

Short Note

# Assessing the Influence of Roads on Fire Ignition: Does Land Cover Matter?

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**Abstract:** In human-affected fire environments, assessing the influence of human activities on the spatial distribution of wildfire ignitions is of paramount importance for fire management planning. Previous studies have shown that roads have significant effects on fire ignition. However, since different land cover classes are subject to different levels of ignition risk, roads in different land cover classes may differently affect fire ignition. The aim of this paper is thus to assess the influence of roads on fire ignition in selected land cover classes subjected to different levels of anthropogenic pressure in Sardinia (Italy). Our results show that fires are preferentially ignited close to roads in all land cover classes. However, the influence of roads is much stronger in less impacted land uses, where the availability of human-induced ignitions highly depends on the accessibility networks. Our approach represents a first step towards the systematic integration of interacting fire ignition drivers such as roads and land cover into fire risk analysis.

**Keywords:** distance to roads; fire selectivity; human pressure; land use/land cover; Sardinia

## 1. Introduction

Fire characteristics such as type, intensity, spread, and temporal distribution depend on several factors such as climate, fuel load and continuity and ignition sources [1]. In this respect, humans play a manifold role, acting indirectly as fuel shapers through land use and related vegetation, as well as directly by setting and controlling fire [2]. As a consequence, human influence on fire regimes highly varies from region to region worldwide [3].

Several studies demonstrated that in the Mediterranean areas intentional or accidental human-induced fire ignitions account for more than 90% of forest fires [4], thus far exceeding natural fires [5,6]. In such human-affected fire environments, a detailed understanding of the influence of anthropogenic activities on the temporal and spatial distribution of wildfires is a crucial factor for fire management planning [7–9]. Given the complexity of the human component of wildfires, quantitative analyses of fire risk have generally paid less attention to the socioeconomic aspects of fire occurrence compared to the biological and physical factors, such as elevation, slope, temperature, rainfall, or vegetation [10–15]. In recent years, however, fire risk studies have started to incorporate human-related variables that directly or indirectly influence spatial patterns of fire occurrence. At the landscape scale, a number of authors showed that fire ignitions exhibit strong preferences for given land

use/land cover (LULC) types [16–20]. LULCs are the result of human activities, such as agriculture, farming, silviculture, recreation, etc. and are subjected to a different fire ignition risk depending on fuel load and human-induced ignitions. The intrinsic fire selectivity of each LULC is further modulated by the presence of infrastructure such as housing or roads, which are also well known to influence fire ignition [21–29]. However, while LULC data and distance from roads have been already used for modeling fire ignition risk e.g., [30–33], to the best of our knowledge, a detailed analysis of the combined effect of road network and land cover is provided here for the first time.

The objective of this paper is thus to assess the influence of roads on fire ignition in selected land cover classes subjected to different levels of anthropogenic pressure in Sardinia (Italy). To achieve this aim, we referred to a selectivity index according to which certain resources (i.e., land cover classes) are more fire-prone than others, being differentially selected by fire, depending on fuel load, structure, spatial continuity and human activities [5,16–18]. Fire is considered selective towards a resource (i.e., land cover type) when such resource is used disproportionately to their availability [16,17]. By integrating the selectivity behaviour of fire towards LULC and the road network, this paper can provide insights into the interaction between these two fire drivers.

