



Review

Metabolically Active Microbial Communities in Oilfields: A Systematic Review and Synthesis of RNA Preservation, Extraction, and Sequencing Methods

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Abstract: Characterizing metabolically active microorganisms using RNA-based methods is a crucial tool for monitoring and mitigating operational issues, such as oil biodegradation and biocorrosion of pipelines in the oil and gas industry. Our review, a pioneering study, addresses the main methods used to preserve, isolate, and sequence RNA from oilfield samples and describes the most abundant metabolically active genera studied. Using the MEDLINE/PubMed, PubMed Central, Scopus, and Web of Science databases, 2,561 potentially eligible records were identified. After screening, 20 studies were included in our review, underscoring the scarcity of studies related to the subject. Data were extracted and reported according to the Preferred Reporting Items for Systematic Reviews and Meta-Analysis (PRISMA). These studies evaluated different samples, including produced water (PW), injection water (IW), solid deposits (SD), oil (OIL), and oily sludge (OS) collected from oilfields located in Australia, China, India, Mexico, and the United Arab Emirates. Environmental samples accounted for 55% of the studies, while enriched cultures and microbial consortia represented 35% and 15% of studies, respectively. PW was the most frequently studied sample, comprising 72% of all samples. Filtration and centrifugation were the only processes employed to concentrate the biomass present in samples. For RNA preservation, the most used method was a solution composed of 95:5 *v/v* ethanol/TRIZOL, while for RNA isolation, the TRIZOL reagent was the most cited. The Sanger sequencing method was used in all studies evaluating functional genes (*alkB*, *dsrA*, *aprA*, *assA*, and *mcrA*), and the Next-Generation Sequencing (NGS) method was employed in studies for sequencing transcripts of the 16S rRNA gene and metatranscriptomes. *Pseudomonas* (16S rRNA = PW: 2%; IW: 8%; metatranscriptome = PW: 20%) and *Acinetobacter* (16S rRNA = PW: 1%; IW: 4%; metatranscriptome = PW: 17%) were the most abundant genera. This study outlined the primary methods employed in researching metabolically active microorganisms. These data provide a foundation for future research. However, it is essential to note that we cannot yet determine the most effective method. We hope that this study will inspire further research related to the standardization of RNA preservation, extraction, and sequencing methods and significantly contribute to our understanding of active microbial communities in oilfields.

Keywords: oilfields; petroleum; metabolically active microorganisms; hydrocarbon-degrading microorganisms; corrosion-influencing microorganisms; RNA preservation; RNA extraction; 16S rRNA gene transcripts; metatranscriptome; systematic review

1. Introduction

In the oil production process, the fluid that reaches the surface from the subsurface is triphasic, containing oil, water, and gas phases [1]. During the stage of oil primary recovery, the water phase is composed exclusively of formation water that occurs naturally within the pores of reservoir rocks, while waterflood during secondary recovery occurs through the injection of water (or gas) into reservoirs to repressurize the environment and displace the oil to the producing wells [2–5]. As oil production continues, the amount of produced water increases in relation to the oil and gas phases, and, as a consequence, produced water represents the largest volume of waste stream in oil and gas production operations [2].

Microbes occur in all these types of water (formation water, seawater, injection water, and produced water) and they colonize the internal environment of industrial equipment that interacts with water-containing streams, establishing biofilms on their internal surfaces [1–4,6]. The presence and activities of microbes, especially those present in biofilms, may engender a series of problems for the oil industry, such as biofouling, biocorrosion, and biological souring [6–13].

The identification and characterization of microbial communities, including hydrocarbon-degrading microorganisms (HDM) [14–16] and corrosion-influencing microorganisms (CIM) of pipelines [1,10,17,18], are crucial for developing strategies that minimize the biological impacts on oil quality, and transport and storage facilities of fluids resulting from oil production processes [1,10]. Traditionally, culture-dependent methods were used to isolate and identify microbial groups from oil reservoirs [4,11,19,20]. However, the so-called “culturable” strains usually represent about 1–5% of the total species present in a given environmental sample, casting doubt on the representativeness of the results [21]. To overcome these limitations, microbiological molecular methods that are culture-independent have been increasingly used to characterize microbial communities and identify unculturable and rare species present in complex environmental samples from the oil industry [21,22].

The most applied molecular methods involve DNA extraction, polymerase chain reaction (PCR), amplicon sequencing of 16S rRNA, and shotgun metagenomics [10,22–28]. However, RNA-based sequencing approaches that are presented have become alternatives to examine metabolically active microbial communities, including those from oil reservoirs [1,16,17,22].

RNA transcript sequencing is limited by difficulties in preserving and isolating high-quality RNA from environmental samples as RNA is an extremely unstable molecule and susceptible to degradation by the action of RNase enzymes [22]. Moreover, an adequate amount of RNA is critical to successfully carry out all stages of the analytical process, which involves the synthesis of complementary DNA (cDNA), construction of cDNA libraries, sequencing, and bioinformatic processing of obtained sequences [29]. Consequently, there is a lack of studies reporting the microbial diversity and composition in oil industry samples based on RNA transcripts [22].

The existence of standardized and established protocols for nucleic acid preservation ensures that samples collected in remote areas can be successfully and consistently transported to laboratories with suitable infrastructure for RNA extraction without significantly compromising their integrity [30–32]. Therefore, it allows for the characterization of metabolically active microbial communities, as well as the identification of rare, less abundant species [32].

In this context, the present review aimed to describe the methods applied in studies that evaluated metabolically active microbial communities (based on RNA) from oil industry samples. Additionally, the dominant genera of active microorganisms (bacteria and archaea)

in the analyzed samples were identified. It should be noted that no previous review studies were identified in the literature aimed at active microorganisms in oil reservoirs or oil industry facilities. Therefore, this review is innovative in analyzing methodological data of preservation, extraction, and sequencing of RNA from samples related to the oil industry.

2. Materials and Methods

This systematic review was designed and carried out according to the guidelines of the preferred reporting items for systematic reviews and meta-analyses (PRISMA) [33]. The study consisted of the following steps: (1) identification of records in databases; (2) automated and manual screening of records; (3) assessment of document eligibility and inclusion of selected studies; and (4) synthesis and analysis of data. The PRISMA-S checklist [34] helped to describe the items applicable to these four stages (Supplementary Table S1).

2.1. Identification of Records in Databases

The search for records was carried out in the MEDLINE/PubMed (via National Library of Medicine), PubMed Central, Scopus, and Web of Science (Core Collection) databases. Records were identified by searching for combined terms (keywords) using the Boolean operators “OR” and “AND” (Supplementary Table S2). These terms were carefully defined to characterize the sampling points, types of samples, and analyses. The search was carried out in the title or abstract (MEDLINE/PubMed and PubMed Central) and title or abstract or keywords (Scopus and Web of Science) of publications. There were no restrictions on document types, language, and publication date to avoid pre-excluding relevant records. The last access to databases was on 20 January 2023.

2.2. Automated and Manual Screening of Records

Data from identified records were exported from databases (MEDLINE/PubMed, PubMed Central, Scopus, and Web of Science) in .csv file format (Supplementary Table S3). These were converted to the .xlsx format using the `format_input.py` script (<https://github.com/lbmcf/format-input>) (Supplementary Table S4). The `format_input.py` script also identified and removed records without DOI and those with identical titles or DOI.

Using the `remove_duplicates.py` script (<https://github.com/lbmcf/remove-duplicates>), the .xlsx files from databases were unified and duplicate data were removed (Supplementary Table S4). Documents corresponding to records listed in the unified file were downloaded in PDF format using an automated program that downloads scientific articles based on their DOI. This program is restricted to the network of the Federal University of Minas Gerais (UFMG), and inaccessible documents (closed access) were categorized as not available (NA).

2.3. Assessment of Eligibility and Inclusion of Studies

The downloaded PDF documents were converted to TXT format, using the `pdf2txt.py` script (<https://github.com/lbmcf/pdf2txt>). Using the `search_keywords.py` script (<https://github.com/lbmcf/search-keywords>), an automated screening was conducted on the methodology section in the TXT documents (Supplementary Table S4). For selection, terms analogous to those used in the database record search were applied (Supplementary Table S2).

