurema दौरa* [Pieridae], *Danaus eresimus*, *Danaus plexippus*, *Danaus gilippus*, *Euptoieta claudia*, *Anartia fatima* and *Anartia jatrophae* (Nymphalidae). Six species were specific for the reference forest (Zone W), namely *Hamadryas glauconone* (Nymphalidae), *Astraptes fulgerator*, *Heliopetes laviana*, *Synapte syraeces*, *Synapte pecta* and *Piruna purepecha* (Hesperiidae). Finally, the species specific for shrubby areas (Zone B and C) were *Baronia brevicornis* (Papilionidae), *Siproeta stelenes* (Nymphalidae), *Emesia emesia*, and *Cyanophrys miserabilis* (Lycaenidae).

For the dry season, the main indicator species in open-field areas was *Nathalis iole* and for forest, it was *Eurema दौरa* (Pieridae). Concerning this last species, it is very interesting to note that its classification as a full generalist species has to be drastically modified as we realized that it is a specialized taxon depending on season (open field for rain season/conserved-closed forest for dry season). A seasonal migration occurs which was already observed and suggested by some authors [43,62,63].

Some trends in family characteristics appear to be present. Pieridae were more often observed in open habitats including semi-desert and arid grasslands [64] and many Nymphalidae were seen in dry or mesophilic areas basking with their wings wide open [65]. Hesperiidae are considered indicators of the richness and uniqueness of a habitat [66]. Papilionidae were commonly seen in grasslands, plains and hills (even though some, such as Troidini, are specialized for humid forest) [65]. No clear trends were found for Lycaenidae, but we did notice that most of the Theclinae were good markers of conserved forests, while Polyommatainae were mostly from open fields. The subfamily Riodininae (considered to be a distinct family for many authors) was much more diverse with species specialized in various environments.

#### 4.2. Lepidoptera as Bioindicators

Our study showed that Lepidoptera were sensitive to temporal variations in humidity, which was especially evident from the difference between the two seasons, and they had a differential response to the environmental changes. For the data analyses, it was necessary to separate the two seasons of dry tropical forest [7,49]. Our monthly sampling during two successive years took into account the seasonal variation in the density of butterflies which was more relevant in the rainy season. Considering that our four study sites were geographically in close proximity, it is possible that some individuals migrated between them. This could explain some similarities between some of our calculated diversity indices.

The case of the genus *Pyrgus* was especially interesting. The three species seem to be present everywhere, mainly because they look very similar, especially when in flight, which makes identification difficult. However, it was rare to observe two of these species flying together. An important point is that the *Pyrgus* species feed on the same widespread and common plants (Malvaceae, *Sphaeralcea* sp. or *Sida* sp. [43,67]). These three species of the same genus can be compared to those of the European *Pyronia* sp. for which, in a former study [41], we found a very clear specificity of habitat depending on species sharing the same host plant. This illustrates a competition/exclusion process between close species of the same genus.

#### 4.3. Diversity of the Successional Stages

The estimation of species richness using the Chao1 index (Table 2) was congruent with the observed richness for the dry season. For the rainy season, a slight but non-significant undervaluation

of the sampling, when compared to the Chao1 index and  $\alpha$ -diversity, was noticed (especially for Zone A).

The native fauna of the TDF (mainly of Zone W) disappeared or at least were disturbed and a new, so-called "anthropomorphic" or "synanthropic" combination of species replaced the local and specialized fauna. This phenomenon was already described in various studies of butterflies or other arthropods [41,47,68–71]. It was already noted that, during the passive regeneration of TDFs after human activities ceased, a low diversity of plants with very high densities was observed for early successional stages (Zone A) and that for the most developed stages (Zones B, C and W), a larger diversity of plant families and species occurred but with much lower densities [68,72].

Zone A was unexpectedly diverse, which could be explained by the occurrence of two types of species, the true specialists of open fields and generalist species. Most generalist species prefer open fields created by human activities, which are areas more easily colonized by less specialized species [44–46,50]. The transitional state, i.e., the shrub environment (Zones B and C), exhibited a higher proportion of habitat specialists than other sites (73.81% in dry season and 53.76% in rainy season). Most of the species found in the conserved forest (Zone W) were absent from all the other zones demonstrating a high specificity and the important role of the habitat characteristics in terms of herbaceous layer in controlling Lepidoptera community assemblage [73]. A main finding of this study is that almost no species from Zone W were present in the most mature (25 years) zone of our sampling.

When comparing  $\beta$ -diversity values, a decrease of observable imagos was the main factor of change between the rainy (greater number of visible species) and dry season (lower number) while replacement of species between communities was the main factor of change observed between studied sites independent of the season.