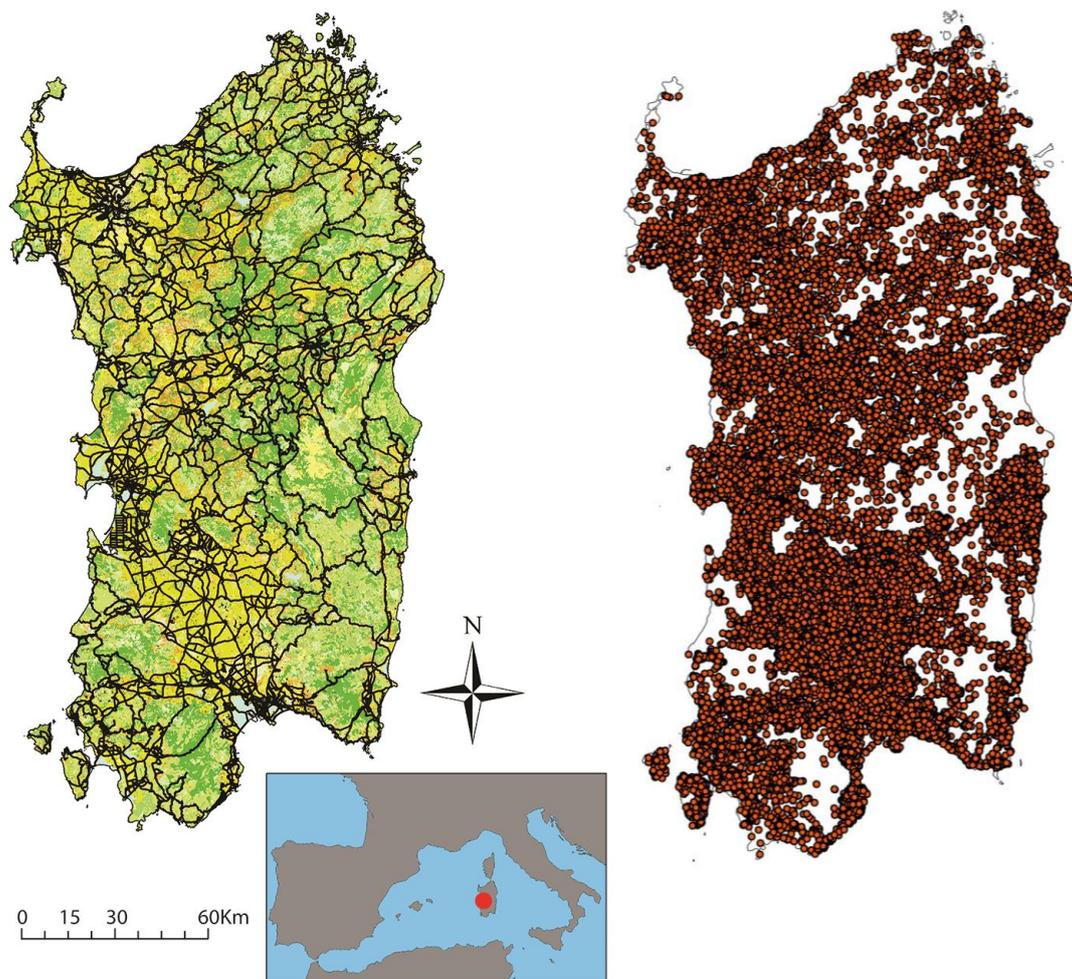
## 2. Materials and Methods

### 2.1. Study Area

Sardinia is the second largest island of the Mediterranean basin, with a total surface area of approximately 24,000 km<sup>2</sup> (Figure 1). The island is characterized by a prevalently hilly topography and extreme heterogeneity in geological and geomorphological features. The highest peak is Punta La Marmora (1834 m a.s.l.), located in central Sardinia. The climate of Sardinia is typically Mediterranean with a mild and rainy period from October to May and a warm and dry period from June to September when temperatures may reach peaks above 30 °C. The average annual amount of precipitation ranges from 500 mm along the coasts to more than 1200 mm in the higher regions [34].

Sardinia hosts a very high ecosystem diversity, with habitats ranging from coastal sand dunes to extensive maquis and forests in the mountains. Land use is shaped by a long human history. Urban areas cover 3% of Sardinia, and the two main cities of Cagliari and Sassari are where most of the 1.7 million inhabitants are concentrated. Agricultural areas represent 25% of the island, whereas approximately 15% is covered by broadleaf forests, mainly composed of Mediterranean oaks, such as *Quercus ilex*, *Q. suber* and *Q. pubescens*. Mediterranean maquis and garrigue cover an additional 23%.

Sardinia is also considered to be a wildfire hotspot, with more than 2500 events per year on average [35]. Like in many other regions of southern Europe, the majority of fires directly originate from human activities. However, in recent decades the pronounced summer aridity combined with the abandonment of the traditional agriculture and the resulting progressive built up of fuel increased the risk of large and severe fires. Mean yearly burnt area is approximately 14,500 ha. More than 86% of fires occur from June to September with a peak in July. Fire size ranges over many orders of magnitude from 0.01 ha to >1000 ha, although less than 1% of total events get larger than 100 ha. However, fires > 100 ha account for roughly 51% of the total area burned in Sardinia between 2000 and 2015. Detailed statistics on fire causes are not available. Like in most Mediterranean regions, we can however assume that almost all fire events are of anthropogenic origin.