The selected documents were independently reviewed by two reviewers (R.F.G. and J.d.C.F.D.). In cases of disagreement, a third reviewer (M.S.C.) was consulted to obtain definitive consensual information. According to eligibility criteria (Table 1), the screening was conducted by checking the title and methodology of the documents. Records without DOI, previously removed by the `format_input.py` script, were recovered, and the corresponding documents were manually downloaded (in PDF format). These records were reviewed by title and methodology (Supplementary Table S4). According to the adopted eligibility criteria, studies that evaluated metabolically active microbial communities based on RNA sequencing from samples collected from reservoirs, pipelines, and tanks in the oil industry were included (Table 1).

Table 1. Eligibility criteria for the inclusion of articles in the systematic review.

Criteria	Description
Sampling location	Reservoir; pipeline; tank
Sample type	environmental (oil, injection water, produced water, biofilm, pig residue, and oil sludge); laboratory (cultures inoculated with environmental samples)
Analysis Study	RNA sequencing Original

Studies were excluded if: (1) they only presented genomic DNA sequencing data; (2) they analyzed samples of soil, fauna and flora (marine and terrestrial) contaminated with oil; (3) they evaluated samples of oil-refined products; (4) they performed experiments with commercial strains or isolated from oilfields in which the isolation conditions and molecular analysis were not specified; and (5) they consisted of a literature review of documents, book chapters, conferences, or similar.

2.4. Synthesis and Analysis of Data of Included Studies

The methods used in different stages of development of the studies included in this review were evaluated. Data were collected on (1) sampling, (2) preservation and isolation of RNA, (3) RNA sequencing, and (4) identification of metabolically active microorganisms by analysis of 16S rRNA gene transcripts and metatranscriptome.

Analysis of Data from 16S rRNA Gene Transcripts and Metatranscriptome

All published data were compiled from studies related to the relative abundance of metabolically active microorganisms (Bacteria and Archaea). These data were obtained from the amplicon analyses of 16S rRNA gene transcripts (Supplementary Table S6). To normalize the data, only relative abundances greater than zero (>0) and with valid scientific nomenclature were considered (Supplementary Table S6).

Raw metatranscriptome data were downloaded from the NCBI SRA database with the following codes: SRR5352268 (W15-5), SRR5352269 (W9-18), and SRR5352270 (W2-71) for study 5 [16] and with codes: SRR11866712 (PW), SRR11866713 (O3), SRR11866714 (E3), and SRR11866715 (A3) for study 8 [35]. For study 19 [14], metatranscriptomes with codes OES118212 (HX_G1), OES053914 (XY_B2), OES053913 (XY_B1), OES118211 (HX_G6a, HX_G6b), OES053915 (XY_P1, XY_P2, XY_P3), and OES118213 (HX_G2) were downloaded from the Bio-Med Big Data Center database.

The analysis of metatranscriptomic data was carried out using the SqueezeMeta pipeline, which performs a complete analysis from quality control to taxonomic and functional annotation [36], and default parameters were used for the co-assembly mode. The pipeline also uses Trimmomatic [37] for quality control (trimming and filtering). Assembly was done using Megahit [38], followed by Prodigal [39] for open reading frame (ORF) prediction. The Diamond tool was used for the taxonomic classification of ORFs against the Genbank NR database [40]. Using the R package SQMtools v1.6.0 [41], a tabulated file containing the results of absolute taxonomic abundances was exported.

The 16S rRNA transcripts data from studies were grouped and analyzed by sample type, while the metatranscriptomic data were evaluated by study. For the analysis of phyla and genera abundance, the ggplot2 package (v3.3.6) was used [42]. To compare the similarity of shared phyla and genera, a Venn diagram was elaborated with the Venn package (v1.11) (<https://github.com/dusadrian/venn>).

3. Results

3.1. Identification and Selection of Studies

Potentially eligible records were identified in the four selected databases: MEDLINE/PubMed ($n = 154$), PubMed Central ($n = 60$), Scopus ($n = 1650$), and Web of Science ($n = 697$), totaling 2561 records (Figure 1). Duplicates by title or DOI identified in the Scopus

database ($n = 7$) and those without a DOI detected in the MEDLINE/PubMed ($n = 10$), Scopus ($n = 156$), and Web of Science ($n = 30$) databases were removed (Figure 1), using the format input.py script. After removal, 2358 records were tracked (Figure 1). These were unified into a single file, and duplicate records ($n = 624$) were removed (Figure 1) using the remove_duplicates.py script. After duplicate removal, 1734 reports were selected for recovery (Figure 1). Using software with permissions to access UFMG's internal network, 1648 reports were retrieved in PDF format (Figure 1). The remaining 86 reports were not available due to access restrictions (Figure 1). In the methodology section of recovered reports, a keyword screening (Supplementary Table S2) was performed using the pdf2txt.py and search_keywords.py scripts. A total of 1457 reports (Figure 1) that did not contain at least one of the determined terms were excluded. Following the eligibility (Table 1) and exclusion criteria, the remaining 191 reports were manually reviewed by title and methodology (Figure 1). Of these, 20 original articles were included in this systematic review for data synthesis and analysis (Figure 1; Supplementary Table S5). It is important to note that the reports corresponding to records without DOI ($n = 196$) removed at the identification stage were downloaded and reviewed. However, none of these reports were considered eligible for inclusion (Figure 1).

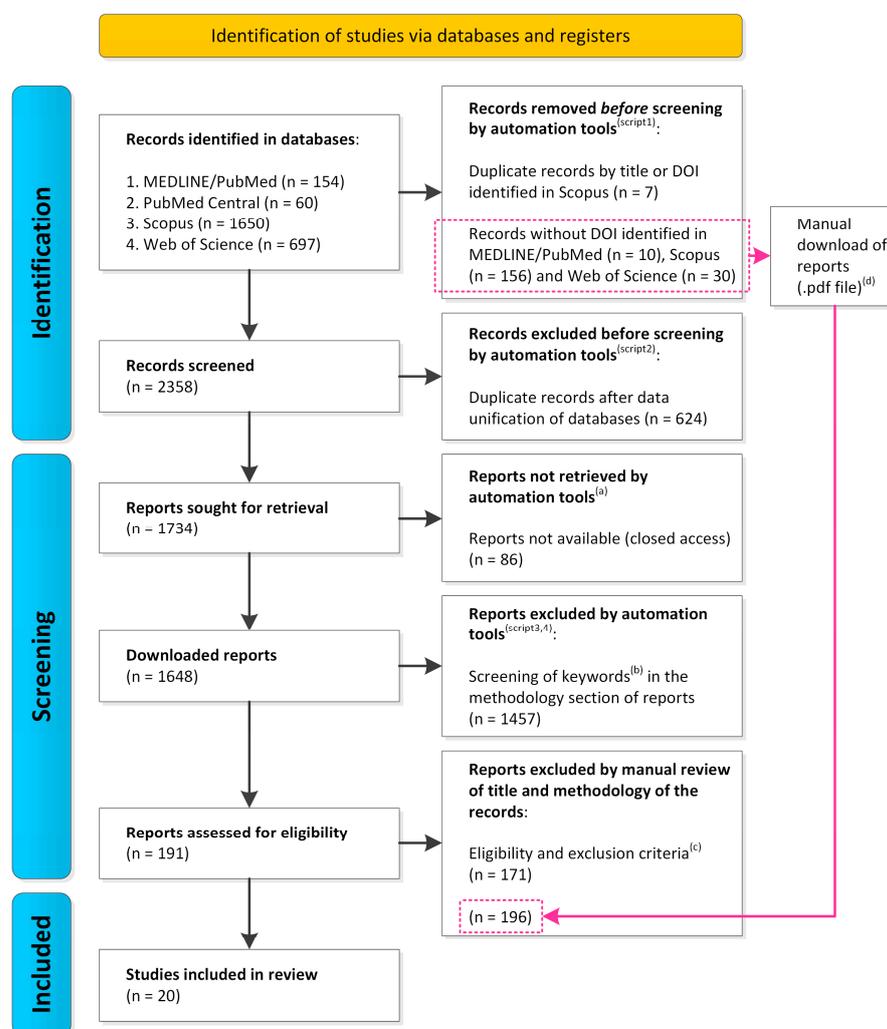


Figure 1. PRISMA flow diagram [33] of identification, screening, and inclusion of studies in the systematic review. ^{Script1} format_input.py; ^{Script2} remove_duplicates.py; ^{Script3} pdf2txt.py; ^{Script4} search_keywords.py (Supplementary Table S4). ^(a) Software based on permissions of the UFMG internal network. ^(b) Keywords (Supplementary Table S2). ^(c) Eligibility and exclusion criteria (Methodology Section 2.3). ^(d) Reports screened manually.

