



Approaches in Design of Laboratory-Scale UASB Reactors

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Abstract: Up-flow Anaerobic Sludge Blanket (UASB) reactors are popular tools in wastewater treatment systems due to the ability to work with high feed rates and wastes with high concentration of organic contaminants. While full-scale industrial applications of UASB reactors are developed and described in the available literature, laboratory-scale designs utilized for treatability testing are not well described. The majority of published studies do not describe the laboratory UASB construction details or do use reactors that already had developed a trophic network in microbial consortia under laboratory environment and therefore are more stable. The absence of defined guidelines for geometry design, selection of materials, construction, operation rules, and, especially, the start-up conditions, significantly hamper researchers who desire to conduct treatability testing using UASB reactors in laboratory scale. In this article, we compiled and analyzed the information available in the refereed literature concerning UASB reactors used in laboratory environment, where information on geometry and/or operational conditions were provided in detail. We utilized the information available in the literature and the experience gained in our laboratory (Sustainable Waste to Bioproducts Engineering Center) to suggest a unified operation flowchart and for design, construction, operation, and monitoring for a laboratory-scale UASB reactors.

Keywords: up-flow anaerobic sludge blanket reactors; anaerobic digestion; laboratory-scale experiment

1. Introduction

Up-flow Anaerobic Sludge Blanket (UASB) reactor is an anaerobic digester for wastewater treatment, and its operational concept can be described as a vertical up-flow pumping of liquid substrate, including wastewater or growth media, through a layer of anaerobic sludge [1–6]. Microbial consortia inside the sludge layer consume digestible components as substrate and decompose them into smaller chemical compounds [7]. Within the scope of a wastewater treatment, the goal of anaerobic digestion is a complete mineralization of organic compounds combined with the production of biogas for the purpose of energy recovery.

A distinguishing feature of UASB reactors is the formation of microbial conglomerates, where the metabolic product of one microbial group is a consumable substrate for another microbial group [8]. Such microbial conglomerates grow into spherical or bean-shaped granules over time [9–13]. The sizes of granules may vary, but typically are reported in the range 0.5 to 6 mm, where longer operation leads to larger sizes [14–16]. Granulation of sludge is promoted by the presence of microorganisms that are able to produce and secrete Exocellular Polymeric Substances (EPS) [17]. The term "EPS" includes multiple types of compounds, which serve as a glue to agglomerate microorganisms together and to add some mechanical strength to a granule [18].

A combination of developing trophic microbial connections and mechanical cementation with EPS, results in higher resilience of larger granules to sudden changes of operation conditions including a change in pH, temperature mode failure, substrate switch or inconsistency of a substrate strength and content, feeding rate fluctuations, etc. [19,20]. In some cases, granules can be disrupted due to hydrodynamic forces or inner gas pressure into several smaller fragments [12,21,22] which become cores for the formation of new granules [23].

The traditional concept of a UASB reactor, as suggested by Lettinga [10,11,24–26], is represented in Figure 1a. The substrate is pumped to a reactor through the distribution system into a bottom layer of anaerobic sludge. Equally distributed in normal cross-section of the reactor, the substrate is pushed through the sludge layer (called a "digestion zone") creating a vertical up-flow. This process is concurrent with the decomposition of organic compounds of substrate and a formation of gaseous products. Besides feeding the reactor, the continuous vertical up-flow of substrate prevents the sludge layer from clogging, keeping it afloat. However, the up-flow does wash out the unattached biomass (microorganisms, that did not start to form flocs) and small flocs/granules. The liquid part above the sludge layer (called "settling zone") serves as a vertical settler and/or coagulation column to initiate the biomass and solids retention process before the actual separation. The separation process occurs in the compartment called Gas–Liquid–Solids separator (GLSS, a.k.a. three-phase separator). GLSS is traditionally located on top of the reactor column and it starts with a baffle-shaped structure in its bottom part, which serves the purpose of collecting and re-directing the gas bubbles to the main gas collection part and preventing gas bubbles from escaping with effluent.



Figure 1. Operational concept of traditional Up-flow Anaerobic Sludge Blanket (UASB) reactors: (a) traditional; (b) with modified gas collector; and (c) Y-shaped.

The construction concept of the GLSS is shown in Figure 1a, where it's implemented via narrowing the outlet of the reaction tube with baffles. Such baffles are typically referred to as "deflectors" or "collar". The side effect of narrowing the reaction tube outlet is a creation of local velocity gradient (velocity shear), which slightly enhances the formation of granulated particles, their separation from liquid and settling back to the bottom of the reactor. Above the baffles, the GLSS contains the gas collecting structure, where the cross-section looks like a flipped upside-down funnel. In some studies, this funnel is replaced by a tubular structure with diameter larger than the distance between baffles [27,28]. The liquid is forced to flow through the space in between the lower edge of the gas collector and the baffles, to go around the funnel and leave the reactor at the effluent port.

Other existing modifications of GLSS in laboratory-scale reactors can improve the higher solids retention time, such as installing a high rate settler in headspace [29] or modification of three-phase separators [30].

In addition to the operational concept of the UASB reactor shown in Figure 1a, the same authors [11] also describe UASB reactor with a modified gas collector, which is demonstrated in Figure 1b. However, some studies [31] call such a modification of the Up-flow Anaerobic Sludge Baffled Reactor (UASBR). It may also contain the inner mechanical agitation device to prevent foam formation in the gas collecting area [32]. Recently, the Y-shaped variation of UASB reactor also became popular and is depictured in Figure 1c. In the case of the Y-shaped reactor, the GLSS is split into two individual separators: one separator is used to separate gas from the liquid and collect it directly at the top of a main tube, whereas a second collector is a sidearm tube that serves as an inclined settler for separating solids from liquid (similar to a Lamella clarifier). Use of a funnel-shaped gas-collecting element becomes optional in such case, since it serves only the purpose of preventing gas flow to an effluent side-arm.

Considering the concepts described, the optimization goal of a laboratory scale UASB reactor operation is to achieve better performance, where optimization targets for UASB performance include the following:

- Higher removal of contaminants;
- Higher biogas production rate;
- Shortening of adaptation period; and
- Resilience (robustness) of sludge.

To achieve some of those optimizations, the classical UASB concept can be combined with other types of reactors, resulting in a range of composite reactors. Some modifications are found in the literature and are presented in Table 1. This table represents options, where another reactor type is incorporated into the UASB itself, but not a sequence of two consecutive reactors.

Unit to Incorporate into UASB	Resulting Reactor Name	Purpose of Incorporation	Reference
Electrolysis cell	Up-flow Anaerobic BioElectroChemical reactor (UABE)	Increase the methane production via partial capture of dissolved carbon dioxide	[33–36]
Anaerobic Filter	UASB-AF Anaerobic hybrid	Increase retention of solids inside of a reactor and prevent washout	[37] [14,31,38,39]
	reactor (AHR)	of active biolinass	
Lamella settler	No Name	Increase solids retention time (SRT)	[40]

Table 1. Existing hybrid versions of UASB reactors.

In a holistic view, the purpose of UASB reactor optimization is to keep the microorganisms in a stage of maximum substrate consumption and active growth. However, from an operational perspective, the optimization of UASB functioning is achievable via adjusting operational parameters, including, but not limited to:

- Organic Loading Rate (OLR) and Hydraulic Retention Time (HRT);
- Recycle ratio of effluent;
- Regulation of pH;
- Retention of biomass; and
- Granulation enhancement.

Despite the long history since the invention and description of the UASB concept by Lettinga et al. [41] and increasing its application in industry, UASB laboratory scale reactors used for treatability studies are highly variable with regard to terminology, design, construction, and operation

processes. This lack of uniformity leads to different results regarding water quality indicators, for example, Chemical Oxygen Demand (COD), as well as bioenergy production, for example for biomethane and biohydrogen. There is a lack of uniformity with regard to the guidelines for operation of laboratory scale conditions, which is highlighted in this manuscript and recommendation are provided for making UASB laboratory studies and results more uniform with results more transferrable among laboratories and more useful for scale up activities. These lack of uniformity with laboratory scale UASB reactors is addressed in this study and guidelines are provided for increasing the uniformity so that results are comparable across different laboratories and are also more meaningful for scale up applications of the UASB reactor process.

2. Review of Existing Solutions across Various Published Works

Despite a large number of available publications on wastewater treatment involving UASB reactors, a majority of the studies only briefly mentions constructional concepts of the reactor, and dimensional parameters are mentioned even more rarely. We collected available information on physical dimensions among existing studies in Table 2, while Table 3 shows the geometry of either hybrid reactors or where UASB reactors are installed in series with any other reactor. While building those tables we focused on the geometry of the reactor and operational conditions including substrate strength expressed as COD, Biological Oxygen Demand (BOD), etc.; loading rates; volume of reactor; and effluent recycling rates. The type of the substrate used in reported studies is provided for reference purpose only. Where Table 2 does not contain the geometric or operational parameter means that such value was not specified in the reference. Also, Tables 2 and 3 do not provide calculations based on available geometry. All information provided there is information stated in referenced publications, nothing was added. The only modifications were made to units (for COD, BOD, etc.) where they were unified across all publications.

As we can see from Tables 2 and 3, there is no uniformity in parameters of operating the UASB reactor and, and what is in our opinion even more important, information on the start-up of a laboratory UASB reactor. Such inconsistency may complicate the interpretation of results as well as the accuracy and successfulness of an experiment in general. On the larger scale, it also complicates the comparison of results obtained by different laboratories, which creates problems for feasibility studies, if the literature is the primary source of information. To be more precise, in case of a failure, incomplete information does not allow an interpretation of the data and to trace-back the reason for failure, such as unadopted inoculum or its insufficient amount, problem of biomass washout due to the geometry, problematic substrate properties, or wrong OLR or recycle ratio. Inconsistent reporting units (like OLR calculated as per total volume of reactor or per volume of digestion zone) may also harm the attempt to reproduce results of one laboratory in another one, or wrong implementation of a procedure.

Among the inconsistencies found across the studies, we also see terminology problem in the use of terms 'sludge blanket' and 'sludge bed'. The controversy of the terms is in that fact that they are:

- used interchangeably (equally)
- 'sludge bed' refers to a layer of sludge at the bottom, where it is concentrated and visually seems to be a packed layer, while 'sludge blanket' refers to a part of the reactor where sludge is swimming as flocs above the 'sludge bed'
- 'sludge bed' refers to a bottom layer of sludge, and uses 'transition zone' instead of a 'sludge blanket'

Below, in Table 4, we attempted to systematize all parameters we were able to identify in the publications reviewed. Information in Table 4 does not intend criticize, but instead the intent is to generalize and categorize information from publications referenced above.

