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Ozone and Temperature May Hinder Adaptive Capacity of Mediterranean Perennial Grasses to Future Global Change Scenarios

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Abstract: Climate warming is recognized as a factor that threatens plant species in Mediterranean mountains. Tropospheric ozone (O₃) should also be considered as another relevant stress factor for these ecosystems since current levels chronically exceed thresholds for plant protection in these areas. The main aim of the present study was to study the sensitivity of four Mediterranean perennial grasses to O₃ and temperature based on plant growth, gas exchange parameters (photosynthesis—A, stomatal conductance—g_s, and water use efficiency—WUE), and foliar macro- (N, K, Ca, Mg, P, and S) and micronutrients (B, Cu, Fe, Mn, Mo, and Zn) content. The selected species were grasses inhabiting different Mediterranean habitats from mountain-top to semi-arid grasslands. Plants were exposed to four O₃ treatments in Open-Top chambers, ranging from preindustrial to above ambient levels, representing predicted future levels. Chamber-less plots were considered to study the effect of temperature increase. Despite the general tolerance of the grasses to O₃ and temperature in terms of biomass growth, WUE and foliar nutrient composition were the most affected parameters. The grass species studied showed some degree of similarity in their response to temperature, more related with phylogeny than to their tolerance to drought. In some species, O₃ or temperature stress resulted in low A or WUE, which can potentially hinder plant tolerance to climate change. The relationship between O₃ and temperature effects on foliar nutrient composition and plant responses in terms of vegetative growth, A, g_s, and WUE constitute a complex web of interactions that merits further study. In conclusion, both O₃ and temperature might be modifying the adaptation capacity of Mediterranean perennial grass species to the global change. Air pollution should be considered among the driving favors of biodiversity changes in Mediterranean grassland habitats.

Keywords: *Agrostis castellana*; *Festuca iberica*; *Festuca indigesta* subsp. *curvifolia*; *Stipa tenacissima*; gas exchange; plant growth; plant nutrients; O₃-sensitivity; climate warming



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1. Introduction

Tropospheric ozone (O₃) is one of the most relevant air contaminants due to its high toxicity, wide distribution, and its greenhouse effect [1–4]. Transportation and industrial activity are the main anthropogenic sources of O₃ precursors. However, these pollutants are transported by air masses and can react with precursor emissions from natural sources causing large O₃ surface levels in rural and forested areas far away from anthropic activities [5,6].

The photochemical formation and persistence of O₃ is favored under Mediterranean climatic conditions, causing an important air pollution problem in the Mediterranean Basin [4,6–8]. The current O₃ levels in the Iberian Peninsula chronically exceed the threshold levels for plant protection established by the UNECE Air Convention [9] and the EU Air Quality Directive (2008/50/EC) [10]. This problem is even more acute in the Mediterranean mountains [11,12].

Normal plant development is greatly impaired due to ozone [13–15]. Its oxidative capacity affects plants at different scales, from metabolic to ecosystem, causing effects on growth and biomass production [16,17], nutrient imbalances [18–20], or reproductive capacity [3,21], and finally affecting the structure and biodiversity of plant communities [22,23].

The higher levels of this pollutant in the Mediterranean area contrast with the fact that the Mediterranean Basin is considered a global biodiversity hotspot [24–26]. At the Spanish Central System mountain range, the spring–summer snow-free zones above the tree line harbor high levels of taxonomic diversity [27–29], including plant communities where grasses are dominant. Furthermore, in the Mediterranean Basin, grasses dominate large arid or semiarid areas [30].

The experimental evidence on the O₃ sensitivity of Mediterranean grassland species varies greatly between community types. Mediterranean annual pasture species growing at lower altitudes have been the most extensively explored. In terms of biomass growth, Fabaceae species are generally more sensitive to O₃ than Poaceae [31,32] but tolerant legume species have also been identified [33]. Annual grasses also responded to O₃ by increasing leaf senescence or decreasing their life-span [34,35]. Regarding perennial grass species, most studies have been focused on species from more temperate climates [36–38], with less attention to species characteristics of the Mediterranean area [39]. Alpine and mid-elevation perennial temperate grass species have shown a wide variety of O₃ responses, from tolerant to carry over O₃-effects on above-ground biomass [38,40–42]. However, there has been no study addressing the O₃ effects on Mediterranean perennial alpine species.

Plant communities at the top of Spanish Central System Mountains are very valuable for their high biodiversity, with a great number of endemic and relict species, and for being the southern limit of other more widely distributed species. In addition, within these communities, there are grass species which are of great economic interest in the region, since they are the basis for livestock feeding, especially during the summer season [43,44]. These summits are characterized by poor acid soils covered by snow during winter and with a high soil moisture deficit during the summer season, inhabited by the xerophytic grass species *Festuca indigesta* subsp. *curvifolia* (hereafter *F. indigesta*) [44]. The *Cervunal*, a neighboring community dominated by *Nardus stricta*, occupies patches with higher soil humidity [44]; *Festuca iberica* and *Agrostis castellana* are two other grass species characteristic of the *Cervunal* habitat. There is another interesting grassland community dominated by the esparto grass *Stipa tenacissima*, that in the Central Iberian Peninsula lives at a lower elevation but it can be found at 1400 m.a.s.l. in the southern mountain ranges. This “esparto” grass lives in drier and warmer habitats, facilitates the biological soil crusts and vascular plant establishment in semiarid habitats, and it is an important species for the local economy due to its use as raw material for the manufacture of clothing, footwear, basketry, or paper pulp production [45,46].

In this study, the O₃ sensitivity of *A. castellana*, *S. tenacissima*, *F. iberica*, and *F. indigesta* was tested in an O₃ fumigation experiment using Open Top Chambers (OTCs). The four species, belonging to three different tribes or subtribes within the *Pooideae* subfamily [47], together with their different habitats and water requirements, make them a good representation for testing the tropospheric O₃-response of Mediterranean perennial grasses.

The main aim of the present study was to assay the O₃-effect on Mediterranean perennial grasses according to plant growth, gas exchange (photosynthesis (A), stomatal conductance (g_s), and water use efficiency (WUE)), and leaf composition considering macro (N, K, Ca, Mg, P, and S) and micronutrients (B, Cu, Fe, Mn, Mo, and Zn). Based on previous results on the O₃-response of Mediterranean annual grasses growing at lower elevations, we hypothesize that Mediterranean perennial grasses would present some tolerance to the pollutant based on growth parameters. However, we expected more subtle effects at a physiological scale on gas exchange parameters and foliar nutrients that might challenge their competitive capacity and compromise their tolerance to other stresses frequent in their habitat.

Taking advantage of the OTC experimental design, the effects of temperature increase were also tested at the same time as the O₃ effect following [48]. Climate warming is one of the most important threats for plant species from the Mediterranean basin mountains, including the Spanish Central System mountain range [49]. We expect that species adapted to drier environments (*S. tenacissima* and *F. indigesta*) will be more tolerant to both, O₃ and temperature increase, than species from moister areas. Specifically, we aimed to answer the following questions: (i) can we find any indicator of O₃ effects on Mediterranean perennial grasses considering plant growth, leaf nutrients, or gas exchange?; (ii) might any of the O₃ and temperature sensitive parameters compromise the long-term survival of the species under the global change scenarios for the Mediterranean Basin?; (iii) is water stress tolerance linked to O₃ tolerance or temperature resistance?; and (iv) is the O₃ sensitivity of species more related to water deficit tolerance or to their phylogenetic signature?

2. Results

2.1. Ozone Exposure and Temperature Increase

Accordingly with the experimental design, the lowest O₃ concentration (nL L⁻¹) was found in the CFA treatment (Table 1), with an O₃ filtration efficiency of 47.5% compared with NFA. Non-filtered air OTCs and AMB plots showed comparable O₃ levels throughout the fumigation period, the pollutant levels inside the OTC being only 5% lower than the AMB plots. Plants under the NFA++ treatment were the most exposed to the pollutant, with average O₃ concentrations 37.5% higher than NFA levels, which caused the highest AOT40 values close to 15,000 nL L⁻¹ h, above 10,000 nL L⁻¹ h higher than AMB and NFA. NFA+ O₃ concentration values fell in between NFA and NFA++ treatments, with cumulative AOT40 values in the range of 8500 nL L⁻¹ h. During the daylight hours from 7:00 to 15:00 (GMT), OTCs compared with AMB plots presented an increase in the mean air T by 20.6% which implied an increment of around 5 °C inside the OTCs, also a reduction in RH by 4% was found; both caused an increase in the water Vapor Pressure Deficit (VPD) by 30.5% (Table 2).

Table 1. Ozone exposure during the whole experimental period for the different grasses: accumulated O₃ exposure index AOT40 and O₃ 8h-mean. DaS, days after the start of the O₃ exposure.

	DaS	AOT40 (nL L ⁻¹ h)					O ₃ 8h-Mean (nL L ⁻¹)				
		AMB	CFA	NFA	NFA+	NFA++	AMB	CFA	NFA	NFA+	NFA++
<i>S. tenacissima</i>	68	5780	189	4285	9340	16,677	43.4	21.1	40.7	50.8	63.3
<i>A. castellana</i>	63	4622	187	3399	8466	15,807	42.3	20.9	39.8	50.7	64.1
<i>F. iberica</i>	62	4456	187	3286	8349	15,692	42.2	20.8	39.7	50.8	64.4
<i>F. indigesta</i>	57	3964	187	2951	8072	15,321	41.9	20.9	39.8	51.6	66.2
Mean	63	4705	187	3480	8557	15,875	42.5	20.9	40.0	51.0	64.5

Table 2. Mean meteorological conditions during the experimental assay for the 8-h daylight period 7:00–15:00 (GMT) inside the OTCs and in the AMB plots. Mean and maximum air temperature (mean T, max T), air relative humidity (RH), photosynthetic active radiation (PAR) and vapor pressure deficit (VPD, kPa).

	Mean T (°C)	Max T (°C)	RH (%)	PAR (μmol m ⁻² s ⁻¹)	VDP (kPa)
AMB	20.4	37.2	49.4	1149	1.21
OTCs	25.7	43.2	47.4	991	1.74

2.2. Vegetative Growth

The vegetative growth of all the species was relatively O₃ tolerant (Figure 1A). Table S3 shows the lineal effect (LE), quadratic effect (QE), and cubic effect (CE) contrasts for each species. A marginally CE of O₃ was found in *F. indigesta* ($t_6 = 2.22$, $p = 0.07$) where NFA and NFA++ had the highest growth. However, in *A. castellana* ($t_{66} = -1.18$, $p = 0.24$ for

a QE), *S. tenacissima* ($t_{64} = -1.12$, $p = 0.27$ for a LE), and *F. iberica* ($t_6 = 0.37$, $p = 0.48$ for a QE) no significant O_3 effect was detected. The temperature increase did not affect vegetative growth for either *F. indigesta* ($F_{1,2} = 0.11$; $p = 0.77$), *F. iberica* ($F_{1,2} = 0.01$; $p = 0.94$), or *S. tenacissima* ($F_{1,32} = 0.1$; $p = 0.76$) (Figure 1B, Table S4). However, the vegetative growth was marginally higher in NFA than AMB in *A. castellana* ($F_{1,32} = 3.95$; $p = 0.055$). The structure of the random part of the model for each variable and species is presented in the Table S5.