3.2. Included Studies

The included studies were published between 2011 and 2023 (Figure 2). These studies evaluated metabolically active microorganisms capable of degrading hydrocarbons and organic compounds and/or influencing corrosion processes in oil industry facilities. In 2011, only one study was published on the subject (Figure 2). However, since 2016, there has been a considerable increase in the number of publications, ranging from one to three articles per year, except in 2020, when seven studies were identified (Figure 2).

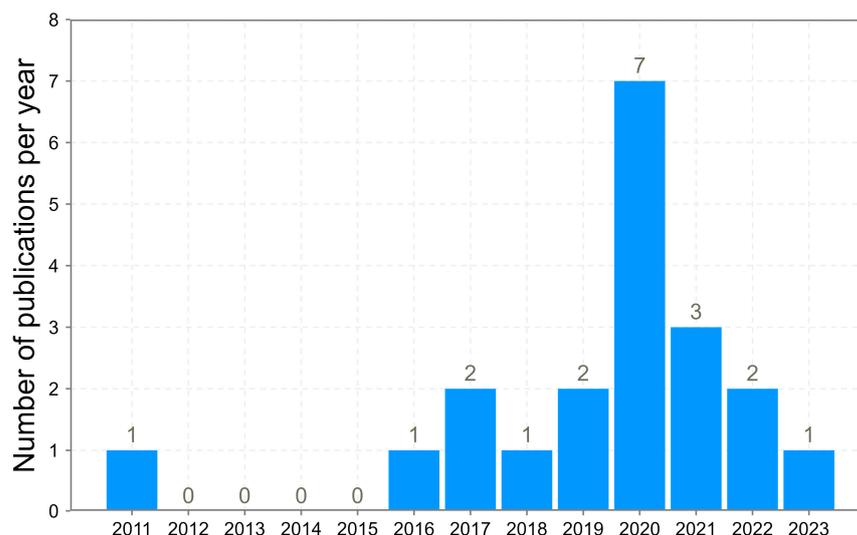


Figure 2. Number of publications per year of studies that evaluated metabolically active microorganisms in oilfields.

The studies evaluated samples collected from oilfields located in Australia, China, India, Mexico, and the United Arab Emirates (UAE) (Table 2). Notably, most of these studies ($n = 13$; 65%) were carried out with samples from Chinese fields, followed by Australia ($n = 4$), Mexico ($n = 1$), India ($n = 1$), and UAE ($n = 1$) (Table 2). Different types of environmental samples were evaluated, including produced water (PW), injection water (IW), solid deposits (SD), oil (OIL), and oily sludge (OS) (Table 2). The PW, IW, and OIL samples were mostly collected from production and injection wells, and surface pipelines [1,3,4,15,17–19,43–46]. The OS samples were obtained from storage tanks and fluid passages originating from oil reservoirs [13,14], while the DS was acquired from the internal scraping of pipelines [1]. Salgar-Chaparro and Machuca (2019) [1] were the only ones who proposed to evaluate planktonic microbiota circulating in fluids (PW and IW) and sessile microbiota adhered (SD) to the inner walls of the pipelines (Table 2).

3.3. Methods of Preprocessing, Preservation, and Extraction of RNA

In total, 50 samples were collected and analyzed in the 20 included studies (Table 2). The PW was the most studied sample ($n = 36$; 72%), followed by IW ($n = 6$), SD ($n = 4$), OS ($n = 2$), and OIL ($n = 2$) (Table 2). Altogether, 11 studies focused on the analysis of environmental samples (PW, IW, and SD) to investigate microorganisms degrading organic compounds (OCDM) and hydrocarbons (HDM) and/or corrosion-influencing microorganisms (CIM) (Table 2). The other seven studies analyzed cultures enriched with PW (PW.HDM and PW.SRB), IW (IW.HDM), OIL (OIL.HDM) and OS (OS.OCDM and OS.MET-HDM) (Table 2). Liu et al. (2020) [35] were the only ones who evaluated the environmental sample (PW) and methanogenic cultures (MET-HDM) in the same study (Table 2). Finally, three studies focused on the analysis of CIM consortia (PW.CIM, ND.CIM) (Table 2; Supplementary Table S6). Albahri et al. (2021) [18] evaluated a microbial consortium recovered from an oil production facility, but the authors did not declare the type of sample used as inoculum so this sample was named ND (no data).

Table 2. Reference, country, and samples studied in the articles included in this systematic review.

Reference	Sampling Country	Sample Type	N	Sample Studied	ID
Shestakova et al. (2011) [15]	Hebei, China	PW	1	Culture	PW.HDM ^(a)
		IW	1	Culture	IW.HDM ^(a)
Zapata-Peñasco et al. (2016) [4]	Mexico	PW	2	Culture	PW.SRB ^(a)
Nazina et al. (2017) [3]	Hebei, China	IW	1	Environmental	IW ^(a)
Li et al. (2017) [17]	Jiangsu, China	PW	6	Environmental	PW ^(a)
Liu et al. (2018) [16]	Jiangsu, China	PW	3	Environmental	PW ^(a)
Salgar-Chaparro e Machuca (2019) [1]	Australia	PW	6	Environmental	PW ^(a)
		IW	1	Environmental	IW ^(a)
		SD	4	Environmental	SD ^(a)
Zheng et al. (2019) [19]	China	OIL	1	Culture	OIL.HDM ^(a)
Liu et al. (2020) [35]	Shandong, China	PW	1	Environmental Culture	PW ^(a) PW.HDM ^(a)
Liu et al. (2020) [47]	China	PW	2	Environmental	PW ^(a)
Liu et al. (2020) [48]	Jiangsu, China	PW	1	Environmental	PW ^(a)
Zhou et al. (2020) [22]	China	PW	2	Environmental	PW ^(a)
Liu et al. (2020) [49]	China	PW	1	Environmental	PW ^(a)
Salgar-Chaparro et al. (2020) [43]	Australia	PW	1	Consortium	PW.CIM ^{(a),(b)}
Salgar-Chaparro et al. (2020) [44]	Australia	PW	2	Consortium	PW.CIM ^{(a),(b)}
Alhefeiti et al. (2021) [13]	United Arab Emirates	OS	1	Culture	OS.OCDM ^(a)
Zhou et al. (2021) [45]	Jiangsu, China	PW	6	Environmental	PW ^(a)
Albahri et al. (2021) [18]	Australia	OIL	1	Consortium	ND.CIM ^{(a),(b),(c)}
Su et al. (2022) [20]	China	PW	1	Culture	PW.HDM ^(a)
Zhou et al. (2022) [14]	Shengli, China	OS	1	Culture	OS.MET-HDM ^(a)
Prajapat et al. (2023) [46]	Rajasthan, India	PW	1	Environmental	PW ^(a)
		IW	3	Environmental	IW ^(a)

Produced water (PW); injection water (IW); solid deposits (SD); oil (OIL); oily sludge (OS); corrosion-influencing microorganisms (CIM); hydrocarbon-degrading microorganisms (HDM); organic compounds degrading microorganisms (OCDM); sulfate-reducing bacteria (SRB); methanogenic hydrocarbon-degrading microorganisms (MET-HDM). Sample number (n). Identification code (ID). ^(a) Samples used for RNA sequencing: (1) environmental and/or microbial, (2) microbial culture and/or (3) consortium enriched with environmental samples. ^(b) Description of CIM groups (Supplementary Table S7). ^(c) Sample type (inoculum) not available in the study, no data (ND).

The filtration process of water samples (PW and IW) was described in only four studies, three of which chose field preprocessing with sample filtration immediately after collection (Table 3). For biomass concentration, these studies used filter membranes with pore sizes of 0.1 µm [1,17,45]. In two studies, the microbial cells retained on filters were preserved with a 95:5 *v/v* ethanol/TRIZOL solution, and in one study with RNAProtect Bacteria Reagent (Table 3). In the study by Nazina et al. (2017) [3], the IW sample was preserved with ethanol reagent during collection and, unlike the other three articles, the filtration process with a 0.22 µm diameter pore membrane filter took place in the laboratory (Table 3).

