

Article Application of Phytosociological Information in the Evaluation of the Management of Protected Areas

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Abstract: The classification system of plant communities using phytosociological methods can be applied to their conservation in protected areas, as well as in establishing adequate protections and granting legal status to such areas. A new integrative index is developed to classify plant communities for the evaluation of the conservation status of protected areas, obtained from the product of three statistical indices of diversity: Syntaxonomic Distinctness, Rarefaction and Areas Prioritisation, which has been named DRA (acronym of the three indices used). The DRA is used to assess whether the status granted to Protected Areas matches the values provided by the plant communities within them and which were the basis for the identification and description of the Habitats of Community Interest (Habitats Directive—92/43/CEE). The proposed method was applied to the network of protected natural areas on the Andalusian coast, including 14 areas with different protection status, where, once the plant communities they contain were identified, the DRA index was applied to each of them and compared with the Legal Protection Index, i.e., the current protection regime; it becomes clear, objectively, that not all the statuses assigned, whether the IUCN criteria or those of the Andalusian government, correspond to the real levels of protection they should have on the basis of their plant communities.

Keywords: plant communities; syntaxonomic distinctness; rarefaction; conservation priority; legal protection; protected area

1. Introduction

The analysis of landscape from the phytosociological point of view is a valuable tool for its comprehensive study, including its dynamism and heritage value. The common methodologies used in phytosociology are considered as an optimal choice in environmental management assessments of habitats, as has been recognized for decades [1–8]. It is based on floristic inventories of homogeneous areas and the evaluation of the taxa present according to their abundance and dominance. This method has proven very useful in obtaining knowledge on vegetation and its dynamics over increasingly large territories [9,10]. Despite the implicit subjectivity of such information, the enormous amount that has accumulated over the past century is currently viewed as an extraordinary database susceptible to statistical and multivariate analysis, using the inventory as a working unit. Researches have shown that these observations can be treated with a high degree of confidence [5,11,12]. The use of taxa and plant communities as indicators in land-use planning and their application in natural environment conservation policies is accepted in several countries, insofar as they are in themselves the object of such protection [13,14]. Some studies have applied the information contained in the study of vegetation (habitats) and their cartographic representation to territorial biological assessment criteria for natural areas [15–17]. Loidi (2008) [18] estimated the environmental or naturalistic value of natural



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Copyright: © 2023 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (https:// creativecommons.org/licenses/by/ 4.0/). spaces, for which he proposed an evaluation methodology based on the application of indices that respond to the following premises: they have to be elaborated with analytical criteria, be quantifiable and accepted by the scientific community, as well as using fundamental aspects such as naturalness, resilience, threat, floristic-phytocoenotic value and rarity. Other complementary criteria would represent the services that terrestrial plant communities provide to ecosystems and human societies, such as protection of the soil or water resources, as well as a coefficient that considers those ecosystems located near densely populated sites and that reflects a priority in their conservation. The sum of each of these fundamental values gives the biological value of each site. Additionally, each of the complementary criteria used as factors can be used to obtain the conservation interest of the site.

Bio-evaluation proposals and the application of indices of naturalness, risks and patrimonial value have existed since the 1980s. Lalanne, Bioret and Boullet (2016) [8], compiled them in their publication, which was additionally a tribute to Professor Jean-Marie Géhu, pioneer in the evaluation of the vegetation across the landscape as a characterization of the territory [19].