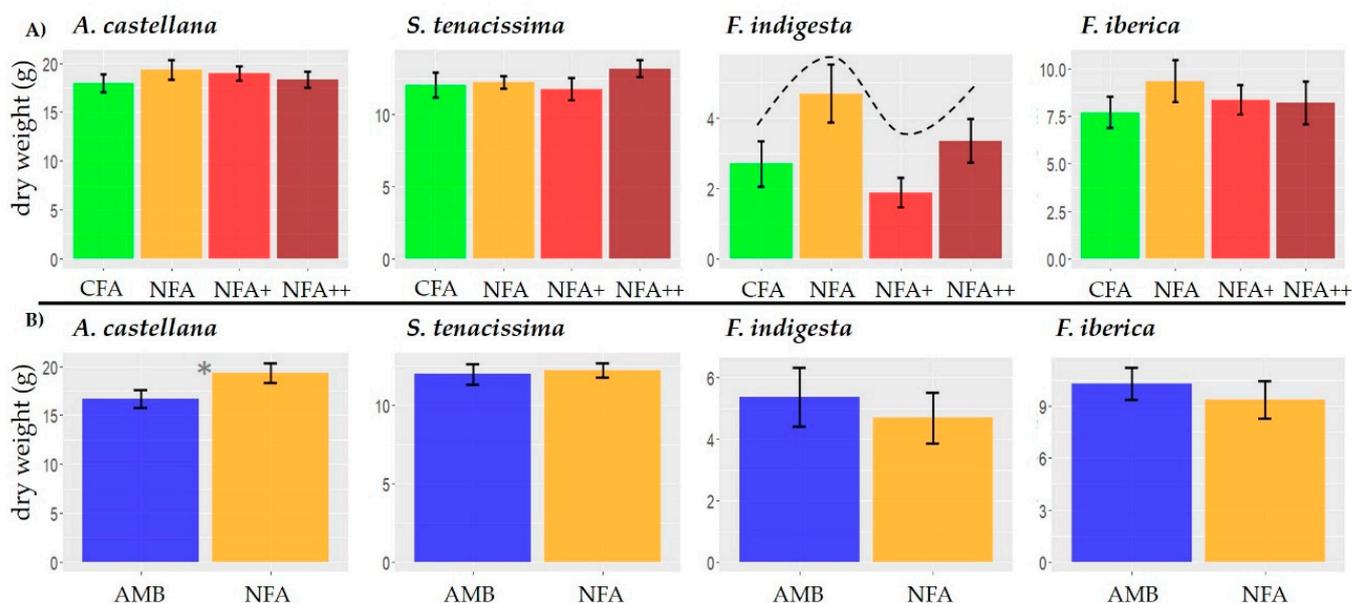


Figure 1. Vegetative aboveground growth response to O_3 (A) and temperature increase (B) for *A. castellana*, *S. tenacissima*, *F. indigesta*, and *F. iberica*. Bars denote mean \pm SE. Dotted lines denote marginally significant trends. Grey asterisk denotes marginally significant differences.

2.3. Gas Exchange

Table S3 shows the lineal effect (LE), quadratic effect (QE), and cubic effect (CE) contrast for each species. Considering the O_3 effects on net photosynthesis (Figure 2A), a marginally significant LE was observed in *A. castellana* ($t_{20} = 1.93$, $p = 0.07$). However, *F. indigesta* presented a marginally significant CE with maximum values under NFA+ ($t_{16} = -2.03$, $p = 0.06$); a similar response was found in *F. iberica* although it was not significant ($t_{26} = -1.29$, $p = 0.21$). The observed response of A in *S. tenacissima* was neither significant ($t_{30} = 1.61$, $p = 0.12$ for a QE).

Regarding g_s (Figure 2B), three of the four species presented a similar concave trend in response to O_3 , with a greater stomatal opening under the lowest and highest O_3 concentrations (CFA and NFA++ treatments). In *A. castellana*, NFA and NFA+ treatments were marginally lower than CFA and NFA++ ($t_{20} = 1.79$, $p = 0.09$); but even though *S. tenacissima* and *F. indigesta* showed the same concave pattern, they were not significant ($t_{30} = 1.42$, $p = 0.17$; $t_{16} = 1.52$, $p = 0.15$, respectively, for a QE). Contrary to conductance response in the other species, a marginally CE was observed in *F. iberica* ($t_{26} = -1.75$, $p = 0.09$).

Consistently with g_s values, the most observed WUE response pattern to the increased O_3 levels was a convex QE (Figure 2C). In *A. castellana* WUE was higher in NFA and NFA+ than in CFA and NFA++ (QE $t_{20} = -2.22$, $p = 0.04$); a similar effect was observed in *F. indigesta* ($t_{16} = -2.37$, $p = 0.03$). Although in *S. tenacissima* the observed trend approached a convex response, the CE obtained the lowest p -value and was not significant ($t_{30} = -0.97$, $p = 0.34$). As happened with g_s , the WUE behavior in *F. iberica* was different from the rest of species with a significantly increasing LE ($t_{26} = 2.26$, $p = 0.03$).

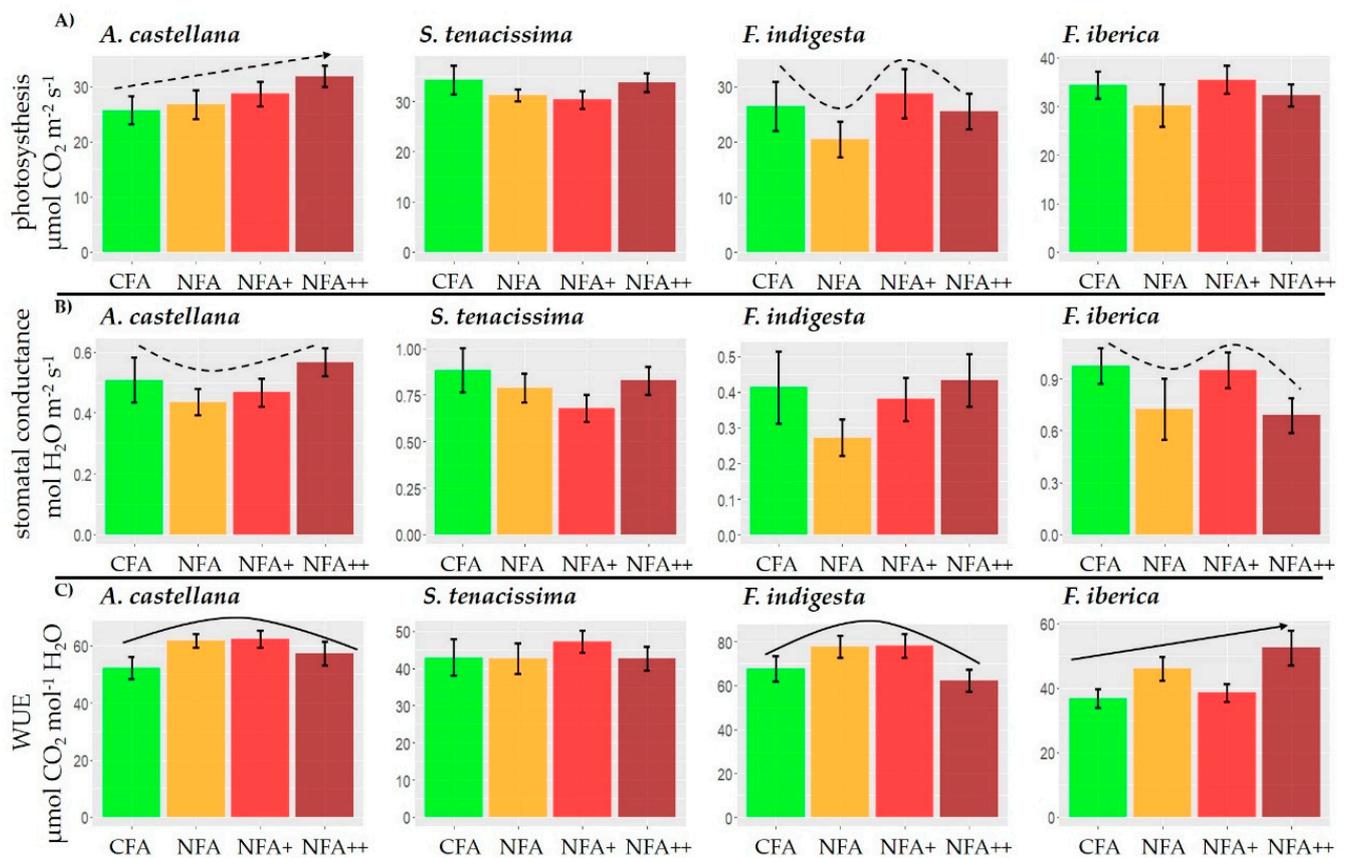


Figure 2. Ozone effect on gas exchange parameters: photosynthesis (A, $\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$) (A), stomatal conductance (g_s , $\text{mol H}_2\text{O m}^{-2} \text{ s}^{-1}$) (B) and water use efficiency (WUE, $\mu\text{mol CO}_2 \text{ mol}^{-1} \text{ H}_2\text{O}$) (C) for *A. castellana*, *S. tenacissima*, *F. indigesta*, and *F. iberica*. Bars denote mean \pm SE. O₃ treatments: charcoal-filtered air (CFA), non-filtered air (NFA), non-filtered air + 20nL L⁻¹ of O₃ (NFA+), and non-filtered air + 40nL L⁻¹ of O₃ (NFA++) treatments. Solid and dotted lines denote significant and marginally significant trends, respectively.

The temperature increase caused different effects on the gas exchange parameters depending on the species (Figure 3, Table S4). Photosynthetic activity was reduced in *S. tenacissima* plants grown inside the NFA OTCs compared with the AMB plots ($F_{1,7} = 7.24$, $p = 0.02$), but no temperature effect was detected in *F. iberica* ($F_{1,7} = 1.73$, $p = 0.23$) and *A. castellana* ($F_{1,9} = 0.53$, $p = 0.48$). Regarding g_s , there was no difference between AMB and NFA in *F. iberica* ($F_{1,7} = 1.75$, $p = 0.23$) and *S. tenacissima* ($F_{1,14} = 0.17$, $p = 0.69$). However, in *A. castellana* g_s was marginally higher under the lower temperatures of the AMB plots ($F_{1,9} = 3.59$, $p = 0.09$). On the contrary, in the latter species, WUE was larger in NFA under the higher temperature values than in the AMB plots ($F_{1,9} = 5.65$, $p = 0.04$). No differences due to temperature were found for WUE in *F. iberica* ($F_{1,7} = 0.75$, $p = 0.41$) and *S. tenacissima* ($F_{1,14} = 0.49$, $p = 0.49$). The structure of the random part of the model for each variable and species is presented in the Table S5.

