



Article

Selective Recovery of Manganese from Anode Sludge Residue by Reductive Leaching

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Received: 26 March 2020; Accepted: 15 June 2020; Published: 17 June 2020



Abstract: Manganese-containing anode sludge is a common side-product in the electrowinning of zinc. The anode sludge consists mainly of oxidized manganese, but also lesser amounts of lead, calcium, and other minor metals. The impurities present in the anode sludge mandate new recycling strategies for its efficient use. This work demonstrates a novel method for selective manganese recovery from lead- and calcium-bearing manganese oxide solid residue. Leaching with sulfuric acid in the presence of a selected reducing agent, such as hydrogen peroxide or citric acid, yields a concentrated MnSO₄ solution with high selectivity over calcium and lead. Manganese yields up to 98% can be obtained. Minimization of calcium and lead in final manganese product can be accomplished with the correct choice of leaching conditions. Alongside manganese sulfate solution, leaching residue with high content of lead and silver was also formed.

Keywords: recycling; manganese; reductive leaching; anode sludge; impurity control

1. Introduction

Manganese is the 12th most abundant element in Earth's crust, but utilizable resources are unevenly distributed [1–3]. South Africa has the largest reserves and it is producing most of the world's manganese. Main uses for manganese are steel production, non-ferrous metallurgy, batteries, dietary additives, and fertilizers [2]. The amount of high-grade ores has been diminishing, making low-grade and secondary resources more important in the future [4]. One potential secondary manganese source is the anode sludge formed during zinc electrolysis [5].

EU Waste Directive [6] and corresponding national legislation in Finland [7] aim to prevent waste and by-product formation and to promote the utilization of secondary resources. The potential re-use of anode sludge supports these intentions. The use of secondary resources and the aim to create a circular economy is very topical [6,7].

Divalent manganese is present in zinc electrowinning feed solution in order to preserve the surface of lead anodes from oxidative corrosion during the electrowinning process. Even a small concentration of manganese can increase the operation time of the lead anodes significantly [8,9]. Lead anodes can be also alloyed with small amounts of silver and calcium in order to further increase the corrosion resistance [8]. The corrosion of lead anodes results in the formation of lead sulfate (Equation (1)) and oxide (Equations (2) and (3)).

$$Pb(s) + SO_4^{2-}(aq) = PbSO_4(s) + 2e^- E^0 = 0.361 V$$
 (1)

$$Pb(s) + 2H_2O = PbO_2(s) + 4H^+(aq) + 4e^- E^0 = 0.610 V$$
 (2)

$$PbSO_4(s) + 2H_2O = PbO_2(s) + 2H^+(aq) + 2e^- + H_2SO_4(aq) E^0 = 1.685 V$$
 (3)

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Manganese oxidizes on the anode and in the bulk of the process solution, which leads to precipitation of different manganese oxides [9]. Amount of lead in the electrolyte solution is limited by adding water-soluble strontium compounds into the electrowinning cells [8,10,11]. Lead is incorporated into precipitating strontium sulfate crystals [10].

Precipitated oxidation products settle on the bottom of the electrolysis cells and form a mixture called anode sludge. Along with manganese and lead, anode sludge contains lesser amounts of calcium, zinc, strontium and sodium [8].

Anode sludge mostly contains tetravalent manganese, which has a limited solubility even in acidic conditions. Pyrometallurgical pre-treatment in elevated temperatures and the presence of a reducing agent are required to convert manganese to a readily dissolving state. Various reducing agents such as coal [12], graphite [13], carbon monoxide [14], various carbohydrates [15], sulfur dioxide [16], hydrogen [14], sulfur [17], and pyrite [18], have been reportedly tested. Reduction is followed by mineral acid leaching with sulfuric acid, nitric acid or hydrochloric acid [2]. These two steps can be carried out in one step introducing suitable reducing agent together with the mineral acid into the leaching solution [19].