#	Substrate	Operating Conditions	Used Type of Reactor, Material, Inoculum, and Seeding	Constructional Geometry	Reference
1.	Hydrous ethanol vinasse COD: 121,000 $\frac{mg}{L}$ pH: 4 Ethanol: 21,007 $\frac{mg}{L}$ Acetic acid: 2237 $\frac{mg}{L}$ Propionic acid: 4304 $\frac{mg}{L}$	HRT: 6 15 days OLR: 7.27 22.16 $\frac{k_{gCOD}}{m^3.day}$ Start-up OLR: Days 1–6: 0.34 $\frac{k_{gCOD}}{m^3.day}$ with synthetic wastewater Days 7–8: 5.9 $\frac{k_{gCOD}}{m^3.day}$ with substrate	<i>Type</i> : Figure 1b with added extra high rate settler above gas collecting part <i>Material</i> : Acrylic <i>Inoculum</i> : taken from already functioning UASB reactor treating vinasse of banana waster.	Cylindrical part diameter: 11 cm Cylindrical part height: 35 cm Settler basement square side: 17 cm Settler height: 21 cm Settler is installed on top of cylindrical part. Settler plates incline: 45° Operational Volume: 3 L No sampling ports	[29]
2.	Distillery Spentwash pH: 3.8 4.2 COD: 122,000 $\frac{mg}{L}$ TS: 121,020 $\frac{mg}{L}$	HRT: 10 days OLR: 11.75 $\frac{kgCOD}{m^3 \cdot day}$	<i>Type</i> : Figure 1a <i>Material</i> : Acrylic <i>Inoculum</i> : Laboratory enriched sludge from ongoing reactor by cow dung slurry. <i>Seeding</i> : Seeding by filling the 50% of volume with sludge mixture and multiple dilution by wastewater sample.	Digestion zone: $10 \times 10 \times 98$ cm Transition zone: $10 \times 10 \times 6$ cm GLS zone: $19.2 \times 19.2 \times 25$ cm Digester volume: 16.75 L Settler volume: 7.15 L GLS opening angle: 53 °C 8 sampling ports with 10 cm spacing	[42,43]
3.	Spent wash of distillery plant COD: 90,000 100,000 $\frac{mg}{L}$ BOD: 30,000 50,000 $\frac{mg}{L}$ pH: 3.5 4.5	pH is adjusted to $6.5 \dots 7.5$ with lime (Ca(OH) ₂) Dilution of substrate applied OLR: $5.63 \dots 9.5 \frac{k_{\rm g}{\rm COD}}{m^3 \cdot day}$ Temperature: $36 \dots 40 ^{\circ}{\rm C}$ Suggests to adjust the ratio COD:N:P as 300:5:1 with urea and diammonium phosphate	<i>Type</i> : Figure 1a <i>Material</i> : Acrylic <i>Inoculum</i> : active sludge from anaerobic reactor <i>Inoculation</i> : 3 L of sludge per reactor	Operational Volume: 10 L Tube I: 11.7 cm Full height: 97 cm Digestion Zone: 78 cm Several sampling ports as 5, 19, and 57 cm levels Extra Sludge washing port Deflectors angle: 55 °C GLS opening angle: 55 °C	[44]
4.	Municipal sewage	Temperature: 9 32 °C HRT: 6 h OLR: 2.4 ^{kgCOD} m ^{3.day}	<i>Type</i> : Figure 1a Pre-existing functioning UASB reactor	Full Volume: 1148 L Height: 4 m	[45]
5.	Synthetic wastewater based on unbleached pulp mill COD: 1400 $\frac{m_g}{L}$ pH: 6.3 8.3	Temperature: 30 ± 1°C HRT: 30 h	<i>Type</i> : Figure 1b <i>Inoculum</i> : granulated sludge from UASB reactor treating poultry slaughterhouse effluent	Total volume: 15 L Digestion compartment: ID: 15 cm Height: 52 cm Settler cylindrical and conical compartment: ID: 15 cm Height: 30 cm	[46]

Table 2. Overview of available information of UASB reactors used in laboratory studies.

#	Substrate	Operating Conditions	Used Type of Reactor, Material, Inoculum, and Seeding	Constructional Geometry	Reference
6.	Vanderbilt mineral medium with tetrachloroethylene COD: 3500 $\frac{mg}{L}$	Temperature: $35 \pm 2 ^{\circ}\text{C}$ OLR: $10.5 \frac{gCOD}{m^3 \cdot day}$ HRT: 0.4 day	<i>Type</i> : Figure 1a <i>Material</i> : Stainless steel <i>Inoculant</i> : Flocculent anaerobic biomass from anaerobic Continuous Stirring-Tank Reactor (CSTR) <i>Seeding</i> : 350 mL of sludge, equal to 8 g/L of TSS	Total Volume: 2 Liters ID: 9 cm Total Height: 100 cm GLS height: 15 cm Equipped with 5 sampling ports and inner heater	[47]
7.	Municipal sewage COD: 97 196 $\frac{mg}{L}$ pH: 6.8 7.2	Ambient temperature: 24 28 °C	<i>Type</i> : Figure 1a <i>Material</i> : Acrylic Sheets	Total volume: 62 L Total height: 270 cm Non-cylindrical form Sludge bed: height: 80 cm Square cross-section: 16 cm Gas collector slope: 60°	[48]
8.	Municipal landfill leachate COD: 1.5 3.2 $\frac{g}{L}$ pH: 6.5 7.0	Ambient temperature: 13–23 °C OLR: 1.2 4 $\frac{kgCOD}{m^3 day}$ HRT: 35 15 h Added NaHCO ₃ as 0.5 g/L for neutralization purpose and no extra pH adjustment was done Recycle ratio: 3.5:1 Reports escape of methane with effluent	<i>Type</i> : Figure 1a <i>Material</i> : Stainless steel with PVC tubing and insulated with poly-urethane sheets <i>Inoculum</i> : Mesophilic anaerobic sludge from sewage treatment plant	Height: 295 cm Diameter: 13.5 GLS height: 50 cm Total Volume: 40 L Recycle ratio: 3.5:1 (feed to recycle) 2 sampling ports Contained the heater	[49]
9.	Grey water from sewer pipe COD: 647 681 $\frac{mg}{L}$	HRT: 8 20 h Ambient temperature: 14 24.5 °C	<i>Type</i> : Figure 1a <i>Material</i> : not specified <i>Inoculum</i> : sludge from anaerobic digester treating primary and secondary sludge	Full volume: 7 L Diameter: 7 cm Total height: 200 cm GLS height: 50 cm Reactor sludge filling: 100 cm	[50]
10.	Municipal wastewater COD: 672 698 $\frac{mg}{L}$	HRT: 2.4 4 h Temperature: set of reactors operating in range 12 25 °C as water bath made of PVC pipe Ø30 cm	<i>Type</i> : Figure 1c <i>Material</i> : PVC <i>Inoculum</i> : not specified	Full volume: 25 L Height: 1.35 m ID: 15 cm Inclined arm angle: 45° 4 sampling ports	[51]
11.	Sanitary waste + aerated filter effluent COD: $351 \pm 166 \frac{mg}{L}$	HRT: 6 h Experiment duration of 120 days	<i>Type</i> : Figure 1a <i>Material</i> : not specified, but either PVC or PMMA, based on provided images <i>Inoculum</i> : not specified	Cylindrical (tubular) shape Full volume: 7.8 L Total height: 60 cm Diameter: 14.8 cm 2 sampling ports GLSS opening angle: ~60° Height from top to baffles: 15 cm	[52]

#	Substrate	Operating Conditions	Used Type of Reactor, Material, Inoculum, and Seeding	Constructional Geometry	Reference
12.	Municipal wastewater COD: 176 224 $\frac{mg}{L}$	Temperature: 20 28 °C HRT: 3 h OLR: 0.014 $\frac{m_gCOD}{Lday}$ or 0.009 $\frac{m_gVS}{Lday}$	<i>Type</i> : Figure 1a <i>Material</i> : Not specified Already existing and functioning reactors	Total Height: 3.85 m Total volume: 2.5 m ³ 3 sampling ports Separate preheater of substrate before inlet point	[15]
13.	Sugar cane vinasse COD: 19,220 $\frac{mg}{L}$ sCOD: 15,300 $\frac{mg}{L}$ pH: 5.2	Temperature: $22 \pm 3 \text{ °C}$ OLR: 0.5 $32.4 \frac{kgCOD}{m^3 \cdot day}$ Up-flow velocities: 0.008 0.292 $\frac{m}{h}$ HRT: $33.33 \dots 0.86$ days Recycling ratio: 1:3 Added 0.3 g NaHCO ₃ per 1 g of COD to adjust the pH and alkalinity.	<i>Type</i> : Figure 1c <i>Material</i> : PVC <i>Inoculum</i> : Granular sludge from UASB treating poultry slaughterhouse <i>Seeding</i> : 60 L of granular sludge of VVS content 37 g/L	Total volume: 120 L Reaction zone volume: 60 L Total height: 4 m Reaction Zone Height: 2 m Diameter: 19.5 mm 8 sampling ports	[53]
14.	Mix of domestic waste with molasses (0.5:785 mix ratio) COD: 6597 $\frac{mg}{L}$ BOD: 3197 $\frac{mg}{L}$ TSS: 4500 $\frac{mg}{L}$	OLR: 6 $\frac{kgCOD}{m^3 \cdot day}$ (1.5 start-up) Temperature: 15 25 °C HRT: 10–12 h Vertical velocity: 0.5–0.7 m/h	<i>Type</i> : Figure 1a <i>Material</i> : UPVC Pre-existing reactors	Diameter: 25 cm Height: 2 m 4 sampling ports Volume: 98 L	[54]
15.	Pre-digested chicken manure pH: ~8.0 COD: 807 ± 215 $\frac{mg}{L}$ sCOD: 295 ± 46 $\frac{mg}{L}$	Feed rate; 500 mL/day/reactor + dilution with tap water HRT: 13 days Semi-continuous operation	<i>Type</i> : Figure 1a <i>Material</i> : plexiglass <i>Inoculum</i> : sludge for internal circulation reactor treating paper/cardboard industry waste <i>Seeding</i> : 1.3 L sludge per reactor (20% of working volume)	Digestion zone height: 1 m Diameter: 90 mm Volume: 6.5 L Extra sampling ports	[55]
16.	Synthetic wastewater with butyrate as a main substrate pH: 6.0–6.5 COD: 2100–15,500 $\frac{mg}{L}$	Temperature: 37 °C OLR: 4–83 $\frac{kgCOD}{m^3 day}$ HRT: 12.5–4.5 With water-jacket pH: 7.1–7.9 by addition of NaHCO ₃	<i>Type</i> : Figure 1a <i>Inoculum</i> : flocculant sludge from anaerobic sludge digester, partially granulated in pilot-scale reactor for 2 month growing on sucrose <i>Seeding</i> : 1.5 L of adapted sludge per reactor	Digestion zone height: 50 cm Digestion zone diameter: 8.4 cm Digestion zone volume = 2.8 L Settler zone height: 25 cm Settler zone diameter: 11.4 cm Settler zone volume: 2.0 L 5 sampling ports	[27]

#	Substrate	Operating Conditions	Used Type of Reactor, Material, Inoculum, and Seeding	Constructional Geometry	Reference
17.	Synthetic wastewater COD: 6000–20,000 $\frac{mg}{L}$ pH: 7.1–7.8 (caused by buffers in WW)	HRT at beginning: 12–1.8 h OLR: 18–260 $\frac{k_SCOD}{m^3.day}$ Increasing OLR by 50% after each achieving of removal rate of 80% Preheating of substrate: 37 °C Alkalinity spiked with NaHCO ₃ Volumetric loads calculated per digestion zone volume only	<i>Type</i> : Figure 1a <i>Inoculum</i> : anaerobic digester treating municipal wastewater <i>Seeding</i> : 6.5 L of inoculum (1.% VSS and 1.3 TSS)	Volume: 8.5 L Digestion (+ 5 L of GLS) Digestion zone ID: 104 mm Digestion zone H: 1000 mm GLS: ID 144 mm GLS: H 300 mm 7 evenly distributed sampling ports	[56]
18.	Municipal sewage pH: 4.4 COD: 531 $\frac{mg}{L}$ BOD: 359 $\frac{mg}{L}$	Temperature: 25 35 (ambient) Feed rate: 28 L/day Up-flow velocity: 0.116 m/h OLR: 1.062 $\frac{kgCOD}{m^3 day}$ pH adjusted with NaOH up to 6.7 ± 0.1 Reported granulation on 20 th day for main experiment.	<i>Type</i> : Figure 1a <i>Material</i> : Glass sheets <i>Inoculum</i> : adjusted cow dung manure Inoculum adaptation: 9 L of inoculum + 1 L of nutrients, grow for 120 days growing on sucrose with (NH ₄) ₂ HPO ₄ . Remove undigested residuals via filtering through the 3 mm mesh. <i>Seeding</i> : 4 L of filtered sludge from adaptation per reactor	Working volume: 14 L Length to height ration: 1:14 Height: 140 cm (it is not cylindrical) Length of base: 10 cm Area of reactor: 100 cm ²	[57]
19.	Distillery effluent from fermentation-based vitamin C production plant COD: 6000 38,000 $\frac{mg}{L}$ BOD ₅ : 2000 14,000 $\frac{mg}{L}$ pH: 4.5–6.2	$35 \pm 2 \ ^{\circ}C \ (constant \ temperature \ room) \\ OLR: 6 \dots 11.8 \ \frac{kgCOD}{m^3 \cdot day} \\ Upflow \ velocity: \ 0.52 \ m/h \\ HRT < 10 \ h \\ 252 \ days \ of \ total \ experiment \\ 100 \ days \ of \ start-up \\ Adjusted \ COD:N:P \ as \ 300-600:5:1 \\ with \ urea \ and \ KH_2PO_4 \\ pH \ adjusted \ with \ NaOH \ up \ to \ 7.2 \\ \end{cases}$	<i>Type</i> : Figure 1a <i>Material</i> : stainless steel <i>Inoculum</i> : sludge (VSS 31.0 g/L) from anaerobic digester treating the municipal wastewater	Active volume: 2.3 m ³ Height: 5.90 m Inner diameter: 0.8 m Conical shape of bottom 5 sampling ports Recycle line active	[58]
20.	High-strength distillery wastewater pH: 3.42–5.84 TS: 31,520–126,240 $\frac{mg}{L}$ TSS: 1040–26,640 $\frac{mg}{L}$ COD: 68,000–100,000 $\frac{mg}{L}$ BOD: 21,600–35,000 $\frac{mg}{L}$	HRT: 4 2d Temperature: 37 °C with water jacket OLR: 15.34 $\frac{k_gCOD}{m^3 \cdot day}$ Experiment duration: 635 days Start-up: 65 days with HRT: 47.11 h	<i>Type</i> : Figure 1a <i>Material</i> : borosilicate glass <i>Inoculum</i> : the sludge from UASBR treating distillery wastewater <i>Seeding</i> : ~30% of reactor volume	Inner diameter: 92.1 mm Total Height: 79.6 cm Digestion zone: 59.97 cm Digestion volume: 5 L Water jacket ID: 132.10 mm Sampling ports: 6 GLSS opening angle: 70° (flipped funnel)	[59]