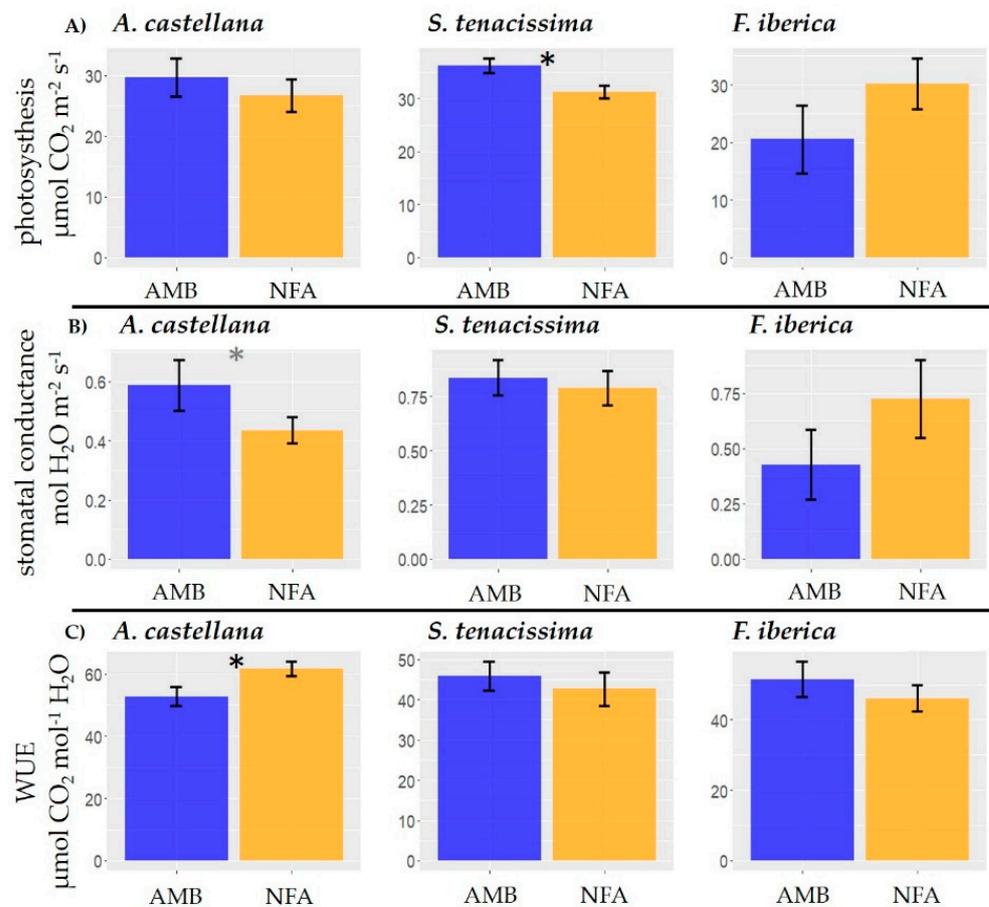


Figure 3. Temperature effects on gas exchange parameters: photosynthesis (A, $\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$) (A), stomatal conductance (g_s , $\text{mol H}_2\text{O m}^{-2} \text{ s}^{-1}$) (B) and water use efficiency (WUE, $\mu\text{mol CO}_2 \text{ mol}^{-1} \text{ H}_2\text{O}$) (C) for *A. castellana*, *S. tenacissima*, and *F. iberica*. Bars denote mean \pm SE. Temperature treatments: low temperature (AMB plots), high temperature (NFA OTC). Black and grey asterisk denote significant and marginally significant differences, respectively.

2.4. Nutrients and Nutrient Ratios

Nutrients (including C and H content) and nutrient ratios were differentially spatially distributed among the treatments following MNDS (Non-metric multidimensional scaling). MNDS results are presented for *A. castellana* and *S. tenacissima* in Figure 4, and for both *Festuca* species, *F. indigesta* and *F. iberica*, in Figure 5. The statistical differences among treatments that are shown below are based in PERMANOVA analyses (Table S6). Regarding foliar nutrient content, *A. castellana* plants grown under the lowest O_3 exposure, the CFA treatment, had more Mn and P than plants grown under the highest O_3 exposure of the NFA++ ($F_{3,19} = 2.46$, $p = 0.02$). In *S. tenacissima*, the concentration of S, Ca, Mg, Mn, B, and Zn was higher in CFA than NFA ($F_{3,19} = 2.34$, $p = 0.03$). PERMANOVA analysis also found significant differences among treatments in *F. iberica* ($F_{3,12} = 1.86$, $p = 0.048$), with NFA having lower Fe, Mg, and Mo values than CFA. Although NFA+ was also significantly different from NFA, the dispersion of the three NFA+ points calls for caution in the interpretation of these results. However, there were no differences among treatments in *F. indigesta* for foliar nutrient concentration ($F_{3,14} = 0.81$, $p = 0.64$).

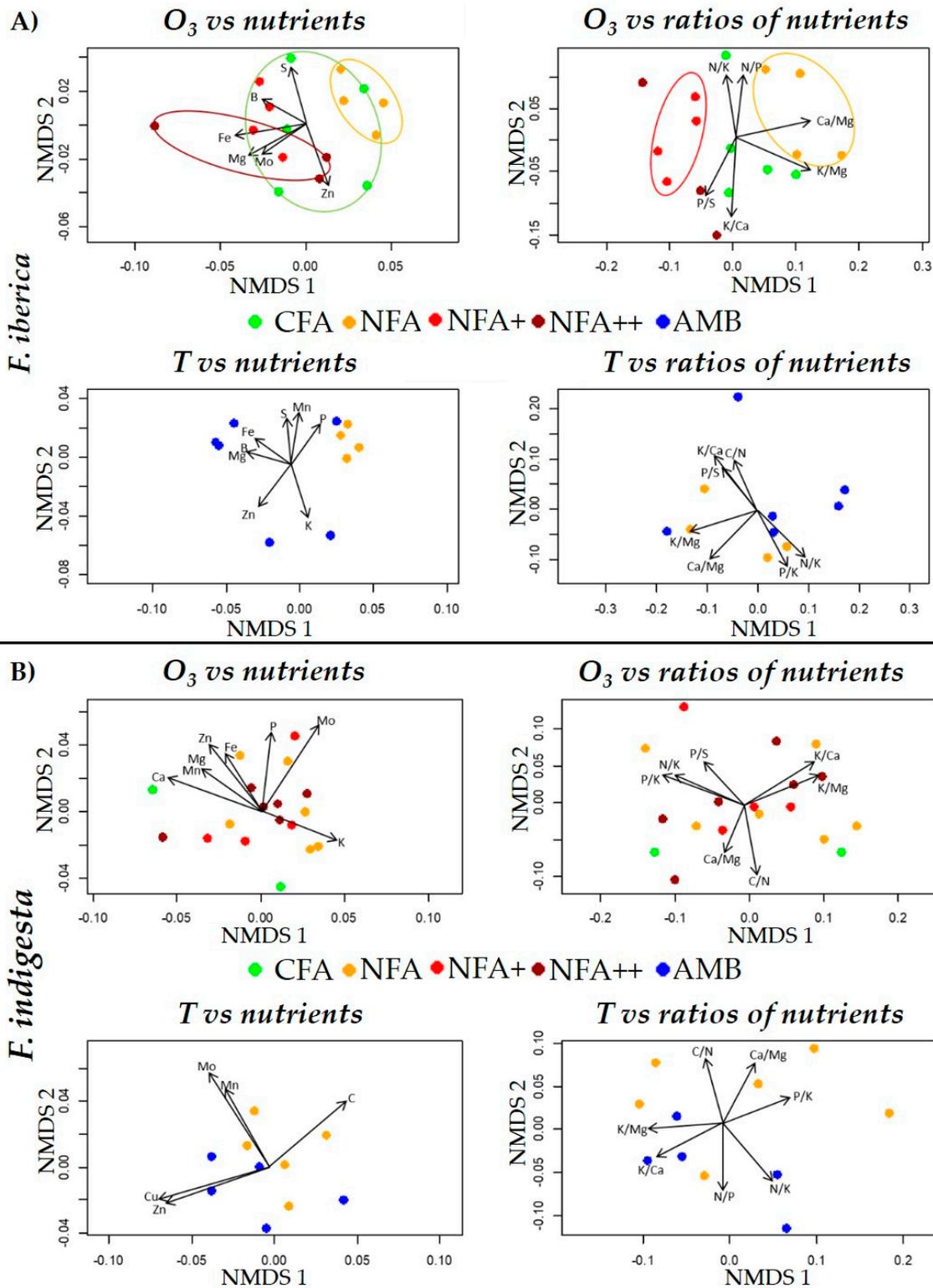


Figure 5. Non-metric multidimensional scaling (NMDS) of nutrients (left) and nutrient ratios (right) found in *F. iberica* (A) and *F. indigesta* (B) in relation to ozone and temperature (T) factors. The presence of colored ellipse represents significant differences among treatments following PERMANOVA results. Arrows denote significant correlations of nutrient or nutrient ratios with the NMDS axis.

P/K and P/S nutrient ratios were different in CFA compared with the rest of the O_3 treatments in *A. castellana* ($F_{3,19} = 3.02, p = 0.01$). In *S. tenacissima*, only CFA and NFA were marginally different ($F_{3,19} = 2.06, p = 0.07$) due to different P/S, N/K, P/K, Ca/Mg, C/N,

K/Ca, and K/Mg ratios. NFA had lower Ca/Mg and K/Mg ratios than NFA+ in *F. iberica* ($F_{3,12} = 2.8, p = 0.01$). There were no differences among O_3 treatments in *F. indigesta* due to nutrient ratios ($F_{3,14} = 0.7, p = 0.68$).

The lower temperature of AMB compared with NFA resulted in a higher P, Ca, Mg, and Fe content for *A. castellana* ($F_{1,9} = 3.68, p = 0.01$). AMB had higher amounts of Mg, Ca, Mn, and Zn than NFA ($F_{1,10} = 5.15, p = 0.01$) in *S. tenacissima*. Temperature caused no difference in nutrients content between AMB and NFA for either *F. iberica* ($F_{1,8} = 2.31, p = 0.12$) or *F. indigesta* ($F_{1,9} = 1.68, p = 0.18$). AMB plants presented higher ratios of P/K, P/S, and N/K and lower N/P than NFA plants in *A. castellana* ($F_{1,9} = 5.53, p = 0.01$). In *S. tenacissima*, AMB was significantly different from NFA due to different ratios of N/K, P/K, K/Ca, and K/Mg ($F_{1,10} = 5.24, p = 0.02$). There were no differences between AMB and NFA treatments in nutrients ratios in *F. iberica* ($F_{1,8} = 1.68, p = 0.21$) and *F. indigesta* ($F_{1,9} = 1.57, p = 0.23$).

2.5. Correlations

Figure 6 summarizes the significant correlations ($p < 0.05$) of nutrients content (Figure 6A) and nutrient ratios (Figure 6B) with gas exchange parameters (A, g_s , and WUE) and aboveground vegetative biomass (leaf DW) for the four grasses studied. Photosynthetic activity was positively correlated with Mn and Mg, and g_s increased with the amount of K. WUE was correlated positively with Zn, N/P, K/Ca, and K/Mg. On the other hand, higher K, Ca, Mg, Mn, and S were correlated with lower WUE values. Leaf DW was positively correlated with C/N, N/P, and K/Ca, but negatively with N, K, Ca, Cu, Fe, Mn, P, S, N/K, P/K, P/S, K/Mg, Ca/Mg.

