



Article Mechanical Properties and Water Resistance of Magnesium Oxychloride Cement–Solidified Residual Sludge

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Abstract: As a solid waste, the amount of residual sludge produced by the municipal wastewater treatment process is escalating. How to dispose it properly is attracting much attention in society. Herein, solidifying residual sludge using magnesium oxychloride cement (MOC) is promising for converting it into building materials. Various factors of mass ratio ($R_{W/S}$) of liquid to solid, molar ratio (R_n) of MgO to MgCl₂ in MOC, mass ratio (R_m) of residual sludge to MOC, the mass concentration of Na_2SiO_3 ($D_{Na_2SiO_3}$), and dosage of fly ash (D_F) influenced the unconfined compression strength (R_C) of the as-obtained MOC-solidified residual sludge, and it was characterized using SEM and XRD analysis. The results show that the value of R_C for MOC-residual sludge solidified blocks increased initially and then decreased as R_n and R_m increased, respectively, for 60–day curing. At 10–day curing, equilibrium $R_{\rm C}$ was reached at all $R_{\rm W/S}$ values except 1.38, and at 60–day curing, $R_{\rm C}$ decreased with $R_{W/S}$ increasing. The maximum R_C of 60 days of 20.90 MPa was obtained at $R_{W/S} = 0.90$, $R_n = 5.0$, and $R_m = 1.00$. Furthermore, adding Na₂SiO₃ or fly ash in the solidifying process could improve R_C. The water resistance test showed that SM13 and NF5 samples exhibited good alkaline resistance after immersion for 7 and 14 days in an aqueous solution with pH = 7.0-11.0. The water resistance of MOC-residual sludge solidified blocks decreased with increase in immersion duration in aqueous solutions. The fly ash could also help improve water resistance of MOC-solidified residual sludge in neutral and basic aqueous solutions. This work provides an important theoretical basis and possibility for the efficient disposal and comprehensive utilization of residual sludge through solidification/stabilization technology using MOC from the perspective of mechanics and water resistance.

Keywords: magnesium oxychloride cement; residual sludge; solidification; unconfined compression strength

1. Introduction

With the social development and progress of human civilization, more emphasis is placed on environmental health, and safe and beneficial disposal of wastewater. As a byproduct, the residual sludge is a dilute suspension of solids generated commonly in screening, coagulation–flocculation, sedimentation, and granular filtration procedures of the municipal wastewater treatment process [1–4]. The amount of residual sludge is increasing worldwide at an alarming rate because of the increased amount of municipal sewage and rapid development of population, urbanization, industrialization, sewage drainage systems, and wastewater treatment facilities [1,5–12]. The heavy metals, organic and inorganic compounds, and microorganisms contained in residual sludge are harmful to soil, vegetation, animals, and humans [6–9,12–23]. Some sludge was transported for treatment as a hazardous waste before discharging into the environment [24]. Due to the potential toxicity and ecotoxicological risks of untreated sludge to the environment, conventional methods such as discarding in landfills and water bodies and composting are being eliminated [6,8,9,12,21]. This has given rise to three types methods, namely, thermal treatment, biological treatment, and chemical treatment [7,9,12,18,21–23,25–32].



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Copyright: © 2023 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (https:// creativecommons.org/licenses/by/ 4.0/). These methods are, however, not widely applied in China, and only a small proportion of residual sludge has been used for composting and material recovery [6,7,29]. Thus, the proper recycling of residual sludge is needed to alleviate a number of environmental problems and reduce operating costs of water treatment plants.

Recycling residual sludge as the traditional raw ingredients in the construction industry is a promising approach, which is used mainly in forms of dewatered sludge, dry sludge, or incinerated sludge ash in different construction applications such as cementbased materials production, ceramic products, lightweight construction materials, soil stabilization, and other civil engineering applications (such as landfill lining) [33–57], while in the literatures, using residual sludge in a form of semiliquid state (with high content of water) as a construction material is rarely reported.

Solidification/stabilization (S/S) methods can prevent harmful chemicals from being released into the environment though chemical or mechanical binding using curing agents [6,58–60]. Magnesium oxychloride cement (MOC), known as Sorel cement, comprises MgO and MgCl₂; it is typically used as a curing agent for S/S of solid waste, because of its good engineering and mechanical properties, fast curing, air hardening, good resistance to abrasion and fire, and so on [61–67]. MgO and MgCl₂ react with water to form MgO–MgCl₂–H₂O systems such as Mg(OH)₂·MgCl₂·8H₂O (phase 3) and 5 Mg(OH)₂·MgCl₂·8H₂O (phase 5), which are the main substances to form strength of MOC–solid waste in the solidifying process [64,67–75]. Using MOC to solidify the residual sludge in a form of semiliquid state could replace a part of water; the containing contaminants would be bound in MOC–sludge solidified blocks to alleviate some environmental problems from residual sludge.