In the present study, we extended these management tools and assessed the management of protected areas by applying a new index, which we named DRA (acronym for Distinctiveness, Rarefaction and Area prioritization), and questioned whether the range of protection granted is in accordance with the conservation interest of the plant communities that these areas hold. Thus, we used three indices to obtain the DRA value: (1) habitat distinction, an estimator of syntaxonomic richness that can be used to measure impacts on the phytocoenotic diversity of a territory [20], in a way that greater richness and diversity of plant communities implies higher value of distinction; (2) degree of habitat rarefaction, used to compare habitat richness between two protected areas of different sizes [21]. This reduced the samples to a standard size, i.e., interpolating to the same number of plant communities: that of the protected area with the lowest abundance [22,23], thus making it possible to compare the conservation status of different natural areas with different surface areas, using the diversity and uniqueness of their plant communities; and (3) the level of prioritization of the protective measures which highlights those elements that are of greater importance for granting a level of protection to a natural area, considering the anthropic impacts on them [24]. This would allow for the checking of the degree of importance of vegetation in the assignment of a category of legal protection. More specifically, we assessed whether the status granted to these natural spaces is supported by the DRA values obtained for the plant communities they include and the values of the three described indices.

Thus, while the habitat distinction strengthens the exclusivity of certain plant communities of conservation interest, the degree of rarefaction complements it by incorporating the valuation of those elements that, being of high interest, are also less frequent (scarce). Finally, the level of prioritisation in the conservation of the area allows the identification of those that stand out because they incorporate the greatest number of communities. All of these considerations were integrated into the new proposed index (DRA), which therefore provides all the necessary information to question whether there is coherence with level of protection assigned and their plant communities value.

The usefulness and functionality of the DRA index were put to the test in the network of protected natural areas along the Andalusian coast.

2. Results

2.1. Syntaxonomic Distinctness (STD)

Based on the STD values obtained (Table S1), four quartiles were differentiated, distinguishing the first and the last, and integrating in a single group the intermediate ones, ordered from lowest to highest, including the expected probability ranges. This allowed for the distinguishing of three types of protected areas (PAs) (Figure 1): (a) Enebrales de Punta Umbría Natural Site (EPU), Tómbolo de Trafalgar (TTR), and Dunas de Artola (DUA) Natural Monuments. Lower values ranged between 4.6 and 4.7, and wide ranges between 4.4 and 5, the STD value below the median (dashed horizontal line); (b) Doñana National Park (DOÑ), Breña y Marismas del Barbate (BMB), Bahía de Cádiz (BCA) and Estrecho (EST) Natural Parks and Marismas del Odiel (MOD), Marismas de Isla Cristina (MIC), Marismas del Río Piedras y Flecha del Rompido (MRF) and Punta Entinas-Sabinar (PES) Natural Sites, with intermediate values between 4.74 and 4.84 and narrow ranges between 4.6 and 4.95; and (c) Cabo de Gata-Níjar Natural Park (CGN) and Desembocadura del Guadalhorce (DEG) and Acantilados de Maro-Cerro Gordo (AMC) Natural Sites, with very high distinction values between 4.87 and 5, very wide ranges between 4.3 and 5, and an STD value above the median. The communities found in CGN are grouped in a few syntaxons of higher hierarchical rank (order or alliance), hence the reduction in the probability rank.



Figure 1. STD values (points) and the predicted interval between the maximum and minimum range (vertical lines) for each PA. AMC, Acantilados de Maro-Cerro Gordo; BCA, Bahía de Cadiz; BMB, La Breña y Marismas del Barbate; CGN, Cabo de Gata-Níjar; DEG, Desembocadura del Guadalhorce; DOÑ, Doñana; DUA, Dunas de Artola; EPU, Enebrales de punta Umbría; EST, Estrecho; MIC, Marismas de Isla Cristina; MOD, Marismas del Odiel; MRF, Marismas del Río Piedras y Flecha del Rompido; PES, Punta Entinas-Sabinar; TTR, Tómbolo de Trafalgar.

2.2. Rarefaction Index (RRF)

The results for the RRF index (Table S1) rendered four groups (ranges), considering the mean and the standard deviation (Figure 2). The first range contained AMC and DEG, which had a low RRF and few communities; the second range contained TTR, DUA and EPU, which had a moderate RRF and also consisted of few communities; and the third range contained EST, which had a high RRF but few communities, and BMB, MIC, PES, CGN, MOD and MRF, which also had a high RRF as well as many communities. Finally, the fourth range contained BCA and DOÑ, which had a very high RRF and many communities.