#### 4.4. Passive Regeneration of the Dry Tropical Forest

Ecosystem resilience [74] is the ability to return to a stable state following disturbance. Monitoring the effects of restoration on the ecosystem can easily be performed through insects because their populations depend on the plants that grow in that zone [36,37,39]. In our study, even though no replication of zones within age classes across the landscape were possible to perform, 25 years of passive restoration was not sufficient to reach a totally restored TDF, comparable to the reference Zone W. Resilience (and passive regeneration) of the dry tropical forest seemed to be extremely variable depending on local conditions. Madagascar has shown a fast regeneration capacity (around 10 years) [75], while other studies have found that a minimum of 25 years (but mostly up to 50 years) was necessary [76–78]. Recently, it was considered that the time of plant community recovery in terms of biomass, species richness and structural composition ranges from 30 to 50 years [79], but when precipitation is less than 1000 mm per year, this time range may increase to 80 years [27,80]. In our study, a single butterfly species (*Myscelia cyananthe*) was found to be equally common in Zones C and W, showing a possible starting point to original climax. However, in the specific case of the Mexican TDF, we observed that the stable and natural vegetation state can be similar to the one observed in Zone C. Such breakpoints during succession result in almost pure acacia settlements (locally called "cubateras") and represent the biotope for ultra-specialized species such as the Mexican endemic *Baronia brevicornis* [14,81].

Up to now, such breakpoints in succession have not been clearly explained, but they are certainly caused by several factors such as the local soil pedology, geology, topography, availability of water and possibly allelopathic effects due to secondary metabolites produced by the roots of Mimosoideae. The slow passive regeneration of TDF compared to humid tropical forests [24,82] may be due to its high average temperature and low precipitation. As previously mentioned, with only 2% of intact forest and 8% legally protected, the TDF is probably (with the cloud forest) the most endangered ecosystem in the world and has to be prioritized in terms of management and conservation effort [8,10,26,79]. Lepidoptera showed sufficient resilience which should be considered as a good marker of passive regeneration (and restoration) in the context of the Mexican tropical dry forest, which was also found in temperate areas [41,44–47]. Despite the limitations inherent to biodiversity records in

remote locations, our dataset is substantial with more than 5000 butterflies counted. Even though our study needs more replicates to secure our conclusions, it seems that plants, which are a classical bioindicator, are following a similar succession and recovery rate in the Mexican TDFs (Dorado and Jesus-Almonte, unpublished results). A similar observation applies to some Coleoptera families (Toledo-Hernandez and Corona-Lopez for Cerambycidae, unpublished results). In future work, a meta-analysis of various group of easily observable organisms could therefore be relevant to demonstrate convergent patterns among coexisting taxa (including the Lepidoptera considered in our work) with the aim to bring out efficient surrogate taxa for the survey of TDF status and recovery.

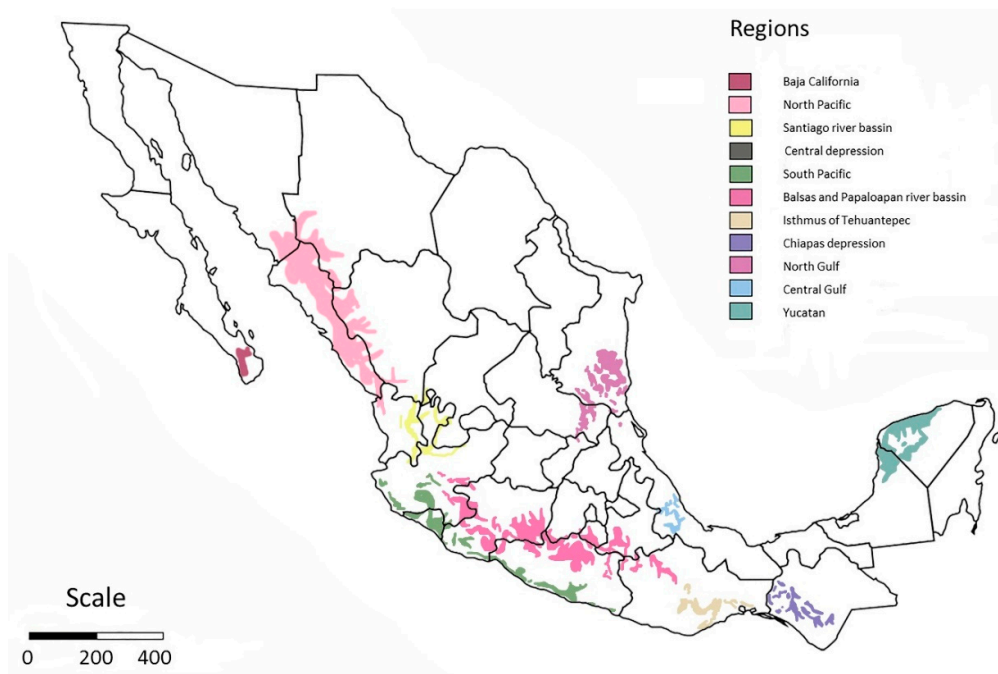
**Author Contributions:** LL and OD conceived the research; LL and JMJ performed field work; LL, MV, KL, RC analyzed the data; LL, MV, KL, RC wrote and edited the manuscript. All authors have read and agreed to the published version of the manuscript.