**Figure 1.** The study area of Sardinia with the road network used in this paper and the fire ignitions recorded during 2000–2015.

## 2.2. Data Preprocessing

A dataset of 45,976 wildfire events that occurred in Sardinia between 2000 and 2015 was extracted from the digital database of the Regional Forest Service (Figure 1). For each record, the dataset includes the date and the geographic coordinates of fire ignition and a field estimate of the burned area. The ignition points of the most recent fires (from 2006 onwards) are usually accurately geolocated with GPS, while the positional accuracy of the older ignition points is more uncertain.

For analyzing the relationship between fires, land cover and road network, we used a road network map containing all state, regional and county roads of Sardinia, together with a LULC map of the island at scale 1:25,000. The LULC map was built by the Regional Administration of Sardinia according to the standard classification methodology of the CORINE Land Cover legend [36] based on 2003–2004 digital orthophotos and high-resolution Ikonos satellite images (2005–2006) combined with field surveys. Both maps were downloaded from the Regional geographic information portal of Sardinia ([www.sardegnageoportale.it](http://www.sardegnageoportale.it)).

We next selected eight land cover classes with a sufficient number of fire records for studying fire incidence patterns at the landscape scale: (i) discontinuous urban areas, (ii) arable land, (iii) mixed agriculture, (iv) olive groves, (v) deciduous forests, (vi) maquis, (vii) garrigue, (viii) pastures. The selected classes were grouped into two main categories with different levels of anthropogenic pressure according to the CORINE Land Cover classification [36]: Artificial and agricultural classes (i–iv) and Natural and semi-natural classes (v–viii).

### 2.3. Data Analysis

We first created a sequence of overlapping buffers around the road network, ranging from 0 to 200 m in four steps of 50 m each (i.e., 0–50 m, 0–100 m, 0–150 m and 0–200 m). The buffers were then intersected with the selected LULCs to calculate the area  $A_{kb}$  of each land cover class  $k$  within each buffer  $b$ . Next, the number of fires  $N_{kb}$  within each land cover class in each buffer were counted (Table 1).

**Table 1.** Number of fires and areal extent of the selected land cover classes within each buffer.

Land Cover Class		Number of Fires within Each Buffer				Total Number of Fires within Each Class
		0–50 m	0–100 m	0–150 m	0–200 m	
<b>Artificial and agricultural</b>	Discontinuous urban areas	365	624	793	931	1366
	Arable land	1829	3332	4507	5429	10,599
	Mixed agriculture	1023	1888	2496	3025	5889
	Olive groves	279	512	711	893	2012
<b>Natural and semi-natural</b>	Deciduous forests	320	581	789	943	2113
	Maquis	332	596	788	955	2443
	Garrigue	230	442	609	755	1916
	Pastures	278	526	684	827	1951
<b>Total number of fires within each buffer</b>		7326	13,499	18,272	22,234	
Land Cover Class		Class Area within Each Buffer (ha)				Total Class Area (ha)
		0–50 m	0–100 m	0–150 m	0–200 m	
<b>Artificial and agricultural</b>	Discontinuous urban areas	4536	8143	10,983	13,238	29,899
	Arable land	29,714	57,249	82,805	106,326	415,700
	Mixed agriculture	14,791	28,168	40,264	51,162	199,000
	Olive groves	4353	8458	12,290	15,803	48,780
<b>Natural and semi-natural</b>	Deciduous forests	9493	18,703	27,653	36,509	351,400
	Maquis	9052	17,666	25,998	34,235	345,716
	Garrigue	5760	11,356	16,872	22,457	220,000
	Pastures	5544	10,794	15,725	20,477	144,800
<b>Total area of each buffer (ha)</b>		123,721	238,490	345,267	444,967	