Table 3. Methods of preprocessing, preservation, and extraction of RNA from the studied samples in the articles included in this systematic review.

Reference	Sample	Preprocessing	RNA Preserving Agent	RNA Extraction Method
Shestakova et al. (2011) [15]	PW.HDM	Centrifugation ^(b)	Not declared	TRIzol reagent ^(d)
	IW.HDM	Centrifugation ^(b)	Not declared	TRIzol reagent ^(d)
Zapata-Peñasco et al. (2016) [4]	PW.SRB	Centrifugation ^(b)	RNAprotect Bacteria Reagent ^(e)	RNeasy Protect Bacteria kit ^(e)
Nazina et al. (2017) [3]	IW	Filtration ^(b)	Ethanol reagent ^(h)	TRIzol reagent ^(d)
Li et al. (2017) [17]	PW	Filtration ^(a)	95:5 v/v ethanol/trizol ^(h)	High Pure RNA Isolation Kit ^(f)
Liu et al. (2018) [16]	PW	Centrifugation ^(b)	95:5 v/v ethanol/trizol ^(h)	PowerMicrobiome RNA Isolation kit ^(g)
Salgar-Chaparro e Machuca (2019) [1]	PW	Filtration ^(a)	RNAprotect Bacteria Reagent ^(e)	RNeasy PowerWater kit ^(e)
	IW	Filtration ^(a)	RNAprotect Bacteria Reagent ^(e)	RNeasy PowerWater kit ^(e)
	SD	Not declared	RNAprotect Bacteria Reagent ^(e)	RNeasy PowerSoil kit ^(e)
Zheng et al. (2019) [19]	OIL.HDM	Not declared	Not declared	TRIzol reagent ^(d)
Liu et al. (2020) [35]	PW	Centrifugation ^(b)	95:5 v/v ethanol/trizol ^(h)	PowerMicrobiome RNA Isolation kit ^(g)
	PW.HDM	Centrifugation ^(b)	Not declared	PowerMicrobiome RNA Isolation kit ^(g)
Liu et al. (2020) [48]	PW	Centrifugation ^(b)	95:5 v/v ethanol/trizol ^(h)	TRIzol reagent ^(d)
Liu et al. (2020) [47]	PW	Centrifugation ^(b)	95:5 v/v ethanol/trizol ^(h)	PowerMicrobiome RNA Isolation kit ^(g)
Zhou et al. (2020) [22]	PW	Centrifugation ^(b)	95:5 v/v ethanol/trizol ^(h)	TRIzol reagent ^(d)
Liu et al. (2020) [49]	PW	Centrifugation ^(b)	95:5 v/v ethanol/trizol ^(h)	TRIzol reagent ^(d)
Salgar-Chaparro et al. (2020) [43]	PW.CIM ^(c)	Centrifugation ^(b)	Not declared	RNeasy PowerBiofilm kit ^(e)
Salgar-Chaparro et al. (2020) [44]	PW.CIM ^(c)	Centrifugation ^(b)	Not declared	RNeasy PowerBiofilm kit ^(e)
Alhefeiti et al. (2021) [13]	OS.OCDM	Centrifugation ^(b)	20% glycerol reagent ^(h)	TRIzol reagent ⁽ⁱ⁾
Zhou et al. (2021) [45]	PW	Filtration ^(a)	95:5 v/v ethanol/trizol ^(h)	High Pure RNA Isolation kit ^(f)
Albahri et al. (2021) [18]	ND.CIM ^{(c),(i)}	Centrifugation ^(b)	Not declared	RNeasy PowerBiofilm kit ^(e)
Su et al. (2022) [20]	PW.HDM	Centrifugation ^(b)	Not declared	TRIzol reagent ^(d)
Zhou et al. (2022) [14]	OS.MET-HDM	Centrifugation ^(b)	Liquid nitrogen (n ₂)	acid phenol chloroform/ isoamyl alcohol reagent ^(h)
Prajapat et al. (2023) [46]	PW	Not declared	Not declared	RNeasy plant mini kit ^(e)
	IW	Not declared	Not declared	RNeasy plant mini kit ^(e)

Produced water (PW); injection water (IW); solid deposits (SD); oil (OIL); oily sludge (OS); corrosion-influencing microorganisms (CIM); hydrocarbon-degrading microorganisms (HDM); organic compounds degrading microorganisms (OCDM); sulfate-reducing bacteria (SRB); methanogenic hydrocarbon-degrading microorganisms (MET-HDM). Preprocessing: ^(a) before preserving the samples; ^(b) after preserving or not preserving the samples. ^(c) Description of CIM groups (Supplementary Table S7). Manufacturer of RNA extraction reagent and kits: ^(d) Thermo Fisher Scientific/Invitrogen; ^(e) QIAGEN; ^(f) Roche. ^(g) Mobi. ^(h) Not declared. ⁽ⁱ⁾ Sample type (inoculum) not available in the study, no data (ND).

In the laboratory, 15 studies used centrifugation as preprocessing to concentrate microbial cells, of which six were environmental samples (PW), six were microbial cultures (PW.HDM, IW.HDM, PW.SRB, OS.MET-HDM, and OS.OCDM), and three were consortia (PW.CIM and ND.CIM) (Table 3). To preserve the pellets generated from PW samples, a 95:5 v/v ethanol/ TRIzol solution was used (Table 3). On the other hand, studies of cultures PW.SRB, OS.OCDM, and OS.MET-HDM used different methods to preserve the pellets: RNAprotect bacteria reagent, 20% glycerol, and liquid nitrogen (N₂), respectively (Table 3). It is noteworthy that eight studies do not specify the use of RNA preservatives (Table 3). Furthermore, two studies [19,46] did not specify preprocessing and nucleic acid preservative agents (Table 3).

It is highlighted that SD samples [1] and OS samples [13] may present characteristics that make filtration and/or centrifugation a difficult process. As reported by Salgar-Chaparro and Machuca (2019) [1], deposit samples were collected from pipelines covered by approximately 3 cm of schmoos material. Zhou et al. (2022) [14] mention that the OS sample evaluated in their study was a mixture of water (27–46% *w/w*), crude oil (35–59% *w/w*), and sand (13–19% *w/w*).

For RNA extraction, studies used different methods, including commercial kits, TRIzol reagent, and acid phenol chloroform/isoamyl alcohol reagent (Table 3). Most studies ($n = 11$; 55%) opted for the use of a kit, describing seven types (Table 3). According to information provided by the manufacturers, these kits are specific for extracting RNA from bacteria and samples of water, soil, biofilm/microbiomes, and plants (Table 3). TRIzol reagent was used in four environmental sample studies (PW and IW) and four culture studies (PW.HDM, IW.HDM, OIL.HDM, and OS.OCDM) (Table 3). Zhou et al. (2022) [14] was the only one that used acid phenol chloroform/isoamyl alcohol reagent for RNA isolation from OS.MET-HDM samples (Table 3).

After RNA extraction, five studies did not specify the method of removing residual DNA [13,16,19,35,47]. Depletion of ribosomal RNA (rRNA) for enrichment of messenger RNA (mRNA) was described in six studies [4,13,14,16,19,35]. The synthesis of complementary DNA (cDNA) from total RNA or from mRNA was carried out in all studies. Liu et al. (2020) [47] was the only study that did not specify the preprocessing of RNA samples before sequencing.

3.4. Methods for Amplification and Sequencing RNA

Among the eligibility criteria considered, studies were included that carried out the amplicon sequencing of 16S rRNA and functional gene transcripts, as well as transcriptome and metatranscriptome of studied samples (Table 4). It was observed that amplicon sequencing of 16S rRNA transcripts was the most applied ($n = 8$; 40%). This sequencing was used to identify metabolically active microorganisms (bacteria and archaea) from environmental samples (PW and IW), cultures (PW.HDM and IW.HDM), and CIM consortia (PW.CIM and ND.CIM) (Table 4).

Primers directed at the Bacteria domain (827F-519R, 515F-907R, and 341F-806R) and the Archaea domain (341F-806R, A109F-A1041R, 344F-915R, and 524F10extF-Arch958RmodR) were used for 16S rRNA gene amplification (Table 4). The pair of primers 341F (5'-CCTAYGGGRBGCASCAG-3') and 806R (5'-GGACTACNNGGGTATCTAAT-3') were used in four of the eight studies that sequenced 16S rRNA gene (Table 4). It should be highlighted that in one study [46], isolated RNA from PW and IW samples was used solely for quantitative PCR (qPCR) based on the 16S rRNA gene (EUB341F-EUB534R) and functional genes *narG* (narG1575F/1748R) and *nirS* (nirS1189F/1376R) (Table 4).