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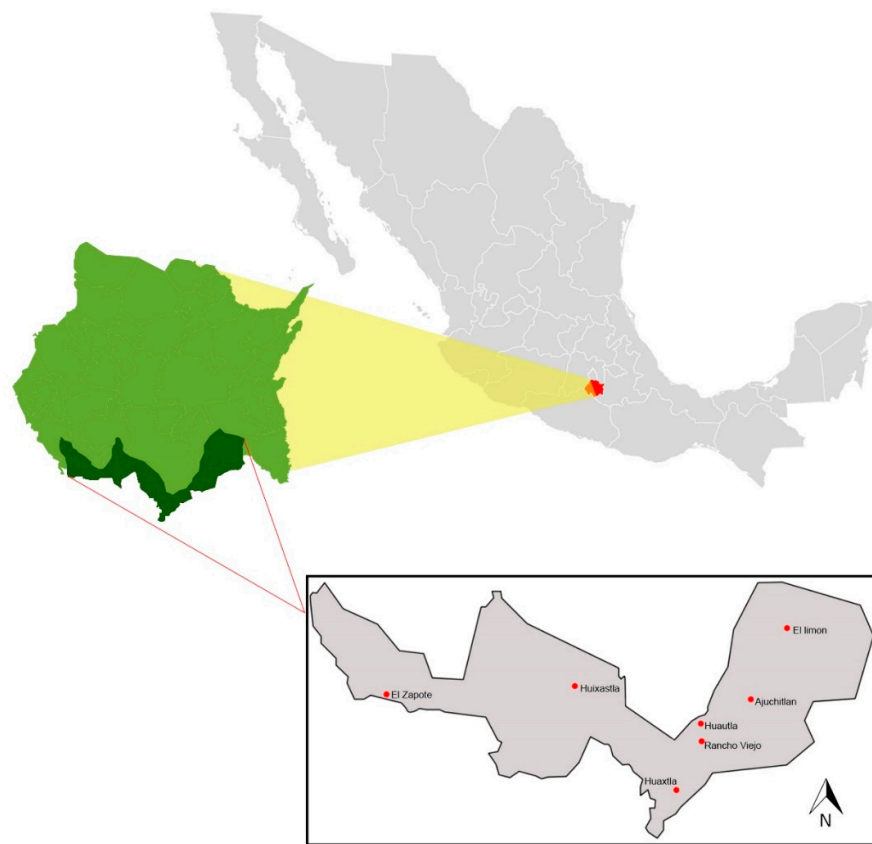
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**Conflicts of Interest:** One of us (LL) is section editor in chief of this journal.

## Appendix A



**Figure A1.** Main subdivisions of the tropical dry forest of Mexico according to [26]. Scale in km.



**Figure A2.** Location of REBIOSH in Mexico (Biosphere reserve of the Sierra de Huautla, Morelos State).

**Table A1.** Complete list of Lepidoptera collected during this study, including their distributions. Note that descriptions correspond to the local subspecies [64] including the name change of *Papilio cresphontes* for *Papilio rumiko* (see description of this species). Highlighted in gray are the species with only a few individuals sampled, which were excluded from statistical analyses.

Family	Subfamily	Genus	Species	Description	Distribution
Papilionidae	Baroniinae	<i>Baronia</i>	<i>brevicornis</i>	Salvin, 1893	W Mexico
	Papilioninae				
	Troidini	<i>Parides</i>	<i>montezuma</i>	Westwood, 1844	Mexico to Costa Rica
		<i>Parides</i>	<i>photinus</i>	Doubleday, 1844	Mexico to Costa Rica
	Leptocircini	<i>Eurytides</i>	<i>epidaus</i>	Salvin and Godman, 1868	W. Mexico
	Papilionini	<i>Papilio</i>	<i>rumiko</i>	Shiraiwa and Grishin, 2014	S USA to Panama
		<i>Papilio</i>	<i>thoas</i>	Rothschild and Jordan, 1906	S USA to Panama
<b>6 species</b>					
Pieridae	Coliadinae	<i>Abaeis</i>	<i>nicippe</i>	Cramer, 1779	S USA to Costa Rica
		<i>Eurema</i>	<i>boisduvaliana</i>	Felder and Felder, 1865	S USA to Costa Rica
		<i>Eurema</i>	<i>daira</i>	Felder, 1869	S USA to W Mexico
		<i>Eurema</i>	<i>mexicana</i>	Boisduval, 1836	S USA to Panama
		<i>Nathalis</i>	<i>iole</i>	Boisduval, 1836	Canada to Honduras
		<i>Phoebis</i>	<i>agarithe</i>	Boisduval, 1836	S USA to Peru
		<i>Phoebis</i>	<i>philea</i>	Linnaeus, 1763	S USA to Brazil
		<i>Phoebis</i>	<i>sennae</i>	Cramer, 1777	S USA to S America