The reduction of MnO_2 in acidic media proceeds via steps given in Equations (4) and (5) [20–22]. The overall reaction and the corresponding standard reduction potential are given in Equation (6). In addition, various side-reactions and alternative steps have been proposed between different manganese species [22].

$$MnO_2(s) + H^+(aq) + e^- \rightarrow MnOOH(s)$$
 (4)

$$MnOOH(s) + 3H^{+}(aq) + e^{-} \rightarrow Mn^{2+}(aq) + 2H_{2}O$$
 (5)

$$MnO_2(s) + 4H^+(aq) + 2e^- \rightarrow Mn^{2+}(aq) + 2H_2O E^0 = -1.28 V$$
 (6)

Reducing agents such as hydrogen peroxide [19,23–25], calcium sulfide [26], EDTA [27], ferrous ions [28], citric acid, oxalic acid, ascorbic acid, carbohydrates [29,30], sodium sulfide [31] and sulfur dioxide [32] have been reported in literature.

Manganese dioxide dissolves in sulfuric acid media by citric acid according to Equation (7) [29]. Although hydrogen peroxide is widely known as a strong oxidizing agent, it can act as reducing agent towards very strong oxidizing agents [23]. In sulfuric acid media, manganese dioxide dissolves by hydrogen peroxide reduction according to Equation (9) [19].

$$9MnO_2(s) + 9H_2SO_4 + C_6H_8O_7 \rightarrow 9MnSO_4 + 13H_2O + 6CO_2$$
 (7)

$$O_2 + H^+ + e^- \rightarrow H_2O_2 E_0 = 0.68V$$
 (8)

$$MnO_2(s) + H_2O_2 + H_2SO_4 \rightarrow MnSO_4 + 2H_2O + O_2$$
 (9)

The aim of this study was to leach manganese selectively over calcium, lead and strontium from anode sludge by using sulfuric acid and a reducing agent. Citric acid and hydrogen peroxide were compared as possible reducing agents. The focus of this work was to determine the effect of leaching conditions on concentrations of sparingly soluble impurities present in anode sludge. Even though manganese has low economic value, the leaching process simultaneously concentrates lead and silver to leaching residue. Possible recovery of lead and silver increases viability of the process, alongside a reduction of landfill costs.

2. Experimental Methods

2.1. Materials

Real anode sludge from an operating zinc electrowinning process was used for this work. Excess solution was filtrated from anode sludge batch by vacuum filtration. Filtrated anode sludge was dried before use. Anode sludge sample was decomposed by microwave in presence of hydrochloric acid,

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nitric acid and tetra fluoroboric acid. Elemental analysis for anode sludge sample was measured with ICP-OES (Thermo Fisher). Table 1 shows the average chemical composition of dried anode sludge batch. Figure 1 shows the X-ray powder diffraction pattern of the anode sludge sample as measured by PANalytical X'Pert Pro (Malvern, UK) using Cu-K α radiation. It can be seen that the anode sludge consists mainly of manganese dioxide and lead sulfate. Gypsum and strontium sulfate were also identified in the sample. Sulfuric acid (95.0–97.0%, puriss, p.a., Sigma-Aldrich, St. Louis, MO, USA) was used as a leaching agent. Citric acid monohydrate (>99.5%. ACS Reagent, Sigma-Aldrich) and H_2O_2 (30%, puriss, p.a., Sigma-Aldrich) were used as reducing agents.

Mn (wt%)	Pb (wt%)	Ca (wt%)	S (wt%)	Sr (wt%)	Zn (wt%)
41.9	6.5	1.2	4.2	0.92	1.00
K (wt%)	Mg (wt%)	Na (wt%)	Fe (wt%)	Al (wt%)	Si (wt%)
1.4	0.17	0.21	0.015	0.04	0.23

Table 1. Average chemical composition of the main elements in anode sludge.