The actual number of fires  $N_{kb}$  was then compared with the expected number of fires assuming complete spatial randomness in fire distribution within each land cover class ( $CSR_k$ , implying that distance to roads does not influence the spatial distribution of fires within single LULC classes) and within each buffer ( $CSR_b$ , implying that distance to roads does not influence the relative proportion of fires among different LULC classes). For land cover class  $k$ , the number of fires in a given buffer  $b$  expected under  $CSR_k$  is obtained by multiplying the total number of fires in that land cover class  $N_{k+}$  by the proportional area of that class within the buffer relative to the total area of  $k$  over the entire study area ( $p_{k'b}$ ). The expected number of fires  $E_{k'b}$  in class  $k$  within buffer  $b$  is then

$$E_{k'b} = p_{k'b} \times N_{k+} \tag{1}$$

Likewise, for a given land cover class  $k$ , the number of fires in a given buffer  $b$  expected under  $CSR_b$  is obtained by multiplying the total number of fires in that buffer  $N_{+b}$  by the proportional area of that class within the buffer relative to the total area of buffer  $b$  ( $p_{kb'}$ ). The expected number of fires  $E_{kb'}$  in class  $k$  under  $CSR_b$  is then

$$E_{kb'} = p_{kb'} \times N_{+b} \tag{2}$$

To compare the actual number of fires  $N_{kb}$  in class  $k$  within buffer  $b$  with the corresponding expected values  $E_{k'b}$  and  $E_{kb'}$ , we finally calculated the normalized selectivity indices

$$\sigma_{k'b} = (N_{kb} - E_{k'b}) / (N_{kb} + E_{k'b}) \tag{3}$$

And

$$\sigma_{kb'} = (N_{kb} - E_{kb'}) / (N_{kb} + E_{kb'}) \tag{4}$$

$\sigma_{k'b}$  thus summarizes the degree of fire clustering of a given land cover class  $k$  in a given buffer  $b$  compared to the fire distribution within the whole land cover class, while  $\sigma_{kb'}$  summarizes the degree of fire clustering of class  $k$  in buffer  $b$  compared to the fire distribution within the whole buffer. Both indices range from  $-1$  to  $+1$ . Values of  $\sigma > 0$  represent areas where fire occurs more frequently than expected under complete spatial randomness in fire distribution within each land cover class ( $\sigma_{k'b}$ ) or within each buffer ( $\sigma_{kb'}$ ), meaning that these areas are positively selected by fire. By contrast, values of  $\sigma < 0$  represent areas where ignitions are less frequent than expected under complete spatial randomness, meaning that these areas are negatively selected by fires [18].

### 3. Results

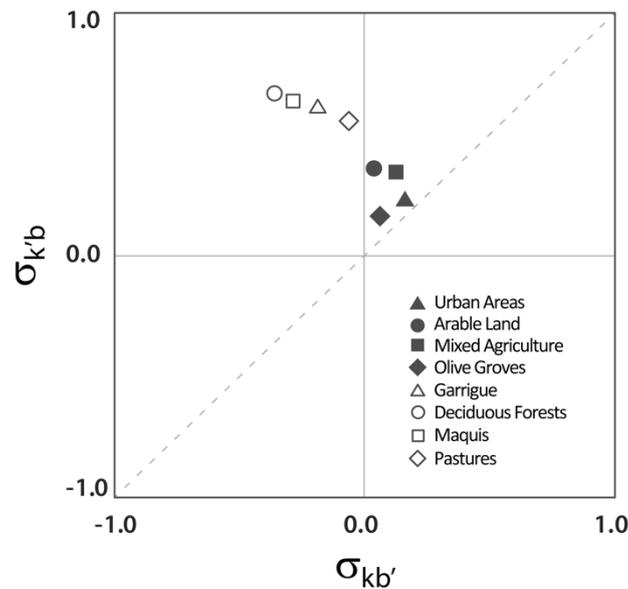
The selectivity values of  $\sigma_{k'b}$  and  $\sigma_{kb'}$  for all buffers and land cover classes are shown in Table 2, while Figure 2 shows the plot of  $\sigma_{k'b}$  vs.  $\sigma_{kb'}$  for the 100-m buffer only. Note that, as shown in Table 2, the results obtained for the 100-m buffer are qualitatively and quantitatively consistent with the results obtained for the other buffers.

According to Figure 2, the values of  $\sigma_{k'b}$  show that relative fire occurrence close to roads is higher than expected by chance alone in all LULC classes, meaning that for every land cover class, the proportion of fires close to roads is significantly higher compared to with respect to more interior areas of the same class. Nonetheless, this ‘clustering effect’ tends to be much lower for classes with high human pressure (i.e., urban areas, arable land, mixed agricultural land, and olive groves) than for natural and semi-natural classes (i.e., pastures, deciduous forests, garrigue, maquis) where the occurrence of human-related ignitions highly depends on the proximity to accessibility networks.