#	Substrate	Operating Conditions	Used Type of Reactor, Material, Inoculum, and Seeding	Constructional Geometry	Reference
21.	Distillery wastewater COD: 107,000 $\frac{mg}{L}$ TOC: 39,200 $\frac{mg}{L}$	Temperature: 37 °C Flow rate: 2.2 3.3 L/d Recycle: 50% of influent flow NaHCO ₃ added as 3 g/L to adjust pH to 7 OLR: up to 3 $\frac{kgTOC}{m^3.day}$	Type: Figure 1aSeeding: 500 mL of inoculum per reactor, + 3 Lof glucose-based synthetic wastewater andnutrients, including Ca^{2+} and PO_4^{3-} tostimulate granulation	Total Height: 1.35 m Fluidization part volume: 3 L Fluidization part H: 1.05 m Fluidization part ID: 5.9 cm Settling part volume: 3 L	[28]
22.	Cane molasses vinasse COD: 10 $\frac{g}{L}$ pH: 4.1 COD: 120 $\frac{g}{L}$ BOD: 30 $\frac{g}{L}$ TS: 100 $\frac{g}{L}$ TS: 50 $\frac{g}{L}$	Temperature: 55 °C with water jacket OLR: up to 28 $\frac{kgCOD}{m^3 \cdot day}$ Experiment duration: 430 days Added 5 g/L of NaHCO ₃ to maintain 7.3 pH all-over experiment.	<i>Type</i> : Figure 1b with extra settler above gas collector <i>Material</i> : Stainless steel <i>Inoculum</i> : sludge (12 g VS/L) from suspended growth type digester treating distillery wastewater <i>Seeding</i> : 87 L of sludge per reactor	Volume: 140 L (126 L digestion + extra for GLSS) Digestion par: 20 cm ID × 4 m height Solids separator was made of inclined plates: 60°	[60]
23.	Wastewater with high corn-starch content pH: 6.8–7.9 COD: 3000–75,000 $\frac{mg}{L}$	Temperature: 37 °C with pre-heater HRT: 24–12 <i>h</i> and OLR: 3 150 $\frac{kgCOD}{m^3 \cdot day}$ Experiment duration: 510 days OLR is calculated on the volume of digestion zone only. pH adjusted with NaHCO ₃ equal to COD, but < 8 g/L to prevent toxicity of Na ⁺ . Reports pH of effluent as 6.8 at the highest OLR	<i>Type</i> : Figure 1a <i>Inoculum</i> : Sludge from anaerobic digester treating sewage wastewater <i>Seeding</i> : 6.5 L of sludge per reactor	Volume: 8.5 L of digestion zone + 5.0 L GLSS Digestion ID: 104 mm Digestion Height: 1000 mm GLS ID: 144 mm GLS Height: 300 mm	[61]
24.	Recycled paper mill wastewater pH: 7.4 COD: 5330.5 $\frac{mg}{L}$ TS: 32.99 $\frac{g}{L}$ VS: 27.28 $\frac{g}{L}$	Temperature: 37 ± 2 °C with helix heat exchanger Feed: 0.5–4.5 l/h, increment by 0.5 l/d OLR: 1–10 $\frac{kgCOD}{m^3 \cdot day}$ Load calculations per digestion zone!	<i>Type</i> :Figure 1a <i>Inoculum</i> : sludge from full-scale UASB <i>Seeding</i> : 25 L of sludge per reactor	Volume: 70 L (digestion zone: 53 L) Height: 1 m (30 cm of which is GLS) Diameter: 30 cm	[62]

#	Substrate	Operating Conditions	Used Type of Reactor, Material, Inoculum, and Seeding	Constructional Geometry	Reference
25.	Distillery wastewater COD: 40.389 $\frac{g}{L}$ pH: 3.2 3.8	Flowrate: $18 L/d$ Vertical up-flow velocity: $0.0925 \frac{m}{h}$ HRT: $15.6 h$ ORL: $53.75 \frac{kgCOD}{m^3 \cdot day}$ (digestion zone) pH: 6.7 ± 0.1 with NaOH Extra mixing pump inside of reactor	<i>Type</i> : Figure 1a <i>Material</i> : PVC <i>Inoculum</i> : 18 L of cow dung and 2 L of substrate and aged for 3 weeks and filtered through 3 mm mesh.	ID: 10.16 cm (4in) Height: 142.24 cm (56 in) + 14.2 cm of GLSS Effective volume: 15.4 L D: H ratio: 1:14 5 sample ports Gas collection funnel opening angle: 55°	[63]
26.	Vinasse cane alcohol wastewater pH 4.03 4.44 COD: 57.59 128.63 $\frac{g}{L}$ TS: 17.85 113.98 $\frac{g}{L}$ VS: 11.81 58.11 $\frac{g}{L}$	Temperature: $35 \pm 2 ^{\circ}\text{C}$ OLR: varied $1 \dots 6 \frac{kgCOD}{m^3 \cdot day}$ HRT: $109 \dots 25 \text{days}$ Up-flow velocity: $2 \dots 3 \frac{m}{h}$ Biogas cleaned with $3N$ NaOH solution	<i>Type:</i> Figure 1a <i>Material:</i> Glass <i>Inoculum:</i> sludge from wastewater treatment plant treating mix of urban and industrial wastewater <i>Seeding:</i> 600 mL of inoculum resulting in 10.63 g VS/L in reactor	Digestion part: 53 cm H × 7.5 cm ID Digestion part volume: 2.3 L 6 sampling ports	[64]
27.	Distiller's grains wastewater COD 16,500–22,520 $\frac{mg}{L}$ pH 3.3–4.3 VFA: 3000–3600 $\frac{mg}{L}$ VSS 190–640 $\frac{mg}{L}$	pH: ~7 with NaHCO ₃ OLR: 3.2 48.3 $\frac{kgCOD}{m^3 \cdot day}$ (33.3 was optimal) No reactor heater, the substrate was preheated to 37 °C before entering the reactor Start-up OLR: Linear increase from 0.42 to 5.6 $\frac{kgCOD}{m^3 \cdot day}$ for 27 days	<i>Type</i> : Figure 1a with second level of gas collectors as on Figure 1b <i>Material</i> : acrylic Inoculum: Sludge from mesophilic anaerobic digester in sewage treatment plant <i>Seeding</i> : Seeded with 5.2 L of sludge with VSS content of 12.3 g/L, degassed by auto-incubation at 37 °C for three weeks.	6 sampling ports with spacing of 20 cm in between. Inner diameter of Tube: 8.2 cm Height: 190 cm (total), 155 cm (reaction zone) Total volume: 8.18 L Inner diameter of GLS: 14 cm OD of gas harvesting funnel: 10 cm Funnel opening angle (60°) Duration of experiment: 420 days	[65]

#	Substrate	Operating Conditions	Type, Material, Inoculum, Seeding	Geometry	Reference
1.	Distillery spent wash pH: 4 4.5 COD: 80,000 12,000 $\frac{mg}{L}$ TS: 60,000 85,000 $\frac{mg}{L}$ BOD ₅ : 35,000 45,000 $\frac{mg}{L}$	Temperature: 20 40 °C (ambient) pH: ~7 with NaHCO ₃ Substrate COD:N:P as 100:5:1 with NaH ₂ PO ₄ and Urea OLR: 1.0 8.0 $\frac{k_{S}COD}{m^3 \cdot day}$ (start-up), 36 $\frac{k_{S}COD}{m^3 \cdot day}$ HRT: 6 48 h Observed granulation at day 50.	Type: Figure 1a with packing materials. So called Hybrid UASB reactor Material: PMMA Inoculum: flocculent sludge from anaerobic digester of sewage treatment plant Seeding: sieved through 1 mm mesh, loaded as 15 g VSS/L (2.5 L per reactor)	Operational liquid volume—5 L (45 cm of total height) Diameter: 10 cm Overall height: 77 cm GLS separator was replaced with packing, taking 19 cm of total height (volume 1.5 L)	[14]
2.	Tannery wastewater COD: $8600 \dots 14,100 \frac{mg}{L}$ pH: 2.8 3.7	Temperature: 17 38 °C (ambient) Substrate was diluted to COD value of 5400 9400 $\frac{mg}{L}$ Experiment duration of 52 weeks Equalization tank (600 L) prior to 1 st stage UASB Start-p OLR: 24 <i>h</i> HRT: 5 24 <i>h</i>	Type: two reactors as Figure 1a in line Material: UPVC	Volume: 94 L Total height: 325 cm Digestion zone height: 240 cm Tube ID: 20 cm 5 sampling ports every 55 cm Funnel overlap on baffles: 1.5 cm per side	[66]
3.	Molasses-based ethanol distillery wastewater	HRT: 70 h, treating as 2 nd stage after CSTR Feed flow: 3.4 L/d Temperature: ambient	Type: Figure 1a	Digestion Volume: 10 L Digestion ID: 0.08 m Digestion Height: 1.5 m	[67]

 Table 3. Information on UASB modifications of multi-step reactors involving UASB.