		<i>Pyrisitia</i>	<i>dina</i>	Boisduval, 1836	S USA to Panama
		<i>Pyrisitia</i>	<i>lisa</i>	Herrich-Schäffer, 1865	S USA to Panama
		<i>Pyrisitia</i>	<i>nise</i>	Felder, 1869	S USA to Panama
		<i>Pyrisitia</i>	<i>proterpia</i>	Fabricius, 1775	S USA to Peru
		<i>Zerene</i>	<i>cesonia</i>	Stoll, 1790	S USA to Colombia
	Pierinae	<i>Ascia</i>	<i>monuste</i>	Linnaeus, 1764	S USA to W Mexico
		<i>Melete</i>	<i>lycimnia</i>	Boisduval, 1836	S USA to Costa Rica
<b>15 species</b>					
Nymphalidae	Libytheinae	<i>Libytheana</i>	<i>carinenta</i>	Michener, 1943	W Mexico to Panama
	Danaidae				
	Danaini	<i>Danaus</i>	<i>eresimus</i>	Talbot, 1943	S USA to Columbia
		<i>Danaus</i>	<i>gillippus</i>	Bates, 1863	USA to Columbia
		<i>Danaus</i>	<i>plexippus</i>	Linnaeus, 1758	Canada to Mexico
	Heliconiinae				
	Heliconiini	<i>Agraulis</i>	<i>vanillae</i>	Riley, 1926	S USA to Panama
		<i>Dione</i>	<i>juno</i>	Reakirt, 1866	Mexico to Panama
		<i>Dione</i>	<i>moneta</i>	Butler, 1873	S USA to Panama
	Argynnini	<i>Euptoieta</i>	<i>claudia</i>	Cramer, 1775	USA to Panama
		<i>Euptoieta</i>	<i>hegesia</i>	Stichel, 1938	S USA to Argentina
	Biblidinae	<i>Hamadryas</i>	<i>amphinome</i>	Jenkins, 1983	W Mexico
		<i>Hamadryas</i>	<i>glauconome</i>	Bates, 1864	C Mexico to Costa Rica
		<i>Hamadryas</i>	<i>guatemalena</i>	Fruhstorfer, 1916	S USA to Mexico
		<i>Mestra</i>	<i>amymone</i>	Menestries, 1857	S USA to Costa Rica
		<i>Myscelia</i>	<i>cyananthe</i>	Felder and Felder, 1867	C Mexico
	Nymphalinae				
	Nymphalini	<i>Smyrna</i>	<i>blomflidia</i>	Fruhstorfer, 1908	S USA to Panama
	Victorinini	<i>Anartia</i>	<i>jatrophae</i>	Fruhstorfer, 1907	S USA to Panama
		<i>Anartia</i>	<i>fatima</i>	Lamas, 1995	W Mexico
		<i>Siproeta</i>	<i>stelenes</i>	Frustorfer, 1907	S USA to Brazil
	Junoniini	<i>Junonia</i>	<i>coenia</i>	Hübner, 1822	S USA to Guatemala
		<i>Junonia</i>	<i>evarete</i>	Barnes and McDunnough, 1916	S USA to Mexico
	Melitaeini	<i>Anthanassa</i>	<i>ardys</i>	Hewitson, 1864	C Mexico
		<i>Anthanassa</i>	<i>texana</i>	Edwards, 1863	S USA to Guatemala
		<i>Chlosyne</i>	<i>lacinia</i>	Geyer, 1837	C Mexico to Panama
		<i>Chlosyne</i>	<i>rosita</i>	Bauer, 1961	W Mexico
		<i>Microtia</i>	<i>elva</i>	Bates, 1864	S USA to Costa Rica
		<i>Phyciodes</i>	<i>graphica</i>	Felder, 1869	S USA to C Mexico
		<i>Phyciodes</i>	<i>pallenscens</i>	Felder, 1869	S USA to C Mexico
		<i>Texola</i>	<i>elada</i>	Hewitson, 1868	C Mexico
	Charaxinae				
	Anaeini	<i>Anaea</i>	<i>aidea</i>	Guerin-Meneville, 1844	S USA to Costa Rica
	Satyrinae	<i>Cissia</i>	<i>similis</i>	Butler, 1867	W Mexico to Columbia
		<i>Cissia</i>	<i>themis</i>	Butler, 1867	W Mexico to Nicaragua
		<i>Hermeuptychia</i>	<i>hermes</i>	Fabricius, 1775	S USA to Argentina
		<i>Pindis</i>	<i>squamistriga</i>	Felder, 1869	C Mexico to Guatemala
<b>33 species</b>					
Lycaenidae	Theclinae	<i>Arawacus</i>	<i>jada</i>	Hewitson, 1867	S USA to Coasta Rica
		<i>Calycopis</i>	<i>isobea</i>	Butler and Druce, 1872	C Mexico to Brazil