At the same time, the values of  $\sigma_{kb'}$  show that land uses with high anthropogenic pressure experience a disproportionately higher number of ignitions than expected under complete spatial randomness in fire distribution within each buffer ( $\sigma_{kb'} > 0$ ). By contrast, in forests and semi-natural LULC classes, the number of ignitions is generally lower than expected by chance alone ( $\sigma_{kb'} < 0$ ). As a result, all natural and semi-natural land uses are located in the upper-left portion of the plot of  $\sigma_{k'b}$  vs.  $\sigma_{kb'}$ , which is characterized by positive values of  $\sigma_{k'b}$  and negative values of  $\sigma_{kb'}$ , whereas all anthropogenic land uses are located in the upper-right quadrant in which both variables assume positive values.

**Table 2.** Values of the selectivity indices  $\sigma_{k'b}$  and  $\sigma_{kb'}$  for the selected land cover classes within each buffer.

Land Cover Class	Selectivity Related to Land Cover $\sigma_{k'b}$				Selectivity Related to Buffer Size $\sigma_{kb'}$				
	0–50 m	0–100 m	0–150 m	0–200 m	0–50 m	0–100 m	0–150 m	0–200 m	
<b>Artificial and agricultural</b>	Discontinuous urban areas	0.276	0.253	0.225	0.212	0.152	0.150	0.154	0.169
	Arable land	0.414	0.391	0.362	0.334	0.019	0.014	0.014	0.011
	Mixed agriculture	0.401	0.387	0.354	0.333	0.078	0.084	0.079	0.084
	Olive groves	0.217	0.190	0.168	0.156	0.040	0.034	0.045	0.061
<b>Natural and semi-natural</b>	Deciduous forests	0.697	0.676	0.652	0.622	−0.274	−0.291	−0.299	−0.318
	Maquis	0.677	0.654	0.622	0.596	−0.235	−0.253	−0.272	−0.283
	Garrigue	0.642	0.634	0.611	0.589	−0.194	−0.185	−0.189	−0.196
	Pastures	0.576	0.567	0.527	0.500	−0.083	−0.075	−0.098	−0.106



**Figure 2.** Plot of  $\sigma_{k'b}$  (fire selectivity related to land cover) vs. ( $\sigma_{kb'}$  fire selectivity related to buffer size) for the 100-m buffer. Gray symbols represent artificial and agricultural Classes, open symbols represent natural and semi-natural classes.

#### 4. Discussion

There has been a long research tradition focusing on the socio-economic aspects of wildfire [17,30,33,37–40]. In this study, we analyzed existing interactions between roads and land cover in driving fire ignition. Despite existing uncertainties in terms of the geolocation of a subset of ignition points, the very consistent patterns of the influence of roads and LULC classes on fire ignitions obtained throughout the buffer sizes allow us to assume a strong consistency of the overall results. According to our results, fires are preferentially ignited close to roads in all LULC classes. This effect is consistent with a number of previous studies in various ecosystems, which showed a high spatial clustering of human-caused fire ignitions around accessibility networks [8,9,22,25,41–43]. However, the observed influence of roads on the spatial pattern of fire ignitions is not the same in all LULC classes. It is particularly noticeable for the natural and semi-natural land uses, which usually experience a lower number of fires compared to more human-impacted land uses and for which roads represent the major vector of anthropogenic ignition energy [29]. By contrast, in LULC classes with high anthropogenic pressure, the ubiquity of the ignition energy related to the constant and widespread presence of humans reduces the clustering effect of the road network on fire ignitions.

In LULC classes with coarser fuels, such as deciduous forests or maquis, road corridors may directly influence ignition risk by reducing canopy closure, thus affecting the environmental conditions along road edges through increased solar radiation, temperature and wind speed, and reduced fuel moisture content [25]. These edge effects are not limited to the immediate neighborhoods of roads, but extend over a wider area in which fuel conditions are altered compared to continuous forest sites. The peculiar fuel characteristics of road edges have led [25]) to consider them as a separate and specific landscape component in terms of fuel proprieties that differ from the adjacent interior areas.

These results are of relevance for land and forest managers who have to be aware that planned roads and extensions of the road network may have different impacts on the potential of fire ignitions according to the land uses in the concerned area. In this sense, the increased ignition risk related to increasing housing and road density as postulated by [44] may be additionally modulated by the vegetation types crossed by roads.