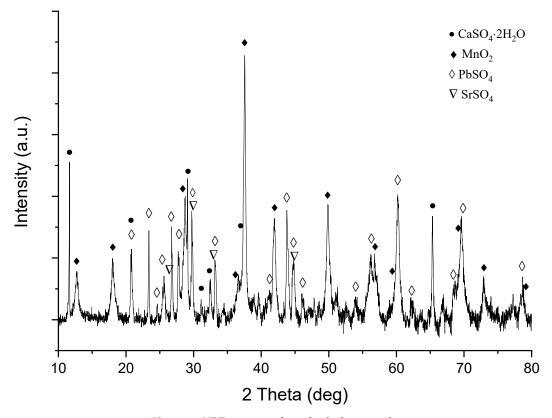


Figure 1. XRD pattern of anode sludge sample.

2.2. Leaching Experiments

The anode sludge was leached in a 1000 mL round-bottom flask. The reaction temperature was adjusted using a water bath on a hot plate. The evaporation of water was prevented by a reflux-condenser. The reaction mixture was magnetically stirred at a constant speed of 500 rpm. The anode sludge and the sulfuric acid diluted with water was added to the flask. The volume of the solution was adjusted so that the final volume of the reaction mixture after reducing agent addition was approximately 500 mL in each experiment.

Citric acid (1.23 M) or hydrogen peroxide (30%) were fed during first 30 min of reaction time by motorized pump (Watson-Marlow 520u). The reducing agents were fed over an extended period of time as both reactions are fast, exothermic, and lead to the formation of gases. Table 2 shows experimental

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parameters for each solid-to-liquid ratio used in leaching experiments. Citric acid experiments with 420 g/L solid-to-liquid ratio were repeated using hydrogen peroxide as reducing agent.

S/L (g/L)	Anode Sludge (g)	H ₂ SO ₄ (M)	H ₂ SO ₄ (mL)	Reducing Agent (mL)
420	210	5.18	332	168
360	180	4.14	356	144
300	150	3.22	380	120

Table 2. Experimental parameters for each solid-to-liquid ratio.

5 mL samples were taken at predetermined intervals (5, 10, 20, 30, 60, 120 and 180 min). Each sample was filtered and quickly diluted with a dilute nitric acid solution, in order to prevent precipitation of sparingly soluble sulfates. After each experiment, the solid residue was recovered by filtration, washed, and dried. Residue solids were decomposed by microwave in presence of hydrochloric acid, nitric acid and tetra fluoroboric acid. All samples were analyzed with ICP-OES (Thermo Fisher, Waltham, MA, USA).

3. Results and Discussion

3.1. Effect of Leaching Parameters on Manganese Concentration and Yield

The stirring rate and stoichiometric ratios of chemicals (leaching agent and reducing agent) were kept constant during the set of experiments. The effects of temperature (50, 60, 70, and 80 $^{\circ}$ C) and solid-to-liquid ratio (320, 360, and 420 g/L) were examined. The volumes of each batch and reagent solution were adjusted to yield a final volume of 500 mL.

Figures 2 and 3 show the effect of leaching time on manganese concentration at different temperatures with a constant solid-to-liquid ratio of 420 g/L using citric acid and hydrogen peroxide as reducing agents. The results show that the effect of temperature on the leaching efficiency is not significantly different between temperatures 50 and 70 $^{\circ}$ C. Most of the significant changes in manganese concentration happened during the first 30 min of the reaction, which is the feeding time of reducing agent. This indicates that the fed reducing agent reacted very quickly with manganese dioxide. Manganese concentrations ranging from 170 to 190 g/L were obtained.

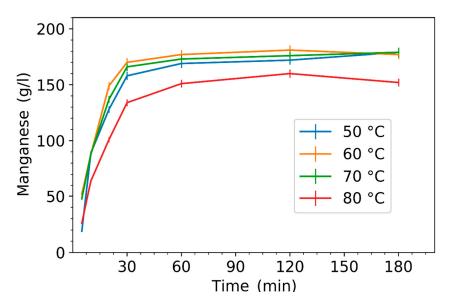


Figure 2. Manganese concentration as a function of time at 50, 60, 70 and 80 $^{\circ}$ C (solid-to-liquid (S/L) ratio 420 g/L, Citric acid).