Criteria	Options/Area of Application/Observations
Height	No constraints on height. The smallest found reactor was 30 cm tall, the largest as above 4 m. Perhaps, limited only by the available space in
	a laboratory.
Volume	Small volumes are 0.5, 0.75, 1, and 2–2.4 L. Larger volumes of 14 and 55 L
	were also found. Usually, reactors with volume greater than 1 m ³ are
	referred as pilot-scale.
Height: Diameter (H:D) ratio	Since the substrate has an up-flow velocity, the reaction part of UASB
	reactor in some degree functions as a sedimentation or coagulation
	column, where ratio H:D should help preventing the biomass
	washout [19].
	This parameter is very rarely reported, and reporting of it can be
	confusing due to not clear geometry reference. There are reports of H:D
	height of a reactor without CLS. We see reasonable to calculate it as a
	diameter of reactor to the height without CLSS since the goal of CLSS is
	to create a chamber for gas capture above the reaction tube of a reactor
	From review studies, such ratio for majority of cases is in range from 8 to
	14. However, there are also extreme cases as 3.5 or 23.
Construction material	For small volumes (up to couple liters): Borosilicate glass
	For small and medium sizes: PVC and PMMA
	For pilot scale: Stainless steel.
Gas-Solid-Liquid Separator	The particular design varies with the concept of the reactor itself, and
(GLSS, Three-Phase Separator)	options can be split into:
	Implementation of baffles
	Gas collection
	Origon with a triangular cross-section. For rectangular reactors, a series
	of inclined baffles are installed to parrow a main liquid flow
	For smaller reactors, baffles are sometimes omitted, probably, because it's
	difficult to implement those in smaller volumes. Another case when
	deflectors were noticed to be omitted is when GLSS is represented by a
	separate part (either tube or funnel), wider than the major reaction tube,
	and a diameter of a gas collector is close to a diameter of a reaction tube.
	Gas collector is usually represented by a flipped upside-down funnel for
	smaller reactors. For larger ones, it can be a separate compartment.
	Y-shaped reactors do not have any specific structure inside.
Heating	Among the reviewed designs the following heating systems were noticed:
	Heating pads of tapes
	No beating
	Inner heaters (helix shaped)
	Water jacket is the most common option for smaller designs but it
	complicates the placement/insertion of sensors (like pH, ORP,
	temperature, etc.) into a reactor. Larger reactors usually use heating pads
	or a combination of heating pads with thermal insulation material.
Temperature ranges	Mesophilic: $35 \pm 2 \degree C$ or $37 \pm 1 \degree C$
	Thermophilic: $55 \pm 1 ^{\circ}\text{C}$
	Ambient temperature
Inoculum material	Ambient temperature with thermostat to prevent overcooling
moculum material	Cranular or flocculated sludge from another UASB
	Non-granulated anaerobic or active sludge
	Adjusted inoculum from non-sludge sources, like animal manure
Seeding (inoculating)	Across the reviewed studies, this was the most inconsistent parameter,
	which was not even always reported. The process was reported as:
	(a) filling reactor with raw sludge up to a certain percentage of height;
	(b) volumetric load of sludge per reactor, sometimes mentioning its VSS
	and/or TSS equivalent; and (c) final concentration of sludge in reactor as
	TSS or VSS. Also, few studies suggested to sieve the sludge through
	1–3 mm mesh to remove any undigested particulate or residuals
	before seeding.

Table 4. Summary of the geometry and operational parameters for existing UASB reactor designs.

Criteria	Options/Area of Application/Observations
Substrate preparation, feeding and	Few studies considered the adjustment of substrate based on ratio
pH management	COD:N:P. However, the final ratio does not match across publications and
	varies for COD parameter 300-600:5:1. Surprisingly, no-one mentioned
	adjusting the C:N ratio, which is recommended for anaerobic treatment in
	general. Only one publication mentioned the addition of compounds to
	stimulate granulation.
Substrate pH management	Researchers use either pH adjustment in substrate directly or pumping
	pH adjusting solution to the reactor. Used adjusting compounds are either
	hydroxides or bicarbonates. Interesting fact: addition of 0.5–3 g of
	NaHCO ₃ per 1 L of substrate was sufficient to maintain a stable effluent
	pH around 7. In some extreme cases 8 g per 1 L of substrate were
	sufficient to work with OLR 150 $\frac{kgCOD}{m^3 \cdot day}$.
OLR and HRT	HRT and OLR are interdependent values and both are optimization points
	in research. Researchers aim to increase OLR and decrease HRT.
	These parameters are points of inconsistent reporting:
	Some sources report OLR and HRT as referred to the total volume of the
	reactor (both reaction tube and GLSS)
	Some sources report OLR and HRT as referred to the volume of the reactor
	without the volume of GLSS
	Higher limit for OLR is not specified, since it depends on chemical
	composition of influent wastewater and its strength.
Substrate distribution system	Typically is not reported, but where it is mentioned it's either:
	a circular tube with evenly distributed outlet holes and an inlet from the
	side through the wall of reactor
	an inlet into conical-shaped bottom of reactor
	a side injet through the wall into bottom compartment with
	inclined bottom

3. Discussion

Studies, involving the UASB trials, are usually purposed for: (a) treatability testing and energy recovery estimation; (b) microbiology studies on changes in microbial consortia during adaptation to new substrates or long-term operation for further modeling of trophic network; or (c) toxicity and granulation process studies. In the scope of this manuscript, we would like to identify the common needs of such research and point out the differences, where it is important. Here, we would like to focus on experimental aspects, which are needed to pay attention to, while designing the reactors and its infrastructure.

3.1. Volume of Reactor

The first thing that affects the final volume of a designed reactor is the available amount of sample/substrate. Some samples of substrates are available in very limited quantities due to the policies of supplier companies or may be a subject of special regulations preventing the dumping of effluent to a sewer (Ex. industrial wastewater). Depending on the complexity of substrate and potential inhibitory effects, the reactor start-up period might occupy a substantial period up to 120 days [14,57,59,65,68], thus the volume of a reactor should allow to utilize the available sample volume for both start-up period and experiment duration.

3.2. Material of the Reactor

Due to the specifics of laboratory studies, the reactor needs to be constructed with the feature to visually inspect the content. It allows one to: (a) confirm the fact of granulation and (b) inspect the foam or scum formation, etc. This significantly narrows the selection choice of materials, limiting it to (a) polymethyl methacrylate (a.k.a., PMMA, Plexiglas, Perspex, acrylic glass), (b) borosilicate glass, and (c) clear polyvinyl chloride. Each of the mentioned materials can be used, and in our opinion it is more of a question of budget and available stock parts. We compare pros and cons of each material in Table 5.

Material	Pros	Cons
PMMA	Less expensive than glass	Needs machining equipment
	Almost no film formation (unless scratched)	If sterilization is needed: consider chemical
	Optically clear, may have some UV-protective coating	sterilization
	Stronger that glass	Easily scratchable
	Machinable with mechanical tools	
	(CNC/lathe/mill/drill)	
	Cracks can be fixed with either solvent treatment,	
	epoxy of UV-curable resin in short time frame	
Glass	Optically clear	Expensive
	Non-UV degradable, chemically inert (under	In case of cracks becomes sensitive to vibrations and
	conditions of AD)	not usable
Washable		Requires specialist (glass blower) to build or repair/fix
	Autoclavable	Fragile
PVC	Clear, but not optically. Has blueish color	Degrades over time, becomes fragile
	Machinable, but melts easily	Non-UV-stable, becomes yellowish over long-term
	Relatively cheap and available on the market, has a	expose to light containing UV spectrum (sunlight)
	wide set of existing fittings for quick assembly	Microorganisms form biofilm on its surface

Table 5. Comparison of materials used for UASB reactors in various studies.

Borosilicate glass is an excellent option if used for studies with sterile cultures, since it can be autoclaved. However, in the author's opinion, the ideal reactor must be manufactured of stainless steel and be featured with an inspection window, a water jacket and multiple sampling ports. Such a design would be chemically resistant under conditions of anaerobic digestion, autoclavable, and meet multiple research needs. However, such construction complicates the customization and should be done for optimized and fully tested design after confirmation of its efficiency. The authors currently use PMMA due to machinability of this material, its transparency, and stability under conditions of anaerobic digestion (AD).

3.3. Heating of Reactor

Heating of reactor under laboratory conditions is defined by: (a) actual need for heating and (b) necessity of sampling the content of reactor and location of sampling. If no sampling of reactor content is needed, the water jacket would be the most suitable option. Otherwise, sampling ports complicate the construction of water jacket. Without a water jacket, consider: (a) use of heating tapes or flexible heating pads or (b) preheating of substrate and thermal insulation of reactor to keep the temperature.

Heating tape on the outer surface of PMMA or PVC reactor is not recommended, since it could cause local damage, when the contact point of wall material and heating tape is locally overheated, causing melting or other types of damage. Our laboratory experienced problems with heating tapes even under mesophilic conditions. The reactor that got damaged, was controlled by thermostat with an external submergible temperature sensor. The damage consisted of the tape melting through the wall of the reactor causing leakage. Thus, we moved to a water jacket in our projects.

Perhaps, the use of heating pads would be more secure due to a larger area of contact and, hence, more uniform heating. Extra uniformity may be added by use of heat-transfer pastes, but they will decrease the observability of the process in reactors. However, it is still a viable option when there is a need for the presence of sampling ports on various levels or there is no way to implement a water jacket due to other reasons.

3.4. Inoculum: Preparation, Adaptation, and Seeding

While the granulated anaerobic sludge is the desirable inoculum, authors fully realize the probability of a situation when researchers do not have a source for granulated sludge. In such a case, the manure sample of animal origin could be a source of methanogenic microbial consortia, and referenced studies [27,42,43,57,63,65] suggest self-digestion of such sample or mixing it with a substrate and conditioning for up to 3 weeks. The presence of methanogenic microorganisms is

required for the generation of methane, but not every manure contains methanogens. The most typical confirmed cases of manure containing methanogens are cattle and swine manures. The presence of methanogenic bacteria could be confirmed by conducting specific methanogenic activity test [69,70], which is very close in technique to a popular Bio-Methane potential (BMP) test [71], but conducted on a nutrient media containing acetate as the only source of carbon [72].

Some studies suggest the sieving of inoculum through a 1–3 mm mesh to remove undigested or large inert material. It is reasonable, if the inoculum originated from manure, since manure samples may contain some animal bedding, or sewage wastewater treatment facility, which may contain hair, etc. However, if the sample originated from an industrial wastewater treating facility, such sieving could be optional, especially if sludge is already granulated and granules are large. Also, the effect of exposing sludge or granules to air during the sieving is not clear. Perhaps, the sieving process should be done in anaerobic chamber.

The seeding of reactor must be calculated and expressed as Volatile Suspended Solids (VSS), introduced with the inoculum, per working volume of reactor according to [73,74] and seeding should be in the range 10 to $20 \frac{kg_{VSS}}{m^3}$, (however, it also could be up to $25 \frac{kg_{VSS}}{m^3}$) [10]. Inoculum should be analyzed for Total Solids, (TS), Volatile Solids (VS), Total Suspended Solids (TSS), and Volatile Suspended Solids (VSS) since sludge is also characterized by VS:TS and VSS:TSS ratio, as criteria of alive biomass if condition of sludge is tracked over time [75] and ratio VSS:TSS of sludge in range of 0.7 to 0.85 is likely to cause granulation [59]. The recommended method for solids content analysis is specified in Standard Methods 2540 [76].

3.5. Substrate Adjustment

Before any adjustments is done to a substrate, the treatability can be roughly characterized by the ratio BOD₅:COD, which is referred to as a biodegradability index (BI) [77,78]. For municipal raw wastewater BI is usually in the range 0.4 to 0.8 [79,80] and it is considered to indicate good treatability. Greater index means better bio-treatability and pretreatment can increase the value of BI [81,82]. Estimation of sample degradability based on BOD₅:COD is inconsistent, but can be generalized as: (a) highly bio-treatable if greater than 0.5; (b) bio-treatable if greater than 0.3; and (c) not bio-treatable when lower than 0.2 [81–85].

One of the primary adjustments for substrate is the C:N ratio [86] by mass, with the optimum in the range 25 to 30 [87–89] or 20 to 30 [90] and higher temperature ranges require higher C:N ratio. However, it also could be a substrate-specific optimization parameter [91–93]. Some authors also consider C:P ratio for methanogens between 16:1 and 75:1 [94,95] as optimal, while C:N:P ratio is considered to be favorable in the range 400:5:1 to 100:28:6 [95,96]. Some inconsistence to in attempt to meet those ranges may come from measurement techniques, where various authors use either: (a) elemental analyzers [97] or (b) total carbon and Total Kjeldahl Nitrogen (TKN) [90]. Across referenced in this manuscript studies, following compounds were used to correct ratio: KH_2PO_4 , NaH_2PO_4 , $CO(NH_2)_2$ (urea or carbamide), and $NH_4H_2PO_4$.

Individual studies stated the need to account for sulphur [98], and report C:N:P:S ratio as 600:15:5:3 to be the optimal for methanization [99]. Perhaps, such increase of considered elements is reasonable, since elemental composition of anaerobic biomass is reported as $C_5H_7O_2NP_{0.06}S_{0.1}$ according to [100–102], but not yet widely used in anaerobic digestion studies.