To conclude, the proposed approach represents a first step towards the integration and the detailed assessment of two major fire ignition drivers, such as land cover and road network, which represent a

basic input to fire risk analysis and modern fire management planning [45,46]. A thorough knowledge of the joint effect of LULC and roads on fire ignition patterns is a key point, yet future studies should also take into account fire size and severity in order to fully understand the role of anthropogenic presence in fire regimes.

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## References

1. Krebs, P.; Pezzatti, G.B.; Mazzoleni, S.; Talbot, L.M.; Conedera, M. Fire regime: History and definition of a key concept in disturbance ecology. *Theory Biosci.* **2010**, *129*, 53–69. [[CrossRef](#)] [[PubMed](#)]
2. Pyne, S.J. *World Fire: The Culture of Fire on Earth*; Henry Holt and Company: New York, NY, USA, 1995.
3. Bowman, D.M.J.S.; O'Brien, J.A.; Goldammer, J.G. Pyrogeography and the global quest for sustainable fire management. *Annu. Rev. Environ. Resour.* **2013**, *38*, 57–80. [[CrossRef](#)]
4. FAO. *Fire Management—Global Assessment 2006*; FAO Forestry Paper 151; FAO: Rome, Italy, 2007.
5. Ganteaume, A.; Camia, A.; Jappiot, M.; San-Miguel-Ayanz, J.; Long-Fourne, I.M.; Lampin, C. A review of the main driving factors of forest fire ignition over Europe. *Environ. Manag.* **2013**, *51*, 651–662. [[CrossRef](#)] [[PubMed](#)]
6. Moreno, M.V.; Chuvieco, E. Characterising fire regimes in Spain from fire statistics. *Int. J. Wildland Fire* **2013**, *22*, 296–305. [[CrossRef](#)]
7. Cardille, J.A.; Ventura, S.J.; Turner, M.G. Environmental and social factors influencing wildfires in the Upper Midwest, United States. *Ecol. Appl.* **2001**, *11*, 111–127. [[CrossRef](#)]
8. Syphard, A.D.; Radeloff, V.C.; Keeley, J.E.; Hawbaker, T.J.; Clayton, M.K.; Stewart, S.I.; Hammer, R.B. Human influence on California fire regimes. *Ecol. Appl.* **2007**, *17*, 1388–1402. [[CrossRef](#)] [[PubMed](#)]
9. Syphard, A.D.; Radeloff, V.C.; Keuler, N.S.; Taylor, R.S.; Hawbaker, T.J.; Stewart, S.I.; Clayton, M.K. Predicting spatial patterns of fire on a southern California landscape. *Int. J. Wildland Fire* **2008**, *17*, 602–613. [[CrossRef](#)]
10. Vázquez, A.; Moreno, J.M. Sensitivity of fire occurrence to meteorological variables in Mediterranean and Atlantic areas of Spain. *Landsc. Urban Plan.* **1993**, *24*, 129–142. [[CrossRef](#)]
11. Viegas, D.X.; Viegas, M.T. A relationship between rainfall and burned area for Portugal. *Int. J. Wildland Fire* **1994**, *4*, 11–16. [[CrossRef](#)]
12. Vázquez, A.; Pérez, B.; Fernández-González, F.; Moreno, J.M. Recent fire regime characteristics and potential natural vegetation relationships in Spain. *J. Veg. Sci.* **2002**, *13*, 663–676. [[CrossRef](#)]
13. Lloret, F.; Pausas, J.G.; Vilà, M. Response of Mediterranean plant species to different fire regimes in Garraf Natural Park (Catalonia, Spain): Field observations and modelling predictions. *Plant Ecol.* **2003**, *167*, 223–235. [[CrossRef](#)]
14. Flannigan, M.D.; Krawchuk, M.A.; de Groot, W.J.; Wotton, B.M.; Gowman, L.M. Implications of changing climate for global wildland fire. *Int. J. Wildland Fire* **2009**, *18*, 483–507. [[CrossRef](#)]
15. Wotton, B.M.; Nock, C.A.; Flannigan, M.D. Forest fire occurrence and climate change in Canada. *Int. J. Wildland Fire* **2010**, *19*, 253–271. [[CrossRef](#)]
16. Moreira, F.; Ferreira, P.G.; Rego FC Bunting, S. Landscape changes and breeding bird assemblages in northwestern Portugal: The role of fire. *Landsc. Ecol.* **2001**, *16*, 175–187. [[CrossRef](#)]
17. Nunes, M.C.S.; Vasconcelos, M.J.; Pereira, J.M.C.; Dasgupta, N.; Alldredge, R.J. Land cover type and fire in Portugal: Do fires burn land cover selectively? *Landsc. Ecol.* **2005**, *20*, 661–673. [[CrossRef](#)]
18. Bajocco, S.; Ricotta, C. Evidence of selective burning in Sardinia (Italy): Which land-cover classes do wildfires prefer? *Landsc. Ecol.* **2008**, *23*, 241–248. [[CrossRef](#)]
19. Conedera, M.; Torriani, D.; Neff, C.; Ricotta C Bajocco, S.; Pezzatti, G.B. Using Monte Carlo simulations to estimate relative fire ignition danger in a low-to-medium fire-prone region. *For. Ecol. Manag.* **2011**, *261*, 2179–2187. [[CrossRef](#)]
20. Guglietta, D.; Migliozi, A.; Ricotta, C. A multivariate approach for mapping fire ignition risk: The example of the National Park of Cilento (Southern Italy). *Environ. Manag.* **2015**, *56*, 157–164. [[CrossRef](#)] [[PubMed](#)]