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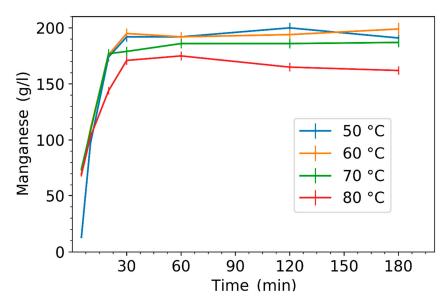


Figure 3. Manganese concentration as a function of time at 50, 60, 70 and 80 °C (S/L ratio 420 g/L, H₂O₂).

Experiments at 50 and 60 $^{\circ}$ C produced quite similar leaching behaviors. At 80 $^{\circ}$ C, the dissolution rate of manganese is the slowest and yield is lowest. During the first 30 min, manganese dissolution is slower at 80 $^{\circ}$ C probably due to partial thermal decomposition of reducing agents. At the later stages of leaching reaction, probably limitations in solubility of the manganese sulfate takes place. Figures 4 and 5 show the effect of leaching time on manganese yield at different temperatures with constant solid-to-liquid ratio of 420 g/L using citric acid and hydrogen peroxide as reducing agents, respectively. All significant changes in manganese dissolution were observed during the first 30 min of reaction time.

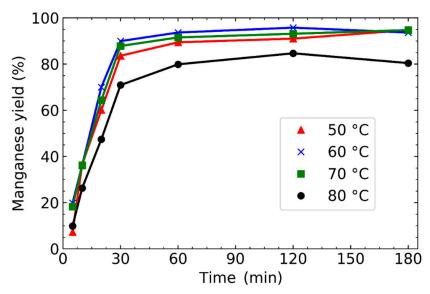


Figure 4. Manganese yield as a function of time at 50, 60, 70 and 80 °C (S/L ratio 420 g/L, Citric. acid).

Figure 6 compares manganese leaching efficiency between the two reducing agents, hydrogen peroxide and citric acid, at different temperatures while using a solid-to-liquid ratio of 420 g/L. The final concentrations of each leaching solution were used for comparison. The manganese concentration is approximately 6–7% higher when hydrogen peroxide was used as the reducing agent. Manganese dissolution kinetics are almost similar at 50 and 60 $^{\circ}$ C when using citric acid as the reductant. While using hydrogen peroxide, 50 and 60 $^{\circ}$ C give almost similar leaching kinetics. At 70 $^{\circ}$ C the leaching rate of manganese is almost as high as at 50 and 60 $^{\circ}$ C. Manganese dissolution slows down greatly at

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 $80\,^{\circ}\text{C}$ regardless of the choice of reducing agent. This limited manganese concentration is possibly due to the lowered solubility of manganese sulfate at $80\,^{\circ}\text{C}$ [33] or thermal decomposition or consumption of the reducing agent on side reactions.

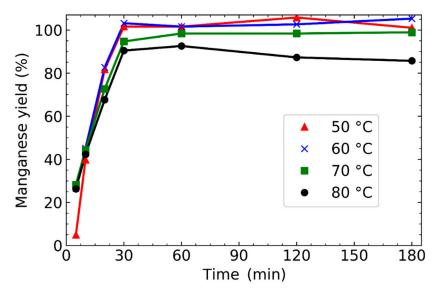


Figure 5. Manganese yield as a function of time at 50, 60, 70 and 80 °C (S/L ratio 420 g/L, H₂O₂).

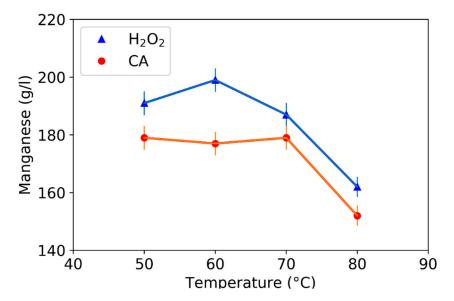


Figure 6. Comparison of reducing agents on manganese concentration at different temperatures after 180 min of reaction time using S/L 420.

Figure 7 shows the effect of the solid-to-liquid ratio and reducing agent at 60 °C on manganese concentration as a function of time. Manganese concentration was highest when 420 g/L solid-to-liquid ratio was used. Using hydrogen peroxide as the reducing agent produced higher manganese concentrations than citric acid.

