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Economic MPC of Wastewater Treatment Plants Based on Model Reduction

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Abstract: In this paper, we consider the problem of economic model predictive control of wastewater treatment plants based on model reduction. We apply two model approximation methods to a wastewater treatment plant (WWTP) described by a modified Benchmark Simulation Model No.1 to overcome the intensive computation associated with economic model predictive control (MPC). Two computationally efficient models are obtained based on trajectory piecewise linearization (TPWL) and reduced order TPWL. To obtain the reduced order TPWL model, a proper orthogonal decomposition (POD)-based method is utilized. Further, the reduced order model is linearized to obtain a TPWL-POD model. The objective is to design controllers which minimize the overall economic cost. Accordingly, we design economic MPC (EMPC) controllers based on each of the models. The economic control cost can be described as a weighted summation of effluent quality and overall operating cost. We compare the accuracy of the two proposed approximation models with different linearization point numbers. We evaluate the average evaluation time for the two proposed EMPC controllers and make comparisons with the EMPC based on the original nonlinear model. We also investigate how the number of linearization points involved in the TPWL model and TPWL-POD model affects the control performance in terms of average performance cost and the average evaluation time.

Keywords: economic model predictive control; wastewater treatment plant; model reduction; trajectory piecewise linearization

1. Introduction

Wastewater treatment plants (WWTPs) have been widely used to recycle wastewater in order to minimize its adverse environmental impacts. A typical WWTP is a large-scale nonlinear process that consists of a series of interconnected biological reactors and a secondary settler. While meeting the strict requirements in environmental regulations, ensuring the process safety and minimizing the cost of operation, a wastewater treatment plant should be monitored and regulated [1]. However, significant variability of inlet flow rates and wastewater compositions leads to the increased complexity in the design of advanced control and monitoring schemes for WWTPs [2].

In the literature, there are many existing resutls on the control of WWTPs. For example, in [3], proportional-integral (PI) was used to control a nutrient removal WWTP. In [4], the tuning of proportional-integral-derivative control was considered in the control of the dissolved oxygen concentration in the activated sludge process. Advanced control methods such as model predictive control (MPC) have also attracted much research attention in the control of WWTPs. In [5–7], simultaneous design and control of WWTPs were considered with MPC. In [8], an MPC controller was developed for the Benchmark Simulation Model No.1 (BSM1). MPC was also used to in the control of oxygen concentration in [9,10]. In [11,12], MPC was applied to a WWTP to increase the

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plant efficiency, and in [13] it was applied to reduce the power usage. In one of our previous works, we applied economic MPC (EMPC) to explicitly optimize the operating economics of WWTPs [14]. While the EMPC can lead to improved effluent quality and improved overall operating cost, its high evaluation time renders the online implementation of this economic MPC intractable.

When considering economic MPC, it in general requires more time to evaluate the corresponding optimization problem compared with the traditional tracking MPC. One way to address the high evaluation time of economic MPC of nonlinear system is to use a distributed computing framework. Following this idea, in a recent work [1], we developed distributed EMPC schemes for WWTPs to reduce the evaluation time of the centralized EMPC. In this work, we consider another approach to address the high evaluation time of economic MPC; that is, to design economic MPC based on approximated models. This is based on the consideration that when solving the nonlinear optimization problem of an economic MPC, the model of the system is evaluated many times. If we can use an approximated model that can be evaluated faster, then the economic MPC can be evaluated faster.

The objective of this work is to study how two model approximation methods can help speed up the evaluation of the optimization problem of economic MPC. Specifically, we propose to develop linear approximated models of the original nonlinear system and use the approximated models in EMPC design. A lot of achievements have been made on the use of approximated models in MPC. The work of [15] applied feedback linearization to the plant and then used MPC in a cascade arrangement for the resulting linear system. The nonlinear programming problem was reduced to a quadratic optimization problem which can improve the computational efficiency. The implementation of MPC with a different linear model at each time step derived from a local (Jacobian) linearization of the nonlinear plant was first proposed in [16]. An extended Kalman filter was proposed to be added to this approach to deal with unstable nonlinear process and ensure a better disturbance rejection in [17,18]. This idea was further developed in [19], contraction constraints were applied and the explicit stability conditions were derived. In [20], one technique was proposed to approximate the nonlinear system with a linear time varying (LTV) model, which is obtained from a linearization of the system along the predicted system trajectory. In [21,22], a novel MPC algorithm was proposed which can significantly reduce the online evaluation time. The approach is to only compute the first control move while approximating the rest of control moves by using a linear controller.

Pertaining to model approximation, model order reduction techniques are also popular and can significantly abate the complexity of the nonlinear system while conserving the dominant dynamics of the process [23]. The proper orthogonal decomposition (POD) method is extensively used in data analysis for approximating the high-dimensional process by low-dimensional descriptions [24]. The POD method has been widely applied in control in chemical engineering [25–29]. In [30], a state estimation scheme was established for WWTPs using POD-based model approximation. In [31,32], a reduced order model based on the POD-Galerkin projection method was constructed for economic MPC. With the reduced order model, the evaluation time of the economic MPC was significantly reduced. In [33], a nonlinear system was represented by a piecewise-linear system and each of the pieces were reduced with a Krylov projections.

In this work, we apply the trajectory piecewise linearization and a combination of trajectory piecewise linearization and proper orthogonal decomposition method to WWTPs to derive low-order linear models, which are subsequently used for EMPC controller design. The model accuracy, economic MPC evaluation time, and economic control performance of the WWTP process under these EMPC controllers are compared. Note that we investigate the performance of these model approximation and reduction methods in a centralized control framework. These methods may also be used together with distributed control to further reduce the evaluation time of the centralized EMPC. The main contributions of this work include the following:

 The investigation of two model approximation methods in EMPC evaluation time reduction in the context of wastewater treatment control. Processes 2019, 7, 682 3 of 21

 The detailed model reduction procedures for wastewater treatment processes and the design of EMPC based on the reduced models.

• Extensive simulation results and analysis on the performance of EMPC based on reduced models in terms of economic performance and controller evaluation time.