In this study, residual sludge from sewage treatment plants was solidified using MOC as a building material. Fly ash and sodium silicate (Na₂SiO₃) were added during the solidification process for improving unconfined compressive strength ($R_{\rm C}$) and moisture absorption. $R_{\rm C}$ tests were used to investigate the influences of major factors, namely, mass ratio ($R_{W/S}$) of liquid to solid, molar ratio (R_n) of MgO to MgCl₂ in MOC, mass ratio (R_m) of residual sludge to MOC, the mass concentration of Na₂SiO₃ ($D_{Na_2SiO_3}$), and dosage of fly ash $(D_{\rm F})$. The microtopography of MOC–solidified residual sludge was observed using scanning electron microscopy (SEM), and crystal composition of crystal was analyzed using X-ray diffraction (XRD) analysis to reveal intrinsic mechanisms. To investigate the water resistance properties of MOC-sludge solidified block, and effects of Na₂SiO₃ and fly ash on the water resistance properties, SM13, NG7, and NF5 samples were used as representatives to measure the unconfined compression strength (R_{AC}) after immersion in aqueous solutions with varying pH values for 7 or 14 days. The coefficient of corrosion resistance (K_F) was used to evaluate the water resistance in acidic, neutral, and basic aqueous solutions. We hope the MOC-solidified residual sludge could satisfy the mechanical and water resistance requirements as a building material though this study, providing a theoretical basis for utilization of residual sludge as a resource.

2. Experiments

2.1. Raw Materials

The residual sludge (aqueous, dark brown) used in this study was taken from a wastewater treatment plant in Liaoning Province. MgO and MgCl₂ were analytical grade and purchased from Bidepharm Medical Technology Co., Ltd. (Shanghai, China). Analytical–grade sodium silicate (Na₂SiO₃; modulus = 2.43, apparent density = $0.76 \text{ g} \cdot \text{cm}^{-3}$) was purchased from Shanghai Rhawn Chemical Technology Co., Ltd. (Shanghai, China). Industrial-grade fly ash was purchased from Henan Gongyi Longze Water Purification Material Co., Ltd. (Gongyi, China). The main contents of the fly ash were SiO₂, Al₂O₃, Fe₂O₃, CaO and so on [6]. All chemicals were used as received. Deionized water was obtained from a Hitech–Kflow water purification system (Hitech Co., Ltd., Beijing, China).

2.2. Solidification Experiment of Residual Sludge

The residual sludge was mixed with specific amounts of MgO, MgCl₂, Na₂SiO₃, or fly ash using a slightly modified form of the method reported in our previous study [6]. Then, deionized water was added to the solid at a liquid to solid ratio, $R_{W/S}$, of 0.9–1.38, and then mixed using a YD90S-8/4 cement mortar mixer (Wuxi Construction Engineering Test Equipment Co., Ltd., Wuxi, China). R_{W/S} was calculated using Formula (1). The value of R_n of MgO to MgCl₂ varied in the range of 2.0–8.0, and the value of R_m of residual sludge to MOC varied in the range of 0.50-1.50. The dosage of residual sludge (D_S) in the mixed mortar varied in the range of 19.83 wt%–53.33 wt%. $D_{\text{Na}_2\text{SiO}_3}$ and D_F were calculated using Formulas (2) and (3), respectively. The process of solidification of residual sludge using MOC is shown in Table 1. In Table 1, M1–M12 samples were designed as comparison for investigating the effects of R_n and $R_{W/S}$ on R_C , respectively; SM1–SM7, SM8–SM12, and SM13–SM16 were designed to investigate the effects of R_n, R_m, and R_{W/S} on R_C of MOC-residual sludge solidified blocks, respectively; NG1-NG7 and NF5-NF20 were designed to investigate the effects of $D_{Na_2SiO_3}$ and D_F on R_C of MOC–residual sludge solidified blocks, respectively. The MOC-residual sludge mixing mud was shaken in a shaking table for 20 min to remove air bubbles, and then it was injected into a cubic Plexiglass mold (dimensions = $40 \text{ mm} \times 40 \text{ mm}$). The MOC–residual sludge mixing mud in cubic Plexiglass mold was treated in the curing box at 20 ± 1 °C with relative humidity of \geq 95% for 24 h, forming solidified blocks. These solidified blocks were then demolded and sealed with plastic cling film, and placed under natural maintenance conditions to be cured for 3–60 days [10]. The values of $R_{\rm C}$ (calculated using Formula (4)) of 3–60 days were measured using a YAW-1000A Electro-hydraulic servo pressure testing machine controlled by microcomputer (Jinan Dong Fang Test Instrument Co., Ltd., Jinan, China). The testing machine loads a specimen at a speed of 2400 ± 200 N/s until it is destroyed, and the maximum load is F_{C} .