		<i>Cyanophrys</i>	<i>miserabilis</i>	Clench, 1946	C Mexico to Honduras
		<i>Lamprospilus</i>	<i>sethon</i>	Godman and Salvin, 1887	C Mexico to Costa Rica
		<i>Panthiades</i>	<i>bathildis</i>	Felder and Felder, 1865	C Mexico to Brazil
		<i>Parrhasius</i>	<i>moctezuma</i>	Clench, 1971	C Mexico
		<i>Rekoa</i>	<i>zebina</i>	Hewitson, 1869	C Mexico to Costa Rica
		<i>Strymon</i>	<i>bazochii</i>	Godart, 1824	S USA to Argentina
		<i>Strymon</i>	<i>bebrycia</i>	Hewitson, 1868	S USA to Costa Rica
	Polyommatainae	<i>Echinargus</i>	<i>isola</i>	Reakirt, 1867	S USA to Guatemala
		<i>Hemiargus</i>	<i>ceraunus</i>	Lucas, 1857	S USA to Panama
		<i>Leptotes</i>	<i>cassius</i>	Boisduval, 1870	S USA to Panama
		<i>Leptotes</i>	<i>marina</i>	Reakirt, 1868	S USA to Nicaragua
	Riodininae	<i>Apodemia</i>	<i>walkeri</i>	Godman and Salvin, 1886	C Mexico to Costa Rica
		<i>Calephelis</i>	<i>matheri</i>	McAlpine, 1971	W Mexico
		<i>Calephelis</i>	<i>yauatepecensis</i>	Maza and Turrent, 1977	W Mexico
		<i>Caria</i>	<i>ino</i>	Godman and Salvin, 1878	W Mexico
		<i>Caria</i>	<i>stillaticia</i>	Dyar, 1912	W Mexico
		<i>Emesis</i>	<i>emesia</i>	Hewitson, 1867	C and S Mexico
		<i>Melanis</i>	<i>acroleuca</i>	Felder, 1869	W Mexico
20 species					
Hesperiidae	Eudaminae	<i>Achalarus</i>	<i>toxus</i>	Plötz, 1881	S USA to Panama
		<i>Astraptus</i>	<i>"fulgurator"</i>	Reakirt, 1867	S USA to S America
		<i>Chiodes</i>	<i>albofasciatus</i>	Hewitson, 1867	S USA to Costa Rica
		<i>Chiodes</i>	<i>zilpa</i>	Butler, 1872	S USA to Venezuela
		<i>Codatractus</i>	<i>bryaxis</i>	Hewitson, 1867	N Mexico to Honduras
		<i>Codatractus</i>	<i>hyster</i>	Dyar, 1916	Mexico
	These 2 sp	<i>Codatractus</i>	<i>melon</i>	Godman and Salvin, 1893	N Mexico to Costa Rica
	were pooled	<i>Codatractus</i>	<i>sallyae</i>	Warren, 1995	W Mexico to Guatemala
		<i>Cogia</i>	<i>cajeta</i>	Godman and Salvin, 1894	W Mexico to Costa Rica
		<i>Cogia</i>	<i>calchas</i>	Herrich-Schäffer, 1869	S USA to Argentina
		<i>Thessia</i>	<i>jalapus</i>	Plötz, 1881	S USA to Colombia
		<i>Urbanus</i>	<i>belli</i>	Hayward, 1935	S USA to W Mexico
		<i>Urbanus</i>	<i>dorantes</i>	Stoll, 1790	S USA to Argentina
		<i>Urbanus</i>	<i>procne</i>	Plötz, 1881	S USA to Argentina
		<i>Urbanus</i>	<i>teleus</i>	Hübner, 1821	S USA to Argentina
		<i>Urbanus</i>	<i>viterboana</i>	Ehrmann, 1907	C Mexico to Ecuador
	Pyrginae				
	Carcharodini	<i>Bolla</i>	<i>litus</i>	Dyar, 1912	W Mexico
		<i>Bolla</i>	<i>subapicatus</i>	Schaus, 1902	W Mexico to Guatemala
		<i>Nisoniades</i>	<i>rubescens</i>	Möschler, 1877	S USA to Bolivia
		<i>Staphylus</i>	<i>iguala</i>	Williams and Bell, 1940	SW Mexico
		<i>Staphylus</i>	<i>tierra</i>	Evans, 1953	W Mexico
	Erynnini	<i>Chiomara</i>	<i>georgina</i>	Reakirt, 1868	S USA to Costa Rica
		<i>Erynnis</i>	<i>funeralis</i>	Scudder and Burgess, 1870	S USA to Argentina
		<i>Erynnis</i>	<i>tristis</i>	Edwards, 1883	S USA to Colombia
		<i>Gesta</i>	<i>invisus</i>	Butler and Druce, 1872	S USA to Costa Rica
		<i>Gorgythion</i>	<i>begga</i>	Möschler, 1877	S USA to Bolivia
	Pyrgini	<i>Antigonus</i>	<i>emorsa</i>	Felder, 1869	S USA to W Mexico
		<i>Antigonus</i>	<i>funebis</i>	Felder, 1869	SW Mexico
		<i>Heliopetes</i>	<i>laviana</i>	Hewitson, 1868	S USA to Venezuela
		<i>Heliopetes</i>	<i>macaira</i>	Reakirt, 1867	S USA to Panama
		<i>Pyrgus</i>	<i>albescens</i>	Plötz, 1884	S USA to Oaxaca