2. Preliminaries

2.1. WWTP Process Description and Modeling

The WWTP process considered in this work is a modified version of the Benchmark Simulation Model No.1 (BSM1). A schematic of the process is shown in Figure 1. The process consists of a sludge reactor consisting of five compartments and an ideal separator [34]. In the sludge reactor, the first two compartments are anoxic compartments and the remaining three compartments are aerobic compartments. During the operation, denitrification reactions occur in the anoxic compartments where nitrate is converted to nitrogen, while nitrification reactions take place in three aerobic compartments where ammonium is oxidized to nitrate. In the process, the inlet stream of the first compartment of the reactor includes a feed stream, an internal recycle stream and an external recycle stream. The wastewater enters the first compartment of the reactor in the feed stream at flow rate Q_0 and concentration Z_0 . The internal recycle stream is a portion of the effluent from the last aerated compartment of the reactor at flow rate Q_a and concentration Z_a . The other portion of the effluent enters the ideal separator at flow rate Q_f and concentration Z_f . The external recycle stream is a portion of the underflow of the separator and it is fed back into the first compartment of the reactor at flow rate Q_r and concentration Z_r . The remaining portion of the underflow which is the generated sludge leaves the separator at flow rate Q_w and concentration Z_w . The purified water leaves the process via the overflow of the separator at flow rate Q_e and concentration Z_e . According to [35], eight basic biological reactions are taken into account, and the concentrations of 13 major reaction-involved compounds defined in Table 1 are the state variables of this model.

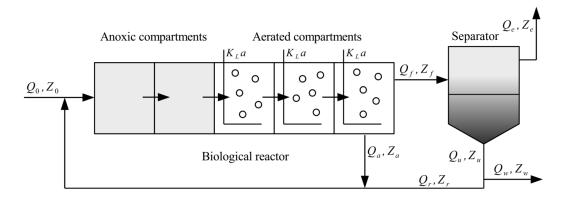


Figure 1. A schematic of the wastewater treatment plant.

The dynamics of WWTP based on the modified BSM1 model can be described by 78 ordinary differential equations. The dynamics of each compartment of the reactor and the ideal separator can be described by 13 differential equations according to 13 state variables defined in Table 1. The parameter values of this process model are reported in [2] and Table 2 shows the model parameters adopted in this work. The dynamics of the system are described based on mass balance as follows:

For compartment k (k = 1) of the biological reactor:

$$\frac{dZ_1}{dt} = \frac{1}{V_1} (Q_a Z_a + Q_r Z_r + Q_0 Z_0 + r_1 V_1 - Q_1 Z_1)$$
 (1)

$$Q_1 = Q_a + Q_r + Q_0 (2)$$

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For compartment k (k = 2, ..., 5) of the biological reactor:

$$\frac{dZ_k}{dt} = \frac{1}{V_k} (Q_{k-1} Z_{k-1} + r_k V_k - Q_k Z_k)$$
(3)

$$Q_k = Q_{k-1} \tag{4}$$

Special case for concentration $S_{O,k}$, k = 1, ..., 5, (the concentration of dissolved oxygen in compartment *k* of the biological reactor):

$$\frac{dS_{O,k}}{dt} = \frac{1}{V_k} (Q_{k-1}S_{O,k-1} + r_k V_k + K_L a_k V_k (S_O^* - S_{O,k}) - Q_k S_{O,k})$$
 (5)

In Equations (1), (3) and (5), Z_k is the concentration of the compounds defined in Table 1 in kth compartment of the biological reactor, V_k represents the volume of kth compartment, r_k denotes the observed conversion rates of the compound in kth compartment, S_O^* is the saturation concentration for oxygen and is equal to $8 \text{ g} \cdot \text{m}^{-3}$, and $K_L a_k$ represents the oxygen transfer coefficient in kth compartment, since compartment 1 and 2 of the reactor are anoxic compartments, $K_L a_1 = K_L a_2 = 0 \text{ d}^{-1}$ [2].

Definition	Notation	Unit
inert soluble organic matter	S_I	$g COD \cdot m^{-3}$
inert particulate organic matter	X_I	$g COD \cdot m^{-3}$
readily biodegradable and soluble substrate	S_s	$g COD \cdot m^{-3}$
slowly biodegradable and soluble substrate	X_s	$g COD \cdot m^{-3}$
biomass of active autotrophs	X_{B_A}	$g COD \cdot m^{-3}$
biomass of active heterotrophs	X_{B_H}	$g COD \cdot m^{-3}$
particulate generated from decay of organisms	X_P	$g COD \cdot m^{-3}$
particulate biodegradable organic nitrogen	X_{ND}	$g N \cdot m^{-3}$
nitrite nitrogen and nitrate	S_{NO}	$g N \cdot m^{-3}$
free and saline ammonia	S_{NH}	$\mathrm{gN\cdot m^{-3}}$
biodegradable and soluble organic nitrogen	S_{ND}	$g N \cdot m^{-3}$
dissolved oxygen	S_{O}	$g(-COD) \cdot m^{-3}$
alkalinity	S_{ALK}	$\text{mol}\cdot\text{m}^{-3}$
total sludge concentration in settler	X	$g COD \cdot m^{-3}$

Table 1. Definition and notation of the process variables in the modified BSM1 model.

The ideal separator is assumed to be a membrane filtration unit and no biological activity exists in it [34]. All soluble compounds (i.e., S_I , S_S , S_O , S_{NO} , S_{NH} , S_{ND} , and S_{ALK}) are assumed to be well-mixed in the separator, and all solid compounds (i.e., X_I , X_S , $X_{B,H}$, $X_{B,A}$, X_P , and X_{ND}) are assumed to precipitate in the bottom of the separator [36]. Based on mass balance, the dynamics of the separator unit (k = 6) can be described as follows:

For soluble compounds in the separator unit (k = 6):

$$\frac{dZ_k}{dt} = \frac{1}{V_k} (Q_f Z_f + Q_e Z_e + Q_r Z_r + Q_w Z_w)$$
 (6)

$$Z_e = Z_r = Z_w = Z_k \tag{7}$$

$$Z_f = Z_{k-1} = Z_k \tag{8}$$

For solid compounds in the separator unit (k = 6):

$$\frac{dZ_k}{dt} = \frac{1}{V_k} (QfZ_f + Q_rZ_r + Q_wZ_w) \tag{9}$$

$$Z_r = Z_w = Z_k \tag{10}$$

$$Z_f = Z_{k-1} = Z_k (11)$$

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The flow rates of each stream in Figure 1 can be described as follows:

$$Q_1 = Q_2 = Q_3 = Q_4 = Q_5 (12)$$

$$Q_1 = Q_a + Q_r + Q_0 (13)$$

$$Q_f = Q_5 - Q_a = Q_e + Q_r + Q_w = Q_e + Q_u$$
 (14)

$$Q_0 = Q_e + Q_w \tag{15}$$

Table 2. Parameter value of the modified BSM1 model.