Figure 8 shows the effect of the leaching conditions on manganese yield. Yields have been calculated from washed solid leaching residues. On average, the highest yields were obtained at $60\,^{\circ}$ C. Even though the highest concentrations were obtained from experiments with a solid-to-liquid (S/L) ratio of 420 g/L using citric acid, these experiments had on average lower yields compared to experiment with lower S/L ratio. At 80 $^{\circ}$ C and S/L 420 g/L yields were higher than in yields calculated from liquid samples (Figures 4 and 5). This confirms that solid samples contained manganese sulfate,

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which dissolved during the washing. Hence manganese yields were lower due to crystallization of MnSO₄ from supersaturated solution.

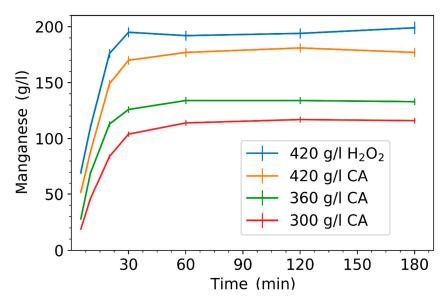


Figure 7. Effect of S/L on manganese concentration as a function of time at 60 °C.

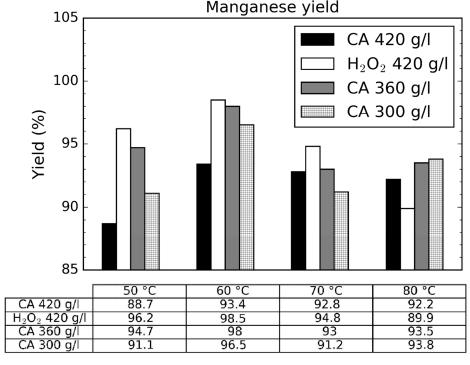


Figure 8. Effect of leaching conditions on manganese recovery rate after 180 min of reaction time.

3.2. Effect of Leaching Parameters on Calcium and Lead Concentrations

Solid-to-liquid ratio played an important role in the selectivity of manganese over lead, calcium and strontium. Figure 9 shows the effect of the solid-to-liquid ratio and temperature on the calcium (g/L) to manganese (g/L) ratio. On average solid-to-liquid ratio 420 g/L improved selectivity of manganese over calcium versus lower solid-to-liquid ratios. Temperature changes between 50 and 70 °C do not seem to have clear pattern. At 80 °C Ca/Mn ratio increases due to lowered manganese

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concentration as shown on Figure 6. This is probably a result of the lowered solubility of manganese at higher temperature.

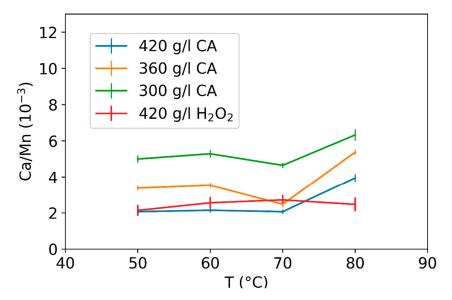


Figure 9. Effect of S/L and temperature on calcium (g/L) to manganese (g/L) ratio after 180 min of reaction time.

Figure 10 shows the effect of the solid-to-liquid ratio and temperature on lead (g/L)-to-manganese (g/L) ratio. By taking uncertainty into account, lead-to-manganese ratio does not show any obvious trend.

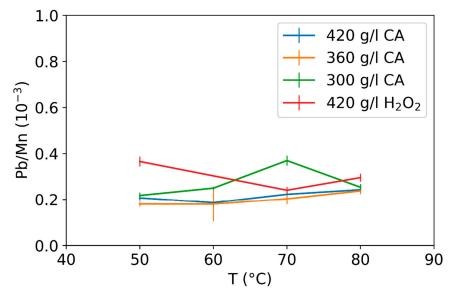


Figure 10. Effect of S/L and temperature on lead (g/L) to manganese (g/L) ratio after 180 min of reaction time.