The ratio of COD:N:P of 250:5:1 is generally suggested for anaerobic treatment [103–105], however, some variation exist between 900:5:1.7 and 150:5:1 [104,106,107] and could be even 300:1:0.1 [108]. Other studies recommend 300:5:1 as start-up conditions specifically for UASB [75,96,109,110]. Important to mention, that "N" in such proportion refers to the total nitrogen [108]

pH adjustment for methanogenic bacteria should bring the pH in the optimal range 6.8 to 7.5, while outside of the 5 to 8.5 range, methanogenesis is fully suppressed [111–113]. However, for anaerobic digester the range of 6.8 to 7.2 is recommended, due to widely used in wastewater

treatment lime as pH adjusting chemical [114]. Across referenced studies we noticed NaOH and NaHCO₃ as widely usable compounds to adjust pH, however, the choice is wider [115].

3.6. Granulation Stimulation

If granulation enhancement is needed, the Ca^{2+} in concentration 100–200 mg/L of substrate can be added [116], or even 150... 300 mg/L at the start-up [117,118]. The role of calcium in granulation process is not clear, but it is assumed to form precipitates with carbonate and phosphate [21,119]. Use of Mg^{2+} is not recommended, since it causes disaggregation of granules [120], even though it is expected to precipitate as $MgNH_4PO_4$ [10]. Normally, granulation should be observed within 4–6 weeks after the start of the experiment [73].

3.7. Start-up Feeding

The original research of [73] recommends the OLR as $0.05 \dots 0.10 \frac{k_{g_{COD}}}{k_{g_{sludge VSS} \times day}}$ for the start-up period and increasing of OLR after achieving the removal rate of at least 80%, however, the increment values are not specified. The expression of COD load per VSS of sludge per day is called "sludge load", but in studies OLR is usually reported as $\frac{k_{g_{COD}}}{m^3 \times day}$, which is called "space load". Based on the previously suggested inoculum seeding as $10 \dots 20 \frac{k_{g_{VSS}}}{m^3}$, the start-up space load should be in range $0.5 \dots 2 \frac{k_{g_{COD}}}{m^3 \times day}$, however exact calculation based on the loaded VSS of inoculum must be done. The vertical velocity of the substrate is suggested not to exceed $0.5 \dots 1 \frac{m}{h}$ [10] in general, but minimal values are not reported and no details were found for the start-up period.

3.8. Infrastructure of UASB Reactor

Based on our experience and referenced here studies, we want to suggest a unified operational process flow diagram as in Figure 2, where we would like to emphasize several aspects.



Figure 2. Recommended flowchart of UASB infrastructure set-up.

3.8.1. pH Adjusting and Alkalinity

As mentioned above, the composition of pH adjusting solution is a point of choice [115], but, regardless, the solution should be pumped directly into the reactor feeding line (the mixing manifold on schematic of Figure 2.

Otherwise there is a potential for growth of competitive microorganisms in the substrate feeding tank, leading to chemical changes of substrate. The concentration of pH adjusting solution should be balanced based on: (1) daily amount of solution needed to pump (pumps may have lower limit of pumping speed) and (2) minimizing substrate dilution. The interim decision could be to use concentrated solution that is pumped and dosed on a timer, if calculated flowrates are below the limits of a pump. High concentrations of pH adjusting solutions can be chemically aggressive. To prevent the contact between the substrate and parts of pumping mechanism, the peristaltic pumps are recommended for use.

Referenced here studies dissolve or add 0.5–3 g of NaHCO₃ per 1 L of substrate and achieve the stable pH in favorable methanogenic range. However, there are more general suggestions to maintain the ratio between alkalinity of substrate (expressed as CaCO₃) and its COD as 1.2–1.6 $\frac{g CaCO_3}{g COD}$ [121]. This value can be used as a reference to calculate the dosage portions and intensity for pH adjusting solution, but should be optimized later [122] downwards. Methanogenesis could occur until ratio of 0.8 $\frac{g CaCO_3}{g COD}$ with some extreme cases of 0.3 $\frac{g CaCO_3}{g COD}$, but lower values should inhibit methanogenesis and stimulate the formation of hydrogen [123,124].

Another reference value for regulation of alkalinity is ratio of Volatile Fatty Acids (VFA) to Total Alkalinity (TA). Industrial guidelines [114] recommends the ratio VFA:TA to be below 0.35 and consider the value below 0.25 as best for anaerobic digesters, and below 0.15 as safe against pH changes in substrate. The VFA should be expressed as equivalent concentration of acetic acid in $\frac{mg}{L}$ and TA as equivalent of CaCO₃ in $\frac{mg}{L}$. The determination of alkalinity by titration is described in Standard methods 2320 [76], as well as appropriate methods for VFA with gas chromatography is covered by Standard methods 5560. However, there were attempts to substitute VFA determination by titration [125–127], to avoid using of gas chromatograph.

3.8.2. Feeding and Recycling

The feeding of reactor, based on OLR and HRT is the point of optimization targeting the maximum achievable loads, however, the [10] recommends to limit the vertical velocity of liquid depending on the type of sludge and type of waste:

- Granular sludge + soluble wastewater: $3 \frac{m}{h}$ continuous, up to $6 \frac{m}{h}$ peak for a couple of hours per day;
- Granular sludge + partially insoluble water: 1–1.25 $\frac{m}{h}$, up to 2 $\frac{m}{h}$ peak for a couple of hours per day; and
- Flocculent sludge: $0.5 \frac{m}{h}$ continuous, up to $2 \frac{m}{h}$ peak.

After sludge matured and granulated, the flow could be increased by 50%. In the case of insufficient vertical velocity and to prevent clogging, the recycle line can be used to manage it and (a) to dilute substrate with treated wastewater, (b) to reuse of alkalinity [28], or (c) enhance the granulation by increasing the vertical up-flow [128].

Important remark: effluent recycle port must be separate and located below the effluent discharge port. It is made to prevent back-pumping of air from the effluent discharge line. In our laboratory set-up, we used the flow splitters on effluent port to obtain a recycle line, and we noticed some gas bubbles in it.

3.8.3. Manual Injection Port

A manual injection is strongly recommended and is purposed for:

- urgent (emergency) injection of solution for managing pH, coagulation/flocculation, or granulation agent problems;
- testing an enhancement of inoculant via injection of specific microbial culture(s); and
- sampling of substrate which is supplied to a reactor after all mixing procedures.

3.8.4. Biogas Collection and Counting

Notice the installed one-way valves in the gas line in Figure 2.

Check valves are important to prevent the back flow of gas and there are several reasons for that particular phenomenon:

- drop of the ambient temperature and, consequently, gas compression in gas lines according to the combined gas law;
- at the beginning of UASB operation, when substrate gradually fills the reactor and gas tubes have residual air. The oxygen from residual air is consumed and thus the volume shrinks.

Any of those reasons can lead to one of the two undesirable consequences:

- ingress of liquid from reactor to a gas line, which potentially grabs the foam and clogs the pipeline.
- if water displacement gas counter is used: backflow of liquid from counter back to a reactor.

Important clarification is to use check valves with low cracking pressure. 'Cracking pressure' is a pressure value when check valve starts opening (passing gas through itself) and this pressure (converted into inches of water column) must be taken into consideration when designing the gas separator, specifically, the height and the level difference between gas collecting part and the effluent release port. Usage of valve with high cracking pressure result in need to build tall GLSS, increasing the material needed to build reactor and its dead volume.

We also want to stress that the gas counter working on the water displacement principle is the only option for raw biogas. There are gas counters working on heat transfer principle (similar to thermal conductivity detectors of gas chromatographs), which seem to be cheaper options, but they should not be used. Those counters can be calibrated for gas flow with constant content only, which is not a case for biogas. However, they can be theoretically applied if biogas was stripped with alkaline solution to remove acidic gasses (carbon dioxide, hydrogen sulfide, etc.) and assumed to be upgraded to bio-methane. We do not recommend the use of that.

3.9. Tracking Operational Parameters

The exact set of trackable parameters depend on the purpose of a particular study, but for general cases, we listed those parameters in Table 6.

Parameter	Measurement For	Used For
COD	Influent	Calculate the degradability rate of substrate
	Effluent	Calculate corrected OLR
		Reference for energy yield calculation and substrate utilization rate
pН	Influent	Estimation how favorable are conditions for methanogenesis
-	Effluent	Tracking the changes of substrate
	Reactor entrance	: The actual pH value in a sludge layer if recycle line is used
Gas	Yield volume	Estimation of yield per unit of substrate
	Content	Calculation of energy recovery
		Balancing COD on biomass growth
Flowrate	Feed rate	Calculation of OLR, HRT, up-flow velocity
	Recycle rate	- · ·

Table 6. Minimal list of parameters for tracking during UASB experiments.

These parameters are already enough to calculate the main operational parameters specified in Table 7 [10,91,129,130]:

Parameter	Equation	
Substrate utilization rate	$U = \frac{COD_{influent} - COD_{effluent}}{HRT \times VSS_{sludge in reactor}}$	(1)
Removal efficiency	$E = \frac{COD_{influent} - COD_{effluent}}{COD_{influent}} \times 100\%$	(2)
Hydraulic retention time (HRT)	$ heta=rac{working volume of reactor}{volumetric flowrate of influent}$	(3)
Organic Loading rate (space load)	$OLR_{space} = \frac{volumetric \ flowrate \ of \ influent \times COD_{influent}}{working \ volume \ of \ reactor}$	(4)
Organic loading rate (sludge load)	$OLR_{sludge} = \frac{volumetric flow rate of influent \times COD_{influent}}{volatile suspended solids of sludge in reactor}$	(5)
Up-flow velocity	$\nu = \frac{\text{influent flowrate+recycle flowrate+adjusting flowrate}}{\text{area of horizontal crosssection of reactor}}$	(6)

Table 7. Main operational parameters of UASB.

In addition, the track of biogas composition during the UASB experiments, the total gas yield and methane yield should be logged. Mentioned above parameters for logging and calculations on their basis do provide a basic understanding of ongoing process inside of UASB reactors, while interpretation of calculations result are not the scope of this manuscript to avoid swelling of it. However authors feel also a need to mention, that if some deeper understanding of chemical process or COD balancing is needed, other researchers [59,131–133] suggest calculation of what part of metabolism is presented by certain process according to the equations, collected in Table 8:

Table 8. Equations for metabolic ratios estimation.

Parameter	Equation	
Hydrolysis	$H = \frac{COD_{CH_4} + sCOD_{effluent} - sCOD_{influent}}{COD_{influent} - sCOD_{influent}} \times 100\%$	(7)
Acidification	$A = \frac{COD_{CH_4} + COD_{VFA-ef fluent} - COD_{VFA-influent}}{COD_{influent} - COD_{VFA-influent}} \times 100\%$	(8)
Methanogenesis	$M = \frac{COD_{CH_4}}{COD_{influent}} \times 100\%$	(9)
COD mass balance	$COD_{influent} = COD_{accumulated} + COD_{biogas} + COD_{effluent}$	(10)

Other parameters, not included here, belong to some partial cases of UASB experiments and are subjects of individual consideration. Examples for a category of such specialized studies could be effects of salinity or metal ions on the process inside of UASB, which would require extra electrical conductivity, ion-selective electrodes, or other quantitative measurements for both influent and effluent [134,135]. If the study is dedicated to toxicity or biodegradation of particular compound, the appropriate assay tests for that compound or its metabolites should be added [136,137], etc.

4. Conclusions

With this article we would like to draw the researcher's focus towards the need to report in their publications more information on materials and methods, including specifically sketches/operational flowcharts, seeding conditions, inoculum sources and pre-treatments, and all adjustments to the substrate and feeding equipment. The consideration and addition of these details will help to facilitate a strong scientific and engineering community with comparable research results and conditions. Such detailed data and methods reporting will also significantly propel modeling studies that aim to realistically predict bioreactor behavior in various process conditions.