	<i>Pyrgus</i>	<i>oileus</i>	Linnaeus, 1767	S USA to Costa Rica
	<i>Pyrgus</i>	<i>philetas</i>	Edwards, 1881	S USA to Chiapas
	<i>Zopyrion</i>	<i>sandace</i>	Godman and Salvin, 1896	C Mexico to Costa Rica
Heteropterinae	<i>Piruna</i>	<i>penaea</i>	Dyar, 1918	W Mexico
	<i>Piruna</i>	<i>purepecha</i>	Warren and Gonzalez, 1999	SW Mexico
Hesperiinae				
Thymelicini	<i>Ancyloxypha</i>	<i>arene</i>	Edwards, 1871	S USA to Costa Rica
	<i>Copaedes</i>	<i>minima</i>	Edwards, 1870	S USA to Costa Rica
Anthoptini	<i>Synapte</i>	<i>pecta</i>	Evans, 1955	S USA to Costa Rica
	<i>Synapte</i>	<i>syraces</i>	Godman, 1901	W Mexico
Moncini	<i>Amblyscirtes</i>	<i>tolteca</i>	Scudder, 1872	S USA to Nicaragua
	<i>Callimormus</i>	<i>saturnus</i>	Herrich-Schäffer, 1869	C Mexico to Paraguay
	<i>Cymaenes</i>	<i>trebius</i>	Mabille, 1891	S USA to Colombia
Hesperiini	<i>Polites</i>	<i>vibex</i>	Scudder, 1872	Canada to Argentina
	<i>Pompeius</i>	<i>pompeius</i>	Latreille, 1824	C Mexico to Argentina
<b>45 species</b>				
<b>Total</b>	<b>119 species</b>	<b>26 discarded</b>	<b>93 species, 92 taxa analyzed</b>	

**Table A2** Theoretical versus observed distributions (SOM analysis) of the 93 species sampled in the rainy season. “G” is generalist species and “S” is specialist species. Degree of openness: “A” is open, B and C are shrubs and W is trees. Humidity gradient: \* is xeric, \*\* moderately xeric, \*\*\* is humid. Evaluation of the congruence: Y (yes): 51/93 species (54.8%); N (no): 5/93 species (5.4%); 0: compatible but not positive or negative. Tendency: congruence not at 100%. Bolt species are considered bioindicators. Note: some generic names were abbreviated to fit the table.

Species	Theoretical			SOM result observed		
	Degree of openness	Humidity gradient	G or S	Degree of openness	Humidity gradient	Congruence
<i>B.brevicornis</i>	Shrubs	*	S	Shrubs	*	Y
<i>Papilio rumiko</i>	Open-Shrubs	**	G	Shrubs	**	Tendency
<i>Eurytides epidaus</i>	Shrubs	*	G	Shrubs	*	Y
<i>Ascia monuste</i>	Open	**	G	Open	**	Y
<i>Abaeis nicippe</i>	Shrubs	*	G	Open	**	0
<i>Eurema arbella</i>	Trees	***	S	Trees	**	Tendency
<i>Eurema daira</i>	Open	*	S	Open	**	Tendency
<i>Eurema mexicana</i>	Open-Shrubs	**	G	Open	**	Y
<i>Nathalis iole</i>	Open-Shrubs	**	G	Open	*	0
<i>Phoebis agarithe</i>	Open	**	G	Open	*	Tendency
<i>Phoebis philea</i>	Trees	***	G	Trees	***	Y
<i>Phoebis sennae</i>	Shrubs-Trees	**	G	Trees	**	Y
<i>Pyrisitia dina</i>	Shrubs	**	G	Shrubs	**	Y
<i>Pyrisitia lisa</i>	Open-Shrubs	**	G	Open	***	Tendency
<i>Pyrisitia nise</i>	Shrubs	**	G	Shrubs	**	Y
<i>Pyrisitia proterpia</i>	Shrubs	**	G	Shrubs	***	Tendency
<i>Zerene cesonia</i>	Open-Shrubs	**	G	Open-Shrubs	**	Y
<i>L.carinenta</i>	Shrubs-Trees	*	G	Shrubs	**	Tendency
<i>D.eresimus</i>	Open	*	S	Open	*	Y
<i>D.plexippus</i>	Open	**	S	Open	*	Y
<i>Danaus gillipus</i>	Open	*	S	Open	*	Y
<i>Agraulis vanillae</i>	Open-shrub	**	G	Open	**	Y
<i>E. claudia</i>	Open	**	S	Open	**	Y
<i>Euptoieta hegesia</i>	Open	**	S	Open-Shrubs	**	Y
<i>H. glaucomone</i>	Trees	***	S	Trees	***	Y
<i>M.cyananthe</i>	Trees	***	G	Trees	***	Y