$$R_{\rm W/S} = \frac{m_{\rm S} \times 87.75\% + m_{\rm W}}{m_{\rm S} \times 12.25\% + m_{\rm MgO} + m_{\rm MgCl_2} + m_{\rm Na_2SiO_3} + m_{\rm F}}$$
(1)

$$D_{\rm F} = \frac{m_{\rm F}}{m_{\rm F} + m_{\rm MOC}} \times 100\% \tag{2}$$

$$D_{\text{Na}_2\text{SiO}_3} = \frac{m_{\text{Na}_2\text{SiO}_3}}{m_{\text{S}} \times 87.75\% + m_{\text{w}} + m_{\text{Na}_2\text{SiO}_3}} \times 100\%$$
(3)

$$R_{\rm C} = F_{\rm C} / A \tag{4}$$

where m_S , m_w , m_F , m_{MgO} , m_{MgCl_2} , and $m_{Na_2SiO_3}$ are the masses (gram) of surplus sludge, added water, fly ash, MgO, MgCl₂, and Na₂SiO₃, respectively; m_{MOC} is the total mass of MgO and MgCl₂. F_C is the maximum load (newton) at failure. A is the pressure area (mm²).

Table 1. Experiment scheme of MOC solidifying residual sludge.

Samples	D _S (wt%)	D _{MgO} (wt%)	D _{MgCl2} (wt%)	D _{MOC} (wt%)	D _W (wt%)	R _{W/S}	R _n	R _m	D _{Na2SiO3} (wt%)	D _F (wt%)
M1	-	19.31	22.78	42.09	57.91	1.38	2.0	-	-	-
M2	-	23.55	18.54	42.09	57.91	1.38	3.0	-	-	-
M3	-	26.44	15.65	42.09	57.91	1.38	4.0	-	-	-
M4	-	28.57	13.52	42.09	57.91	1.38	5.0	-	-	-
M5	-	30.20	11.89	42.09	57.91	1.38	6.0	-	-	-
M6	-	31.46	10.63	42.09	57.91	1.38	7.0	-	-	-
M7	-	32.48	9.61	42.09	57.91	1.38	8.0	-	-	-
M8	-	39.47	18.67	58.14	41.86	0.72	5.0	-	-	-
M9	-	37.93	17.94	55.87	44.13	0.79	5.0	-	-	-
M10	-	36.70	17.36	54.06	45.94	0.85	5.0	-	-	-
M11	-	30.86	14.60	45.46	54.54	1.20	5.0	-	-	-
M12	-	27.15	12.85	40.00	60.00	1.50	5.0	-	-	-

Samples	D _S (wt%)	D _{MgO} (wt%)	D _{MgCl2} (wt%)	D _{MOC} (wt%)	D _W (wt%)	R _{W/S}	R _n	R _m	D _{Na2SiO3} (wt%)	D _F (wt%)
SM1	37.50	17.20	20.30	37.50	25.00	1.38	2.0	1.00	-	-
SM2	37.50	20.98	16.52	37.50	25.00	1.38	3.0	1.00	-	-
SM3	37.50	23.56	13.94	37.50	25.00	1.38	4.0	1.00	-	-
SM4	37.50	25.46	12.04	37.50	25.00	1.38	5.0	1.00	-	-
SM5	37.50	26.90	10.60	37.50	25.00	1.38	6.0	1.00	-	-
SM6	37.50	28.03	9.47	37.50	25.00	1.38	7.0	1.00	-	-
SM7	37.50	28.93	8.57	37.50	25.00	1.38	8.0	1.00	-	-
SM8	19.83	22.19	17.47	39.66	40.51	1.38	3.0	0.50	-	-
SM9	25.94	21.77	17.14	38.91	35.15	1.38	3.0	0.67	-	-
SM10	32.40	21.33	16.79	38.12	29.48	1.38	3.0	0.85	-	-
SM11	44.03	20.53	16.17	36.70	19.27	1.38	3.0	1.20	-	-
SM12	53.33	19.89	15.67	35.56	11.11	1.38	3.0	1.50	-	-
SM13	46.89	31.83	15.05	46.88	6.23	0.90	5.0	1.00	-	-
SM14	43.46	29.50	13.95	43.45	13.09	1.05	5.0	1.00	-	-
SM15	40.49	27.49	13.00	40.49	19.02	1.20	5.0	1.00	-	-
SM16	35.63	24.20	11.44	35.64	28.73	1.50	5.0	1.00	-	-
NG1	46.89	31.83	15.05	46.88	6.23	0.90	5.0	1.00	1.0	-
NG3	46.89	31.83	15.05	46.88	6.23	0.90	5.0	1.00	3.0	-
NG5	46.89	31.83	15.05	46.88	6.23	0.90	5.0	1.00	5.0	-
NG7	46.89	31.83	15.05	46.88	6.23	0.90	5.0	1.00	7.0	-
NF5	46.89	30.24	14.30	44.54	6.23	0.90	5.0	1.00	-	5.0
NF10	46.89	28.65	13.55	42.20	6.23	0.90	5.0	1.00	-	10.0
NF15	46.89	27.06	12.80	39.86	6.23	0.90	5.0	1.00	-	15.0
NF20	46.89	25.47	12.04	37.51	6.23	0.90	5.0	1.00	-	20.0

Table 1. Cont.