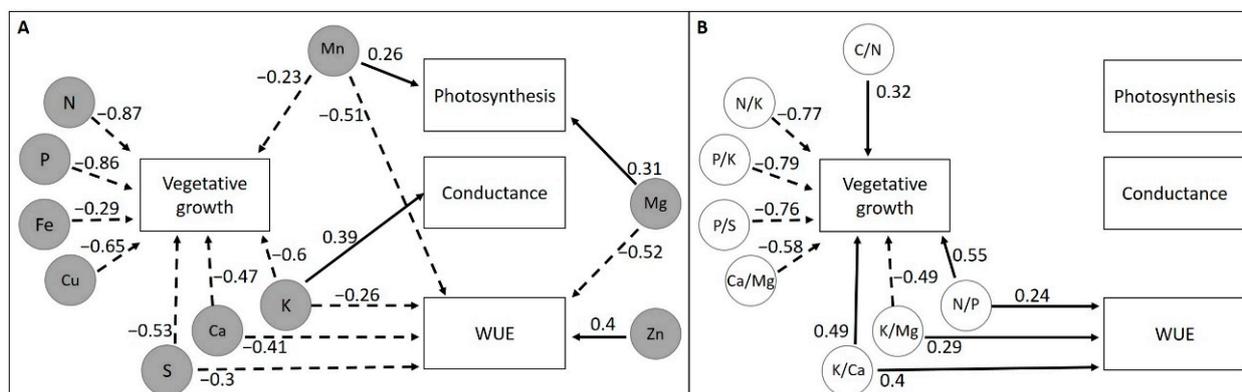


Figure 6. Correlations of (A) nutrients and (B) nutrient ratios with photosynthesis (A), stomatal conductance (g_s), water use efficiency (WUE), and vegetative growth of the four perennial Mediterranean grasses assayed. Only significant correlations are shown. Solid and dotted arrows denote positive and negative correlations, respectively. Values next to arrows are the correlation values.

3. Discussion

The present study reproduced the current and foreseen O_3 levels and temperature increases on the pasture habitats of Mediterranean mountains. Ozone concentrations recorded at a mountain summit of the Spanish Central System yielded AOT40 values, cumulated over a 3-month period, ranging from 13,900 to 19,100 nL L⁻¹ h depending among years [12]. These values greatly exceeded the current objectives for plant protection established in the EU directive of air quality (2008/50/EU), and the critical levels of the Air Convention for perennial pastures [9]. The equivalent estimated 3-month AOT40 index for NFA+ and NFA++ treatments in the present assay, comprising a 63-day fumigation period, would be about 12,000 and 22,000 nL L⁻¹ h, respectively. Therefore, the O_3 treatments in the experiment captured the interannual variability of O_3 in the Iberian mountains. Furthermore, O_3 concentrations in the NFA++ were also in the range of expected future O_3 levels in Mediterranean mountain areas by 2050 [24].

Regarding the increase in temperature, previous studies have detected a temperature increment in the Central System of 0.5 °C in the last decade, and projected temperatures in the Mediterranean mountains for 2085 could rise 5 degrees [50,51]. Therefore, the increase in temperature reached in the OTCs with respect to the AMB plots mimicked the expected increase caused by climate warming in the area. Nonetheless, the mean daylight temperatures in this study are within the range of the maximum values recorded in the natural environment of the species tested in the same year of the experiment (25.6 °C; [51,52]). The effect of the temperature increase found in this study could be slightly affected by the small differences between OTC and AMB in the other meteorological factors.

The O₃ and temperature sensitivity of the Mediterranean perennial grass species tested in this study does not present a single response or pattern. Alternatively, response patterns between species will be discussed in the light of their phylogenetic proximity and their adaptation to drought.

3.1. Effects on Plant Growth and Gas Exchange Parameters

As expected, the vegetative growth of Mediterranean perennial grasses showed a tolerant response to O₃ except in *F. indigesta*. This general lack of response to O₃ in growth parameters is consistent with previous results with grasses from natural ecosystems. Both Mediterranean annual and temperate perennial grasses have shown none or minor effects on growth under increasing O₃ exposure, as compared with O₃ sensitive legumes and forbs [31,33,34,41,42,53]. Despite this fairly widespread O₃-tolerance of perennial grasses, some species such as *Nardus stricta* responded negatively to the pollutant [54]. In the present study, *F. indigesta* responded to the O₃ increase following a non-linear pattern: plants developed the highest biomass under the NFA and NFA++ treatments while minimum values were registered in the NFA+ with biomass reductions of 30% compared with CFA. *F. indigesta* showed a greater O₃-sensitivity than its close phylogenetic relative, *F. iberica*, despite the larger growth observed in the latter. *F. iberica* shows preference for wetter soils under natural conditions which would have been favored, as compared with *F. indigesta*, by the growing conditions in this study, where plants were kept with full water availability. A higher growth rate is usually associated with more intense physiological activity, greater gas exchange rates, and O₃ uptake [9]. Therefore, this study shows that the growth is not a reliable indicator of O₃ sensitivity for *Festuca* species. Interestingly, the O₃ levels in the natural mountain habitats of the *Festuca* species at the Spanish Central System during the spring season, when plants do not suffer drought stress and are fully physiologically active [12], are closely reproduced by the NFA+ conditions. Therefore, these results should warn of the current potential risk of O₃ negative effects in *F. indigesta*-dominated natural habitats, which are widespread in the Spanish Central System [28,55].

Only *A. castellana* responded positively to temperature increases. Considering the absence of soil moisture limitation in the present study, a higher growth under warmer conditions for all species could be expected [56]. During daylight hours, the mean temperature inside the OTCs was 26 °C during the experimental period, which is a common value for the natural habitat of this species in late spring and summer months. However, during the last 10 days of the fumigation experiment, temperatures rose inside the OTCs reaching values up to 40 °C. This late condition could hardly have affected the growth responses, since most of the growth occurred before the heat wave, and plants remained healthy without symptoms of heat stress. Therefore, *A. castellana* could potentially benefit, in the absence of water limitation, from the expected warming in its natural habitat in terms of aboveground biomass growth.

Gas exchange parameters were more sensitive to O₃ than biomass yield. *F. iberica*, *F. indigesta*, and *A. castellana* were affected differently according to the gas exchange parameter considered, while *S. tenacissima* showed no effect. Previous studies have found that O₃ frequently inhibits A as a broad physiological response to the pollutant [57–59], response that has also been reported for grass species, and similarly this pollutant can also induce stomatal closure in O₃-sensitive species [13,22,54,60,61]. In agreement with biomass growth

responses, no species showed reductions of A in response to O₃ except *F. indigesta* and a marginal increase was even recorded in *A. castellana*. Both *Festuca* species showed a similar non-linear response pattern, but this effect was only marginally significant for *F. indigesta*. Strikingly, the highest A values measured in NFA+ were associated with the maximum growth loss in *F. indigesta*. A possible explanation is that photosynthate production under NFA+ was allocated to oxidative damage repair and defence [62–65], below ground biomass, or reproductive output [60,66]. However, similar A values in the NFA++ treatment would again be allocated to aboveground biomass growth. This kind of hormetic trend has been studied as an O₃-triggered biphasic response under an oxidative stress [67]. A marginally linear positive response was found in *A. castellana*, not associated with higher aboveground biomass. In this case, like in *F. indigesta*, results suggest that higher amounts of photosynthates may have been allocated to other sinks under increasing O₃ stress.

The O₃ tolerance of plant species have been frequently related with stomatal behavior, since higher g_s favors O₃ absorption and greater oxidative damage [68–70]. Of the species analyzed here, *S. tenacissima* and *F. iberica* would be the species with the maximum potential g_s (based on CFA values). However, both species were tolerant in terms of growth and A. The O₃-tolerance of the assayed grasses could then be related with a greater plant capacity for detoxification and repair, or a greater allocation to aboveground biomass, which has been considered among the main mechanisms underlying the tolerance of some species [60,66,71].

WUE was the gas exchange parameter most significantly affected by O₃ exposure in the present study. Ozone can unbalance WUE by inducing changes of different intensity in A and g_s [72,73]. Ozone affected WUE in both *Festuca* species, although with different patterns. In *F. iberica*, the pollutant induced a reduction of stomatal opening maintaining A, resulting in a significant increase in WUE with increasing O₃ exposure. *F. iberica* can be classified as a “water use opportunistic” species [74], based on high g_s and low WUE observed under the CFA treatment. However, under scenarios of high O₃ pollution, *F. iberica* would turn on a more “conservative” water use strategy. Therefore, the physiological response of *F. iberica* to O₃ levels expected in the future could help to simultaneously limit the effects of drought. Moreover, the WUE of this species was not affected by temperature. Therefore, the results would point towards a good performance of *F. iberica* under future combined scenarios of warming and O₃ pollution. *F. indigesta* showed the highest WUE under current O₃ levels in mountain areas (NFA+ treatment) of the species tested in this study, which is in concordance with its better adaptation to low soil water availability growing conditions [44]. However, increasing O₃ levels might affect the plant adaptation to these growing conditions, as O₃ induced a quadratic (hormetic) response in WUE, with the lowest WUE found under NFA++. The observed response of WUE to O₃ in *A. castellana* can also be explained as another hormetic response explained by effects on g_s: first the rise of O₃ would induce a stomatal closure and then the excess of oxidative stress by higher O₃ levels would increase this parameter. This pattern of response could be related with a sluggish response of the stomata under O₃ stress [13,67,69]. The WUE response of *A. castellana* and *F. indigesta* pointed out in the same direction: both species maintained elevated A values under higher O₃ exposure, but at the cost of losing efficiency in the gas exchange process. Thus, the projected O₃ concentrations in mountain areas [24] could limit the tolerance of these species to the expected increases in drought [49] in these areas.

Contrarily to O₃ effects, the increase in temperature improved WUE rate in *A. castellana* because of the decrease in g_s while maintaining A. The higher WUE was in agreement with a greater biomass growth. *S. tenacissima*, an O₃ tolerant species in terms of biomass growth and gas exchange, was the only species that lowered A values when grown at higher temperatures. *S. tenacissima* was sensitive to warming despite the fact that this species usually inhabits the driest and hottest environments of all the species tested [45,46]. Frequently, under well-watered conditions, a moderate increase in air temperature is accompanied by an increase in A [75,76]. However, the opposite response observed in *S.*

tenacissima could be more related with the hot temperatures reached at the late exposure period which can be affected the functioning of the photosystem II, accordingly with [77] for some herbaceous forbs.

3.2. Ozone and Temperature Effects on Foliage Nutrient Content

The sensitivity of foliar nutrient content and nutrient ratios to O₃ and temperature varied among the species and nutrients, or the nutrient ratios considered, in agreement with the results already published in the literature. *A. castellana* and *S. tenacissima* were the most sensitive species based on the number of affected nutrients (Figures S2 and S3), although O₃ sensitivity in *S. tenacissima* was based on the small increase in O₃ from CFA to NFA. *A. castellana* and *S. tenacissima* shared the same P/S, P/K, N/K, Ca, Mg, and Mn responses against the rise of O₃ and/or temperature. Among *Festuca* species, only *F. iberica* showed some response to O₃ in terms of nutrient content (Figures S4 and S5). Beyond the species-specific responses, the lack of similarity in the sensitivity of *A. castellana* and *F. iberica* to O₃ and temperature do not support the hypothesis that species sharing the same habitat would respond similarly to the abiotic factors tested in this study. On the other hand, the lack of effects on nutrient content and ratios in the *Festuca* species suggest that temperature sensitivity is phylogenetically related. Nevertheless, more studies involving more species are needed to unravel whether the nutrient response to O₃ and temperature is related to habitat adaptation or phylogeny.