17 / 1 / () ()	1000 3
V_1 (volume of compartment 1)	1000 m^3
V_2 (volume of compartment 2)	1000 m^3
V_3 (volume of compartment 3)	1333 m ³
V_4 (volume of compartment 4)	1333 m ³
V_5 (volume of compartment 5)	1333 m^3
Q_w (Underflow discharge flow rate)	$385 \text{ m}^3 \cdot d^{-1}$
Q_r (Outer recycle flow rate)	$18,446 \text{ m}^3 \cdot \text{d}^{-1}$
V_s (Volume of separator)	6000 m^3
$K_L a_3$ (Oxygen transfer coefficient of chamber 3)	$240 d^{-1}$
$K_L a_4$ (Oxygen transfer coefficient of chamber 4)	$240 d^{-1}$

2.2. Compact form of the System Model

The modified BSM1 model for the wastewater treatment plant can be described in the following compact form:

$$\dot{x}(t) = f(x(t), u(t)) \tag{16}$$

where $x \in \mathbb{R}^{78}$ is the vector of process states, $u \in \mathbb{R}^3$ represents the input vector consisting manipulated inputs and the uncontrolled inputs. The manipulated inputs are the flow rate of the recirculation stream (i.e., Q_a) and the oxygen transfer rate in the fifth compartment of the biological reactor (i.e., $K_L a_5$), respectively. The uncontrolled inputs contain the influent information under different weather conditions.

2.3. Economic Control Objective

The economic control objective is defined as follows:

$$l_{eco}(x(\tau), u(\tau)) = \alpha_{EQ} \widehat{EQ}(\tau) + \alpha_{OCI} \widehat{OCI}(\tau)$$
(17)

where \widehat{EQ} is the economic index of effluent quality (EQ) which is calculated as the average amount of the pollutants discharged and is defined in Equation (18), \widehat{OCI} represents the average amount of the overall cost index (OCI) which contains the factors that affect the operating costs significantly and it is defined in Equation (23), α_{EQ} and α_{OCI} are two weighting coefficients for EQ and OCI, respectively.

2.3.1. Effluent Quality

In BSM1, EQ (kg pollution unit \cdot day⁻¹) represents the daily average of a weighted summation of the effluent concentrations of several compounds, which significantly affect the quality of the processed water according to regional regulations. Specifically, the EQ index is evaluated as follows:

$$EQ = \frac{1}{T \cdot 1000} \int_{t_0}^{t_f} \begin{pmatrix} \beta_{SS} \cdot SS_e(t) + \beta_{COD} \cdot COD_e(t) + \beta_{NKj} \cdot S_{NKj,e}(t) \\ + \beta_{NO} \cdot S_{NO,e}(t) + \beta_{BOD} \cdot BOD_e(t) \end{pmatrix} Q_e(t) dt$$
(18)

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where

$$SS_{e} = 0.75(X_{S,e} + X_{B_{A},e} + X_{B_{H},e} + X_{I,e} + X_{P,e})$$

$$COD_{e} = S_{I,e} + S_{S,e} + X_{S,e} + X_{I,e} + X_{B_{H},e} + X_{B_{A},e} + X_{P,e}$$

$$BDO_{e} = 0.25(S_{S,e} + X_{S,e} + (1 - f_{p})(X_{B_{A},e} + X_{B_{H},e}))$$

$$S_{NKI,e} = S_{ND,e} + S_{NH,e} + X_{ND,e} + i_{XB}(X_{B_{A},e} + X_{B_{H},e}) + i_{XP}(X_{I,e} + X_{P,e})$$

In Equation (18), t_0 and t_f denote the initial and final time instants of the evaluation horizon; $T:=t_f-t_0$ is the length of the horizon, and is usually selected to be T=7 days; SS represents the concentration of suspended solids, COD is chemical oxygen demand, BOD denotes the biological oxygen demand, and $S_{\rm NKj}$ denotes the concentration of Kjeldahl nitrogen; $\beta_{\rm SS}$, $\beta_{\rm COD}$ β_{NKj} , β_{NO} and $\beta_{\rm BOD}$ represent the weighting coefficients used to calculate EQ; f_p , i_{XB} , and i_{XP} are stoichiometirc parameters. The corresponding values are listed in Table 3 [37] and the subscript "e" refers to the effluent of the secondary settler.

Table 3. Values of the weighting coefficients of EQ Index.

Weighting Coefficient	$\beta_{\rm SS}$	β_{COD}	β_{NKj}	β_{NO}	β_{BOD}	f_p	i_{XB}	i_{XP}
Value	2	1	30	10	2	0.08	0.08	0.06

2.3.2. Overall Cost Index

The total operating cost required for processing wastewater is another important factor in performance evaluation. Factors that have major effects on the operating cost include the sludge production that needs to be disposed, the energy required for aerating and pumping, external carbon consumption as well as mixing energy.

The sludge production (SP) is defined as the average amount per day $(kg \cdot day^{-1})$ of the solids produced in the process over the evaluation horizon T, including the solids discharged through the wastage flow Q_w from the settler and the sedimentary solids in a WWTP plant.

$$SP = \frac{0.75}{T \cdot 1000} \int_{t_0}^{t_f} \left(X_{S,w}(t) + X_{I,w}(t) + X_{B_A,w}(t) + X_{B_H,w}(t) + X_{P,w}(t) \right) Q_w(t) dt + \frac{1}{T \cdot 1000} \left(SS(t_f) - SS(t_0) \right)$$
(19)

In Equation (19), the subscript "w" refers to the wastage outlet of the secondary settler.

The aeration energy (AE) (kWh · day⁻¹) is associated with several plant characteristics, including the diffuser type, the submersion depth, the bubble size, etc. AE is calculated based on the oxygen transfer rates (i.e., $K_L a_i$, i = 1, ..., 5) by considering an immersion depth of 4 m as follows:

$$AE = \frac{S_o^{sat}}{T \cdot 1800} \int_{t_0}^{t_f} \sum_{i=1}^{5} V_i \cdot K_L a_i(t) dt$$
 (20)

where $i \in \{1\ 2\ 3\ 4\ 5\}$, V_i denotes the volume of the i-th chamber of the sludge reactor, and $K_L a_i$ denotes the oxygen transfer rate in the i-th chamber. S_o^{sat} denotes the saturation concentration of oxygen and its value is $8\ g/m^3$.