 D_5 , D_{MgC0} , D_{MgCl_2} , $D_{Na_2SiO_3}$, and D_W are the weight percentage of residual sludge, MgO, MgCl_2, Na_2SiO_3, and added water in MOC-residual sludge mixing mortar; D_{MOC} is the total percentage of D_{MgO} and D_{MgCl_2} .

2.3. Water Resistance Test of Solidified Blocks

After 28–day curing, the solidified block samples of SM13, NG7, and NF5 were immersed into aqueous solutions with different pH values (3.0–11.0), respectively, for 7 or 14 days. The pH of the aqueous solution was monitored using the electrode method every 24 h and adjusted using sulfuric acid or sodium hydroxide to keep the pH constant. The unconfined compression strength of solidified blocks after immersion in aqueous solutions with different pH values was denoted as R_{AC} . The unconfined compression strength of the solidified block samples without immersion in aqueous solution was denoted as R_{C} . The coefficient of corrosion resistance (K_{F}) was given as the ratio of R_{AC} to R_{C} , and it was calculated using Formula (5).

$$K_{\rm F} = \frac{R_{\rm AC}}{R_{\rm C}} \times 100\% \tag{5}$$

All tests described above were conducted in triplicate, and the final $R_{\rm C}$ or $R_{\rm AC}$ values were an average of the three measurements.

2.4. Characterization

The composition and pH value of residual sludge were determined using a mass-loss technique and the electrode method according to a standard of determination method for municipal sludge in wastewater treatment plants (CJ/T 221–2005) reported in our previous work [6,30]. The MOC–solidified residual sludge prepared was crushed using a compression–testing machine (Jinan Dong Fang Test Instrument Co., Ltd., Jinan, China), and the collected specimen was ground to powder. The powdered sample was passed through a 100–mesh sieve; the filtered fine powder was characterized using SEM (JSM–7610FPlus, Jeol Japan) and XRD to observe the morphology of solidified sludge samples, and the crystal structures were analyzed using a D/max–rA model diffractometer (Bruker Co., Karlsruhe, Germany) with Cu K α radiation (λ = 1.54184 Å) operating at 40 kV and 40 mA.

3.1. Characterization of Solidified Blocks

The residual sludge used in this work comprised 87.75 wt% water, 7.74 wt% residues, and 4.54 wt% organic material. In dry surplus sludge, organic material content was 37.06 wt%, and the pH was 7.2.

3.1.1. SEM Images of Cemented Blocks

Figures 1 and 2 show the SEM images of MOC–solidified blocks at different R_n of MOC and $R_{W/S}$. Figure 1 shows that the needle–like and rod–like structures were formed in the MOC–solidified blocks. Sample M1 had thick rod–like structures (Figure 1a), which may be phase 5 or phase 3 [31,32]. With increase in R_n , there were more fine rod– or needle–like structures in samples M4 and M6. As shown in Figure 2, the MOC–solidified blocks contained the long rod–like structures, which were mainly formed when $R_{W/S}$ was 0.72 or 0.85.



Figure 1. SEM images of MOC–solidified blocks at different R_n when $R_{W/S}$ was 1.38. Panel (**a**) is sample M1 ($R_n = 2.0$); (**b**) is sample M4 ($R_n = 5.0$); (**c**) is sample M6 ($R_n = 7.0$).



Figure 2. SEM images of MOC–solidified blocks with different $R_{W/S}$ when R_n was 5.0: (**a**) is M8 ($R_{W/S} = 0.72$); (**c**) is sample M10 ($R_{W/S} = 0.85$); (**b**) and (**d**) are the enlarged images of (**a**) and (**c**), respectively.

SEM images of MOC–sludge solidified blocks with different R_n are shown in Figure 3. As reported in our previous work, the surplus sludge had a clumped granular structure [6]. Figure 3a–c illustrate the influence of R_n on microstructure of the MOC–residual sludge solidified blocks, showing rod–like structures encasing the granular sludge and the morphologies of rod–like structures has hardly changed with R_n increasing from 2.0 to 5.0, which indicated successful solidification of the residual sludge by MOC, and R_n only slightly influenced the structural morphology of MOC–sludge solidified block. While, the rod–like structures in Figure 3b is more compact. Figure 3b,d,e show the SEM images of MOC–surplus sludge solidified samples SM2, SM9, and sample SM12 with R_m values of 1.00, 0.67, and 1.50, respectively at $R_n = 3.0$ and $R_{W/S} = 1.38$. Structural morphologies of these samples did not change significantly with increasing R_m . Rod–like structures encased the granular sludge. Figure 3f shows the structural morphology of sample SM13. Compared with sample SM4, the rod–like structure encasing the granular sludge formed a net–like structure when $R_{W/S}$ decreased to 0.90. When $R_{W/S}$ was 1.50, an amorphous gel structure appeared in sample SM16 (Figure 4).