Some of the responses found in the present work have been previously described for other species. Subtle effects of O₃ on the foliar C/N ratio in *S. tenacissima* and the absence of effects in the other species tested, contrast with O₃-induced effects in foliar N observed in trees, pastures, and crop species, with both increasing and decreasing foliar N trends with O₃ exposure [16,78,79]. However, this result is in agreement with other studies reporting no effect in foliar N [80,81]. *Beta vulgaris* plants showed reductions in leaf Fe content in response to O₃ [18], like in *F. iberica*. Again, the O₃ effect on *S. tenacissima* promoting an increase in K/Ca, was also observed in the stomatal guard cells of the O₃-injured leaves of *Betula pendula* [82]. Ozone induced decreases of foliar Mg and changes in the K/Mg ratio in *F. iberica* and *S. tenacissima* have also been described for *Solanum tuberosum* and *Beta vulgaris* [18,19,83]. Other studies, however, do not show O₃ effects on foliar nutrient concentration in species such as *Lolium perenne* or *Trifolium repens*, or a mixture of *Trifolium pratense*, *Phleum pratense*, and *Festuca pratensis* [80,84]. Temperature effects on foliar P and Fe in *A. castellana* have also been described in *Coffea canephora* and in plants from a dwarf shrub-dominated ecosystem [85,86]. However, the negative temperature effect on Mg in *A. castellana* was not found in Mediterranean scrub [87]. These results show that O₃ and temperature effects on foliar nutrient content and ratios seem to be species-specific and variable between nutrients. More studies would be needed for understanding general response patterns.

Nutrient and nutrient ratios were more correlated with vegetative growth than with gas exchange parameters. The influence of O₃ and temperature on leaf nutrient concentration and the relationship with vegetative growth, A, g_s, and WUE constitute a complex web of interactions (Figures S2–S5). Considering all four species together, the vegetative growth was negatively correlated with most of the macro- and micronutrients and the different ratios considered (Figure 6). The dilution of nutrient content related with biomass growth has been shown for N, P, and K [88,89]. These nutrients are directly or indirectly related to plant development [90,91]. In general, N- and P-deficiency inhibits leaf growth and Fe is mostly located in the chloroplast of growing leaves [90,91]. In *A. castellana*, the temperature effect on leaf nutrients could be explained by a dilution effect due to increases in vegetative growth under the warmest treatment. The N/P ratio was also positively correlated with vegetative growth in this species, which is in agreement with temperature-induced increases in this ratio. None of effects of O₃ on foliar nutrient content could be related with effects on biomass growth in any of the four species. However, O₃ effects on foliar nutrient

contents in *A. castellana*, *F. iberica*, or *S. tenacissima* found in this short-term exposure study may indicate that effects on plant growth could appear under long-term high O₃ exposures.

In the present study, Mn and Mg were positively correlated with A. In fact, Mn and Mg act synergistically in the basal metabolism of electron transport reactions in photosynthesis [92,93]. An ozone-induced decrease in foliar Mg has been related with chlorophyll loss and A decline [18,19,83]. In contrast to this, O₃ exposure increased A in *A. castellana* associated with a decrease in leaf Mn content (Figure S2). However, in *S. tenacissima* the decrease in Mn and Mg under the high temperature treatment could be associated with the decrease in A (Figure S2). Foliar K was positively correlated with g_s across species. This correlation is in accordance with the stomatal hydration regulatory function of K that produce the stomata opening due to increased turgor in guard cells [90,92]. Despite O₃ and temperature effects found on the g_s of *A. castellana* and *F. iberica*, leaf K content was not affected by the experimental treatments in any of the four species studied (Figures S2 and S5). Previous studies neither found significant effects on the K-levels of potato or wheat exposed to increased levels of O₃ [19,94]. S, Ca, K, Mg, and Mn foliar concentrations were negatively correlated with WUE while Zn and N/P, K/Ca, and Ca/Mg ratios were positively correlated (Figure 6). These relationships can be explained through the importance of some of these elements on A or g_s processes. Stomatal opening by K accumulation is promoted by Ca signaling [90–92] which could result in decreased WUE. Zn-mediated stability of membrane integrity [91,93] may explain the higher WUE under high foliar Zn levels while decreases in foliar Mg could reduce WUE via reductions in A. The O₃-induced decreases in Mn and P and in the P/K and P/S ratios in *A. castellana* can be associated directly or indirectly with the observed quadratic effect on WUE. However, the quadratic effect of O₃ on WUE in *F. indigesta* was not associated with any significant change in foliar nutrient concentration. With respect to the temperature effects in *A. castellana*, the increase observed in WUE was associated with decreases in Ca and Mg concentration and a reduction in the N/P ratio.

4. Material and Methods

4.1. OTC Experiment

Plant exposure to O₃ treatment levels was performed in an NCLAN-type OTC facility (adapted from the original design by [95]) located in central Spain (450 m.a.s.l., 40°3' N, 4°26' W). Plants were exposed to four O₃ treatments: charcoal-filtered air (CFA), non-filtered air (NFA) reproducing ambient levels, non-filtered air supplemented with 20 nL L⁻¹ of O₃ (NFA+), and non-filtered air supplemented with 40 nL L⁻¹ of O₃ (NFA++). Each O₃ treatment was replicated 3 times in 12 equally built OTCs randomly distributed in 3 lines (blocks) avoiding shading effects between OTCs. One chamber-less plot (AMB) per block was considered to control the chamber effect and to study the consequences of temperature increase on species.

Ozone was produced from pure O₂ through an O₃ generator (A2Z Ozone, Inc., Louisville, KY, USA) and supplied to the NFA+ and NFA++ plots 8 h day⁻¹ (7:00 to 15:00 GMT) and 7 days week⁻¹. Ozone concentrations inside each chamber and AMB plots were monitored continuously above plant canopy using a UV-absorbance O₃ monitor (ML[®] 9810B, Teledyne, Thousand Oaks, CA, USA) with an automated time-sharing system sampling all the OTCs and AMB plots of each line sequentially. Another O₃ monitor within the fumigation system registered continually ambient O₃ levels to contrast with the NFA and AMB treatments, as a double check of the accuracy of O₃ values. Both O₃ monitors were calibrated at the start of the fumigation treatments following the recommended company protocols. More information about the facility can be found in [22]. Figure S1 shows a picture of the OTCs and AMB plots.

The AOT40 index (nL L⁻¹ h) was used to describe the O₃ exposure. This index was calculated as the accumulated hourly O₃ concentration over the threshold of 40 nL·L⁻¹ during daylight (PAR > 100 μmol·m⁻² s⁻¹) hours through the O₃-fumigation period [96]. The 8h-mean O₃ concentration (nL·L⁻¹) from 7:00–15:00 h GMT was also considered. Air

temperature (T, °C), air relative humidity (RH, %), and photosynthetic active radiation (PAR, $\mu\text{mol}\cdot\text{m}^{-2}\cdot\text{s}^{-1}$) inside one OTC were continuously measured above canopy using AM2315 (T and RH; Adafruit Industries LLC, New York, NY, USA) and Apogee SQ 110 (PAR; Apogee Instruments, Inc., Logan, UT, USA) sensors. Ambient T, RH, and PAR were continuously measured using STH-5031 (T and RH; GEONICA, S.A., Madrid, Spain) and LI-200SZ (PAR; GEONICA, S.A., Madrid, Spain) sensors situated in the meteorological station of the OTC facility.

4.2. Plant Material

Established seedlings of *A. castellana*, *S. tenacissima*, *F. indigesta*, and *F. iberica* grown from seeds collected from natural populations of Spanish Central System area were transplanted to 2 L pots during 2017 spring. A mix of peat, vermiculite, and perlite (60:20:20) was used as plant substrate. The total number of plants varied among species, with 90, 88, 77, and 51 for *A. castellana*, *S. tenacissima*, *F. iberica*, and *F. indigesta*, respectively. Plants were kept outdoors since transplantation. They were frequently irrigated and fertilized to meet plant demand, until the start of the experiment. On 15 April 2018, 3 days prior to the start of the O₃-exposure, plants were cut to 5 cm from the substrate. These species, adapted to cattle browsing, present a good regrowth after biomass cut. On 18 April, plants were randomly allocated to OTCs or AMB plots to start the O₃ fumigation experiment. The 4 species were exposed to the same O₃-treatments that lasted 68, 63, 62, and 57 days after the start of the exposure (DaS) for *A. castellana*, *S. tenacissima*, *F. indigesta*, and *F. iberica*, respectively.

4.3. Vegetative Growth

At the end of the O₃-fumigation period, plants were cut again to 5 cm from the substrate to obtain the final aboveground biomass. Samples were dried at 60 °C until constant weight. The O₃ effect on the dry weight biomass (biomass DW, g) was considered to discuss the O₃-effect on the vegetative growth of the species. The number of samples per treatment and species are shown in Table S1.

4.4. Gas Exchange

Leaf-level net photosynthesis ($\mu\text{mol CO}_2 \text{ m}^{-2}\text{s}^{-1}$) and stomatal conductance to water vapor ($\text{mol H}_2\text{O m}^{-2}\text{s}^{-1}$) were measured using a portable LICOR-6400 infra-red gas analysis system (LiCor Inc., Lincoln, NE, USA). The Water Use Efficiency (WUE; $\mu\text{mol CO}_2 \text{ mol}^{-1} \text{ H}_2\text{O}$) was calculated from the A/g_s ratio. In order to increase the leaf area for the gas exchange measurements, a group of intact leaves was introduced in the LICOR chamber. Each individual measurement was corrected per projected leaf area based on the scanned image of leaf sections measured with ImageJ software [97]. The number of samples per treatment and species are shown in the Table S1. Gas exchange measurements were evenly distributed between 9 and 13 h (GMT) among all treatments, from 28 to 35 days after the start of the fumigation date depending on the phenological stage of each species (Table S2). Air humidity and temperature during the measurements ranged 36–49% and 20–25 °C, respectively, and PAR was maintained at 1000 $\mu\text{mol m}^{-2} \text{ s}^{-1}$, allowing maximum g_s according to previous measurements (Table S2). One measurement per plant was taken on the representative biomass of the plant, avoiding the youngest leaves in growth and the oldest senescent leaves. The arrival of a heat wave towards the end of the ozone fumigation experiment prevented the measurement of gas exchange on ambient *F. indigesta* plants.

4.5. Macro- and Micronutrient Composition

Total aerial biomass DW from the different plants was individually milled and pooled to obtain a minimum of 2 independent samples (no plant was repeated in the mixtures) per OTC and AMB plot for elemental content analysis. The number of samples per treatment and species are shown in Table S1. Total carbon and hydrogen content (C and H), macronutrients (N, K, Ca, Mg, P, and S) and micro-nutrients (B, Cu, Fe, Mn, Mo, and Zn) were analyzed and the ratios C/N, N/K, N/P, P/K, P/S, K/Ca, K/Mg, and Ca/Mg were

also calculated. The total content of C, H, and N were determined by combustion using an elemental analyzer. Samples were digested under controlled conditions for putting quantitatively into solution these macronutrients (K, Ca, Mg, P, and S) and micronutrients (B, Cu, Fe, Mn, Mo, and Zn). Concentrations of these elements were determined by inductively coupled plasma optical emission spectrometry (ICP-OES). More details on sample preparation and elemental analyses are provided in the supplementary information.

4.6. Statistical Analysis

To analyze the effect of O₃ and temperature factors separately on vegetative growth, A, g_s, and WUE, two sets of linear mixed models (LMM) were performed. The first set of LMM considered the O₃ treatments (CFA, NFA, NFA+, and NFA++). In the second set, the comparison between AMB and NFA was considered to study the effect of temperature. For each species, the individual plants were the units of replication. Moreover, to control the possible influence of the experimental design in plants growth, A, g_s, and WUE responses to the fixed factors (O₃ and temperature), block (line), and O₃ treatment nested within line were considered as random factors. For each analysis, the AIC (Akaike information criterion) value was used to choose the most parsimonious option between the models including only line as a random factor or the nested model. Linear O₃ effects were evaluated using linear a priori contrasts to test the stated hypothesis. Moreover, quadratic and cubic responses were also tested. A priori contrasts in linear mixed models were based on [98]. Linear mixed models were performed using lme function (nlme package; [99]). To test the hypotheses of a linear, quadratic, or cubic effect the contr.poly function [100] was used. The model with the lowest *p*-value was chosen among lineal, quadratic, and cubic options for each analysis. Temperature effects were tested using ANOVA function.