2.3. Priority Conservation Areas (PCA)

In the distribution (straight line) of PAs by Priority Conservation values (PCA; Table S1) and number of communities (Figure 3), four groups of PAs were distinguished: DOÑ and BCA, with the highest PCA values and many plant communities; CGN and PES, which had relatively high values; MIC, MRF, EST, DUA, and AMC, which had moderate PCA values and small surface areas. BMB, DEG, EPU, MOD, and TTR presented PCA values of 0 because their plant communities are represented in other PAs; i.e., the 16 BMB and eight TTR communities were also represented in BCA (23); the 10 EPU and 21 MOD communities



are also present in DON (23); DEG is a very small area, with only six communities, which are also present in many of the other PAs.

Figure 2. Ranking from the lowest to the highest rarefaction (RRF) value and number of communities in the Protected Areas. AMC, Acantilados de Maro-Cerro Gordo; BCA, Bahía de Cádiz; BMB, La Breña y Marismas de Barbate; CGN, Cabo de Gata-Níjar; DEG, Desembocadura de Guadalhorce; DOÑ, Doñana; DUA, Dunas de Artola; EPU, Enebrales de Punta Umbría; EST, Estrecho; MIC, Marismas de Isla Cristina; MOD, Marismas del Odiel; MRF, Marismas del Río Piedras y Flecha del Rompido; PES, Punta Entinas-Sabinar; TTR, Tómbolo de Trafalgar.



Figure 3. Distribution of Protected Areas (PAs) based on the values of the conservation priority index (PCA) and the number of communities (No. COM). PCmax, Highest priority; PCalt, High priority; PCmod, Moderate priority; AMC, Acantilados de Maro-Cerro Gordo; BCA, Bahía de Cadiz; CGN, Cabo de Gata-Nijar; DOÑ, Doñana; DUA, Dunas de Artola; EST, Estrecho; MIC, Marismas de Isla Cristina; MRF, Marismas de Río Piedra and Flecha del Rompido; PES, Punta Entina-Sabinar.

Principal component analysis of the STD, RRF and PCA showed that component 1 has a variance of 91.8% and that rarefaction is the variable with the highest weight (Figure 4 and Table 1). Five groups of PAs were clearly distinguished. Group A included DOÑ, BCA (closely linked by its communities to DOÑ) and CGN, in addition to PES (closely linked by its communities to CGN). Group B included EST and BMB, while Group C included MFR,

MIC, and MOD, all of which had lower RRFs and STD, as they are closely linked to DOÑ. Group D included TTR and DUA, in addition to EPU. Finally, group E included AMC and DEG, both with a low number of communities (five and six, respectively), although in the case of AMC those present are almost exclusive to this area, hence its high STD and low RRF. In the case of DEG, the few communities present are found almost throughout the entire study area.



Figure 4. Principal Component Analysis of STD, RRF and PCA. National Parks are shown in red, Natural Parks in orange, Natural Areas in yellow, and Natural Monuments in green. Protected Areas considered Biosphere Reserves are shown in squares.

Table 1. Principal components analysis result	ts
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РС	Eigenvalue	% Variance		PC1
1	0.287115	91.78	STD	-0.0061
2	0.018848	6.02	RRF	0.9869
3	0.006894	2.20	CPA	0.1610

2.4. Correlation of the Legal Protection Index and DRA

The integration of the three indices into the new index DRA (acronym of STD, RRF and PCA) is shown in Table S1. We found a high correlation between the DRA and Legal Protection Index (LPI), which refers to the current protection status of a protected area (Section S1 and Table S1). However, it is worth noting the higher weight of the PCA and RRF with respect to it, with the STD being undervalued with the lowest correlation value (Table 2).

Table 2. Spearman correlation values.