Figure 3. SEM images of MOC–residual sludge solidified samples. Panel (**a**) is sample SM1 ($R_n = 2.0$, $R_m = 1.00$, $R_{W/S} = 1.38$); (**b**) is sample SM2 ($R_n = 3.0$, $R_m = 1.00$, $R_{W/S} = 1.38$); (**c**) is sample SM4 ($R_n = 5.0$, $R_m = 1.00$, $R_{W/S} = 1.38$); (**d**) is sample SM9 ($R_n = 3.0$, $R_m = 0.67$, $R_{W/S} = 1.38$); (**e**) is sample SM12 ($R_n = 3.0$, $R_m = 1.50$, $R_{W/S} = 1.38$); (**f**) is SM13 ($R_n = 5.0$, $R_m = 1.00$, $R_{W/S} = 0.90$).



Figure 4. SEM images of MOC–residual sludge solidified samples of SM16 ($R_n = 5.0$, $R_m = 1.00$, $R_{W/S} = 1.50$); (**b**,**c**) are the enlarged images of (**a**).

Figure 5 shows the SEM images of MOC–residual sludge solidified blocks with Na₂SiO₃ or fly ash. Compared with sample SM13 (Figure 3f), Figure 5a,b show that samples NG1 and NG7 mainly comprised short and thick rod–like structures with relatively compact structures. Figure 5c,d show that samples NF5 and NF20 had a rod–like and thin rod–like and flaky structures, respectively.



Figure 5. SEM images of MOC–residual sludge solidified samples with Na₂SiO₃ or fly ash: (a) NG1 ($D_{\text{Na}_2\text{SiO}_3} = 1.0 \text{ wt\%}$); (b) NG7 ($D_{\text{Na}_2\text{SiO}_3} = 7.0 \text{ wt\%}$); (c) NF5 ($D_F = 5.0 \text{ wt\%}$); (d) NF20 ($D_F = 20.0 \text{ wt\%}$).

3.1.2. XRD Analysis of Cemented Blocks

The XRD spectra of MOC–solidified blocks are shown in Figure 6. The phase composition of MOC–cemented blocks mainly comprised phase 5, unreacted MgO, and Mg(OH)₂. A small amount of phase 3 was observed only in sample M1 ($R_n = 2.0$) and M8 ($R_{W/S} = 0.72$). In the XRD spectra, phase 3, phase 5, and Mg(OH)₂ were the typical hydration products [64]. Figure 7 shows XRD spectra of a blank MOC–solidified sample, sludge, and MOC–surplus sludge solidified blocks. Figure 7a shows that the sludge and MOC–sludge solidified samples contained SiO₂ which was from sludge, indicating that the surplus sludge was solidified by MOC. As R_n increased, the peak intensity of phase 5 weakened. Phase 5 could be observed in SM1 and SM2, as shown in Figure 7a. The peak of Mg(OH)₂ at $2\theta = 17.5^{\circ}$ was weakened at $R_m = 1.00$, $R_{W/S} = 1.38$, and $R_n = 3.0$ (Figure 7b). When $R_{W/S}$ increased to 1.50, the peaks of Mg(OH)₂, SiO₂, phase 5, and MgO weakened significantly and even disappeared (Figure 7b).



Figure 6. XRD spectra of blank MOC–solidified blocks at different R_n (**a**) and $R_{W/S}$ (**b**).



Figure 7. The XRD spectra of a blank MOC–solidified block (sample M4), sludge and MOC–solidified sludge samples (**a**); the XRD spectra of MOC–sludge solidified blocks with different R_m or $R_{W/S}$ (**b**).

3.2. Effects of R_n and Mass Ratio of $R_{W/S}$ on R_C of MOC–Cemented Blocks

Figure 8 shows the images of prepared MOC–solidified blocks for R_n varying from 2.0 to 8.0 at $R_{W/S} = 1.38$, and Figure 9 shows the photos of the MOC–solidified blocks with different values of $R_{W/S}$ at $R_n = 5.0$. All MOC–solidified blocks were white. Water droplets appeared on the surfaces of the MOC–solidified block when R_n was 7.0 at $R_{W/S} = 1.38$ (Figure S1 in Supporting Information). Cracking was observed on the surface of the MOC–solidified blocks when R_n was 8.0 at $R_{W/S} = 1.38$. At lower $R_{W/S}$ (<1.0), the MOC–solidified block was crunchy and easy to crack.



Figure 8. The photos of the MOC–solidified blocks with different R_n at $R_{W/S}$ = 1.38.



Figure 9. The photos of the MOC–solidified blocks with different $R_{W/S}$ at $R_n = 5.0$.