To evaluate the O₃ and temperature effect on leaf nutrient composition, PERMANOVA and Non-Metric Multidimensional Scaling (NMDS) analyses were performed using leaf nutrient concentration and nutrient ratios as a multivariate trait among treatments. To prevent PERMANOVA from being largely influenced by the high levels of some elements, the data were transformed using the square root, and a similarity matrix was made by Bray–Curtis approximation [101]. A total of 9999 permutations were chosen, and PERMANOVA was performed using vegan package (Oksanen et al., 2020). Pairwise analyses were performed when PERMANOVA found differences among O₃-treatment using pairwise.adonis2 function (package pairwiseAdonis; [102]). Two-dimension NMDS were used to represent the elemental composition multidimensional data. Function metaMDS was used to perform the NMDS (vegan library; [103]). Correlations of nutrients and nutrient ratios with vegetative growth, A, g_s, and WUE were also performed using vegan package [103]. All the statistical analyses described were carried out in R software [100]. Significant statistical differences were considered at *p*-values lower than 0.05. Marginally statistical differences were considered at *p*-values from 0.1 to 0.05.

5. Conclusions

Gas exchange parameters, in particular WUE, and leaf nutrient concentration were more sensitive to O₃ and temperature increases than aboveground biomass growth in the Mediterranean perennial grass species assayed in this experiment.

The O₃ sensitivity classification of grasses based on growth and gas exchange does not match with that based on foliar nutrients. *S. tenacissima* was the most tolerant to O₃ in terms of growth and gas exchange. However, *A. castellana* were the most sensitive to the pollutant in terms of nutrients. In both *Festucas*, gas exchange, growth, and leaf nutrient content had different responses to O₃. The O₃-effects on foliar nutrients and their ratios, although substantial, did not have an early impact on plant growth. Longer term exposures will be needed to understand the potential consequences of these changes for the growth and survival of these species.

S. tenacissima and *A. castellana* were the two most sensitive grasses to temperature increase. *S. tenacissima* was the only one that decreased A at high temperatures, despite the

fact that it usually inhabits the driest and hottest environments of all the species tested. *A. castellana* was the only species that responded positively in terms of aboveground biomass and WUE when grown in a warmer environment. This species could benefit from climate warming when water is not a limiting factor. However, it was the most temperature-sensitive in terms of foliar nutrients, so the long-term effects may differ from the ones found in this experiment. Only leaf nutrient content responses to temperature seem to be phylogenetically constrained in *Festuca* species.

The results showed that responses to O₃ and temperature can have different species-specific effects on plant physiology, potentially altering the ability of plants to cope with environmental stresses and changing the relationships among species that share the same habitat. Therefore, climate warming and O₃ pollution should be considered as two important threats to Mediterranean perennial pastures in the framework of the Global Change. More experimental work would be needed to better understand the behaviour of these complex Mediterranean pastures in response to combinations of stress factors.

Supplementary Materials: The following supporting information can be downloaded at: <https://www.mdpi.com/article/10.3390/plants12030664/s1>, Figure S1: Picture of two OTCs and one AMB plot of the OTC facility; Description of the description of the total content of C, H, macro- and micro-nutrient determination; Table S1: Number of samples per treatment and specie for Vegetative growth (dry weight), Gas Exchange (A, g_s and WUE) and Nutrient (contents and rates) measurements; Table S2. LICOR measurements conditions for photosynthesis and conductance; Table S3: A priori contrast results of the lineal effect, quadratic effect and cubic effect for each specie in vegetative growth, A, g_s and WUE; Table S4: Temperature effect results on vegetative growth, A, g_s and WUE; Table S5: The structure of the random part of the model for each trait (vegetative growth, photosynthesis, conductance and WUE) and species; Table S6: PERMANOVA tables of nutrient and nutrients ratios for O₃ and temperature; Figure S2: Ozone and temperature observed effects on growth and gas exchange measured traits and nutrients in *A. castellana*; Figure S3: Ozone and temperature observed effects on growth and gas exchange measured traits and nutrients in *S. tenacissima*; Figure S4: Ozone and temperature observed effects on growth and gas exchange measured traits and nutrients in *F. indigesta*; Figure S5: Ozone and temperature observed effects on growth and gas exchange measured traits and nutrients in *F. iberica*.

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References

1. Chappelka, A.H.; Samuelson, L.J. Ambient ozone effects on forest trees of the eastern United States: A review. *New Phytol.* **1998**, *139*, 91–108. [CrossRef]
2. Ainsworth, E.A.; Yendrek, C.R.; Sitch, S.; Collins, W.J.; Emberson, L.D. The effects of tropospheric ozone on net primary productivity and implications for climate change. *Annu. Rev. Plant Biol.* **2012**, *63*, 637–661. [CrossRef]
3. Leisner, C.P.; Ainsworth, E.A. Quantifying the effects of ozone on plant reproductive growth and development. *Glob. Chang. Biol.* **2012**, *18*, 606–616. [CrossRef]
4. Mills, G.; Sharps, K.; Simpson, D.; Pleijel, H.; Frei, M.; Burkey, K.; Emberson, L.; Uddling, J.; Broberg, M.; Feng, Z.; et al. Closing the global ozone yield gap: Quantification and cobenefits for multistress tolerance. *Glob. Chang. Biol.* **2018**, *24*, 4869–4893. [CrossRef] [PubMed]
5. Monks, P.S.; Archibald, A.T.; Colette, A.; Cooper, O.; Coyle, M.; Derwent, R.; Fowler, D.; Granier, C.; Law, K.S.; Mills, G.E.; et al. Tropospheric ozone and its precursors from the urban to the global scale from air quality to short-lived climate forcer. *Atmos. Chem. Phys.* **2015**, *15*, 8889–8973. [CrossRef]
6. Querol, X.; Alastuey, A.; Gangoiti, G.; Perez, N.; Lee, H.K.; Eun, H.R.; Park, Y.; Mantilla, E.; Escudero, M.; Titos, G.; et al. Phenomenology of summer ozone episodes over the Madrid Metropolitan Area, central Spain. *Atmos. Chem. Phys. Discuss.* **2018**, *18*, 6511–6533. [CrossRef]
7. Millán, M.; Salvador, R.; Mantilla, E.; Artiñano, B. Meteorology and photochemical air pollution in Southern Europe: Experimental results from EC research projects. *Atmos. Environ.* **1996**, *30*, 1909–1924. [CrossRef]
8. Cristofanelli, P.; Bonasoni, P. Background ozone in the southern Europe and Mediterranean area: Influence of the transport processes. *Environ. Pollut.* **2009**, *157*, 1399–1406. [CrossRef]
9. CLRTAP. *Chapter III. Mapping Critical Levels for Vegetation*; ICP Vegetation Programme Coordination Centre, Centre for Ecology and Hydrology: Bangor, UK, 2017; Volume 2017.
10. MITERD Evaluación de la Calidad del Aire en España. Available online: https://www.miteco.gob.es/es/calidad-y-evaluacion-ambiental/temas/atmosfera-y-calidad-del-aire/SituacióndecalidaddelaireenEspa-na2010_tcm30-182501.pdf (accessed on 1 January 2022).
11. Díaz-de-Quijano, M.; Peñuelas, J.; Ribas, À. Increasing interannual and altitudinal ozone mixing ratios in the Catalan Pyrenees. *Atmos. Environ.* **2009**, *43*, 6049–6057. [CrossRef]
12. Elvira, S.; González-Fernández, I.; Alonso, R.; Sanz, J.; Bermejo-Bermejo, V. Ozone levels in the Spanish Sierra de Guadarrama mountain range are above the thresholds for plant protection: Analysis at 2262, 1850, and 995 m a.s.l. *Environ. Monit. Assess.* **2016**, *188*, 1–20. [CrossRef]
13. Bergmann, E.; Bender, J.; Weigel, H.-J. *Assessment of the Impacts of Ozone on Biodiversity in Terrestrial Ecosystems: Literature Review and Analysis of Methods and Uncertainties in Current Risk Assessment Approaches. Part II: Literature Review of the Current State of Knowledge on the Impact of Ozone on Biodiversity in Terrestrial Ecosystems*; Umweltbundesamt: Dessau-Roßlau, Germany, 2015; Volume 139. [CrossRef]
14. Booker, F.; Muntifering, R.; Mcgrath, M.; Burkey, K.; Decoteau, D.; Fiscus, E.; Manning, W.; Krupa, S.; Chappelka, A.; Grantz, D. The ozone component of global change: Potential effects on agricultural and horticultural plant yield, product quality and interactions with invasive species. *J. Integr. Plant Biol.* **2009**, *51*, 337–351. [CrossRef]
15. Grulke, N.E.; Heath, R.L. Ozone effects on plants in natural ecosystems. *Plant Biol.* **2020**, *22*, 12–37. [CrossRef]
16. González-Fernández, I.; Elvira, S.; Calatayud, V.; Calvo, E.; Aparicio, P.; Sánchez, M.; Alonso, R.; Bermejo Bermejo, V. Ozone effects on the physiology and marketable biomass of leafy vegetables under Mediterranean conditions: Spinach (*Spinacia oleracea* L.) and Swiss chard (*Beta vulgaris* L. var. *cycla*). *Agric. Ecosyst. Environ.* **2016**, *235*, 215–228. [CrossRef]
17. Wittig, V.E.; Ainsworth, E.A.; Naidu, S.L.; Karnosky, D.F.; Long, S.P. Quantifying the impact of current and future tropospheric ozone on tree biomass, growth, physiology and biochemistry: A quantitative meta-analysis. *Glob. Chang. Biol.* **2009**, *15*, 396–424. [CrossRef]
18. Tiwari, S.; Agrawal, M.; Marshall, F.M. Seasonal variations in adaptational strategies of *Beta vulgaris* L. plants in response to ambient air pollution: Biomass allocation, yield and nutritional quality. *Trop. Ecol.* **2010**, *51*, 353–363.
19. Fangmeier, A.; De Temmerman, L.; Black, C.; Persson, K.; Vorne, V. Effects of elevated CO₂ and/or ozone on nutrient concentrations and nutrient uptake of potatoes. *Eur. J. Agron.* **2002**, *17*, 353–368. [CrossRef]
20. Valkama, E.; Koricheva, J.; Oksanen, E. Effects of elevated O₃, alone and in combination with elevated CO₂, on tree leaf chemistry and insect herbivore performance: A meta-analysis. *Glob. Chang. Biol.* **2007**, *13*, 184–201. [CrossRef]
21. Gimeno, B.S.; Bermejo, V.; Sanz, J.; De Torre, D.; Gil, J.M. Assessment of the effects of ozone exposure and plant competition on the reproductive ability of three therophytic clover species from Iberian pastures. *Atmos. Environ.* **2004**, *38*, 2295–2303. [CrossRef]
22. Calvete-Sogo, H.; Elvira, S.; Sanz, J.; González-Fernández, I.; García-Gómez, H.; Sánchez-Martín, L.; Alonso, R.; Bermejo-Bermejo, V. Current ozone levels threaten gross primary production and yield of Mediterranean annual pastures and nitrogen modulates the response. *Atmos. Environ.* **2014**, *95*, 197–206. [CrossRef]
23. Hayes, F.; Mills, G.; Jones, L.; Ashmore, M. Does a simulated upland grassland community respond to increasing background, peak or accumulated exposure of ozone? *Atmos. Environ.* **2010**, *44*, 4155–4164. [CrossRef]
24. Fuhrer, J.; Val Martín, M.; Mills, G.; Heald, C.L.; Harmens, H.; Hayes, F.; Sharps, K.; Bender, J.; Ashmore, M.R. Current and future ozone risks to global terrestrial biodiversity and ecosystem processes. *Ecol. Evol.* **2016**, *6*, 8785–8799. [CrossRef]