	LPI
DRA	0.82
STD	0.50
RRF	0.72
PCA	0.74

The LPI and DRA values for each AP were distributed along a second-order polynomial trend line with R2 = 0.82 (Figure 5). The values for LPI, STD, PCA and DRA are shown in Table S1.



Figure 5. The Legal Protection Index [24] values compared to DRA. National Parks are shown in red, Natural Parks in orange, Natural Areas in yellow, and Natural Monuments in green. Protected Areas considered Biosphere Reserves are shown in squares.

The residuals obtained for each PA after fitting the curve obtained by comparing the LPI with the DRA allowed three categories to be distinguished, based on their absolute values and their graphical interpretation (Figure 6):

- (1) those residuals whose absolute value is >1 (red bars) corresponded to PAs with a significant divergence between LPI and DRA (DEG, BCA, BMB and EPU);
- a central range with absolute values between 0–0.50 (green bars), corresponding to those PAs where LPI and DRA showed a significant convexity (DOÑ, AMC, MRF and MIC);
- (3) finally, those with absolute values between 0.51–1 (orange bars), which required a specific interpretation through the individual analysis of the ARD components (CGN, DUA, PES, MOD, EST and TTR).



Figure 6. Plot of the residuals obtained for each PA after adjusting the curve obtained by comparing the values of the LPI with those of the DRA.

When clustering (Figure 7), the Phenon Line drawn at 80% similarity configured four groups, identified on the basis of DRA values, and consequently assignable to IUCN categories and/or those declared by the regional government of Andalusia (Spain) through the Andalusian Network of Natural Spaces [25].



Figure 7. Comparison of the values of the Legal Protection Index (LPI) and the ARD Index proposed and the clusters derived from DRA values. Framed in red are those PAs whose assigned categories do not correspond to the DRA values obtained: EPU, EST, BMB and PES.

3. Discussion

The World Commission on Protected Areas (IUCN-WCPA) [26], defined "protected area" as: A clearly defined geographical space, dedicated and managed, through legal or other effective means, to achieve the long-term conservation of nature with associated ecosystem services and cultural values, while specifying among other principles that the fundamental objective should be the conservation of nature, as well as maintaining or even increasing the degree of naturalness of the ecosystem being protected. It also establishes seven categories in a non-hierarchical system, to be applied in the context of national protected area systems and as part of the ecosystem approach: Ia-Strict nature reserve; Ib-Wilderness area; II-National park; III-Natural monument or feature; IV Habitat/species management area; V—Protected landscape; VI—Protected areas with sustainable use of natural resources. Each of these categories was described on the basis of a number of criteria (distinctive values, role in the landscape or uniqueness), although it acknowledges that the assignment to any given category depends more on the intended use by the management authority rather than on any fixed and unchangeable set of criteria. Such categorization is important because it works as the basis of the definition of the management objectives. This is even more the casewhen the allocation of categories has traditionally been the responsibility of governments, which has led to this situation being questioned. In any case, the IUCN's own Guidelines for the Application of Protected Area Management Categories [26] recommend that "The only principle that should be applied in assigning categories is the appropriateness of the management objective assigned to the protected area within the system in relation to the ecological needs and threats to the species or ecosystem, in the context of the whole territory or marine environment in which the biodiversity occurs".

More recently, the criteria for determining the areas to be protected and their categorization evolved and advanced with the help of conservation biology, technological tools such as geographic information systems and statistical processing based on biodiversity databases. The quality and beauty of the landscapes are no longer the only criteria for the selection of an area; the representativeness and complementarity that a reserve offers for the protection of biodiversity are now incorporated [27].