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Abbreviations

AD	Anaerobic Digestion	sCOD	Soluble COD
BOD	Biological Oxygen Demand	SRT	Solids Retention Time
COD	Chemical Oxygen Demand	TA	Total Alkalinity
CSTR	Continuous Stirred-Tank Reactor	TKN	Total Kjeldahl Nitrogen
EPS	Exocellular Polymeric Substances	TOC	Total Organic Carbon
GLSS	Gas-Liquid-Solids Separator	TS	Total Solids
HRT	Hydraulic Retention Time	TSS	Total Suspended Solids
ID	Inner diameter	UASB	Up-flow Anaerobic Sludge Blanket
OD	Outer diameter	UPVC	Unplasticized Polyvinyl Chloride
OLR	Organic Loading Rate	VFA	Volatile Fatty Acids
PMMA	Polymethyl methacrylate	VS	Volatile Solids
PVC	PolyVinyl Chloride	VSS	Volatile Suspended Solids

References

- 1. Saleh, M.M.A.; Mahmood, U.F. UASB/EGSB Applications for Industrial Wastewater Treatment. In Proceedings of the Seventh International Water Technology Conference Egypt, Cairo, Egypt, 1–3 April 2003; pp. 335–344.
- 2. Seghezzo, L.; Zeeman, G.; Van Lier, J.B.; Hamelers, H.V.M.; Lettinga, G. A review: The anaerobic treatment of sewage in UASB and EGSB reactors. *Bioresour. Technol.* **1998**, *65*, 175–190. [CrossRef]
- 3. Yoochatchaval, W.; Ohashi, A.; Harada, H.; Yamaguchi, T.; Syutsubo, K. Characteristics of granular sludge in an EGSB reactor for treating low strength wastewater. *Int. J. Environ. Res.* **2008**, *2*, 319–328.
- 4. Mutombo, D.T. Internal circulation reactor: Pushing the limits of anaerobic industrial effluents treatment technologies. In Proceedings of the 2004 Water Institute of Southern Africa (WISA) Biennial Conference, Cape Town, South Africa, 2–6 May 2004; pp. 608–616.
- 5. Xu, F.; Miao, H.-F.; Huang, Z.-X.; Ren, H.-Y.; Zhao, M.-X.; Ruan, W.-Q. Performance and dynamic characteristics of microbial communities in an internal circulation reactor for treating brewery wastewater. *Environ. Technol.* **2013**, *34*, 2881–2888. [CrossRef] [PubMed]
- 6. Habets, L.H.A. *Introduction of the IC Reactor in the Paper Industry;* Technical Report; Paques BV: Balk, The Netherlands, 1999.
- 7. Fang, H.H.P.; Chui, H.K.; Li, Y.Y. Microbial structure and activity of UASB granules treating different wastewaters. *Water Sci. Technol.* **1994**, *30*, 87–96. [CrossRef]
- 8. Fang, H.H.P. Microbial distribution in UASB granules and its resulting effects. *Water Sci. Technol.* **2000**, *42*, 201–208. [CrossRef]
- 9. Look, H.P. The Phenomenon of Granulation of Anaerobic Sludge. Ph.D. Thesis, Agricultural University, Wageningen, The Netherlands, 1989.
- Lettinga, G.; Hulshoff Pol, L.W. UASB-process design for various types of wastewaters. *Water Sci. Technol.* 1991, 24, 87–107. [CrossRef]
- 11. Pol, L.W.H.; Dezeeuw, W.J.; Velzeboer, C.T.M.; Lettinga, G. Granulation in UASB-reactors. *Water Sci. Technol.* **1983**, *15*, 291–304.
- 12. Kosaric, N.; Blaszczyk, R.; Orphan, L.; Valladarfs, J. The characteristics of granules from upflow anaerobic sludge blanket reactors. *Water Res.* **1990**, *24*, 1473–1477. [CrossRef]
- 13. Novaes, R.F.V. Microbiology of anaerobic digestion. Water Sci. Technol. 1986, 18, 1–14. [CrossRef]
- 14. Shivayogimath, C.B.; Ramanujam, T.K. Treatment of distillery spentwash by hybrid UASB reactor. *Bioprocess Eng.* **1999**, *21*, 255–259. [CrossRef]
- 15. Owusu-Agyeman, I.; Eyice, Ö.; Cetecioglu, Z.; Plaza, E. The study of structure of anaerobic granules and methane producing pathways of pilot-scale UASB reactors treating municipal wastewater under sub-mesophilic conditions. *Bioresour. Technol.* **2019**, *290*, 121733. [CrossRef] [PubMed]
- 16. Kong, Z.; Li, L.; Li, Y.-Y. Characterization and variation of microbial community structure during the anaerobic treatment of N, N-dimethylformamide-containing wastewater by UASB with artificially mixed consortium. *Bioresour. Technol.* **2018**, *268*, 434–444. [CrossRef] [PubMed]
- 17. Dolfing, J.; Griffioen, A.; Van Neerven, A.R.W.; Zevenhuizen, L.P.T.M. Chemical and bacteriological composition of granular methanogenic sludge. *Can. J. Microbiol.* **1985**, *31*, 744–750. [CrossRef]

- Schmidt, J.E.; Ahring, B.K. Granular sludge formation in upflow anaerobic sludge blanket (UASB) reactors. *Biotechnol. Bioeng.* 1996, 49, 229–246. [CrossRef]
- 19. Pereboom, J.H.F. Size distribution model for methanogenic granules from full scale UASB and IC reactors. *Water Sci. Technol.* **1994**, *30*, 211–221. [CrossRef]
- Jijai, S.; Srisuwan, G.; O-Thong, S.; Ismail, N.; Siripatana, C. Effect of Granule Sizes on the Performance of Upflow Anaerobic Sludge Blanket (UASB) Reactors for Cassava Wastewater Treatment; Elsevier B.V.: Amsterdam, The Netherlands, 2015; Volume 79.
- 21. Dolfing, J. Granulation in UASB reactors. Water Sci. Technol. 1986, 18, 15–25. [CrossRef]
- 22. Wu, J.; Lu, Z.Y.; Hu, J.C.; Feng, L.; Huang, J.D.; Gu, X.S. Disruption of granules by hydrodynamic force in internal circulation anaerobic reactor. *Water Sci. Technol.* **2006**, *54*, 9–16. [CrossRef]
- 23. Liu, Y.-Q.; Liu, Y.; Tay, J.-H. The effects of extracellular polymeric substances on the formation and stability of biogranules. *Appl. Microbiol. Biotechnol.* **2004**, *65*, 143–148. [CrossRef]
- 24. Sayed, S.; Dezeeuw, W.; Lettinga, G. Anaerobic treatment of slaughterhouse waste using a flocculant sludge UASB reactor. *Agric. Wastes* **1984**, *11*, 197–226. [CrossRef]
- 25. Trulli, E.; Torretta, V. Influence of feeding mixture composition in batch anaerobic co-digestion of stabilized municipal sludge and waste from dairy farms. *Environ. Technol.* **2015**, *36*, 1519–1528. [CrossRef] [PubMed]
- 26. Koster, I.W.; Lettinga, G. Application of the upflow anaerobic sludge bed (UASB) process for treatment of complex wastewaters at low-temperatures. *Biotechnol. Bioeng.* **1985**, *27*, 1411–1417. [CrossRef] [PubMed]
- 27. Fang, H.H.P.; Chui, H.-K.; Li, Y.-Y. Anaerobic degradation of butyrate in a UASB reactor. *Bioresour. Technol.* **1995**, *51*, 75–81. [CrossRef]
- 28. Kida, K.; Tanemura, K.; Sonoda, Y.; Hikami, S. Anaerobic treatment of distillery wastewater from barley-Shochu making by UASB. *J. Ferment. Bioeng.* **1994**, 77, 90–93. [CrossRef]
- 29. España-Gamboa, E.I.; Mijangos-Cortés, J.O.; Hernández-Zárate, G.; Maldonado, J.A.D.; Alzate-Gaviria, L.M. Methane production by treating vinasses from hydrous ethanol using a modified UASB reactor. *Biotechnol. Biofuels* **2012**, *5*, 82. [CrossRef]
- 30. Caixeta, C.E.T.; Cammarota, M.C.; Xavier, A.M.F. Slaughterhouse wastewater treatment: Evaluation of a new three-phase separation system in a UASB reactor. *Bioresour. Technol.* **2002**, *81*, 61–69. [CrossRef]
- 31. Hutňan, M.; Drtil, M.; Mrafková, L.; Derco, J.; Buday, J. Comparison of startup and anaerobic wastewater treatment in UASB, hybrid and baffled reactor. *Bioprocess Eng.* **1999**, *21*, 439–445. [CrossRef]
- 32. Ten Brummeler, E.; Hulshoff Pol, L.W.; Dolfing, J. Methanogenesis in an upflow anaerobic sludge blanket reactor at pH 6 on an acetate-propionate mixture. *Appl. Environ. Microbiol.* **1985**, *49*, 1472–1477. [CrossRef] [PubMed]
- Feng, Q.; Song, Y.C.; Yoo, K.; Kuppanan, N.; Subudhi, S.; Lal, B. Polarized electrode enhances biological direct interspecies electron transfer for methane production in upflow anaerobic bioelectrochemical reactor. *Chemosphere* 2018, 204, 186–192. [CrossRef]
- Zhao, Z.; Zhang, Y.; Chen, S.; Quan, X.; Yu, Q. Bioelectrochemical enhancement of anaerobic methanogenesis for high organic load rate wastewater treatment in a up-flow anaerobic sludge blanket (UASB) reactor. *Sci. Rep.* 2015, 4, 6658. [CrossRef]
- 35. Alimahmoodi, M.; Mulligan, C.N. Anaerobic bioconversion of carbon dioxide to biogas in an upflow anaerobic sludge blanket reactor. *J. Air Waste Manag. Assoc.* **2008**, *58*, 95–103. [CrossRef]
- 36. Gong, D.; Qin, G. Treatment of oilfield wastewater using a microbial fuel cell integrated with an up-flow anaerobic sludge blanket reactor. *Desalin. Water Treat.* **2012**, *49*, 272–280. [CrossRef]
- De Mendonça, H.V.; Ometto, J.P.H.B.; Otenio, M.H.; Delgado Dos Reis, A.J.; Marques, I.P.R. Bioenergy recovery from cattle wastewater in an UASB-AF hybrid reactor. *Water Sci. Technol.* 2017, 76, 2268–2279. [CrossRef] [PubMed]
- 38. Gupta, S.K.; Gupta, S.K. Morphological study of the granules in UASB and hybrid reactors. *Clean Technol. Environ. Policy* **2005**, *7*, 203–212. [CrossRef]
- Ramakrishnan, A.; Surampalli, R.Y. Comparative performance of UASB and anaerobic hybrid reactors for the treatment of complex phenolic wastewater. *Bioresour. Technol.* 2012, 123, 352–359. [CrossRef] [PubMed]
- Halalsheh, M.M.; Muhsen, H.H.; Shatanawi, K.M.; Field, J.A. Improving solids retention in upflow anaerobic sludge blanket reactors at low temperatures using lamella settlers. *J. Environ. Sci. Heal. Part A Toxic Hazard. Subst. Environ. Eng.* 2010, 45, 1054–1059. [CrossRef]