<b>S. steneles</b>	<b>Shrubs</b>	<b>***</b>	<b>S</b>	<b>Shrubs</b>	<b>***</b>	<b>Y</b>
<i>Smyrna blonfildia</i>	Trees	***	G	Trees	***	Y
<b>Anartia fatima</b>	<b>Open</b>	<b>**</b>	<b>S</b>	<b>Open</b>	<b>**</b>	<b>Y</b>
<b>Anartia jatrophae</b>	<b>Open</b>	<b>**</b>	<b>S</b>	<b>Open</b>	<b>**</b>	<b>Y</b>
<i>A.texana</i>	Shrubs	**	G	Shrubs	**	Y
<i>Chlosyne lacinia</i>	Open-Shrubs	**	G	Shrubs	***	Tendency
<i>Chlosyne rosita</i>	Generalist	***	G	Generalist	**	Tendency
<i>Junonia coenia</i>	Open	**	G	Open	*	Tendency
<i>Microtia elva</i>	Trees	***	G	Trees	***	Y
<i>P.graphica</i>	Trees	**	G	Open-Shrubs	**	Tendency
<i>P.pallescens</i>	Open	**	G	Open-Shrubs	**	Y
<i>Texola elada</i>	Shrubs-Trees	**	G	Trees	**	Y
<i>Anaea aidea</i>	Shrubs	**	G	Shrubs	***	Tendency
<i>Cissia similis</i>	Trees	**	G	Trees	**	Y
<i>Cissia themis</i>	Shrubs-Trees	***	G	Trees	**	Tendency
<i>P.squamistriga</i>	Open-Shrubs	**	G	Open-Shrubs	**	Y
<i>Apodemia walkeri</i>	Open-Shrubs	**	G	Shrubs	***	Tendency
<i>C.matheri</i>	Open	*	G	Open	**	Tendency
<i>C.yautepequensis</i>	Open	*	G	Open	*	Y
<b>Emesis emesia</b>	<b>Shrubs</b>	<b>**</b>	<b>S</b>	<b>Shrubs</b>	<b>**</b>	<b>Y</b>
<i>Melanis acroleuca</i>	Shrubs	***	G	Shrubs	***	Y
<i>H.ceraunus</i>	Open-Shrubs	**	G	Open	*	0
<i>Hemiargus isola</i>	Open	***	G	Shrubs	**	N
<i>Leptotes marina</i>	Open-Shrubs	**	G	Open	**	Y
<i>Arawacus jada</i>	Shrubs	**	S	Open	**	Tendency
<b>C.miserabilis</b>	<b>Shrubs</b>	<b>**</b>	<b>S</b>	<b>Shrubs</b>	<b>**</b>	<b>Y</b>
<i>P.bathildis</i>	Trees	***	G	Trees	***	Y
<i>Rekoa zebina</i>	Shrubs	***	G	Shrubs	***	Y
<i>Strymon bazochii</i>	Shrubs	**	G	Open	**	Tendency
<i>Achalarus toxus</i>	Open-Shrubs	**	G	Trees	**	Tendency
<b>A.fulgerator</b>	<b>Trees</b>	<b>***</b>	<b>S</b>	<b>Trees</b>	<b>***</b>	<b>Y</b>
<i>Chioides zilpa</i>	Open-Shrubs	**	G	Trees	**	Tendency
<i>C.bryaxis</i>	Trees	***	G	Shrubs	***	Tendency
<i>C.hyster</i>	Shrubs	**	G	Shrubs-Trees	***	Tendency
<i>C.melon/sallyae</i>	Shrubs	***	G	Shrubs	***	Y
<i>Cogia cajeta</i>	Open-Shrubs	*	G	Shrubs	**	0
<i>Cogia calchas</i>	Shrubs	**	G	Shrubs	**	Y
<i>Thessia jalapus</i>	Shrubs	**	G	Open	*	0
<i>Urbanus belli</i>	Trees	**	G	Shrubs	**	Tendency
<i>Urbanus dorantes</i>	Shrubs	**	G	Shrubs	***	Tendency
<i>Urbanus procne</i>	Open-Shrubs	**	G	Open-Shrubs	**	Y
<i>Antigonus emorsa</i>	Open-Shrubs	*	G	Shrubs-Trees	***	N
<i>A.funebris</i>	Shrubs	**	G	Shrubs	***	Tendency
<i>Bolla litus</i>	Open-Shrubs	**	G	Open-Shrubs	**	Y
<i>Bolla subapicatus</i>	Shrubs	**	G	Shrubs	**	Y
<i>C.georgina</i>	Open-Shrubs	**	G	Shrubs	***	Y
<i>Erynnis funeralis</i>	Open	**	G	Open-Trees	**	Y
<i>Gesta invisus</i>	Open-Shrubs	**	G	Trees	*	0
<b>H.laviana</b>	<b>Trees</b>	<b>***</b>	<b>S</b>	<b>Trees</b>	<b>***</b>	<b>Y</b>
<i>H.macaia</i>	Open	*	G	Open	*	Y
<i>N.rubescens</i>	Shrubs-Trees	**	G	Shrubs-Trees	**	Y
<i>Pyrgus communis</i>	Open	*	G	Open-Shrubs	*	Y
<i>Pyrgus oileus</i>	Open	**	G	Trees	**	0
<i>Pyrgus philetas</i>	Open	**	G	Open-Shrubs	**	Y
<i>Staphylus iguala</i>	Shrubs-Trees	**	G	Shrubs-Trees	***	Y
<i>Zopyrion sandace</i>	Trees	**	G	Trees	***	Tendency
<i>C.minima</i>	Open	**	G	Shrubs	***	N
<i>A.tolteca</i>	Open-Shrubs	*	G	Shrubs	**	Tendency
<i>A.arene</i>	Open-Shrubs	**	G	Open-Shrubs	**	Y
<i>C.saturnus</i>	Trees	***	G	Trees	***	Y
<i>Cymaenes trebius</i>	Shrubs	***	G	Trees	***	Tendency
<i>C.tripunctus</i>	Shrubs	**	G	Open	**	Tendency
<i>Polites vibex</i>	Open	*	G	Open	**	Tendency
<i>P.pompeius</i>	Generalist	**	G	Generalist	***	Tendency