The pumping energy (PE) (kWh · day⁻¹) represents the energy consumed by the pumps used for inner recycle (i.e., Q_a) and outer recycle (i.e., Q_r). PE is calculated as below:

$$PE = \frac{1}{T} \int_{t_0}^{t_f} (0.004Q_a(t) + 0.05Q_w(t) + 0.008Q_r(t))dt$$
 (21)

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The mixing energy (ME) ($kWh \cdot day^{-1}$) denotes the energy consumed for mixing the compounds in the anoxic chambers to avoid the occurrence of settling. ME is calculated depending on the volume of each chamber and the oxygen transfer rates:

$$ME = \frac{24}{T} \int_{t_0}^{t_f} \left(\sum_{i \in \mathbb{I}} 0.005 \cdot V_i \right) dt$$
 (22)

where $\mathbb{I} := \{i | i \in \{12345\}\}.$

The overall cost index (OCI) approximates the total cost for WWTP operation by taking a weighted summation of the listed major factors as follows:

$$OCI = 5 \cdot SP + AE + PE + ME \tag{23}$$

3. Trajectory Piecewise Linear (TPWL) Model

In this section, the trajectory piecewise linear model approach is introduced and the trajectory piecewise linear model is presented. The steps to generate the piecewise linear model are also shown in this section.

3.1. Piecewise Linear Representation

A linearized model for the nonlinear system can be obtained at a steady-state point (x_s, u_s) as follows:

$$\frac{dx}{dt} = f(x_s, u_s) + A(x - x_s) + B(u - u_s)$$
 (24a)

$$A = \frac{\partial f}{\partial x}|_{x_s, u_s} \tag{24b}$$

$$B = \frac{\partial f}{\partial u}|_{x_s, u_s} \tag{24c}$$

where A is the Jacobian matrix of system f(x, u) evaluated at the steady-state.

The simple linearized model can be used to approximate weakly nonlinear systems with less evaluation time [33]. The approximated result of the linearized model usually depends on the range of inputs. If the system is a highly nonlinear system, the simple linearized model which is only linearized at one point would be less accurate. The main idea of the piecewise linear model approach is to generate a weighted combination of linear models which are linearized at appropriately selected states of the original nonlinear system. Compared with the system which is linearized at one single point, the system consisting of a combination of multiple linearizations would generate a better approximation result for a more complex nonlinear system.

Assuming that s linearized models have been generated for the nonlinear system of Equation (16) at points (x_i, u_i) , i = 0, 1, ..., (s - 1):

$$\frac{dx}{dt} = f(x_i, u_i) + A_i(x - x_i) + B_i(u - u_i)$$
 (25a)

$$A_i = \frac{\partial f}{\partial x}|_{x_i, u_i} \tag{25b}$$

$$B_i = \frac{\partial f}{\partial u}|_{x_i, u_i} \tag{25c}$$

A weighted combination of the linearized models in the form of Equation (25) of the nonlinear system leads to the following representation:

$$\frac{dx}{dt} = \sum_{i=0}^{s-1} w_i(x) (f(x_i, u_i) + A_i(x - x_i) + B_i(u - u_i))$$
(26)

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where the weight $w_i(x)$ is a state dependent variable and it can be computed from the distance between current state x and the linearized point x_i [33].

3.2. Generation of Piecewise Linear Model

The trajectory piecewise linear model is developed based on a fixed trajectory of the entire nonlinear system. The fixed trajectory is generated by simulating the nonlinear system based on a fixed training input u. Let us consider that we have generated a fixed trajectory of the nonlinear system, and the initial state x_0 is given. The selection of linearization points can be shown as in Algorithm 1.

Algorithm 1: Algorithm for finding the linearization points of piecewise linear model

- 1. Define $S = \{0, 1, ..., N\}$ and $S_p = \{\}$.
- 2. Set i = 0 and $S_l = \{x_i\}$.
- 3. If $S \neq S_p$, then:
 - 3.1. Set x_i as one of the linearization points.
 - 3.2. For each $j \in S \setminus S_p$,
 - 3.2.1 Calculate the distance between point x_j and the linearization point x_i ,

$$d_i = ||x_i - x_i||_2.$$

3.2.2 If $d_i \le \delta$ ($\delta > 0$), then:

$$S_p = S_p \cup \{j\}.$$

Else, do:

$$S_p = S_p$$
.

- 3.3. Select k_{\min} such that $k_{\min} = \arg\min_{k} \{d_k | k \in S \setminus S_p\}$.
- 3.4. Set $i = k_{min}$, set $S_l = S_l \cup \{x_i\}$. Go to Step 3.1.

Else, end.

In Algorithm 1, the operator "\" denotes set subtraction such that $A \setminus B := \{x | x \in A, x \notin B\}$. N is the number of sampled points on this fixed trajectory. S_l is the set of the linearization points. The δ value is a pre-determined distance threshold, and it can be determined in Algorithm 2. In Algorithm 2, s is the number of models supposed to be generated.

Algorithm 2: Determination of the pre-determined distance threshold δ value algorithm

- 1. Find the maximum distance between any of the two points on the trajectory, $d_j = ||x_j x_h||_2$, $j, h \in \{0, 1, ..., N\}$.
- 2. Set $\delta = d_{max}/s$.

The state dependent weight $w_i(x)$ shown in Equation (26) can be computed as described in Algrithm 3 [33].

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Algorithm 3: Computation algorithm for the state dependent weight parameter $w_i(x)$

- 1. At each linearization point (x_i, u_i) , compute the distance $d_i = ||x x_i||_2$.
- 2. Find the minimum value among d_i , $m = \min_{i=0,...,(s-1)} d_i$.
- 3. For i = 0, ..., (s-1) compute $\hat{w}_i = e^{-\beta d_i/m}$.
- 4. Compute the summation of \hat{w}_i , $S(x) = \sum_{i=0}^{s-1} \hat{w}_i(x)$.
- 5. Compute the normalized parameter $w_i(x) = \frac{\hat{w}_i(x)}{S(x)}$.