- 41. Lettinga, G.; Van Velsen, A.F.M.; Hobma, S.W.; De Zeeuw, W.; Klapwijk, A. Use of the upflow sludge blanket (USB) reactor concept for biological wastewater treatment, especially for anaerobic treatment. *Biotechnol. Bioeng.* **1980**, *22*, 699–734. [CrossRef]
- 42. Selvamurugan, M.; Doraisamy, P.; Maheswari, M.; Nandakumar, N.B. Comparative study on startup performance of UAHR and UASB reactors in anaerobic treatment of distillery spentwash. *Int. J. Environ. Res.* **2012**, *6*, 235–244.
- 43. Selvamurugan, M.; Doraisamy, P.; Maheswari, M. High-rate anaerobic treatment of distillery spentwash using UASB and UAHR. *Int. J. Environ. Eng.* **2014**, *6*, 273–286. [CrossRef]
- 44. Patyal, V. Study of biogas generation in treatment of distillery wastewater by UASB method. *Int. J. Eng. Res.* **2016**, *V5*, 634–639.
- Tandukar, M.; Ohashi, A.; Harada, H. Performance comparison of a pilot-scale UASB and DHS system and activated sludge process for the treatment of municipal wastewater. *Water Res.* 2007, 41, 2697–2705. [CrossRef]
- 46. Buzzini, A.P.; Patrizzi, L.J.; Motheo, A.J.; Pires, E.C. Preliminary evaluation of the electrochemical and chemical coagulation processes in the post-treatment of effluent from an upflow anaerobic sludge blanket (UASB) reactor. *J. Environ. Manag.* **2007**, *85*, 847–857. [CrossRef] [PubMed]
- 47. Sponza, D.T. Anaerobic granule formation and tetrachloroethylene (TCE) removal in an upflow anaerobic sludge blanket (UASB) reactor. *Enzym. Microb. Technol.* **2001**, *29*, 417–427. [CrossRef]
- 48. Das, S.; Sarkar, S.; Chaudhari, S. Modification of UASB reactor by using CFD simulations for enhanced treatment of municipal sewage. *Water Sci. Technol.* **2018**, *77*, 766–776. [CrossRef]
- 49. Kettunen, R.H.; Rintala, J.A. Performance of an on-site UASB reactor treating leachate at low temperature. *Water Res.* **1998**, *32*, 537–546. [CrossRef]
- 50. Elmitwalli, T.A.; Shalabi, M.; Wendland, C.; Otterpohl, R. Grey water treatment in UASB reactor at ambient temperature. *Water Sci. Technol.* **2007**, *55*, 173–180. [CrossRef] [PubMed]
- 51. Dos Santos, S.L.; Chaves, S.R.M.; Van Haandel, A. Influence of temperature on the performance of anaerobic treatment systems of municipal wastewater. *Water SA* **2018**, *44*, 211–222. [CrossRef]
- 52. Sousa, J.; Santos, K.; Henrique, I.; Brasil, D.; Santos, E. Anaerobic digestion and the denitrification in UASB reactor. *J. Urban Environ. Eng.* **2008**, *2*, 63–67. [CrossRef]
- Del Nery, V.; Alves, I.; Zamariolli Damianovic, M.H.R.; Pires, E.C. Hydraulic and organic rates applied to pilot scale UASB reactor for sugar cane vinasse degradation and biogas generation. *Biomass Bioenergy* 2018, 119, 411–417. [CrossRef]
- 54. El-Seddik, M.M.; Galal, M.M.; Radwan, A.G.; Abdel-Halim, H.S. Fractional-order model (FOM) for high-strength substrate biodegradation in conventional UASB reactor. *Biochem. Eng. J.* **2018**, *133*, 39–46. [CrossRef]
- 55. Yangin-Gomec, C.; Pekyavas, G.; Sapmaz, T.; Aydin, S.; Ince, B.; Akyol, Ç.; Ince, O. Microbial monitoring of ammonia removal in a UASB reactor treating pre-digested chicken manure with anaerobic granular inoculum. *Bioresour. Technol.* **2017**, *241*, 332–339. [CrossRef]
- 56. Fang, H.H.P.; Chui, H.K. Maximum COD loading capacity in UASB reactors at 37 °C. *J. Environ. Eng.* **1993**, *119*, 103–119. [CrossRef]
- 57. Harshan, K.G.; Gana, V.B. Characterization of sewage, design of laboratory scale UASB reactor for its treatment and its performance evaluation. *Int. J. Res. Sci. Innov.* **2018**, *V*, 37–43.
- Shi, R.; Zhang, Y.; Xu, H.; Zhang, Z.; Zhang, C. Pretreatment of distillery wastewater from vitamin C synthesis industry by upflow anaerobic sludge blanket (UASB) reactor. *Environ. Eng. Sci.* 2007, 24, 1333–1337. [CrossRef]
- 59. Saner, A.B.; Mungray, A.K.; Mistry, N.J. Treatment of distillery wastewater in an upflow anaerobic sludge blanket (UASB) reactor. *Desalin. Water Treat.* **2016**, *57*, 4328–4344. [CrossRef]
- 60. Harada, H.; Uemura, S.; Chen, A.-C.; Jayadevan, J. Anaerobic treatment of a recalcitrant distillery wastewater by a thermophilic UASB reactor. *Bioresour. Technol.* **1996**, *55*, 215–221. [CrossRef]
- 61. Kwong, T.S.; Fang, H.H.P. Anaerobic degradation of cornstarch in wastewater in two upflow reactors. *J. Environ. Eng.* **1996**, *122*, 9–17. [CrossRef]
- Bakraoui, M.; Karouach, F.; Ouhammou, B.; Aggour, M.; Essamri, A.; El Bari, H. Biogas production from recycled paper mill wastewater by UASB digester: Optimal and mesophilic conditions. *Biotechnol. Rep.* 2020, 25, e00402. [CrossRef] [PubMed]

- 63. Moe, N.S.; Aung, E.M. A laboratory scale up-flow anaerobic sludge blanket (UASB) reactor for distillery wastewater treatment. *Int. J. Sci. Eng. Technol. Res.* **2014**, *3*, 4050–4055.
- 64. Sosa-Villalobos, C.A.; Rustrián, E.; Houbron, E. Anaerobic digestion of vinasse cane alcohol: The influence of OLR by a UASB reactor. *Int. J. Mod. Eng. Res.* **2014**, *4*, 37–42.
- 65. Gao, M.; She, Z.; Jin, C. Performance evaluation of a mesophilic (37 °C) upflow anaerobic sludge blanket reactor in treating distiller's grains wastewater. *J. Hazard. Mater.* **2007**, *141*, 808–813. [CrossRef]
- 66. El-Sheikh, M.A.; Saleh, H.I.; Flora, J.R.; AbdEl-Ghany, M.R. Biological tannery wastewater treatment using two stage UASB reactors. *Desalination* **2011**, *276*, 253–259. [CrossRef]
- 67. Tejasen, S.; Taruyanon, K. Modelling of Two-stage anaerobic treating wastewater from a molasses-based ethanol distillery with the IWA anaerobic digestion model No.1. *Eng. J.* **2010**, *14*, 25–36. [CrossRef]
- Molina, F.; Ruiz-Filippi, G.; García, C.; Roca, E.; Lema, J.M. Winery effluent treatment at an anaerobic hybrid USBF pilot plant under normal and abnormal operation. *Water Sci. Technol.* 2007, *56*, 25–31. [CrossRef] [PubMed]
- Colleran, E.; Concannon, F.; Golden, T.; Geoghegan, F.; Crumlish, B.; Killilea, E.; Henry, M.; Coates, J. Use of methanogenic activity tests to characterize anaerobic sludges, screen for anaerobic biodegradability and determine toxicity thresholds against individual anaerobic trophic groups and species. *Water Sci. Technol.* 1992, 25, 31–40. [CrossRef]
- 70. Hussain, A.; Dubey, S.K. Specific methanogenic activity test for anaerobic degradation of influents. *Appl. Water Sci.* 2017, 7, 535–542. [CrossRef]
- Angelidaki, I.; Alves, M.; Bolzonella, D.; Borzacconi, L.; Campos, J.L.; Guwy, A.J.; Kalyuzhnyi, S.; Jenicek, P.; Van Lier, J.B. Defining the biomethane potential (BMP) of solid organic wastes and energy crops: A proposed protocol for batch assays. *Water Sci. Technol.* 2009, *59*, 927–934. [CrossRef]
- Borja, R.; Alba, J.; Banks, C.J. Anaerobic digestion of wash waters derived from the purification of virgin olive oil using a hybrid reactor combining a filter and a sludge blanket. *Process Biochem.* 1996, *31*, 219–224. [CrossRef]
- 73. Lettinga, G.; Hobma, S.W.; Hulshoff Pol, L.W.; De Zeeuw, W.; De Jong, P.; Grin, P.; Roersma, R. Design operation and economy of anaerobic treatment. *Water Sci. Technol.* **1983**, *15*, 177–195. [CrossRef]
- Lettinga, G.; Pol, L.W.H.; Koster, I.W.; Wiegant, W.M.; Dezeeuw, W.J.; Rinzema, A.; Grin, P.C.; Roersma, R.E.; Hobma, S.W. High-rate anaerobic wastewater-treatment using the uasb reactor under a wide-range of temperature conditions. *Biotechnol. Genet. Eng. Rev.* 1984, 2, 253–284. [CrossRef]
- 75. Rizvi, H.; Ahmad, N.; Abbas, F.; Bukhari, I.H.; Yasar, A.; Ali, S.; Yasmeen, T.; Riaz, M. Start-up of UASB reactors treating municipal wastewater and effect of temperature/sludge age and hydraulic retention time (HRT) on its performance. *Arab. J. Chem.* **2015**, *8*, 780–786. [CrossRef]
- Baird, R.B.; Eaton, A.D.; Rice, E.W.; Bridgewater, L. Standard Methods for the Examination of Water and Wastewater, 23rd ed.; American Public Health Association: Washington, DC, USA; American Water Works Association: Denver, CO, USA; Water Environment Federation: Alexandria, VA, USA, 2017; ISBN 9780875532875.
- 77. Manyuchi, M.M.; Mbohwa, C.; Muzenda, E. Anaerobic treatment of opaque beer wastewater with enhanced biogas recovery through Acti-zyme bio augmentation. *S. Afr. J. Chem. Eng.* **2018**, *26*, 74–79. [CrossRef]
- 78. Chamarro, E. Use of fenton reagent to improve organic chemical biodegradability. *Water Res.* 2001, 35, 1047–1051. [CrossRef]
- 79. Metcalf and Eddy. *Wastewater Engineering: Treatment and Resource Recovery;* McGraw-Hill: New York, NY, USA, 2014; ISBN 9780073401188.
- 80. Tomašić, V.; Zelić, B. (Eds.) *Environmental Engineering*; De Gruyter: Berlin, Germany; Boston, MA, USA, 2018; ISBN 9783110468038.
- 81. Contreras, S.; Rodríguez, M.; Momani, F.A.; Sans, C.; Esplugas, S. Contribution of the ozonation pre-treatment to the biodegradation of aqueous solutions of 2,4-dichlorophenol. *Water Res.* **2003**, *37*, 3164–3171. [CrossRef]
- 82. Zheng, C.; Zhao, L.; Zhou, X.; Fu, Z.; Li, A. Treatment technologies for organic wastewater. In *Water Treatment*; InTech: London, UK, 2013.
- 83. Li, H.; Zhou, S.; Sun, Y.; Feng, P.; Li, J. Advanced treatment of landfill leachate by a new combination process in a full-scale plant. *J. Hazard. Mater.* **2009**, *172*, 408–415. [CrossRef]
- Czajkowska, J.; Hawer-Strojek, P.; Reczek, L.; Bugajski, P.; Michel, M.; Gajewska, M.; Siwiec, T.; Jóźwiakowski, K.; Gut, B. Correlations between organic pollution indicators in municipal wastewater. *Arch. Environ. Prot.* 2018, 44, 50–57.