The method of inventorying the plant communities should integrate broad ecosystemic information, and it should allow for hierarchization using the vegetation association as a basic unit. Such a hierarchy would include the association in successive higher levels of alliances, orders and classes that would allow a dynamic definition of the territory. For this, we propose an integrated index (DRA). This cumulative index strengthens three variables for the proposal of PA management:

(1) Distinctness, understood as a measure of the diversity of a mosaic of ecosystems that distinguish a PA. It is increasingly recognized that appropriate measures of biodiversity within a particular taxonomic group should not merely be functions of the number of species and their relative abundance, but should also include information on the taxonomic relatedness of the species in question, which has led to the development of taxonomic distinctiveness (TD) [28]. On the same biological basis, syntaxonomy provides, through the hierarchization of communities and their treatment by the same method (STD), complementary information on the relationships of habitat diversity.

(2) Rarefaction, understood as the degree of rarity considering the extent of the PA. One of the challenges in determining the category of protection to be assigned to a territory is to resolve the size dependence and compare richness when the number of plant communities is not equal. To achieve this, one of the possible options [29] is the interpolation of species richness using the rarefaction technique, based on the shape of the species-abundance curve [30,31]. This traditional rarefaction was used to calculate the expected number of plant communities by reducing the samples to a standard size, i.e., interpolating into the same number of communities those areas with the lowest abundance [22].

(3) Prioritization, or understanding that the richness of communities confers a selective value in the process of choosing categories.

It should be noted that according to the results obtained in the clustering, in more than 70% of the cases there is an exact allocation of the categories with the values obtained from the DRA. Only four units differ in the value of the assigned category: PES is a Protected Site declared by the Regional Government, but its geographical proximity and degree of similarity in the cluster imply that it should be integrated as a Natural Park together with CGN, from which it is separated by the city of Almeria. Both BMB and EST are declared as Natural Parks; however, their coastal communities make them more similar to Natural Reserves. Finally, the low STD of EPU led to its integration as a Natural Monument, rather than a Natural Reserve as it appears in the RENPA.

The results of this study confirm that the use of the phytosociological method to describe plant communities in territories provides valuable information for the management of areas and their protection. Thus, statistical analysis using complementary diversity indices—such as STD, RRF and ACP—can be used to compare the status granted to PAs in territorial planning, using a new integrative index (DRA).

4. Materials and Methods

4.1. Materials

We selected the same 14 protected areas chosen by Salvo Tierra et al. (2020) [24] (Figure 8). These areas are included in the Andalusian Network of Natural Spaces. In line with the findings of Pereña (2018) [25], we created a syntaxonomic scheme of the set of plant communities observed that included their corresponding alliance, order, and class (Appendix A). Based on this information, we constructed a Basic Data Matrix (BDM) of the plant communities in each protected area (Table S2), and assigned an identifying abbreviation to each syntaxon.



Figure 8. Location and biogeographical location of the Protected Areas studied on the Andalusian coast (Spain). AMC, Acantilados de Maro-Cerro Gordo; BCA, Bahía de Cádiz; BMB, La Breña y Marismas del Barbate; CGN, Cabo de Gata-Níjar; DEG, Desembocadura del Guadalhorce; DOÑ, Doñana; DUA, Dunas de Artola; EPU, Enebrales de punta Umbría; EST, Estrecho; MIC, Marismas de Isla Cristina; MOD, Marismas del Odiel; MRF, Marismas del Río Piedras y Flecha del Rompido; PES, Punta Entinas-Sabinar; TTR, Tómbolo de Trafalgar.

Salvo Tierra et al. (2020) [24] calculated the Legal Protection Index (LPI) using a methodology based on the Protection Overlap Value. This methodology uses: (1) a protection status scale that is defined by the responsible administrative body and that overlaps in the same PA; and (2) the Legal Rigor, based on the number of overlapping protection regulations (Natural Resources Management Plans) within the same PA and their planning statutes (Master Plans for Use and Management).

Based on the results of the foregoing, numerical values were obtained for each Protected Area (Table 3).