The effects of R_n of MOC and $R_{W/S}$ on R_C of the MOC–residual sludge solidified blocks were investigated first as blank experiments. Figure 10a shows the effect of R_n of MgO to MgCl₂ on R_C of MOC–solidified blocks when $R_{W/S}$ was 1.38. The R_C of MOC– solidified blocks reached the equilibrium strength after 14–day curing when R_n was >2.0. The maximum R_C of 10.20 MPa was reached after 60–day curing when R_n was 2.0. With R_n increasing, the R_C of MOC–solidified blocks decreased generally. That may be because some of phase 3 and phase 5 were formed in M1 ($R_n = 2.0$); the number of lumpy Mg(OH)₂ peaks increased and the number of phase 5 peaks decreased when R_n increased (Figure 6a). Needle– or long–rod–like crystals were observed for phase 3 and phase 5, and the lumpy Mg(OH)₂, respectively (Figure 1). The maximum formation of phase 5 crystals provides the most mechanical strength to the samples [75]. Therefore, the formation of phases 3 and 5 helped increase strength.



Figure 10. The effect of R_n on R_C of MOC–cemented blocks when $R_{W/S}$ was 1.38 (**a**); effect of $R_{W/S}$ on R_C of MOC–cemented blocks when R_n was 5.0 (**b**).

Figure 10b shows the influence of $R_{W/S}$ on R_C of MOC–solidified blocks at $R_n = 5.0$. The value of R_C of MOC first increased when $R_{W/S}$ increased from 0.72 to 0.85, and then it decreased when $R_{W/S}$ increased from 0.85 to 1.50. The maximum R_C reached the equilibrium value of 42.50 MPa at approximately 10 days when $R_{W/S}$ was 0.85. The number of phase 5 peaks decreased when $R_{W/S}$ increased from 0.85 (M10) to 1.38 (M4), resulting in R_C decreasing, as demonstrated in Figure 6b, due to that phase 5 was the main strength phase for MOC–solidified blocks. M8 ($R_{W/S} = 0.72$) contained phase 3 and 5, and the number of Mg(OH)₂ peaks was lower than that in M4 and sample M10. SEM images show needle– or long–rod–like phase 3 and 5 in M8 and M10 (Figure 2). This led to the highest R_C when $R_{W/S}$ was 0.85 (sample M10).

3.3. MOC-Solidified Residual Sludge

Figure 11 shows images of MOC–residual sludge solidified blocks with different values of R_n ($R_{W/S} = 1.38$, $R_m = 1.00$) and R_m ($R_{W/S} = 1.38$, $R_n = 3.0$) after 60–day curing. Compared with MOC–solidified blocks, the color of MOC–residual sludge solidified blocks changed to

light black from dark black with R_n increasing when $R_{W/S}$ was 1.38 and R_m was 1.00. This was because the weight percentage of white MgO (powder) increased from 17.20 wt% to 28.03 wt% with R_n increasing from 2.0 to 7.0. Furthermore, no water droplets appeared on the surfaces of the MOC–sludge solidified specimens. The hygroscopic solidified surplus sludge could imbibe the superfluous water droplets during solidification, which limited the migration of aqueous solution to the surface [76]. No frosting phenomenon was observed on the specimens' surfaces when R_n and R_m increased within the experimental range. Figure 12 shows images of the MOC–residual sludge solidified blocks with different values of $R_{W/S}$ when R_m was 1.00 and R_n was 5.0 after 60–day curing. No frosting phenomenon appeared on the specimens' surfaces with R_n , R_m , and $R_{W/S}$ increasing within the experimental range. The volumes of the solidified blocks were not influenced by R_n , R_m , and $R_{W/S}$.



Figure 11. Images of MOC–residual sludge solidified blocks with different R_n when $R_{W/S}$ was 1.38 and R_m was 1.00 and different R_m when $R_{W/S}$ was 1.38 and R_n was 3.0.



Figure 12. Images of MOC–residual sludge solidified blocks with different $R_{W/S}$ when R_m was 1.00 and R_n was 5.0.

3.3.1. Influence of R_n on R_C of MOC–Solidified Residual Sludge

Figure 13 illustrates the influence of R_n on the R_C of the MOC–residual sludge solidified blocks when $R_{W/S}$ was 1.38 and R_m of residual sludge to MOC was 1.00. The R_C of MOC–solidified residual sludge first increased and then decreased as R_n increased. The maximum R_C of 11.65 MPa was obtained at 60–day curing at $R_n = 3.0$. This was because the main phases of phase 3 and 5 at room temperature (~25 °C) were formed in the hydration process of MOC–residual sludge mixed mortar [77], and were the main sources of strength, as observed in SEM images in Figure 3 and demonstrated in XRD spectra shown in Figure 7a. Due to the morphology of rod–like structures was affected by R_n , the more compact rod–like structure maybe cause the maximum R_C of 60 days obtained at $R_n = 3.0$ (Figure 3a–c). Compared with the blank MOC–solidified sample M2 ($R_n = 3.0, R_{W/S} = 1.38$),



Figure 13. Influence of R_n on the R_C of MOC–residual sludge solidified blocks at $R_{W/S} = 1.38$ and $R_m = 1.00$.