- 85. Govahi, S.; Karimi-Jashni, A.; Derakhshan, M. Treatability of landfill leachate by combined upflow anaerobic sludge blanket reactor and aerated lagoon. *Int. J. Environ. Sci. Technol.* **2012**, *9*, 145–151. [CrossRef]
- 86. Choi, Y.; Ryu, J.; Lee, S.R. Influence of carbon type and carbon to nitrogen ratio on the biochemical methane potential, pH, and ammonia nitrogen in anaerobic digestion. *J. Anim. Sci. Technol.* **2020**, *62*, 74–83. [CrossRef]
- 87. Wang, X.; Lu, X.; Li, F.; Yang, G. Effects of temperature and carbon-nitrogen (C/N) ratio on the performance of anaerobic co-digestion of dairy manure, chicken manure and rice straw: Focusing on ammonia inhibition. *PLoS ONE* **2014**, *9*, e97265. [CrossRef]
- Wang, X.; Yang, G.; Feng, Y.; Ren, G.; Han, X. Optimizing feeding composition and carbon–nitrogen ratios for improved methane yield during anaerobic co-digestion of dairy, chicken manure and wheat straw. *Bioresour. Technol.* 2012, 120, 78–83. [CrossRef]
- Bouallagui, H.; Lahdheb, H.; Ben Romdan, E.; Rachdi, B.; Hamdi, M. Improvement of fruit and vegetable waste anaerobic digestion performance and stability with co-substrates addition. *J. Environ. Manag.* 2009, *90*, 1844–1849. [CrossRef]
- 90. Yen, H.W.; Brune, D.E. Anaerobic co-digestion of algal sludge and waste paper to produce methane. *Bioresour. Technol.* 2007, *98*, 130–134. [CrossRef] [PubMed]
- 91. Soboh, Y.M.; Sorensen, D.L.; Sims, R.C. Upflow anaerobic sludge blanket reactor codigestion of algae and acetate to produce methane. *Water Environ. Res.* **2016**, *88*, 2094–2103. [CrossRef] [PubMed]
- 92. Sievers, D.M.; Brune, D.E. Brune carbon/nitrogen ratio and anaerobic digestion of swine waste. *Trans. ASAE* **1978**, *21*, 537–541. [CrossRef]
- 93. Matin, H.A. The influence of microbial consortium and C/N ratio to biogas production from rice husk waste by using solid state anaerobic digestion (SS-AD). In Proceedings of the E3S Web Conference, Semarang, Indonesia, 15–16 August 2017; Volume 73, p. 01018.
- Scherer, P.; Lippert, H.; Wolff, G. Composition of the major elements and trace elements of 10 methanogenic bacteria determined by inductively coupled plasma emission spectrometry. *Biol. Trace Elem. Res.* 1983, 5, 149–163. [CrossRef]
- 95. Arne Alphenaar, P.; Sleyster, R.; De Reuver, P.; Ligthart, G.-J.; Lettinga, G. Phosphorus requirement in high-rate anaerobic wastewater treatment. *Water Res.* **1993**, *27*, 749–756. [CrossRef]
- Gil, A.; Siles, J.A.; Serrano, A.; Chica, A.F.; Martín, M.A. Effect of variation in the C/[N+P] ratio on anaerobic digestion. *Environ. Prog. Sustain. Energy* 2019, *38*, 228–236. [CrossRef]
- 97. Wilkie, A.; Goto, M.; Bordeaux, F.M.; Smith, P.H. Enhancement of anaerobic methanogenesis from napiergrass by addition of micronutrients. *Biomass* **1986**, *11*, 135–146. [CrossRef]
- Zandvoort, M.H.; Van Hullebusch, E.D.; Gieteling, J.; Lettinga, G.; Lens, P.N.L. Effect of sulfur source on the performance and metal retention of methanol-fed UASB reactors. *Biotechnol. Prog.* 2005, 21, 839–850. [CrossRef] [PubMed]
- 99. Fricke, K.; Santen, H.; Wallmann, R.; Hüttner, A.; Dichtl, N. Operating problems in anaerobic digestion plants resulting from nitrogen in MSW. *Waste Manag.* **2007**, *27*, 30–43. [CrossRef]
- 100. Ranalli, P. (Ed.) *Improvement of Crop Plants for Industrial End Uses;* Springer Netherlands: Dordrecht, The Netherlands, 2007; ISBN 978-1-4020-5485-3.
- Fang, H.H.P.; Zhang, T. Anaerobic Biotechnology: Environmental Protection and Resource Recovery; Imperial College Press: London, UK, 2015; ISBN 978-1-78326-790-3.
- 102. Speece, R.E. *Anaerobic Biotechnology for Industrial Wastewaters*; Archae Press: Nashville, TN, USA, 1996; ISBN 0965022609.
- 103. Metcalf, L.; Eddy, H.P.; Burton, F.L.; Stensel, H.D.; Tchobanoglous, G. *Wastewater Engineering: Treatment and Reuse*; McGraw Hill: New York, NY, USA, 2003; ISBN 0071122508.
- 104. Hamza, R.A.; Iorhemen, O.T.; Tay, J.H. Anaerobic-aerobic granular system for high-strength wastewater treatment in lagoons. *Adv. Environ. Res.* **2016**, *5*, 169–178. [CrossRef]
- Droste, R.L.; Dehr, R.L. Theory and Practice of Water and Wastewater Treatment, 2nd ed.; John Wiley & Sons, Inc.: Hoboken, NJ, USA, 2019; ISBN 9781119312376.
- 106. Araujo, D.J.; Rocha, S.M.S.; Cammarota, M.C.; Xavier, A.M.F.; Cardoso, V.L. Anaerobic treatment of wastewater from the household and personal products industry in a hybrid bioreactor. *Braz. J. Chem. Eng.* 2008, 25, 443–451. [CrossRef]
- Bashaar, Y.A. Nutrients requirements in biological industrial wastewater treatment. *African J. Biotechnol.* 2004, *3*, 236–238. [CrossRef]

- Hussain, A.; Dubey, S.K. Specific methanogenic activity test for anaerobic treatment of phenolic wastewater. Desalin. Water Treat. 2014, 52, 7015–7025. [CrossRef]
- Aiyuk, S.; Amoako, J.; Raskin, L.; van Haandel, A.; Verstraete, W. Removal of carbon and nutrients from domestic wastewater using a low investment, integrated treatment concept. *Water Res.* 2004, *38*, 3031–3042. [CrossRef] [PubMed]
- Annachhatre, A.P. Anaerobic treatment of industrial wastewaters. *Resour. Conserv. Recycl.* 1996, 16, 161–166.
 [CrossRef]
- Gujer, W.; Zehnder, A.J.B. Conversion processes in anaerobic digestion. Water Sci. Technol. 1983, 15, 127–167. [CrossRef]
- 112. Appels, L.; Baeyens, J.; Degreve, J.; Dewil, R. Principles and potential of the anaerobic digestion of waste-activated sludge. *Prog. Energy Combust. Sci.* 2008, 34, 755–781. [CrossRef]
- 113. Suryawanshi, P.C.; Chaudhari, A.B.; Kothari, R.M. Thermophilic anaerobic digestion: The best option for waste treatment. *Crit. Rev. Biotechnol.* **2010**, *30*, 31–40. [CrossRef]
- Zickefoose, C.; Hayes, R.B. Anaerobic Sludge Digestion: Operations Manual; Environmental Protection Agency: Washington, DC, USA, 1976.
- 115. *The Microbiology of Anaerobic Digesters*; Gerardi, M.H. (Ed.) Wastewater Microbiology Series; John Wiley & Sons, Inc.: Hoboken, NJ, USA, 2003; Volume 6, ISBN 0471206938.
- 116. Mahoney, E.M.; Varangu, L.K.; Cairns, W.L.; Kosaric, N.; Murray, R.G.E. The effect of calcium on microbial aggregation during UASB reactor start-up. *Water Sci. Technol.* **1987**, *19*, 249–260. [CrossRef]
- 117. Yu, H.; Tay, J.H.; Fang, H.H.P. The roles of calcium in sludge granulation during uasb reactor start-up. *Water Res.* **2001**, *35*, 1052–1060. [CrossRef]
- 118. Cail, R.G.; Barford, J.P. The development of granulation in an upflow floc digester and an upflow anaerobic sludge blanket digester treating cane juice stillage. *Biotechnol. Lett.* **1985**, *7*, 493–498. [CrossRef]
- Cunha, J.R.; Morais, S.; Silva, J.C.; Van der Weijden, R.D.; Hernández Leal, L.; Zeeman, G.; Buisman, C.J.N. Bulk pH and carbon source are key factors for calcium phosphate granulation. *Environ. Sci. Technol.* 2019, 53, 1334–1343. [CrossRef] [PubMed]
- 120. Schmidt, J.E.; Ahring, B.K. Effects of magnesium on thermophilic acetate-degrading granules in upflow anaerobic sludge blanket (UASB) reactors. *Enzym. Microb. Technol.* **1993**, *15*, 304–310. [CrossRef]
- 121. Speece, R.E. Anaerobic biotechnology for industrial wastewater treatment. *Environ. Sci. Technol.* **1983**, 17, 416A–427A. [CrossRef]
- 122. Isik, M.; Sponza, D.T. Effects of alkalinity and co-substrate on the performance of an upflow anaerobic sludge blanket (UASB) reactor through decolorization of Congo Red azo dye. *Bioresour. Technol.* 2005, 96, 633–643. [CrossRef]
- 123. Bina, B.; Amin, M.M.; Pourzamani, H.; Fatehizadeh, A.; Ghasemian, M.; Mahdavi, M.; Taheri, E. Biohydrogen production from alkaline wastewater: The stoichiometric reactions, modeling, and electron equivalent. *MethodsX* 2019, *6*, 1496–1505. [CrossRef]
- 124. Choi, J.; Ahn, Y. Biohydrogen fermentation from sucrose and piggery waste with high levels of bicarbonate alkalinity. *Energies* **2015**, *8*, 1716–1729. [CrossRef]
- 125. Lützhøft, H.C.H.; Boe, K.; Fang, C.; Angelidaki, I. Comparison of VFA titration procedures used for monitoring the biogas process. *Water Res.* 2014, 54, 262–272. [CrossRef]
- 126. Anderson, G.K.; Yang, G. Determination of bicarbonate and total volatile acid concentration in anaerobic digesters using a simple titration. *Water Environ. Res.* **1992**, *64*, 53–59. [CrossRef]
- 127. Lahav, O.; Morgan, B.E. Titration methodologies for monitoring of anaerobic digestion in developing countries—A review. *J. Chem. Technol. Biotechnol.* 2004, 79, 1331–1341. [CrossRef]
- Arne Alphenaar, P.; Visser, A.; Lettinga, G. The effect of liquid upward velocity and hydraulic retention time on granulation in UASB reactors treating wastewater with a high sulphate content. *Bioresour. Technol.* 1993, 43, 249–258. [CrossRef]
- 129. Visser, A.; Gao, Y.; Lettinga, G. Effects of pH on methanogenesis and sulphate reduction in thermophilic (55 °C) UASB reactors. *Bioresour. Technol.* **1993**, *44*, 113–121. [CrossRef]
- 130. Mahmoud, N.; Zeeman, G.; Gijzen, H.; Lettinga, G. Solids removal in upflow anaerobic reactors, a review. *Bioresour. Technol.* **2003**, *90*, 1–9. [CrossRef]
- Al-Shayah, M.; Mahmoud, N. Start-up of an UASB-septic tank for community on-site treatment of strong domestic sewage. *Bioresour. Technol.* 2008, 99, 7758–7766. [CrossRef] [PubMed]

- 132. Halalsheh, M.; Sawajneh, Z.; Zubi, M.; Zeeman, G.; Lier, J.; Fayyad, M.; Lettinga, G. Treatment of strong domestic sewage in a 96 m UASB reactor operated at ambient temperatures: Two-stage versus single-stage reactor. *Bioresour. Technol.* 2005, *96*, 577–585. [CrossRef] [PubMed]
- 133. Mahmoud, N.; Zeeman, G.; Gijzen, H.; Lettinga, G. Anaerobic sewage treatment in a one-stage UASB reactor and a combined UASB-Digester system. *Water Res.* **2004**, *38*, 2348–2358. [CrossRef] [PubMed]
- Zeng, T.; Rene, E.R.; Hu, Q.; Lens, P.N.L. Continuous biological removal of selenate in the presence of cadmium and zinc in UASB reactors at psychrophilic and mesophilic conditions. *Biochem. Eng. J.* 2019, 141, 102–111. [CrossRef]
- 135. Ismail, S.B.; De La Parra, C.J.; Temmink, H.; Van Lier, J.B. Extracellular polymeric substances (EPS) in upflow anaerobic sludge blanket (UASB) reactors operated under high salinity conditions. *Water Res.* **2010**, *44*, 1909–1917. [CrossRef] [PubMed]
- 136. Wang, W.; Yang, K.; Sierra, J.M.; Zhang, X.; Yuan, S.; Hu, Z. Potential impact of methyl isobutyl ketone (MIBK) on phenols degradation in an UASB reactor and its degradation properties. *J. Hazard. Mater.* **2017**, *333*, 73–79. [CrossRef]
- 137. Díaz-Báez, M.C.; Valderrama-Rincon, J.D. Rapid restoration of methanogenesis in an acidified UASB reactor treating 2,4,6-trichlorophenol (TCP). *J. Hazard. Mater.* **2017**, *324*, 599–604. [CrossRef]



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