25. Agathokleous, E.; Feng, Z.; Oksanen, E.; Sicard, P.; Wang, Q.; Saitanis, C.J.; Araminiene, V.; Blande, J.D.; Hayes, F.; Calatayud, V.; et al. Ozone affects plant, insect, and soil microbial communities: A threat to terrestrial ecosystems and biodiversity. *Sci. Adv.* **2020**, *6*, eabc1176. [CrossRef]
26. Myers, N.; Mittermeyer, R.A.; Mittermeyer, C.G.; Da Fonseca, G.A.B.; Kent, J. Biodiversity hotspots for conservation priorities. *Nature* **2000**, *403*, 853–858. [CrossRef] [PubMed]
27. Gutiérrez-Girón, A.; Gavilán, R.G. Spatial patterns and interspecific relations analysis help to better understand species distribution patterns in a Mediterranean high mountain grassland. *Plant Ecol.* **2010**, *210*, 137–151. [CrossRef]
28. López-angulo, J.; Pescador, D.S.; Sánchez, A.M.; Luzuriaga, A.L.; Cavieres, L.A.; Escudero, A. Science of the Total Environment Impacts of climate, soil and biotic interactions on the interplay of the different facets of alpine plant diversity. *Sci. Total Environ.* **2020**, *698*, 133960. [CrossRef]
29. Stanisci, A.; Laura, M.; Pelino, G.; Chiarucci, A. Assessing the diversity pattern of cryophilous plant species in high elevation habitats. *Plant Ecol.* **2011**, 595–600. [CrossRef]
30. Le Houérou, H.N. Biogeography of the arid steppeland north of the Sahara. *J. Arid Environ.* **2001**, *48*, 103–128. [CrossRef]
31. Gimeno, B.S.; Bermejo, V.; Sanz, J.; De La Torre, D.; Elvira, S. Growth response to ozone of annual species from Mediterranean pastures. *Environ. Pollut.* **2004**, *132*, 297–306. [CrossRef]
32. Bermejo, V.; Gimeno, B.S.; Sanz, J.; De La Torre, D.; Gil, J.M. Assessment of the ozone sensitivity of 22 native plant species from Mediterranean annual pastures based on visible injury. *Atmos. Environ.* **2003**, *37*, 4667–4677. [CrossRef]
33. Calvete-Sogo, H.; González-Fernández, I.; Sanz, J.; Elvira, S.; Alonso, R.; García-Gómez, H.; Ibáñez-Ruiz, M.A.; Bermejo-Bermejo, V. Heterogeneous responses to ozone and nitrogen alter the species composition of Mediterranean annual pastures. *Oecologia* **2016**, *181*, 1055–1067. [CrossRef]
34. Sanz, J.; Bermejo, V.; Muntifering, R.; González-Fernández, I.; Gimeno, B.S.; Elvira, S.; Alonso, R. Plant phenology, growth and nutritive quality of *Briza maxima*: Responses induced by enhanced ozone atmospheric levels and nitrogen enrichment. *Environ. Pollut.* **2011**, *159*, 423–430. [CrossRef] [PubMed]
35. Sanz, J.; Calvete-Sogo, H.; González-Fernández, I.; Lin, J.; García-Gómez, H.; Muntifering, R.; Alonso, R.; Bermejo-Bermejo, V. Foliar senescence is the most sensitive response to ozone in *Bromus hordeaceus* and is modulated by nitrogen input. *Grass Forage Sci.* **2015**, *70*, 71–84. [CrossRef]
36. Nussbaum, S.; Geissmann, M.; Fuhrer, J. Ozone exposure-response relationships for mixtures of perennial ryegrass and white clover depend on ozone exposure patterns. *Atmos. Environ.* **1995**, *29*, 989–995. [CrossRef]
37. Van De Staaij, J.W.M.; Tonneijck, A.E.G.; Rozema, J. The effect of reciprocal treatments with ozone and ultraviolet-B radiation on photosynthesis and growth of perennial grass *Elymus athericus*. *Environ. Pollut.* **1997**, *97*, 281–286. [CrossRef]
38. Bassin, S.; Volk, M.; Suter, M.; Buchmann, N.; Fuhrer, J. Nitrogen deposition but not ozone affects productivity and community composition of subalpine grassland after 3 yr of treatment. *New Phytol.* **2007**, *175*, 523–534. [CrossRef]
39. REILING, K.; DAVISON, A.W. The response of native, herbaceous species to ozone: Growth and fluorescence screening. *New Phytol.* **1992**, *120*, 29–37. [CrossRef]
40. Bassin, S.; Volk, M.; Fuhrer, J. Species Composition of Subalpine Grassland is Sensitive to Nitrogen Deposition, but Not to Ozone, After Seven Years of Treatment. *Ecosystems* **2013**, *16*, 1105–1117. [CrossRef]
41. Bassin, S.; Werner, R.A.; Sörgel, K.; Volk, M.; Buchmann, N.; Fuhrer, J. Effects of combined ozone and nitrogen deposition on the in situ properties of eleven key plant species of a subalpine pasture. *Oecologia* **2009**, *158*, 747–756. [CrossRef]
42. Hayes, F.; Mills, G.; Williams, P.; Harmens, H.; Büker, P. Impacts of summer ozone exposure on the growth and overwintering of UK upland vegetation. *Atmos. Environ.* **2006**, *40*, 4088–4097. [CrossRef]
43. Ruiz-Labourdette, D.; Martínez, F.; Martín-López, B.; Montes, C.; Pineda, F.D. Equilibrium of vegetation and climate at the European rear edge. A reference for climate change planning in mountainous Mediterranean regions. *Int. J. Biometeorol.* **2011**, *55*, 285–301. [CrossRef]
44. Luceño, M.; Vargas, P.; García, B. *Guía de Campo del Sistema Central*; Editorial Raices: Mexico City, Mexico, 2016; ISBN 9788486115852.
45. Anthos Information System of the plants of Spain. Real Jardín Botánico, CSIC—Fundación Biodiversidad. Available online: <http://www.anthos.es> (accessed on 1 March 2022).
46. Cortina, J.; Maestre, F.T.; Ramírez, D. Innovations in Semiarid Land Restoration. The case of *Stipa tenacissima* L. Steppes. In *Land Restoration to Combat Desertification. Innovative Approaches, Quality Control and Project Evaluation*; Fundación CEAM: Valencia, Spain, 2006; pp. 121–144.
47. Soreng, R.J.; Peterson, P.M.; Romaschenko, K.; Davidse, G.; Zuloaga, F.O.; Judziewicz, E.J.; Filgueiras, T.S.; Davis, J.I.; Morrone, O. A worldwide phylogenetic classification of the Poaceae (Gramineae). *J. Syst. Evol.* **2015**, *53*, 117–137. [CrossRef]
48. Prieto-Benítez, S.; Ruiz-Checa, R.; Bermejo-Bermejo, V.; Gonzalez-Fernandez, I. The Effects of Ozone on Visual Attraction Traits of *Erodium paularense* (Geraniaceae) Flowers: Modelled Perception by Insect Pollinators. *Plants* **2021**, *10*, 2750. [CrossRef]
49. Giménez-Benavides, L.; Escudero, A.; García-Camacho, R.; García-Fernández, A.; Iriondo, J.M.; Lara-Romero, C.; Morente-López, J. How does climate change affect regeneration of Mediterranean high-mountain plants? An integration and synthesis of current knowledge. *Plant Biol.* **2018**, *20*, 50–62. [CrossRef]
50. Nogués-Bravo, D.; Araújo, M.B.; Errea, M.P.; Martínez-Rica, J.P. Exposure of global mountain systems to climate warming during the 21st Century. *Glob. Environ. Chang.* **2007**, *17*, 420–428. [CrossRef]