Amplicon sequencing of functional gene transcripts *alkB* (AlkBFB/AlkBRB), *dsrA* (DSRAVibF/DSRAVibR and DSR-1Fdeg/PJdsr853Rdeg), *aprA* (aprA-1-FW/aprA-5-RV), *assA* (assA2F/assA2R) and *mcrA* (mlas-mod-F/mcrA-rev-R) were assessed in six studies, four of which were environmental samples (PW and IW) and two were culture studies (PW.HDM, IW.HDM, and PW.SRB) (Table 4). It is important to note that the clone library of functional gene transcripts was constructed from cDNA libraries obtained from evaluated samples [4,15,17,45,48,49].

In the study by Shestakova et al. (2011) [15], *alkB* gene transcripts were analyzed to investigate the microbial diversity and identify microorganisms responsible for hydrocarbon degradation in cultures enriched with PW (PW.HDM) and IW (IW.HDM) (Table 4). In contrast, two PW studies evaluated the sequencing of *mcrA* gene transcripts to identify methanogens (Table 4). In addition, the biomarker *assA* was assessed in three studies to detect the microorganisms responsible for the activation of alkane during anaerobic degradation processes.

Table 4. Methods for amplification and sequencing of RNA transcripts from the samples studied in the articles included in this systematic review.

Reference	Sample	Gene	Primer	Platform (Sequencer Model)
Shestakova et al. (2011) [15]	PW.HDM and IW.HDM	<i>alkB</i> 16S rRNA 16S rRNA	AlkBFB/AlkBRB Bact-827F/519R Arch-A109F/A1041R	ABI (3730)
Zapata-Penásco et al. (2016) [4]	PW.SRB	<i>dsrA</i>	DSRAVibF/DSRAVibR	ABI (310)
Nazina et al. (2017) [3]	IW	16S rRNA	Bact-827F/519R	ABI (3730)
Li et al. (2017) [17]	PW	16S rRNA 16S rRNA <i>aprA</i> <i>dsrA</i>	Bact-515F/907R Arch-344F/915R aprA-1-FW/aprA-5-RV' DSR-1Fdeg/PJdsr853Rdeg	NGS (MiSeq) NGS (MiSeq) ABI (377) ABI (377)
Liu et al. (2018) [16]	PW	Metatranscriptomic		NGS (MiSeq)
Salgar-Chaparro and Machuca (2019) [1]	PW, IW ^(a) and DS ^(b)	16S rRNA	Bact-341F/806R	NGS (MiSeq)
Zheng et al. (2019) [19]	OIL.HDM	Transcriptomic (<i>Bacillus licheniformis</i>)		NGS (NextSeq)
Liu et al. (2020) [35]	PW and PW.HDM	Metatranscriptomic		NGS (HiSeq X ten)
Liu et al. (2020) [48]	PW	<i>assA</i> <i>mcrA</i>	assA2F/assA2R mlas-mod-F/mcrA-rev-R	Sanger (not specified)
Liu et al. (2020) [47]	PW	Metatranscriptomic ^(c)		NGS (MiSeq)
Zhou et al. (2020) [22]	PW	16S rRNA 16S rRNA	Bact-515F/907R Arch-524F10extF/Arch958RmodR	NGS (MiSeq)
Liu et al. (2020) [49]	PW	<i>assA</i> <i>mcrA</i>	assA2F/assA2R mlas-mod-F/mcrA-rev-R	ABI (377)
Salgar-Chaparro et al. (2020) [43]	PW.CIM ^(f)	16S rRNA	Bact-341F/806R	NGS (MiSeq)
Salgar-Chaparro et al. (2020) [44]	PW.CIM ^(f)	16S rRNA	Bact-341F/806R	NGS (MiSeq)
Alhefeiti et al. (2021) [13]	OS.OCDM	Transcriptomic (<i>Bacillus cereus</i>)		NGS (not specified)
Zhou et al. (2021) [45]	PW	<i>mcrA</i>	mlas-mod-F/mcrA-rev-R	ABI (377)
Albahri et al. (2021) [18]	ND.CIM ^{(e),(f)}	16S rRNA	Bact-341F/806R	NGS (MiSeq)
Su et al. (2022) [20]	PW.HDM	Transcriptomic (<i>Exiguobacterium aurantiacum</i> SW-20)		NGS (HiSeq X ten)
Zhou et al. (2022) [14]	OS.MET-HDM	Metatranscriptomic		NGS (NovaSeq 6000)
Prajapat et al. (2023) [46]	PW and IW	16S rRNA <i>narG</i> <i>nirS</i>	EUB341F/EUB534R narG 1575F/narG 1748R nirS 1189F/nirS 1376R	Unrealized ^(d)

Produced water (PW); injection water (IW); solid deposits (SD); oil (OIL); oily sludge (OS); corrosion-influencing microorganisms (CIM); hydrocarbon-degrading microorganisms (HDM); organic compounds degrading microorganisms (OCDM); sulfate-reducing bacteria (SRB); methanogenic hydrocarbon-degrading microorganisms (MET-HDM). Next Generation Sequencing (NGS), Illumina company. ^(a) Despite several attempts with modified conditions, the extraction of high-quality RNA from the IW sample or cDNA synthesis from SD samples failed [1]. ^(b) Description of CIM groups (Supplementary Table S7). ^(c) Metatranscriptomic data was not disclosed. ^(d) Isolated RNA was used for quantitative PCR (qPCR). ^(e) Sample type (inoculum) not available in the study, no data (ND). ^(f) Description of CIM groups (Supplementary Table S7).

Two studies were identified that investigated sulfate-reducing microorganisms based on *dsrA* and *aprA* transcripts (Table 4). In one of these studies, conducted by Zapata-

Peñasco et al. (2016) [4], the *dsrA* gene was amplified using the pair of primers DSRVibF (5'-CGGCGTTATCGGCCGTTACTG-3') and DSRVibR (5'-GA[A/G]CCCGAACC GCCGAGGTCGG-3'), designed specifically to recover sequences from the Desulfovibrionales order, obtained from SRB cultures (PW.SRB) (Table 4). In another study conducted by Li et al. (2017) [17], the microbial diversity and composition of the sulfate-reducing community were analyzed, using transcripts from the *aprA* and *dsrA* genes of PW samples (Table 4).

In studies of cultures enriched with oil samples (OIL.HDM) [19], oily sludge (OS.OCDM) [13] and PW (PW.HDM) [20], RNA from the bacterial species *Bacillus licheniformis*, *Bacillus cereus*, and *Exiguobacterium aurantiacum* SW-20, respectively, were isolated for transcriptome analysis (Table 4). In four other studies, the metatranscriptomes of environmental samples (PW) and cultures enriched with PW samples (PW.HDM) and oily sludge (OS.MET-HDM) (Table 4) were analyzed. In the study by Liu et al. (2020) [47], the metatranscriptomic data were not disclosed.

The studies reported the use of two sequencing methods, first-generation (Sanger) and next-generation sequencing (NGS) (Table 4). Sanger sequencing was applied in all six studies that assessed the sequences of functional gene transcripts (*alkB*, *dsrA*, *aprA*, *assA*, and *mcrA*) using ABI310, ABI377, and ABI3730 sequencers (Table 4). The studies by Shestakova et al. (2011) [15] and Nazina et al. (2017) [3] also used Sanger sequencing to assess 16S rRNA gene sequences (Table 4). On the other hand, 13 studies used the NGS method to sequence 16S rRNA gene transcripts, transcriptomes, and metatranscriptomes from environmental samples (PW), cultures (OIL.HDM, PW.HDM, OS.OCDM and OS.MET-HDM), and consortia (PW.CIM and ND.CIM), using MiSeq, HiSeq X ten, NextSeq, and NovaSeq 6000 sequencers (Table 4). The NGS MiSeq sequencer was the most frequently used, being mentioned in eight studies that sequenced 16S rRNA gene transcripts, as well as the metatranscriptome (Table 4). In contrast, the NextSeq [19], HiSeq X ten [20,35], and NovaSeq 6000 [14] sequencers were used solely for transcriptome and metatranscriptome sequencing (Table 4).