The effect of temperature on calcium concentration as a function of time using citric acid and hydrogen peroxide was examined. The data was obtained from experiments using a solid-to-liquid ratio 420~g/L, and citric acid as a reducing agent (Figure 11). The solubility reached its peak in the presence of citric acid after 5 min reaction time. It then steadily decreased and achieved a steady state after 60 min. At 80 °C, the solubility was clearly higher than at lower temperatures.

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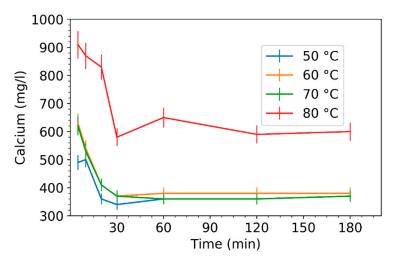


Figure 11. Calcium concentrations as a function of time at 420 g/L S/L using citric acid as reducing agent.

Figure 12 shows calcium concentration as a function of time using hydrogen peroxide as reducing agent. Using hydrogen peroxide as the reducing agent produced low calcium concentrations in the beginning of the leaching reaction, whereas the citric acid experiments displayed the opposite result. Calcium concentrations increased at the beginning of the reaction, but after varying periods of time calcium concentration started to decrease to a level similar to what was observed in presence of citric acid. This difference in calcium leaching behavior is most likely connected to formation of different stable calcium—citrate complexes [34–36]. Complex formation leads to apparent increase in calcium sulfate solubility, which is seen as a large initial total calcium concentration. The concentration falls when the stabilizing citric acid is consumed in the leaching reaction.

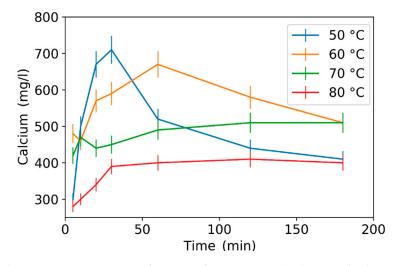


Figure 12. Calcium concentrations as a function of time at 420 g/L S/L using hydrogen peroxide as reducing agent.

The biggest difference between the two reducing agents was that the calcium concentration initially decreased in the presence of citric acid and increased in the presence of hydrogen peroxide. Another difference was observed at 80 °C, where the calcium concentration was the highest when using citric acid and the lowest when using hydrogen peroxide.

3.3. Purity of Leachate

Table 3 shows the compositions of selected leachates obtained from leaching experiments. The leaching method used produced highly concentrated solutions of manganese sulfate. The most

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common impurities are zinc and potassium at approximately 4 g/L concentrations at solid-to-liquid ratio 420 g/L. Lowering the solid-to-liquid ratio lowered concentration of zinc in the leachates. An additional purification step is required to remove zinc completely.

	Mn (g/L)	Zn (g/L)	K (g/L)	S (g/L)	Ca (mg/L)	Pb (mg/L)	Fe (mg/L)	Al (mg/L)
420 g/L, 50 °C, H ₂ O ₂	191	4.06	3.9	119	410	70	44	27
420 g/L, 70 °C, H ₂ O ₂	187	4.09	3.8	120	510	45	46	60
420 g/L, 70 °C, CA	179	4.09	3.7	107	370	40	44	27
360 g/L, 50 °C, CA	139	4.05	2.6	93	470	25	41	11
300 g/L, 50 °C, CA	116	3.07	2.2	79	620	29	30	13

Table 3. Elemental analysis of the purity of leachates after reductive leaching.

Other significant impurities in the filtrates were sodium and calcium. Sodium concentrations were dependent on the solid-to-liquid ratio. The leaching conditions can be used for limiting calcium concentration in leachates. With the choice of proper conditions, the calcium concentration could be limited to around 400 mg/L, while the manganese concentration was closer to 200 g/L.