In Algorithm 3, β is a positive constant value. The weighting parameter $w_i(x)$ changes according to the position of current state x in state space. The exponential term in step 3 to determine the weighting parameter ensures that the distribution of the weight w_i will change immediately close to one if current state x is sufficiently close to the linearization point x_i .

4. TPWL Model Based on POD Method

In this section, the TPWL method is combined with the POD method to further reduce the evaluation time. The POD method is introduced in this section. The steps of establishing the TPWL-POD model are presented and the representation of the model is also shown in this section.

4.1. POD Method Introduction

The POD method can be used to obtain a low-order model but capture the most important dynamics of the original complicated systems [24]. POD method is a SVD-based approximation method to derive the low dimensional system to approximate the large scale system. It can be applied to both high-complexity linear and nonlinear systems [38].

Giving a fixed input u, the trajectory of state $x \in \mathbb{R}^n$ at certain time instances t_k can be measured as:

$$X = [x(t_1) \ x(t_2) \ x(t_3) \ \dots \ x(t_N)] \tag{27}$$

 $X \in \mathbb{R}^{n \times N}$ in (27) can be called as the snapshot matrix of the process data. It should be mentioned that the measured time instant points N should be much greater than the dimension of the system n, i.e., $N \gg n$. By computing the singular value decomposition of the snapshot matrix, X can be decomposed into a product of three matrices [38]:

$$X = U\Sigma V^T \in \mathbb{R}^{n \times N} \tag{28}$$

where $U \in \mathbb{R}^{n \times n}$ and $V \in \mathbb{R}^{N \times N}$ are orthonormal and called the left and right singular vector, respectively. $\Sigma = \operatorname{diag}(\sigma_1, \dots, \sigma_n) \in \mathbb{R}^{n \times N}$ is a diagonal matrix, and each diagonal entry of the matrix is called the singular values. The singular values of X are nonnegative numbers and ordered in a decreasing way, i.e., $\sigma_1 \geq \sigma_2 \geq \cdots \geq \sigma_n$. The greater σ value represents the basis vector captures the more important information present in the data [39]. If the singular values of the matrix drop off rapidly, we can obtain a low-dimensional approximated system [38]:

$$X = U\Sigma V^T \approx U_k \Sigma_k V_k^T, \ k \ll n \tag{29}$$

Let $x(t) \approx U_k z(t)$, $z(t) \in \mathbb{R}^k$, the reduced-order system model can be shown as follows:

$$\dot{z}(t) = U_k^T f(U_k z(t), u(t)) \tag{30}$$

where z(t) is the approximation of states x(t) in a low-dimensional space which is spanned by the first k columns of the left singular vector of X [38], i.e., U_k . Note that POD alone is not effective in reducing the evaluation time of EMPC or other optimization-based control schemes. For nonlinear

systems, it is in general very difficult to express $U_k^T f(U_k z, u)$ explicitly. This makes the evaluation time of $U_k^T f(U_k z, u)$ often even more than the evaluation time of the original nonlinear function f. However, when POD is used together with TPWL, they can lead to significant evaluation time reduction. This method is discussed in the following subsection.

4.2. TPWL-POD Model Representation

Assuming we have generated the reduced order basis U_k via POD method for the process system, and the reduced order model of the system is with k states. The relationship between state x with order N in the original space and projection z with order k ($k \ll N$) in reduced-order space can be represented in the following form:

$$x = U_k z \tag{31}$$

where $U_k \in \mathbb{R}^{N \times k}$ is an orthogonal matrix and represents the projection of x in original space onto z in the reduced-order space.

Combining with Equation (31), the TPWL model of Equation (26) generated in Section 3.1 can be represented as follows:

$$\frac{d(U_k z)}{dt} = \sum_{i=0}^{s-1} w_i(z) (f(U_k z_i, u_i) + A_i (U_k z - U_k z_i) + B_i (u - u_i))$$
(32)

where z_i is the projection of the linearized points x_i in the reduced-order space, and $[z_0, z_1, \dots, z_{s-1}] = [U_k^T x_0, U_k^T x_1, \dots, U_k^T x_{s-1}]$

Multiplying Equation (32) by U_k^T , the model can be shown as:

$$\frac{dz}{dt} = \sum_{i=0}^{s-1} w_i(z) (A_{ir}z + B_{ir}u + \gamma_{ir})$$
(33)

where

$$A_{ir} = U_k^T A_i U_k$$

$$B_{ir} = U_k^T B_i$$

$$\gamma_{ir} = U_k^T (f(x_i, u_i) - A_i x_i - B_i u_i)$$

$$\sum_{i=0}^{s-1} w_i(z) = 1$$

The weight $w_i(z)$ is calculated based on the distance between current projected state z and the linearization point z_i , and it follows the Algorithm 3 shown in Section 3.1.

4.3. Generation Method of TPWL-POD Model

The generation of the reduced order TPWL-POD model consists of two parts: generation of the POD reduced basis and the generation of trajectory piecewise linear model. Figure 2 shows the methodology used in this work.

The strategy to generate the POD basis for WWTP process is described by Algorithm 4.

The strategy to generate the piecewise linear model is the same as the one shown in Section 3.2. After generating the linearization points x_i , these linearization points are projected to the reduced-order space, i.e., $z_i = U_k^T x_i$. The weighting parameter $w_i(z)$ is computed based on the states in reduced-order space, and the strategy to calculate it is the same as the steps in Algorithm 3.

Algorithm 4: Generation of POD basis algorithm

1. Simulate the nonlinear system for $t \in (0, N]$ to generate the snapshot matrix of the process, i.e., X.

- 2. Take singular value decomposition on snapshot matrix X, $X = U\Sigma V^T$.
- 3. Determine the proper reduced order *k* to choose the *k* most relevant basis vectors of the system.
- 4. Generate the POD basis vectors matrix U_k .

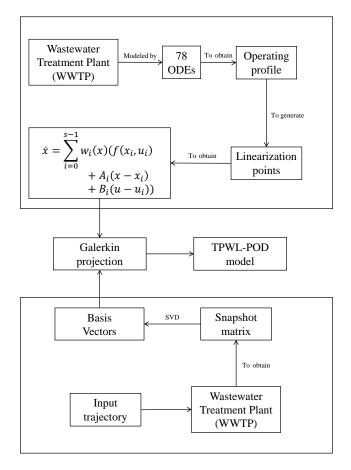


Figure 2. TPWL-POD framework.