3.5. Composition of the Metabolically Active Microbiota

The composition of metabolically active microorganisms based on RNA was described in eight studies through the analysis of 16S rRNA gene transcript amplicons (Supplementary Table S6). These studies covered environmental samples (PW and IW) [1,3,17,22], cultures (PW.HDM and IW.HDM) [15], and a CIM consortium (PW.CIM and ND.CIM) [18,43,44].

After normalization, data from the environmental samples, cultures, and consortia were grouped as follows: (1) PW (PW, PW.HDM, and PW.CIM); (2) IW (IW and IW.HDM); and (3) ND.CIM (Figure 3). This approach was adopted because, in cultures, PW and IW samples were used as an inoculum to maximize the growth of specific microbial groups. Also, different cultures enriched with PW and ND were used to form CIM consortia (Supplementary Table S7). It is important to mention that the ND sample was not grouped with other samples, as it cannot be confirmed whether the fluid was PW, OIL or a PW/OIL mixture.

The aggregated data from PW, IW, and ND (Figure 3, Supplementary Table S6) reveals patterns of sharing and exclusivity, and a Venn diagram analysis was performed to highlight the number of shared and exclusive phyla and genera of metabolically active microorganisms.

The PW sample had a larger number of phyla ($n = 20$) compared to IW samples ($n = 7$) and ND ($n = 3$) (Figure 3). In the comparative analysis, it was observed that PW and IW samples had 13 and two exclusive phyla, respectively. On the other hand, PW and IW shared five bacterial phyla (Bacteroidetes, Deferribacteres, Firmicutes, Ignavibacteriae, and Proteobacteria) (Figures 3 and 4, Supplementary Table S6). Also, the ND sample shares three phyla with PW, two bacterial (Synergistetes and Firmicutes), and one archaeal (Euryarchaeota) (Figures 3 and 4, Supplementary Table S6). No shared phyla were observed between ND and IW (Figure 3).

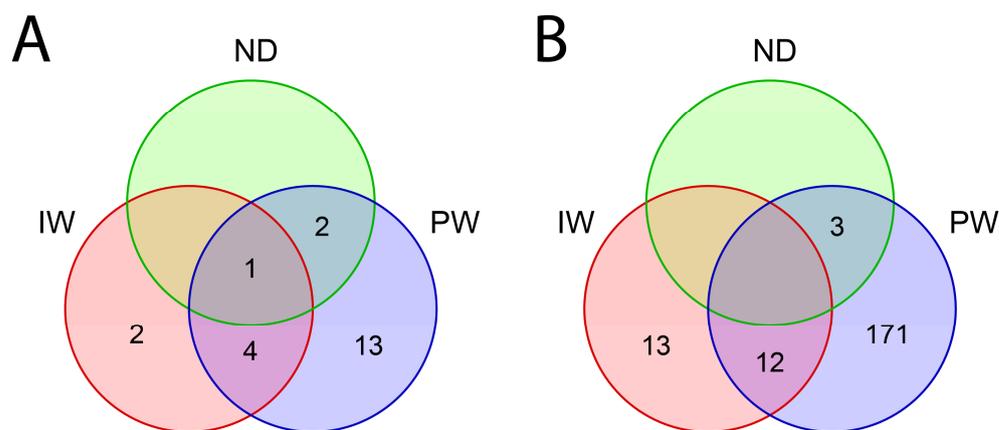


Figure 3. Venn diagram presents the number of shared and exclusive (A) phyla and (B) genera of metabolically active microorganisms based on RNA obtained from the analyses of 16S rRNA gene transcript amplicons. Sample type: produced water (PW) and injection water (IW), no data (ND). ND: not available in the study.

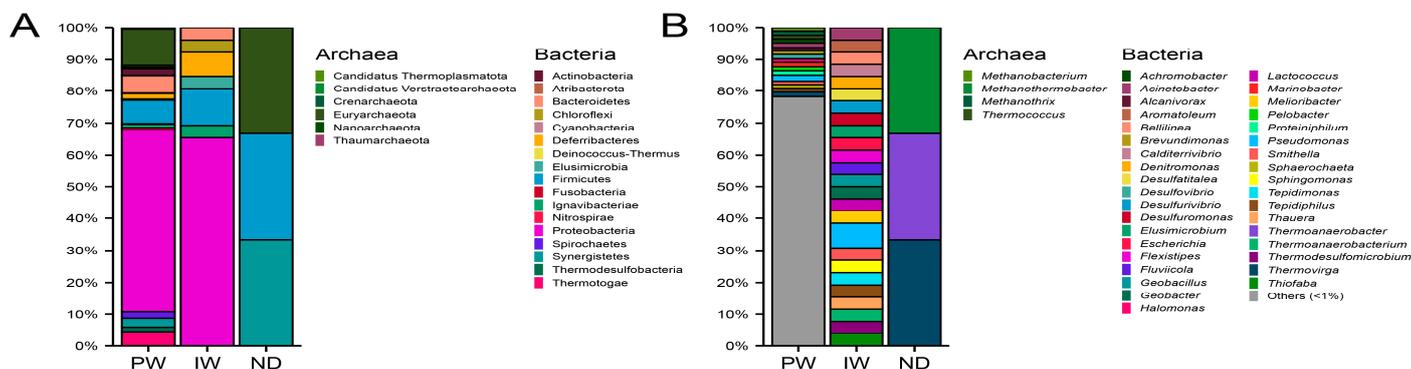


Figure 4. Relative abundance of metabolically active microorganisms based on RNA at the (A) phylum and (B) genus levels obtained from 16S rRNA gene transcript amplicon data. Sample type: produced water (PW) and injection water (IW), no data (ND). ND: not available in the study.

As mentioned earlier, PW samples presented a larger quantity of phyla. Consequently, a larger number of genera was identified in PW ($n = 186$), followed by IW ($n = 25$) and ND ($n = 3$) (Figure 3). PW and IW share 12 bacterial genera (Figures 3 and 4). On the other hand, PW and ND share three genera, two belonging to the Bacteria domain (*Thermoanaerobacter* and *Thermovirga*) and one belonging to the Archaea domain (*Methanothermobacter*) (Figures 3 and 4, Supplementary Table S6). Unique genera were also identified, being 171 in PW and 13 in IW (Figure 3, Supplementary Table S6). On the other hand, no exclusive genera were observed in the ND sample (Figure 3).

The phylum Proteobacteria was predominant in PW (57%) and IW (65%) samples, followed by Euryarchaeota (PW: 11%), Deferribacteres (IW: 8%), Firmicutes (PW: 12% and IW: 7%), and Bacteroidetes (PW: 6% and IW: 4%) (Figure 4, Supplementary Table S6). On the other hand, only the Euryarchaeota, Firmicutes, and Synergistetes phyla were identified in the ND sample, with similar abundances of 33% (Figure 4, Supplementary Table S6).

For the analysis at the genus level, the relative abundances above one (>1%) were considered. Therefore, genera with abundances lower than 1% were grouped and named as "others" (Figure 4). The Bacteria domain was predominant in PW (99%), IW (100%), and ND (67%) samples (Figure 4). The *Acinetobacter*, *Pseudomonas*, *Smithella*, and *Tepidiphilus* genera, all belonging to the Bacteria domain, were identified in PW and IW samples (Figure 4, Supplementary Table S6). The genus *Pseudomonas* showed a higher relative abundance (PW: 2%, IW: 8%) (Figure 4). The *Methanothermobacter*, *Thermoanaerobacter*, and *Thermovirga* genera were observed in the ND sample and demonstrated similar abundances of 33%.

Among them, the genus *Methanothermobacter* is the only representative of the Archaea domain (Figure 4).

Metatranscriptomic data were analyzed in three studies. These evaluated environmental samples (PW) [16] and cultures (PW.HDM and OS.MET-HDM) [14,35]. Metatranscriptome data were downloaded from the NCBI [16,35] and NODE databases [14]. The composition of the metabolically active community was evaluated separately in each study since raw data were used (Figure 5, Supplementary Table S6).