21. Martell, D.L.; Otukol, S.; Stocks, B.J. A logistic model for predicting daily people-caused forest fire occurrence in Ontario. *Can. J. For. Res.* **1987**, *17*, 394–401. [[CrossRef](#)]
22. Yang, J.; He, H.S.; Shifley, S.R.; Gustafson, E.J. Spatial patterns of modern period human-caused fire occurrence in the Missouri Ozark Highlands. *For. Sci.* **2007**, *53*, 1–15.
23. Lampin-Maillet, C.; Jappiot, M.; Long, M.; Morge, D.; Ferrier, J.P. Characterization and mapping of dwelling types for forest fire prevention. *Comput. Environ. Urban Syst.* **2009**, *33*, 224–232. [[CrossRef](#)]
24. Lampin-Maillet, C.; Jappiot, M.; Long, M.; Bouillon, C.; Morge, D.; Ferrier, J.P. Mapping wildland-urban interfaces at large scales integrating housing density and vegetation aggregation for fire prevention in the South of France. *J. Environ. Manag.* **2010**, *91*, 732–741. [[CrossRef](#)] [[PubMed](#)]
25. Narayanaraj, G.; Wimberly, M.C. Influences of forest roads on the spatial patterns of human- and lightning-caused wildfire ignitions. *Appl. Geogr.* **2012**, *32*, 878–888. [[CrossRef](#)]
26. Narayanaraj, G.; Wimberly, M.C. Influences of forest roads and their edge effects on the spatial pattern of burn severity. *Int. J. Appl. Earth Obs. Geoinf.* **2013**, *23*, 62–70. [[CrossRef](#)]
27. Penman, T.D.; Bradstock, R.A.; Price, O. Modelling the determinants of ignition in the Sydney Basin, Australia: Implications for future management. *Int. J. Wildland Fire* **2013**, *22*, 469–478. [[CrossRef](#)]
28. Pezzatti, G.B.; Zumbrennen, T.; Bürgi, M.; Ambrosetti, P.; Conedera, M. Fire regime shifts as a consequence of fire policy and socio-economic development: An analysis based on the change point approach. *For. Policy Econ.* **2013**, *29*, 7–18. [[CrossRef](#)]
29. Conedera, M.; Tonini, M.; Oleggini, L.; Vega Orozco, C.; Leuenberger, M.; Pezzatti, G.B. Geospatial approach for defining the Wildland-Urban Interface in the Alpine environment. *Comput. Environ. Urban Syst.* **2015**, *52*, 10–20. [[CrossRef](#)]
30. Romero-Calcerrada, R.; Novillo, C.J.; Millington, J.D.A.; Gomez-Jimenez, I. GIS analysis of spatial patterns of human-caused wildfire ignition risk in the SW of Madrid (Central Spain). *Landsc. Ecol.* **2008**, *23*, 341–354. [[CrossRef](#)]
31. Ager, A.A.; Preisler, H.K.; Arca, B.; Spano, D.; Salis, M. Wildfire risk estimation in the Mediterranean area. *Environmetrics* **2014**, *25*, 384–396. [[CrossRef](#)]
32. Vilar, L.; Gómez, I.; Martínez-Vega, J.; Echavarría, P.; Riaño, D.; Martín, M.P. Multitemporal Modelling of Socio-Economic Wildfire Drivers in Central Spain between the 1980s and the 2000s: Comparing Generalized Linear Models to Machine Learning Algorithms. *PLoS ONE* **2016**, *11*, e0161344. [[CrossRef](#)] [[PubMed](#)]