5. Centralized EMPC Design Based on TPWL Model and TPWL-POD Model

In this section, we present the proposed centralized EMPC design based on two different models. The two considered models are as follows: TPWL model introduced in Section 3 and TPWL-POD model introduced in Section 4. The control objective is to minimize the economic cost introduced in Section 2.3.

5.1. Centralized EMPC Design Based on TPWL Model

First, the nonlinear model of the WWTP system is linearized using the trajectory piecewise linearization method proposed in Section 3. The EMPC developed based on the TPWL model can be formulated as the following optimization problem:

$$u^*(\tau|t_k) = \arg\min_{u(\tau) \in S(\Delta)} \int_{t_k}^{t_{k+N}} l\left(\tilde{x}(\tau), u(\tau)\right) d\tau \tag{34a}$$

s.t.
$$\dot{\tilde{x}}(\tau) = \sum_{i=0}^{s-1} w_i(\tilde{x}(\tau))(f(x_i, u_i) + A_i(\tilde{x}(\tau) - x_i) + B_i(u(\tau) - u_i))$$
(34b)

$$\tilde{x}(t_k) = x(t_k) \tag{34c}$$

$$u(\tau) \in \mathbb{U}$$
 (34d)

In the above optimization problem of Equation (34), $u^*(\tau|t_k)$ is the optimal solution to the problem. Equation (34a) denotes the objective function for the centralized EMPC controller that minimizes the economic cost l defined in Equation (17), N denotes the control horizon, and $S_i(\delta)$ represents a family of piecewise-constant functions. Equation (34b) is the approximated piecewise linear model of the nonlinear system as introduced in Section 3. The state measurement value at current sampling time t_k (i.e., $x(t_k)$) is used to initialize the predicted state trajectory in Equation (34c). Equation (34d) is the constraint on input u.

The optimal input trajectory is achieved after the optimization problem Equation (34) is solved (i.e., $u^*(\tau|t_k)$). The first step value of the input trajectory is defined to be the manipulated input applied to the operating process at time instant t_k , which can be shown as follows:

$$u(t) = u^*(t|t_k), t \in [t_k, t_{k+1})$$
(35)

5.2. Centralized EMPC Design Based on TPWL-POD Model

In this section, the centralized EMPC is designed based on the TPWL-POD model. The model is linearized using TPWL method shown in Section 3 and the order of the nonlinear system is reduced using POD method proposed in Section 4. The proposed EMPC design can be shown as follows:

$$u^*(\tau|t_k) = \arg\min_{u(\tau) \in S(\Delta)} \int_{t_k}^{t_{k+N}} l(U_k z(\tau), u(\tau)) d\tau$$
(36a)

s.t.
$$\dot{\tilde{z}}(\tau) = \sum_{i=0}^{s-1} w_i(z) (A_{ir}z + B_{ir}u + \gamma_{ir})$$
 (36b)

$$\tilde{z}(t_k) = U_k^T x(t_k) \tag{36c}$$

$$A_{ir} = U_k^T A_i U_k (36d)$$

$$B_{ir} = U_k^T B_i (36e)$$

$$\gamma_{ir} = U_k^T (f(x_i, u_i) - A_i x_i - B_i u_i)$$
(36f)

$$\sum_{i=0}^{s-1} w_i(z) = 1 \tag{36g}$$

$$u(\tau) \in \mathbb{U}$$
 (36h)

In the optimization problem of Equation (36), let $u^*(\tau|t_k)$ accounts for the optimal solution to this problem. The objective function is Equation (36a). Equation (36b) denotes the order reduced linear model using TPWL-POD method introduced in Section 4 for the nonlinear system. A_{ir} , B_{ir} , γ_{ir} , and $w_i(z)$ in model Equation (36b) are shown in Equations (36d)–(36g), respectively. In Equation (36c), the predicted state in the reduced-order space is initialized with the state measurement $x(t_k)$ and it is projected to the reduced-order space by multiplying the projection matrix U_k . Equation (36h) is the input constraint. The manipulated input applied to the control process is the first step value of the optimal input trajectory, i.e., $u^*(t|t_k)$, $t \in [t_k, t_{k+1})$.

6. Simulation Results

In this section, we apply the TPWL method introduced in Section 3 and the TPWL-POD method introduced in Section 4 to the WWTP. The reduced order basis and the linearization points for both TPWL model and TPWL-POD model are obtained based on a given training input signal. The approximated model accuracy for the nonlinear system with TPWL model and TPWL-POD model are discussed. We apply the proposed control strategies introduced in Sections 5.1 and 5.2 to the modified BSM1 model. The performance of these control strategies is compared in terms of effluent quality, operating cost and computational efficiency under dry weather condition. The simulations are all carried out with a computing server with 2.0 GHz CPUs and the EMPC optimization problems are solved using interior point optimizer.

6.1. Simulation Settings

The dry weather condition data file are provided in the International Water Association Web site [40]. The inlet flowrate Q_0 , and the concentration Z_0 of the influent flow can be found in the data file. The wastewater treatment plant is simulated with the average value of Q_0 and Z_0 under dry weather condition to achieve the optimal steady state. We consider the optimal steady state as the initial condition for this wastewater treatment process.

To compare the performance index between different control strategies, the simulation time is set to be 14 days and the simulation results of the last 7 days will be used to evaluate the control performance index. The sampling time is picked as $\Delta=15$ min.

The weighting parameter β in Algorithm 3 is set to be 25. The weighting coefficients α_{EQ} and α_{OCI} in the economic control objective function Equation (17) are considered to be 1 and 0.3. The control horizon is determined to be 25 in all EMPC controller designs. The constraints on manipulated inputs (i.e., \mathbb{U}) for all control designs are considered as follows:

$$0 \le Q_a \le 5Q_{0,\text{stab}} \tag{37}$$

$$0 \le K_L a_5 \le 240 \, \mathrm{day}^{-1} \tag{38}$$

where $Q_{0,\text{stab}}$ is the average dry weather influent flow rate and is equal to 18,446 m³/day.