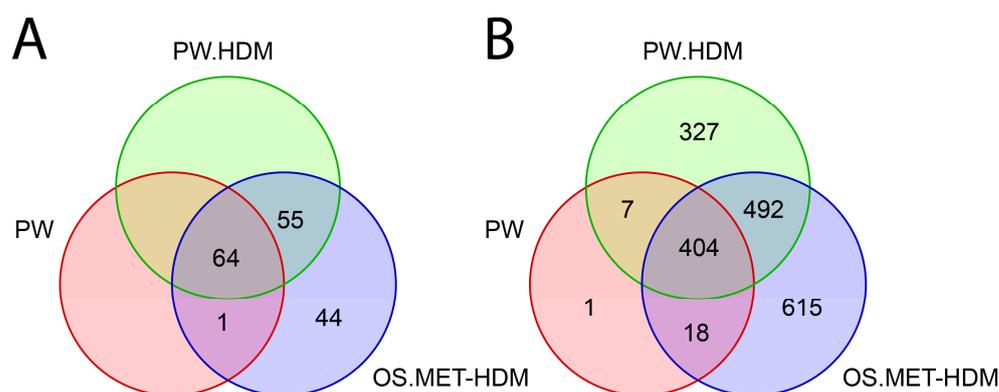


Figure 5. The Venn diagram shows the number of shared and exclusive (A) phyla and (B) genera of metabolically active microorganisms based on RNA obtained from metatranscriptomic data. Produced water (PW); PW hydrocarbon-degrading microorganisms (PW.HDM) and Oily sludge (OS) methanogenic hydrocarbon-degrading microorganisms (OS.MET-HDM).

Using metatranscriptomic data, a Venn diagram was made to show the number of shared and exclusive phyla and genera of metabolically active microorganisms detected in PW, PW.HDM, and OS.MET-HDM samples (Figure 5, Supplementary Table S6). A total of 65, 119, and 164 phyla were identified in the analysis of PW, PW.HDM, and OS.MET-HDM samples, respectively. Of the 65 phyla of PW, 64 are shared with PW.HDM (Figure 5, Supplementary Table S6). The OS.MET-HDM sample presented 44 exclusive phyla. However, it shared 65 phyla with PW and 119 phyla with PW.HDM (Figure 5, Supplementary Table S6). At the genus level, a total of 1864 genera were observed (Figure 5, Supplementary Table S6). Of these, 1864 in PW, 327 in PW.HDM and 615 in OS.MET-HDM samples were exclusive.

The relative abundance of phyla identified in PW, PW.HDM, and OS.MET-HDM samples was also evaluated using metatranscriptome data. The Proteobacteria phylum, belonging to the Bacteria domain, had the highest abundance in the PW sample (59%), while the Euryarchaeota phylum, belonging to the Archaea domain, was more abundant in the PW.HDM (46%) and OS.MET-HDM (27%) samples (Figure 6, Supplementary Table S6).

Similar to what was done with the 16S rRNA gene data, relative abundances above one (>1%) were considered for the analysis at the genus level, while the rest were grouped and named as “others” (<1%) (Figure 6, Supplementary Table S6). In the PW sample, the most abundant genera were *Pseudomonas* (20%) and *Acinetobacter* (17%), whereas in the PW.HDM sample, the predominant genera were *Thermococcus* (29%) and *Methanothermobacter* (11%) (Figure 6, Supplementary Table S6). On the other hand, in the OS.MET-HDM sample, the most abundant genera were “*Candidatus Methanoliparum*” (22%) and *Methanothermobacter* (18%) (Figure 6, Supplementary Table S6).

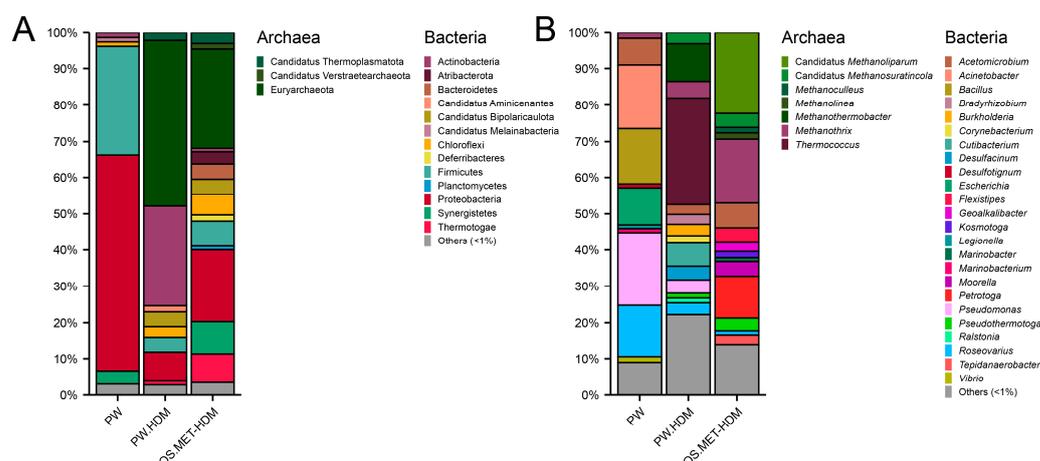


Figure 6. Relative abundance of metabolically active microorganisms based on RNA at the (A) phylum and (B) genus level obtained from metatranscriptomic data. Sample type: Produced water (PW); PW hydrocarbon-degrading microorganisms (PW.HDM) and Oily sludge (OS) methanogenic hydrocarbon-degrading microorganisms (OS.MET-HDM).

4. Discussion

4.1. Identification and Selection of Studies

The present systematic review was based on the PRISMA guidelines and aimed to evaluate the methods applied in studies based on RNA that analyzed metabolically active microbial communities from oilfields. In the four selected databases (MEDLINE/PubMed, PubMed Central, Scopus and Web of Science), 2561 potentially eligible records were identified. After automated and manual screening, only 20 eligible original articles were included, highlighting the scarcity of studies directed at the analysis of 16S rRNA transcripts and functional genes, as well as transcriptome and metatranscriptome of samples of reservoirs and oil industry facilities. It is noteworthy that, in recent decades, several studies have investigated microbial communities from the oil industry [10,23–28,30,31]. However, most of these analyses are primarily focused on DNA sequences.

Probably, the limited number of studies directed at active microorganisms is attributed to the challenges associated with preserving and isolating high-quality RNA [22]. RNA is an inherently unstable molecule and highly vulnerable to degradation caused by RNases [30]. This instability represents a significant challenge in obtaining reliable results for the characterization of active microbial communities.

4.2. Studies Included

The studies analyzed environmental samples (PW, IW, OIL, and SD) collected from oilfields in Australia, China, India, Mexico, and the United Arab Emirates. The majority of evaluated samples were from PW, which is the most abundant waste stream of the oil industry and, consequently, one of the most available sources of samples in oilfield systems [2,31]. Salgar-Chaparro and Machuca (2019) [1] were the only authors who investigated both the planktonic microorganisms in circulating fluids and the sessile microbiota adhered to the inner walls of the pipelines. The comprehensive analysis of both fluids (PW, IW, and OIL) and solids (sediments, sludge, biofilm, pig residue, or similar) provides complementary insights into microbial activities primarily involved in hydrocarbon biodegradation and pipeline biocorrosion processes [21].

Some studies have examined microorganisms involved in the degradation of organic compounds (OCDM) [13] and hydrocarbons (HDM) [14,15,19,20,35], as well as corrosion-influencing microorganisms (CIM) [4,18,43,44], through the analysis of enriched cultures inoculated with environmental samples (PW, IW, OIL, and OS). The oil industry has a vested interest in monitoring HDM and CIM [21] since these groups can cause significant economic and environmental damage.

Therefore, the analysis of metabolically active microbial communities using RNA sequencing provides valuable insights into identifying microorganisms involved in hydrocarbon biodegradation and corrosion in oil industry facilities [16,35,43,44,47]. High-quality RNA sequencing allows for comprehensive coverage and characterization of active microorganisms [32], but the successful application of these methods depends primarily on effective protocols to preserve RNA integrity [30,32].

4.3. RNA Preprocessing, Preservation, and Extraction Method

Despite the implementation of effective protocols for sample collection, transportation, and temporary storage, the use of nucleic acid preservatives is increasingly employed, particularly for samples collected from remote areas such as offshore petroleum platforms [22]. In the studies included in the review, various preservation methods were identified. The 95:5 *v/v* ethanol/TRIZol solution was mentioned in eight out of 20 studies. However, information regarding the impact of preservatives on the RNA integrity of oilfield samples remains limited.