Table 4 shows legislative limit values for heavy metals in fertilizing products in Finland. Based on these limit values, the only troublesome impurity in manganese solution is lead. Zinc is acceptable, if soil analysis indicates a zinc deficiency. The amount of solid substance per liter was calculated from manganese and zinc concentrations. It was also presumed that solid substance would contain a stoichiometric amount of sulfate. The lead concentration per solid substance was calculated according to this assumption. From the solutions presented in Table 3, the lead concentration exceeded limits only in hydrogen peroxide experiments performed at 80 °C. A solution produced at 80 °C is still acceptable for forest fertilization.

Table 4. Limit values for heavy metals according to the Fertilizer Product Decree 24/2011 (Finnish Ministry of Agriculture and Forestry). Limit for copper and zinc can be exceeded if soil analysis indicates deficiency.

	As (mg/kg)	Hg (mg/kg)	Cd (mg/kg)	Cr (mg/kg)	Cu (mg/kg)	Pb (mg/kg)	Ni (mg/kg)	Zn (mg/kg)
Field	25	1	2.5	300	600	100	100	1500
Forest	40	1	2.5	300	700	150	150	4500
This work	-	-	-	-	1.44	76	-	9314

3.4. Leaching Residue

The obtained leaching residue had significant lead content along with manganese and calcium. Lead content in leaching residues varied from 24 to 28 wt%. Manganese composition in leaching residues was on average 12 wt%. With the correct choice of leaching conditions residual manganese could be lowered to as low as 3 wt%, which corresponds to a leaching yield of 99%. The calcium content of leaching residue varied from 5 to 11 wt%. The leaching residue also contains a significant amount of silver (0.1–0.15 wt%). XRD diagram of leaching residue sample, Figure 13, showed that calcium, lead and strontium formed sulfates during the leaching process.

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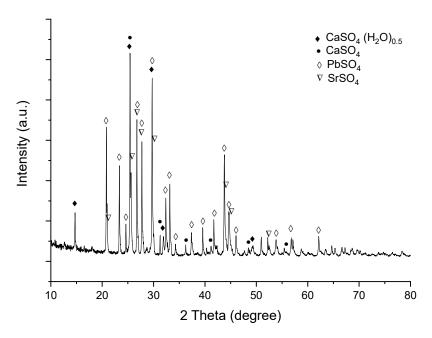


Figure 13. XRD pattern of leaching residue.

4. Conclusions

Manganese is a significant but often underappreciated raw material for ferrous/non-ferrous industry, batteries and fertilizers. In this study selective leaching of manganese from anode sludge was investigated as a single-step process.

Based on the results, concentrated manganese sulfate solution could be obtained from anode sludge with reductive leaching. Careful selection of leaching parameters can be used for lowering ratios of calcium and lead compared to manganese concentration. Using hydrogen peroxide as the reducing agent and 60 °C as reaction temperature favors the manganese concentration. A 98.5% leaching efficiency of manganese was achieved using hydrogen peroxide as a reducing agent at 60 °C. Lower yields were obtained at 420 g/L solid-to-liquid ratio when citric acid was used as a reducing agent. Lowering the solid-to-liquid ratio improved manganese yields when citric acid was used as a reducing agent. Calcium content in the filtrate can be limited to approximately 400 mg/L regardless of reducing agent. The resulting leaching solution still contains approximately 5 g/L of zinc, which is acceptable, if the solution is used as a nutrient solution for plants. High lead and silver content of the leaching residue could make it an attractive raw material for lead smelter. In future work, methods to demonstrate the production of high purity manganese sulfate from anode sludge are worthy of investigation.

Author Contributions: Conceptualization, T.K. and T.V.; Experimental, T.K and T.V. writing—original draft preparation, T.K. and T.V.; writing—review and editing, U.L. and J.S.; supervision, U.L. and J.S.; funding acquisition, U.L. All authors have read and agreed to the published version of the manuscript.

Funding: This research was funded by Central Ostrobothnia Region, Saostus (A71944), Tekes, AdChem (1792/31/2016) and Business Finland, SYMMET (4236/31/2018).

Acknowledgments: Tao Hu is acknowledged for kind help with XRD analyses.

Conflicts of Interest: The authors declare no conflict of interest.

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