6.2. Model Validation and Comparison

6.2.1. Model Validation

To verify the generated TPWL model and TPWL-POD model not only work for the inputs which are very close to the given training input but also for other inputs, the accuracy of the model is investigated under different manipulated input values. For model validation, the number of linearization points s in both TPWL model and TPWL-POD model is set to be 9, and the reduced basis order k is set to be 35 for TPWL-POD model. Note that the number of linearization points s and the order k are determined based on extensive simulations. The guideline is to achieve a balance between model accuracy and evaluation time. In general, with the increase of s and k, the accuracy of the model increases but the time needed in evaluation also increases. When determining s and k, we need to achieve a balance between accuracy and evaluation time, which is indeed case specific.

The approximated state trajectories based on TPWL model, TPWL-POD model and the actual state trajectories based on the nonlinear system model for certain process states in the first week of the operation under the training input signal and fixed input signal are presented in Figures 3 and 4, respectively. For the fixed input signal, the two manipulated inputs are fixed as $K_L a_5 = 83.9405 \, \mathrm{d}^{-1}$ and $Q_a = 37,723.4072 \, \mathrm{m}^3/\mathrm{day}$. As can be seen from Figures 3 and 4, the state trajectories almost overlap each other in all the cases. The results show that TPWL model and TPWL-POD model provide accurate models for the original nonlinear WWTP system.

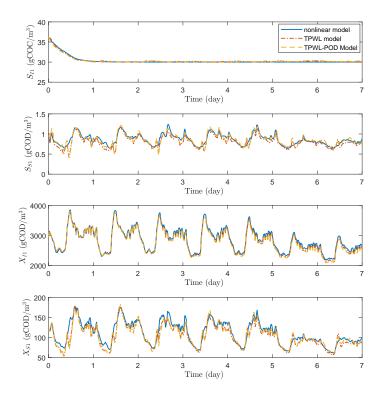


Figure 3. Trajectories of the states based on nonlinear model (blue solid lines), TPWL model (red dash-dot lines), TPWL-POD model (yellow dashed lines) under training input signal in dry weather.

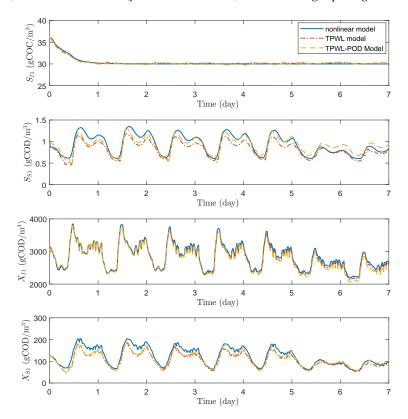


Figure 4. Trajectories of the states based on nonlinear model (blue solid lines), TPWL model (red dash-dot lines), TPWL-POD model (yellow dashed lines) under fixed input signal in dry weather.

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6.2.2. Model Comparison

The model approximation accuracy and computation time for TPWL method and TPWL-POD method are investigated under different number of the linearization points (*s*).

Figure 5 presents the state trajectories for certain process states using TPWL model with different number of linearization points under a given input signal in dry weather condition. Figure 6 shows the state trajectories for those states using TPWL-POD model with different number of linearization points under the same input signal in dry weather condition.

Table 4 presents the root mean square error of TPWL model and TPWL-POD model with different number of the linearization points. The result elucidates that the approximated model accuracy would increase as the number of linearization points increase for both TPWL and TPWL-POD models. As can be seen from the table, under the same linearization point number, TPWL model provides a better approximation than TPWL-POD model.

Table 4. Root mean square error of TPWL model and TPWL-POD model with different number of the linearization points.

Threshold δ	Linearization Point Number	RMSE TPWL Model	RMSE TPWL-POD Model
1000	4	18.3979	48.6197
950	6	10.5667	12.4511
600	9	6.0770	8.6849
570	10	5.5003	6.5913

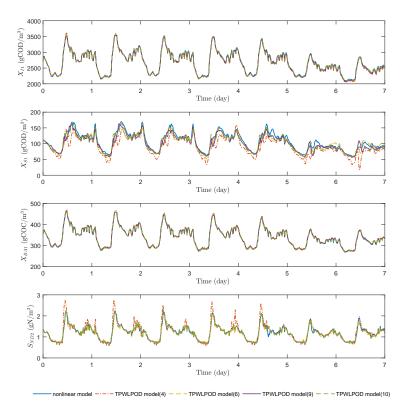


Figure 5. Trajectories of the states based on nonlinear model (blue solid lines), TPWL model with 4 linearization points (red dash-dot lines), TPWL model with 6 linearization points (yellow dashed lines), TPWL model with 9 linearization points (purple solid lines), and TPWL model with 10 linearization points (green dashed lines) under a given input signal in dry weather.

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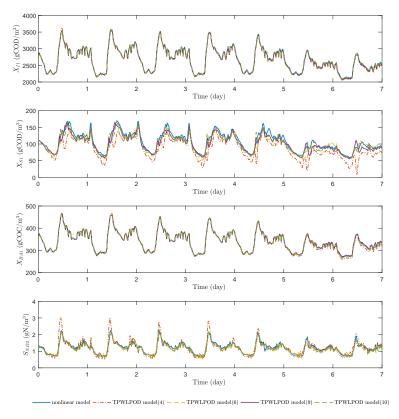


Figure 6. Trajectories of the states based on nonlinear model (blue solid lines), TPWL-POD model with 4 linearization points (red dash-dot lines), TPWL-POD model with 6 linearization points (yellow dashed lines), TPWL-POD model with 9 linearization points (purple solid lines), and TPWL-POD model with 10 linearization points (green dashed lines) under a given input signal in dry weather.

6.3. Simulation Results of EMPC in Dry Weather

We evaluate the performance of EMPC based on the nonlinear model, EMPC based on TPWL model and EMPC based on TPWL-POD model in dry weather condition, respectively.

To study the impacts of the number of linearization points (s) on EMPC design based on TPWL model, we apply the proposed EMPC scheme with s = 4, s = 6, s = 9, and s = 10 respectively. The trajectories of the instantaneous effluent quality level in dry weather condition under EMPC based on the nonlinear model, EMPC based on TPWL model with linearization point number s = 4, s = 6, s = 9, s = 10 are presented in Figure 7. Table 5 presents the calculated control performance index EQ, OCI, average economic cost, and average evaluation time consumed over the full operating period for EMPC based on nonlinear model and EMPC based on TPWL model with s = 4, s = 6, s = 9, s = 10. Based on the average EQ value shown in Table 5 and the instantaneous effluent quality level shown in Figure 7, we can conclude that the effluent quality level is improved as the linearization point number s in TPWL model increases. Our objective is to minimize the average economic cost. Consequently, the smaller average performance cost indicates that the model can give us a better control performance. As can be seen from Table 5, the EMPC with nonlinear model provides us with the best control performance, and the control performance of EMPC based on TPWL model enhances as the linearization point number raises. The EMPC based on TPWL model with s=4 performs 18.24% worse than nonlinear model while the EMPC based on TPWL model with s=10 is only 5.93%lower than nonlinear model. As the linearization point number s increases from 4 to 10, the control performance improves 10.42%.