51. Vegas-Cañas, C.; Fidel González-Rouco, J.; Navarro-Montesinos, J.; García-Bustamante, E.; Lucio-Eceiza, E.E.; García-Pereira, F.; Rodríguez-Camino, E.; Chazarra-Bernabé, A.; Álvarez-Arévalo, I. An assessment of observed and simulated temperature variability in Sierra de Guadarrama. *Atmosphere* **2020**, *11*, 985. [CrossRef]
52. ECA&D European Climate Assessment & Dataset project. Available online: <https://www.ecad.eu/> (accessed on 5 November 2022).
53. Bender, J.; Muntifering, R.B.; Lin, J.C.; Weigel, H.J. Growth and nutritive quality of *Poa pratensis* as influenced by ozone and competition. *Environ. Pollut.* **2006**, *142*, 109–115. [CrossRef]
54. Hayes, F.; Mills, G.; Ashmore, M. Effects of ozone on inter- and intra-species competition and photosynthesis in mesocosms of *Lolium perenne* and *Trifolium repens*. *Environ. Pollut.* **2009**, *157*, 208–214. [CrossRef] [PubMed]
55. García-Romero, A.; Muñoz, J.; Andrés, N.; Palacios, D. Relationship between climate change and vegetation distribution in the Mediterranean mountains: Manzanares Head valley, Sierra De Guadarrama (Central Spain). *Clim. Chang.* **2010**, *100*, 645–666. [CrossRef]
56. Rötter, R.; Van De Geijn, S.C. Climate change effects on plant growth, crop yield and livestock. *Clim. Chang.* **1999**, *43*, 651–681. [CrossRef]
57. Dann, M.S.; Pell, E.J. Decline of Activity and Quantity of Ribulose Bisphosphate Carboxylase/Oxygenase and Net Photosynthesis in Ozone-Treated Potato Foliage. *Plant Physiol.* **1989**, *91*, 427–432. [CrossRef] [PubMed]
58. Lehnerr, B.; Grandjean, A.; Mächler, F.; Fuhrer, J. The Effect of Ozone in Ambient Air on Ribulosebisphosphate Carboxylase/Oxygenase Activity Decreases Photosynthesis and Grain Yield in Wheat. *J. Plant Physiol.* **1987**, *130*, 189–200. [CrossRef]
59. Wittig, V.E.; Ainsworth, E.A.; Long, S.P. To what extent do current and projected increases in surface ozone affect photosynthesis and stomatal conductance of trees? A meta-analytic review of the last 3 decades of experiments. *Plant, Cell Environ.* **2007**, *30*, 1150–1162. [CrossRef]
60. Davison, A.W.; Barnes, J.D. Effects of ozone on wild plants. *New Phytol.* **1998**, *139*, 135–151. [CrossRef]
61. Volk, M.; Bungener, P.; Contat, F.; Montani, M.; Fuhrer, J. Grassland yield declined by a quarter in 5 years of free-air ozone fumigation. *Glob. Chang. Biol.* **2006**, *12*, 74–83. [CrossRef]
62. Halliwell, B. Biochemistry of oxidative stress. *Biochem. Soc. Trans.* **2007**, *35*, 1147–1150. [CrossRef]
63. Sies, H. Biochemistry of oxidative stress. *Angew. Chemie Int. Ed. Engl.* **1986**, *25*, 1058–1071. [CrossRef]
64. Alonso, R.; Elvira, S.; Castillo, F.J.; Gimeno, B.S. Interactive effects of ozone and drought stress on pigments and activities of antioxidative enzymes in *Pinus halepensis*. *Plant, Cell Environ.* **2001**, *24*, 905–916. [CrossRef]
65. Pellegrini, E.; Hoshika, Y.; Dusart, N.; Cotrozzi, L.; Gérard, J.; Nali, C.; Vaultier, M.N.; Jolivet, Y.; Lorenzini, G.; Paoletti, E. Antioxidative responses of three oak species under ozone and water stress conditions. *Sci. Total Environ.* **2019**, *647*, 390–399. [CrossRef] [PubMed]
66. Andersen, C.P. Source-sink balance and carbon allocation below ground in plants exposed to ozone. *New Phytol.* **2003**, *157*, 213–228. [CrossRef]
67. Agathokleous, E.; Belz, R.G.; Calatayud, V.; De Marco, A.; Hoshika, Y.; Kitao, M.; Saitanis, C.J.; Sicard, P.; Paoletti, E.; Calabrese, E.J. Predicting the effect of ozone on vegetation via linear non-threshold (LNT), threshold and hormetic dose-response models. *Sci. Total Environ.* **2019**, *649*, 61–74. [CrossRef]
68. Amiro, B.D.; Gillespie, T.J.; Thurtell, G.W. Response Flux Density. *Atmos. Environ.* **1984**, *18*, 1207–1215. [CrossRef]
69. Paoletti, E.; Grulke, N.E. Ozone exposure and stomatal sluggishness in different plant physiognomic classes. *Environ. Pollut.* **2010**, *158*, 2664–2671. [CrossRef] [PubMed]
70. Sanz, J.; González-Fernández, I.; Elvira, S.; Muntifering, R.; Alonso, R.; Bermejo-Bermejo, V. Setting ozone critical levels for annual Mediterranean pasture species: Combined analysis of open-top chamber experiments. *Sci. Total Environ.* **2016**, *571*, 670–679. [CrossRef] [PubMed]
71. Feng, Z.; Wang, L.; Pleijel, H.; Zhu, J.; Kobayashi, K. Differential effects of ozone on photosynthesis of winter wheat among cultivars depend on antioxidative enzymes rather than stomatal conductance. *Sci. Total Environ.* **2016**, *572*, 404–411. [CrossRef]
72. Calatayud, V.; Cerveró, J.; Sanz, M.J. Foliar, physiological and growth responses of four maple species exposed to ozone. *Water, Air, Soil Pollut.* **2007**, *185*, 239–254. [CrossRef]
73. Reich, P.B.; Lassoie, J.P. Effects of low level O₃ exposure on leaf diffusive conductance and water-use efficiency in hybrid poplar. *Plant, Cell Environ.* **1984**, *7*, 661–668. [CrossRef]
74. Moreno-Gutiérrez, C.; Dawson, T.E.; Nicolás, E.; Querejeta, J.I. Isotopes reveal contrasting water use strategies among coexisting plant species in a mediterranean ecosystem. *New Phytol.* **2012**, *196*, 489–496. [CrossRef]
75. Zhao, C.; Liu, Q. Growth and photosynthetic responses of two coniferous species to experimental warming and nitrogen fertilization. *Can. J. For. Res.* **2009**, *39*, 1–11. [CrossRef]
76. Reich, P.B.; Sendall, K.M.; Stefanski, A.; Rich, R.L.; Hobbie, S.E.; Montgomery, R.A. Effects of climate warming on photosynthesis in boreal tree species depend on soil moisture. *Nature* **2018**, *562*, 263–267. [CrossRef]
77. Loik, M.E.; Redar, S.P.; Harte, J. Photosynthetic responses to a climate-warming manipulation for contrasting meadow species in the rocky mountains, colorado, USA. *Funct. Ecol.* **2000**, *14*, 166–175. [CrossRef]
78. Elvira, S.; Alonso, R.; Inclan, R.; Bermejo, V.; Castillo, F.J.; Gimeno, B.S. Ozone effects on Aleppo pine seedlings (*Pinus halepensis* Mill.) grown in open-top chambers. *Water Air Soil Pollut.* **1995**, *85*, 1387–1392. [CrossRef]

79. Ribas, À.; Peñuelas, J.; Elvira, S.; Gimeno, B.S. Contrasting effects of ozone under different water supplies in two Mediterranean tree species. *Atmos. Environ.* **2005**, *39*, 685–693. [[CrossRef](#)]
80. Pleijel, H.; Pihl Karlsson, G.; Sild, E.; Danielsson, H.; Skärby, L.; Selldén, G. Exposure of a grass-clover mixture to ozone in open-top chambers—Effects on yield, quality and botanical composition. *Agric. Ecosyst. Environ.* **1996**, *59*, 55–62. [[CrossRef](#)]
81. Sanz, J.; González-Fernández, I.; Calvete-Sogo, H.; Lin, J.S.; Alonso, R.; Muntifering, R.; Bermejo, V. Ozone and nitrogen effects on yield and nutritive quality of the annual legume *Trifolium cherleri*. *Atmos. Environ.* **2014**, *94*, 765–772. [[CrossRef](#)]
82. Frey, B.; Scheidegger, C.; Günthardt-Goerg, M.S.; Matyssek, R. The effects of ozone and nutrient supply on stomatal response in birch (*Betula pendula*) leaves as determined by digital image-analysis and X-ray microanalysis. *New Phytol.* **1996**, *132*, 135–143. [[CrossRef](#)] [[PubMed](#)]
83. Elvira, S.; Sanz, J.; Gonzalez-fernandez, I.; Bermejo-bermejo, V. Ozone Effects on the Quality of Swiss Chard. Peri-Urban Crops a Case Study †. 2021. *Biol. Life Sci. Forum* **2021**, *11*, 16.
84. González-Fernández, I.; Bass, D.; Muntifering, R.; Mills, G.; Barnes, J. Impacts of ozone pollution on productivity and forage quality of grass/clover swards. *Atmos. Environ.* **2008**, *42*, 8755–8769. [[CrossRef](#)]
85. Jonasson, S.; Michelsen, A.; Schmidt, I.K.; Nielsen, E.V. Responses in microbes and plants to changed temperature, nutrient, and light regimes in the arctic. *Ecology* **1999**, *80*, 1828–1843. [[CrossRef](#)]
86. Martins, L.D.; Tomaz, M.A.; Lidon, F.C.; DaMatta, F.M.; Ramalho, J.C. Combined effects of elevated [CO₂] and high temperature on leaf mineral balance in *Coffea* spp. plants. *Clim. Chang.* **2014**, *126*, 365–379. [[CrossRef](#)]
87. Sardans, J.; Peñuelas, J.; Prieto, P.; Estiarte, M. Changes in Ca, Fe, Mg, Mo, Na, and S content in a Mediterranean shrubland under warming and drought. *J. Geophys. Res. Biogeosci.* **2008**, *113*, 1–11. [[CrossRef](#)]
88. Greenwood, D.J.; Lemaire, G.; Gosse, G.; Cruz, P.; Draycott, A.; Neeteson, J.J. Decline in percentage N of C3 and C4 crops with increasing plant mass. *Ann. Bot.* **1990**, *66*, 425–436. [[CrossRef](#)]
89. Jarrell, W.M.; Beverly, R.B. The dilution effect in plant nutrition studies. *Adv. Agron.* **1981**, *34*, 197–224. [[CrossRef](#)]
90. Hawkesford, M.; Horst, W.; Kichey, T.; Lambers, H.; Schjoerring, J.; Møller, I.S.; White, P. Functions of Macronutrients. In *Marschner's Mineral Nutrition of Higher Plants*, 3rd ed.; Marschner, P., Ed.; Academic Press: Cambridge, MA, USA, 2012; pp. 135–189. ISBN 9780123849052.
91. Broadley, M.; Brown, P.; Cakmak, I.; Rengel, Z.; Zhao, F. Function of Nutrients: Micronutrients. In *Marschner's Mineral Nutrition of Higher Plants*, 3rd ed.; Marschner, P., Ed.; Academic Press: Cambridge, MA, USA, 2012; pp. 191–248. ISBN 9780123849052.
92. Larcher, W. *Physiological Plant Ecology: Ecophysiology and Stress Physiology of Functional Groups*; Springer: Berlin/Heidelberg, Germany, 2003; ISBN 3-540-43516-6.
93. Welch, R.M.; Shuman, L. Micronutrient Nutrition of Plants. In *Critical Reviews in Plant Sciences*; Taylor and Francis: Oxfordshire, UK, 2014; Volume 14, pp. 37–41.
94. Fangmeier, A.; Grüters, U.; Högy, P.; Vermehren, B.; Jäger, H.J. Effects of elevated CO₂, nitrogen supply and tropospheric ozone on spring wheat - II. Nutrients (N, P, K, S, Ca, Mg, Fe, Mn, Zn). *Environ. Pollut.* **1997**, *96*, 43–59. [[CrossRef](#)] [[PubMed](#)]
95. Heck, W.W.; Taylor, O.C.; Adams, R.; Bingham, G.; Miller, J.; Preston, E.; Weinstein, L. Assessment of Crop Loss from Ozone. *J. Air Pollut. Control Assoc.* **1982**, *32*, 353–361. [[CrossRef](#)]
96. CLRTAP. *Manual on Methodologies and Criteria for Modelling and Mapping Critical Loads & Levels and Air Pollution Effects, Risks and Trends*; Umweltbundesamt: Berlin, Germany, 2004.
97. Abràmoff, M.D.; Magalhães, P.J.; Ram, S.J. Image Processing with ImageJ. *Biophotonics Int.* **2004**, *11*, 36–42.
98. Schad, D.J.; Vasisht, S.; Hohenstein, S.; Kliegl, R. How to capitalize on a priori contrasts in linear (mixed) models: A tutorial. *J. Mem. Lang.* **2020**, *110*, 104038. [[CrossRef](#)]
99. Pinheiro, J.; Bates, D.; DebRoy, S.; Sarkar, D.; Team, R.C. nlme: Linear and Nonlinear Mixed Effects Models. R package version 3.1-155 2022. Available online: <https://cran.r-project.org/web/packages/nlme/nlme.pdf> (accessed on 10 June 2021).
100. Core Team R. *R: A Language and Environment for Statistical Computing*; R Foundation for Statistical Computing: Vienna, Austria, 2022.
101. Clarke, K.R.; Gorley, R.N. *PRIMER v6: User Manual/Tutorial*; PRIMET-E: Plymouth, UK, 2006.
102. Martinez Arbizu, P. pairwiseAdonis: Pairwise multilevel comparison using adonis. *R Packag. version 0.4*. 2020, 1. Available online: <https://github.com/pmartinezarbizu/pairwiseAdonis> (accessed on 28 November 2022).
103. Oksanen, J.; Kindt, R.; Legendre, P.; O'Hara, B.; Stevens, H.H. Vegan: Community ecology package. R package version 2.5-7. *Cran. R-Project. Org/I.* 2020. Available online: <https://cran.r-project.org/web/packages/vegan/index.html> (accessed on 10 June 2021).

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