<i>Synapte syracae</i>	Trees	***	S	Trees	***	Y
<i>Synapte pecta</i>	Trees	***	S	Trees	***	Y
<i>P.purepecha</i>	Shrubs-Trees	***	S	Trees	***	Y

**Table A3.** Theoretical versus observed distributions (SOM analysis) of the 42 species sampled in the dry season. “G” is generalist species and “S” is specialist species. Evaluation of the congruence: Y: 14/42 species (33.3%); N: 7/42 species (16.7%); 0: compatible but not positive or negative, Tendency: congruence not at 100%. Bolt species are considered bioindicators. Note: some generic names were abbreviated to fit the table.

Species	Theoretical			SOM result observed		
	Degree of Openness	Humidity gradient	G or S	Degree of Openness	Humidity gradient	Congruence
<i>Abaeis nicippe</i>	Shrubs	**	G	Open	*	N
<i>Eurema daira</i>	Trees	***	S	Trees	***	Y
<i>E.mexicana</i>	Open-Shrubs	***	G	Open	***	Y
<i>Nathalis iole</i>	Open-Shrubs	*	S	Open	*	Y
<i>Phoebis agarithe</i>	Open	**	G	Open	*	Tendency
<i>Phoebis philea</i>	Trees	***	G	Trees	***	Y
<i>Phoebis sennae</i>	Open-Shrubs	***	G	Open	***	Y
<i>Pyrisitia dina</i>	Shrubs	**	G	Shrubs-Trees	**	Y
<i>Pyrisitia nise</i>	Shrubs	**	G	Shrubs	***	Tendency
<i>Pyrisitia proterpia</i>	Shrubs	**	G	Trees	***	0
<i>Zerene cesonia</i>	Open-Shrubs	**	G	Shrubs	***	Tendency
<i>L.carinenta</i>	Shrubs-Trees	**	G	Trees	***	0
<i>D.plexippus</i>	Open	**	S	Open	*	Tendency
<i>Danaus gilipus</i>	Open	**	S	Shrubs	***	N
<i>M.cyananthe</i>	Trees	***	G	Trees	**	Tendency
<i>S.blonfildia</i>	Trees	***	G	Trees	*	0
<i>A.texana</i>	Shrubs	**	G	Trees	***	0
<i>Junonia coenia</i>	Open	**	G	Open	***	Tendency
<i>Anaea aidea</i>	Shrubs	***	G	Shrubs	***	Y
<i>Cissia similis</i>	Trees	***	S	Trees	**	Tendency
<i>Cissia themis</i>	Shrubs	***	G	Trees	*	N
<i>A.walkerii</i>	Open-Shrubs	**	G	Open-Shrubs	***	Tendency
<i>C.yautepequensis</i>	Open	*	G	Open	*	Y
<i>H.ceraunus</i>	Shrubs-Trees	**	G	Trees	*	0
<i>Hemiargus isola</i>	Open	***	G	Open	**	Tendency
<i>Leptotes marina</i>	Open-Shrubs	**	G	Open	**	Y
<i>Urbanus procne</i>	Open-Shrubs	**	G	Trees	***	N
<i>A.emorsa</i>	Open-Shrubs	*	G	Shrubs-Trees	*	Tendency
<i>Erynnis funeralis</i>	Open	**	G	Open	**	Y
<i>H.laviana</i>	Trees	***	S	Shrubs-Trees	*	N
<i>Heliopterus macaira</i>	Open	**	G	Shrubs	***	N
<i>N.rubescens</i>	Shrubs-Trees	***	G	Open-Shrubs	**	0
<i>P.communis</i>	Open-Shrubs	*	G	Shrubs	***	N
<i>Pyrgus oileus</i>	Open-Shrubs	**	G	Open	**	Y
<i>Pyrgus philetas</i>	Open	*	G	Open	**	Tendency
<i>Staphylus iguala</i>	Open-Shrubs	**	G	Open	*	Tendency
<i>Zopyrion sandace</i>	Trees	*	G	Trees	**	Tendency
<i>C.minima</i>	Open	*	G	Open	*	Y
<i>C.trebius</i>	Shrubs	**	S	Trees	**	Tendency
<i>C.tripunctus</i>	Shrubs	**	G	Open	**	Tendency
<i>Polites vibex</i>	Open	*	G	Open-Trees	*	Y
<i>P.pompeius</i>	Open-Shrubs	**	G	Open	**	Y

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