33. Ye, J.X.; Wu, M.S.; Deng, Z.J.; Xu, S.J.; Zhou, R.L.; Clarke, K.C. Modeling the spatial patterns of human wildfire ignition in Yunnan province, China. *Appl. Geogr.* **2017**, *89*, 150–162. [[CrossRef](#)]
34. Salis, M.; Ager, A.A.; Alcasena, F.J.; Arca, B.; Finney, M.A.; Pellizzaro, G.; Spano, D. Analyzing seasonal patterns of wildfire exposure factors in Sardinia, Italy. *Environ. Monit. Assess* **2015**, *187*, 4175. [[CrossRef](#)] [[PubMed](#)]
35. Bajocco, S.; Dragozi, E.; Gitas, I.; Smiraglia, D.; Salvati, L.; Ricotta, C. Mapping forest fuels through vegetation phenology: The role of coarse-resolution satellite time-series. *PLoS ONE* **2015**, *10*, e0119811. [[CrossRef](#)] [[PubMed](#)]
36. ISPRA. *La Realizzazione in Italia del Progetto Corine Land Cover 2006*; Rapporto 131; Istituto Superiore per la Protezione e la Ricerca Ambientale: Rome, Italy, 2010.
37. Nunes, A.N.; Lourenco, L.; Meira, A.C.C. Exploring spatial patterns and drivers of forest fires in Portugal (1980–2014). *Sci. Total Environ.* **2016**, *573*, 1190–1202. [[CrossRef](#)] [[PubMed](#)]
38. Oliveira, S.; Pereira, J.M.C.; San-Miguel-Ayanz, J.; Lourenco, L. Exploring the spatial patterns of fire density in Southern Europe using Geographically Weighted Regression. *Appl. Geogr.* **2014**, *51*, 143–157. [[CrossRef](#)]
39. Syphard, A.D.; Keeley, J.E. Location, timing and extent of wildfire vary by cause of ignition. *Int. J. Wildland Fire* **2015**, *24*, 37–47. [[CrossRef](#)]
40. Costafreda-Aumedes, S.; Comas, C.; Vega-Garcia, C. Human-caused fire occurrence modelling in perspective: A review. *Int. J. Wildland Fire* **2017**, *26*, 983–998. [[CrossRef](#)]
41. Yang, J.; He, H.S.; Sturtevant, B.R.; Miranda, B.R.; Gustafson, E.J. Comparing effects of fire modeling methods on simulated fire patterns and succession: A case study in the Missouri Ozarks. *Can. J. For. Res.* **2008**, *38*, 1290–1302. [[CrossRef](#)]
42. Vega Orozco, C.; Tonini, M.; Conedera, M.; Kanevski, M. Cluster recognition in spatial-temporal sequences: The case of forest fires. *Geoinformatica* **2012**, *16*, 653–673. [[CrossRef](#)]

43. Ricotta, C.; Di Vito, S. Modeling the Landscape Drivers of Fire Recurrence in Sardinia (Italy). *Environ. Manag.* **2014**, *53*, 1077–1084. [[CrossRef](#)] [[PubMed](#)]
44. Theobald, D.M.; Romme, W.H. Expansion of the US wildland-urban interface. *Landsc. Urban Plan.* **2007**, *83*, 340–354. [[CrossRef](#)]
45. Finney, M.A. The challenge of quantitative risk analysis for wildland fire. *For. Ecol. Manag.* **2005**, *211*, 97–108. [[CrossRef](#)]
46. Curt, T.; Borgniet, L.; Bouillon, C. Wildfire frequency varies with the size and shape of fuel types in southeastern France: Implications for environmental management. *J. Environ. Manag.* **2013**, *117*, 150–161. [[CrossRef](#)] [[PubMed](#)]



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