3.3.2. Influence of R_m on R_C of MOC–Residual Sludge Solidified Blocks

Figure 14 shows the influence of R_m on the R_C of MOC–residual sludge solidified blocks when $R_{W/S}$ was 1.38 and R_n was 3.0. Thus, the R_C of 60–day curing first increased and then decreased as R_m increased. The maximum R_C of 11.65 MPa was obtained at $R_m = 1.00$, which was approximately six times than that (~2.0 MPa) of MOC sample M2. As shown in Figure 3b,d,e, the structural morphologies of these samples did not change significantly with R_m increasing; the rod–like structures in sample SM2 ($R_m = 1.00$, Figure 3b) were more compact compared with those in SM9 ($R_m = 0.67$) and SM12 ($R_m = 1.50$). Formation of compact rod–like structures may be the reason for higher R_C at $R_m = 1.00$.



Figure 14. Influence of R_m on the R_C of MOC–solidified residual sludge when $R_{W/S}$ was 1.38 and R_n was 3.0.

3.3.3. Influence of R_{W/S} on R_C of MOC–Residual Sludge Solidified Blocks

Figure 15 shows the influence of $R_{W/S}$ on the R_C of MOC–residual sludge solidified blocks when R_n was 5.0 and R_m was 1.00. Equilibrium R_C was reached at approximately 10–day curing except at $R_{W/S} = 1.38$. R_C of 60 days decreased when $R_{W/S}$ increased from 0.90 to 1.50 because of the exceeding demand for water dosage. The maximum R_C of 20.90 MPa was obtained at $R_{W/S} = 0.90$. When $R_{W/S}$ was 1.50, water dosage exceeded 91.88% of the demand for phase 5 formation. Therefore, some unstable phase 5 may have been converted to Mg(OH)₂ and soluble ions when water dosage was more than the demand, resulting in strength reduction. This is demonstrated by SEM micrographs and XRD spectra in Figures 4 and 7b. An amorphous gel structure Mg(OH)₂ appeared in sample SM16 (Figure 4). The XRD spectra of MOC–residual sludge solidified blocks with different $R_{W/S}$ values show that SM13 ($R_{W/S} = 0.90$) and SM4 ($R_{W/S} = 1.38$) mainly contained phase 5 and Mg(OH)₂. The intensity of peaks for strength phases (phases 3 and 5) weakened in SM16 ($R_{W/S} = 1.50$), as shown in Figure 7b.



Figure 15. Influence of $R_{W/S}$ on the R_C of MOC–solidified residual sludge at $R_n = 5.0$ and $R_m = 1.00$.

3.4. Influence of Na₂SiO₃ and Fly Ash on R_C of MOC–Solidified Residual Sludge

Figure 16 shows images of MOC–residual sludge solidified blocks with different $D_{\text{Na}_2\text{SiO}_3}$ and D_{F} when $R_{\text{W/S}}$ was 0.90, at R_{n} = 5.0 and R_{m} = 1.00 after curing for 28 days. No water droplets and frosting phenomenon appeared on the surface of the MOC–sludge solidified specimens. The volumes of the solidified blocks were not influenced by $D_{\text{Na}_2\text{SiO}_3}$ and D_{F} .



Figure 16. Images of MOC-residual sludge solidified blocks with different $D_{\text{Na}_2\text{SiO}_3}$ and D_{F} .

Figure 17a shows the influence of $D_{Na_2SiO_3}$ on the R_C of MOC–residual sludge solidified blocks at $R_{W/S} = 0.90$, $R_n = 5.0$, and $R_m = 1.00$. R_C increased as $D_{Na_2SiO_3}$ increased and the maximum R_C of 35.60 MPa was obtained when $D_{Na_2SiO_3}$ was 7.0 wt%. The R_C was higher when Na₂SiO₃ was added than when there was no Na₂SiO₃, indicating that adding Na₂SiO₃ improved R_C , which was reported in our previous studies [6,78,79]. That may be because the Na₂SiO₃ addition accelerated the reaction rate and improved the yield of



short and thick rod–like phase 5 (Figure 5a,b). The short and thick rod-like structure also improved the compressive strength [72].

Figure 17. At $R_{W/S} = 0.90$, $R_n = 5.0$, and $R_m = 1.00$, (**a**) is the effect of $D_{Na_2SiO_3}$ on R_C of MOC–residual sludge solidified blocks; (**b**) is the effect of D_F on R_C of MOC–residual sludge solidified blocks.

Figure 17b shows the influence of D_F on the R_C of MOC–residual sludge solidified blocks at $R_{W/S} = 0.90$, $R_n = 5.0$, and $R_m = 1.00$. At curing for 28 days, R_C increased initially and then decreased as D_F increased. The maximum R_C of 28.40 MPa was obtained at $D_F = 5.0$ wt%. As reported in previous studies [67,80], the SiO₂ and Al₂O₃ in fly ash could react with Mg(OH)₂ to generate a silica–aluminum gel system or magnesium aluminum silicate hydrate during hydration; also, fly ash filled into the MOC hydration products in the form of particles or gel, which decreased the pore space and improved structure density, resulting in strength improvement [81]. The addition of fly ash delayed the curing time compared with that in MOC–sludge solidified blocks without fly ash overview, which is consistent with the literature [82]. When D_F increased to 20.0 wt%, the R_C of 28 days decreased to 19.40 MPa, which was lower than that of MOC–residual sludge solidified blocks without fly ash. The reason may be that the dosage of MOC decreased with increasing D_F .