The evaluated studies employed various methods for RNA isolation, including the use of commercial kits, TRIZol reagent, and acid phenol chloroform/isoamyl alcohol reagent (the TRIZol reagent was utilized in eight studies). While nucleic acid extraction protocols should be tailored to the specific characteristics of each sample [32], the TRIZol method stands out as a rapid, accessible, and cost-effective protocol [50]. Its favorable attributes likely contribute to its widespread use for RNA isolation across diverse sample types, such as medicinal plants [51], human visceral adipose tissue [50], the SARS-CoV-2 virus [52], and fungi [53].

It is worth noting that filtration and centrifugation processes were mentioned in four studies and 14 studies, respectively. While aqueous fluids such as PW and IW are commonly sampled in petroleum systems, these samples typically exhibit low biomass and diversity [31,54]. To address this limitation, larger sample volumes are collected and subjected to preprocessing steps to concentrate biomass and facilitate the isolation of nucleic acids [1,10,17,31,45,54].

4.4. RNA Amplification and Sequencing Method

Among the molecular methods mentioned in the studies, sequencing of 16S rRNA gene transcripts was the most implemented approach for studying the active microbiota. However, there is a challenge in interpreting the data due to variations in the number of copies of the 16S rRNA gene among species, ranging from one to 15 in Bacteria and from one to four in Archaea [55,56]. Some studies suggest that the number of ribosomes increases in actively growing cells, and ribosomal RNA analysis can be used to identify metabolically active cell forms [55,56]. However, evidence indicates that the use of rRNA as an indicator of microbial activity has limitations.

Blazewicz et al. (2013) [57] reviewed several studies and concluded that (a) rRNA concentration and growth rate are not always correlated, (b) the relationship between rRNA concentration and growth rate can vary significantly among taxa, and (c) dormant cells can contain a large number of ribosomes. While there are inherent uncertainties in 16S rRNA gene sequencing, and DNA-based analysis cannot differentiate active species among all those present in the environment, the comparison of DNA and RNA libraries provides a more comprehensive characterization of microorganisms in petroleum environments [1,3,17,22,48,49].

One study in the review utilized the sequencing of the functional gene *alkB*, which codes for the enzyme alkane hydroxylase, to investigate microorganisms involved in the aerobic degradation of hydrocarbons [15]. In contrast, three studies sequenced the functional genes *assA* and *mcrA* to analyze microorganisms associated with the anaerobic degradation of hydrocarbons and methanogenesis, respectively. The *assA* gene has been employed as a biomarker for detecting microorganisms responsible for the initial activation of alkanes during the anaerobic degradation process [48,49]. Meanwhile, the *mcrA* functional gene is widely recognized as a biomarker for identifying methanogens [45,48,49].

The HDM can use hydrocarbons as a source of energy and carbon [58,59], preferably alkanes, cyclic and aromatic, and they metabolize organic and carbon compounds, drastically reducing oil quality [45,60]. With increasing levels of biodegradation, the content of asphaltenes and resins, acidity, and oil viscosity increase, while the content of saturated and aromatic hydrocarbons decreases [60,61]. These alterations have a negative effect on oil production by reducing the flow rates from reservoirs, as well as refining operations, and increasing process costs [61,62].

In turn, CIM represents another significant concern for the oil industry since biocorrosion processes involve the degradation of metallic materials caused by the presence and activity of various microorganisms that are in direct or close contact with the metal surface [43,63]. It is estimated that MIC contributes to nearly 40% of internal corrosion problems and 20–30% of external pipeline corrosion [63]. Corroborating the data obtained in this review, sulfate-reducing bacteria (SRB) have been widely cited in studies related to pipeline corrosion [4,10]. Functional genes such as *aprA* and *drsA*, encoding the enzymes adenosine-5-reductase and dissimilatory sulfate reductase, respectively, are used to track SRB [4,17].

Metatranscriptomic analysis has been evaluated with data from three studies. This analysis is considered a powerful technique that enables the examination of the gene expression profile of microorganisms present in a sample, providing valuable insights into their metabolic activity in a given environment [64]. Despite the advancements in microbiome studies of oilfield samples, there is still a notable scarcity of studies analyzing the metatranscriptome, and this can be attributed to several challenges, including the extraction of high-quality RNA, the presence of inhibitors, limited quantities of genetic material in the samples, and rapid RNA degradation. Additionally, the collection and proper preservation of samples for metatranscriptomic analysis can be complex due to the demanding conditions of the petroleum environment, such as high pressure and temperature, as well as the presence of toxic compounds.

4.5. Composition of the Metabolically Active Microbiota

In total, eight studies performed amplicon sequencing of 16S rRNA gene transcripts [1,3,15,17,18,22,43,44] and three performed metatranscriptome analysis [14,16,35]. The phylum Proteobacteria was predominant in the analysis of 16S rRNA gene data from PW (57%) and IW (65%) samples (Figure 4), as well as in the metatranscriptome analysis of PW samples (59%) (Figure 6). This phylum comprises one of the largest divisions within prokaryotes and represents the vast majority of known gram-negative bacteria, consisting of more than 200 genera [65].

When analyzing data at the genus level, higher relative abundances of the genera *Pseudomonas* (16S rRNA = PW: 2%; IW: 8%; and metatranscriptome = PW: 20%) and *Acinetobacter* (16S rRNA = PW: 1%; IW: 4%; and metatranscriptome = PW: 17%) were observed. These genera belong to the phylum Proteobacteria and are often found in oil environments and can degrade complex hydrocarbons into simpler compounds, reducing oil quality [3]. It is worth mentioning that some genera degrade n-alkanes, while others degrade aromatic compounds [66]. However, *Pseudomonas* can degrade both classes of hydrocarbons [3,67] due to the presence of specific genes, such as *alkB*, *xylXYZ* and *benABCD* [15,68].

Analyzing the metatranscriptome from culture samples, it was observed that the phylum Euryarchaeota showed greater abundance (PW.HDM: 46% and OS.MET-HDM: 27%), a phylum affiliated with the Archaea domain, which harbors hydrogenotrophic and acetotrophic methanogens microorganisms [69]. In oil reservoirs, under methanogenic conditions, hydrocarbon degradation occurs when communities of bacteria and archaea cooperate syntrophically to produce methane through thermodynamically favorable pathways [70].

Analyzing metatranscriptomic data at the genus level, it was observed that in the samples from the PW.HDM cultures, the most abundant genera were *Thermococcus* (29%) and *Methanothermobacter* (11%). While in the study that evaluated OS.MET-HDM cultures,

the most abundant genera were “*Candidatus Methanoliparum*” (22%) and *Methanotherix* (18%). The genera *Thermococcus*, *Methanothermobacter*, and *Methanotherix* belong to the phylum Euryarchaeota and are often detected in oil reservoirs [45].

5. Conclusions

This systematic review reveals the limited research on RNA-based analysis of metabolically active microbial communities in oilfields. Despite substantial prior research on microbial communities in the oil industry, there is a significant gap in the analysis of 16S rRNA transcripts, functional genes, transcriptomes, and metatranscriptomes. It is believed that challenges related to RNA preservation and isolation have hindered progress in this area. The selected studies predominantly focus on produced water samples, highlighting the need for a broader exploration of various sample types within oilfield systems. Additionally, metatranscriptomic analysis remains underutilized due to its technical complexities. The predominance of Proteobacteria and specific genera like *Pseudomonas* and *Acinetobacter* underscores the importance of studying these microorganisms’ roles in hydrocarbon degradation. Furthermore, the presence of Euryarchaeota points to the significance of methanogenic processes in oil reservoirs. This review serves as a call to action for further research in this area, emphasizing the importance of standardizing RNA preservation, extraction, and sequencing methods. Advancements in RNA-based analysis have the potential to significantly enhance our understanding of active microbial communities in oilfields and their impacts on the industry.

Supplementary Materials: The following supporting information can be downloaded at: <https://www.mdpi.com/article/10.3390/applmicrobiol3040079/s1>. Table S1: PRISMA-S checklist. Table S2: Terms used in the search for records in databases and selection of documents by automation tools. Table S3: Records identified in the databases. Table S4: Automated and manual screening of records. Table S5: References, title, and DOI of the studies included in the present systematic review Table S5. Composition of the metabolically active microorganisms. Table S6: References, title, and DOI of the original articles are included in the present systematic review. Table S7: Description of the microbial groups corrosion-influencing microorganisms.

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