To compare the evaluation times of EMPC based on the nonlinear model and EMPC based on TPWL model with s=4, s=6, s=9, s=10, the average computation time consumed over the full operation period is evaluated and shown in Table 5. The average computation time over the full

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operation period is evaluated based on 10 repetitive simulation runs. The evaluation time consumed by EMPC based on TPWL model with s=4, s=6, and s=9 is 30.02%, 27.97%, and 22.05% less than EMPC based on nonlinear model, respectively. However, as the linearization point number s increases, the computation load increases accordingly. The EMPC based on TPWL model with s=10 is 16.29% lower than EMPC with the nonlinear model.

We also evaluate how linearization point number s influences the control performance of EMPC based on TPWL-POD model. Figure 8 presents the trajectories of the instantaneous effluent quality level in dry weather condition under EMPC based on the nonlinear model, EMPC based on TPWL-POD model with linearization point number s=4, s=6, s=9, s=10. Table 6 shows the EQ, OCI, average economic cost and average evaluation time of EMPC based on nonlinear model, EMPC based on TPWL-POD model with s=4, s=6, s=9 and s=10. We can draw the same conclusion as we have for EMPC based on TPWL model. The more linearization points we consider, the better control performance of EMPC based on TPWL-POD model would be. The average evaluation time consumed by TPWL-POD model will raise when the linearization point number s=10 increases. Compared with the evaluation time consumed by EMPC based on the nonlinear model, EMPC based on TPWL-POD model with s=10 is 70.13%, 67.40%, 54.51% and 15.62% less, respectively.

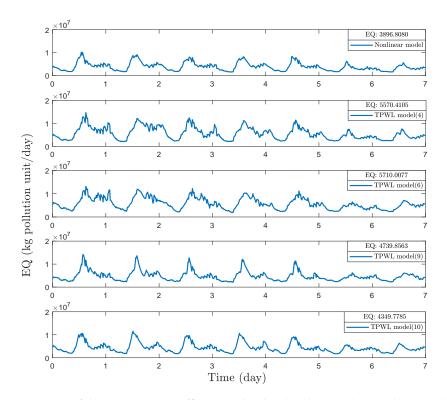


Figure 7. Trajectories of the instantaneous effluent quality level in dry weather condition under EMPC based on nonlinear model, TPWL model with 4 linearization points, TPWL model with 6 linearization points, TPWL model with 9 linearization points and TPWL model with 10 linearization points.

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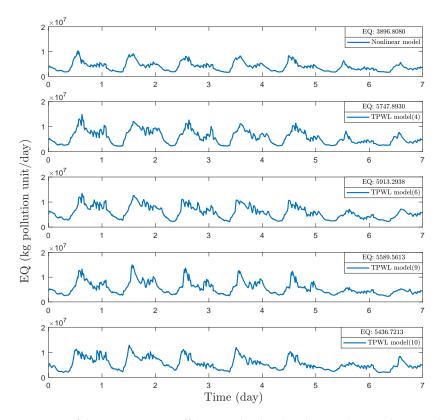


Figure 8. Trajectories of the instantaneous effluent quality level in dry weather condition under EMPC based on nonlinear model, TPWL-POD model with 4 linearization points, TPWL-POD model with 6 linearization points, TPWL-POD model with 9 linearization points and TPWL-POD model with 10 linearization points.

Table 5. Control performance in dry weather condition under EMPC based on nonlinear model, TPWL model with different linearization point numbers (s = 4, s = 6, s = 9, s = 10).

Model	Linearization Point Number	Average Computation Time (s)	EQ (kg Pollution Unit/Day)	OCI	Average Performance Cost
Nonlinear	N/A	$7.34 imes 10^4$	3896.8080	19,592.3016	9774.4985
TPWL	4	5.13×10^{4}	5570.4105	19,958.9991	11,558.1102
	6	5.28×10^{4}	5710.0077	19,404.2978	11,531.2970
	9	5.72×10^{4}	4739.8563	19,654.6556	10,636.2530
	10	8.53×10^{4}	4349.7785	20,014.8107	10,354.2217

By examining Tables 5 and 6, we can see that with the same linearization point number *s*, the average evaluation time used by EMPC based on TPWL-POD model is approximately 58.54% less than that based on the TPWL model. The average evaluation time of the EMPC based on TPWL-POD model is up to 70.13% lower than EMPC based on the nonlinear model. However, it should be noted that the control performance for EMPC based on TPWL-POD model is degraded by 1.39% compared with the TPWL model and 19.89% compared with the original nonlinear model.

Model	Linearization Point Number	Average Computation Time (s)	EQ (kg Pollution Unit/Day)	OCI	Average Performance Cost
Nonlinear	N/A	7.34×10^{4}	3896.8080	19,592.3016	9774.4985
TPWL-POD	4	2.19×10^{4}	5747.8930	19,902.6197	11,718.6789
	6	2.39×10^{4}	5913.2938	19,378.8133	11,726.9378
	9	3.34×10^{4}	5589.5613	19,328.1781	11,388.0147
	10	6.19×10^{4}	5436.7213	19,280.6409	11,220.9136

Table 6. Control performance in dry weather condition under EMPC based on nonlinear model, TPWL-POD model with different linearization point numbers (s = 4, s = 6, s = 9, s = 10).

7. Conclusions

Two model approximation methods are applied to the modified nonlinear WWTP process. In particular, the two model approximation methods are TPWL method and TPWL-POD method. Two centralized EMPC controllers are designed based on the two models correspondingly. We have compared the model accuracy of the TPWL model and the TPWL-POD model with nonlinear model. In this work, our main objective was to investigate how the computational efficiency of EMPC may be improved through model reduction and approximation. While the investigated methods lead to improved computational efficiency, it does come at the cost of reduced control performance. In the future work, we will study how to improve the computational efficiency while remain the control performance close to the one when the original nonlinear model is used.

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