3.5. Water Resistance Test of Solidified Blocks in Different Aqueous Solutions

Figure 18 shows the strength curves of MOC–residual sludge solidified blocks after immersion in aqueous solutions with different pH values. NF5 and SM13 exhibited the best water resistance in aqueous solutions with pH = 3.0–7.0 and pH = 9.0. The corrosion resistance coefficient (K_F) for SM13 and NF5 were above 80.0% after immersion for 7 and 14 days in aqueous solutions with pH = 7.0–11.0, indicating their good alkaline resistance. For the three SM13, NG7, and NF5 samples, K_F was higher in neutral or alkaline aqueous solutions than in acidic aqueous solutions. Compared with Figure 18a, the value of R_{AC} for samples SM13, NG7, and NF5 decreased after immersion in an aqueous solution for 7 and 14 days. K_F was approximately 97.01% and 94.53% when SM13 was immersed in an aqueous solution with pH = 9.0 for 7 and 14 days (Figure 18b). For NG7 (Figure 18c), the K_F was about 83.24% and 80.11% after immersion in an aqueous solution with pH = 11.0 for 7 and 14 days. For NF5 (Figure 18d), the K_F was above 90.0% after immersion in an aqueous solution of pH = 7.0–9.0 for 7 and 14 days. Furthermore, the K_F of 7–day immersion was higher than that for 14–day immersion, indicating that the acid–base resistance property of MOC–residual sludge solidified blocks decreased with increasing immersion time.



Figure 18. The strength curves of MOC–residual sludge solidified blocks after immersion in an aqueous solution with different pH values. Panel (**a**) is the $R_{\rm C}$ of 28 days as blank experiment; (**b–d**) are the $R_{\rm AC}$ and $K_{\rm F}$ of SM13, NG7, and NF5.

4. Conclusions

To investigate the potential application of residual sludge as construction materials, low-carbon and environmentally friendly MOC was used to solidify residual sludge. The effects of R_{W/S}, R_n, R_m, curing time, D_{Na2SiO3}, and D_F on the R_C of MOC-residual sludge solidified blocks were investigated. As determined by SEM micrographs and XRD spectra, the solidified blocks mainly comprised rod-like structure phase 5, unreacted MgO, and Mg(OH)₂. The SEM images show that the cemented blocks did not shrink or dilate much, and no frosting phenomenon appeared on the specimen surface. The maximum values of $R_{\rm C}$ of 10.20 MPa and 42.50 MPa for MOC-solidified blocks were obtained when $R_{\rm n}$ and $R_{W/S}$ were 2.0 and 1.38 and 5.0 and 0.85, respectively. The R_C of MOC–residual sludge solidified blocks increased initially and then decreased as R_n and R_m increased. The maximum $R_{\rm C}$ of 11.65 MPa was reached when $R_{\rm n}$ was 3.0, $R_{\rm W/S}$ was 1.38, and $R_{\rm m}$ was 1.00. The value of $R_{\rm C}$ decreased with $R_{\rm W/S}$ increasing from 0.90 to 1.50. Adding Na₂SiO₃ or fly ash could improve the $R_{\rm C}$ of MOC–residual sludge solidified blocks. With $D_{\rm NaSiO3}$ increasing, the R_C was higher than that of MOC–residual sludge solidified blocks without Na₂SiO₃. With $D_{\rm F}$ increasing, $R_{\rm C}$ increased initially and then decreased. The maximum $R_{\rm C}$ of 35.60 MPa and 28.40 MPa were reached at $D_{\text{NaSiO3}} = 7.0 \text{ wt\%}$ and $D_{\text{F}} = 5.0 \text{ wt\%}$. Waterresistance tests of solidified blocks showed that NF5 exhibited the best water resistance in the aqueous solution with pH = 7.0–9.0. For SM13, NG7, and NF5, water resistance was better in neutral or alkaline aqueous solutions than in acidic aqueous solutions. The waterresistance properties of MOC-residual sludge solidified blocks decreased with increased immersion time. The study found that fly ash could help improve water resistance of MOCsolidified residual sludge in neutral and basic aqueous solutions; Na₂SiO₃ was unfavorable for improving the water resistance. This work provides a new perspective for the efficient disposal and comprehensive utilization of residual sludge. Although the mechanical properties satisfied the application as a building material, the stability of the contaminants, such as metal ions, in sludge still needs to be investigated further to make sure that environmental safety and the acid resistance are improved before practical application.

Supplementary Materials: The following supporting information can be downloaded at: https://www.mdpi.com/article/10.3390/pr11020413/s1, Figure S1: photos of the MOC solidified blocks when the R_n was 7.0 at $R_{W/S}$ was 1.38.

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