4.2. Method

We used three statistical diversity indices to determine whether certain areas have been granted their correct level of protection, based on their plant communities: syntaxonomic distinctiveness (STD), rarefaction (RRF) and level of prioritization of conservation areas (PCA). These were multiplied in order to obtain the DRA index. This was applied to the protected areas on the Andalusian coast and compared with their current category of protection.

So far, these three indices have always been applied to conventional taxonomy, but in this case, we further applied them to syntaxonomy because just as a taxon carries a pool of ecological, biogeographical, evolutionary, etc. information, syntaxons provide equally significant information. In this case, this is relevant because they are communities of plant species that converge in the same territory with identical ecological conditions.

	DESIGNATED STATUS	AMC	BCA	BMB	CGN	DEG	DOÑ	DUA	EPU	EST	MFR	MIC	MOD	PES	TTR
	National Park						+								
	Natural Park		+	+	+		+			+					
Designated status at regional — or national level	Natural Area	+				+			+		+	+	+	+	
	Natural Monument							+							+
	Natural Reserve													+	
Decision at a distance has FU	Special Conservation Area	+	+	+	+		+		+	+	+	+	+	+	+
Designated status by EU -	Special Protection Area for Birds	+	+	+	+		+			+	+	+	+	+	
	Biosphere Reserve				+		+			+			+		
-	RAMSAR site		+		+		+						+	+	
Designated status at international level	Specially Protected Areas of Mediterranean Importance	+			+										
	Geopark				+										
	World Heritage Site						+								
	Legal Protection Index	3.62	5.27	3.85	6.83	0.38	10	0.2	1.22	4.64	2.8	2.74	4.39	5.29	0.69

 Table 3. Protected status of each Protected Area studied and Legal Protection Index value [24].

AMC: Acantilados de Maro-Cerro Gordo; Bahía de Cádiz; La Brea y Marismas del Barbate; Cabo de Gata-Níjar; Desembocadura del Guadalhorce; Doñana; Dunas de Artola; Enebrales de Punta Umbría; Estrecho; Marismas del Río Piedras y Flecha del Rompido; Marismas de Isla Cristina; Marismas del Odiel; Punta Entinas-Sabinar; Tómbolo de Trafalgar. Note: + indicates that the protected area has received this status.

4.3. Syntaxonomic Distinctness (STD)

Taxonomic structures are based on the number of taxa included at the highest levels of the taxonomic hierarchy therefore, in the case of using species as a unit, the higher hierarchical levels would be genus, family, order, subclass/class. They have been widely used in ecology to measure impacts on ecosystem diversity [20,28,32–35]. Among the estimators of taxonomic richness, taxonomic distinctness (TD [35] and references therein) is particularly significant because it is less influenced by sample size than by species diversity. Moreover, TD may be a more sensitive univariate index of community disturbance than species diversity [20]. Syntaxonomic structures provide relevant information, and thus have the same analytical potential as TD [36] for measuring impacts on territorial diversity.

Clarke and Warwick (1998) [28] proposed the Δ + index (Delta-plus) to estimate TD. We extend this index to the syntaxonomic distinction (STD) of PAs. This index measures the overall average length of the syntaxonomic path between two randomly chosen PAs.

The Δ + index for STD would be calculated according to the following equation:

$$\Delta + = \left[\sum_{i < j} w_{ij}\right] / \left[s(s-1)\right] \tag{1}$$

where *s* is the number of phytosociological associations observed in a given PA, and w_{ij} is the weight or distinction value given to each syntaxonomic branch of the hierarchical classification from association *i* to the first common node *j* (alliance), and so on up to the third and fourth level (order and class).

4.4. Rarefaction (RRF)

The rarefaction (RRF) index was used to compare observed diversity between locations that have not been sampled equally [16]. To make a fair comparison of these sites, subsamples (i.e., communities) were drawn from the larger sample, and the expected richness in each unit (i.e., PAs) was calculated based on abundance distributions in the larger sample. The process was repeated for subsamples of different sizes. The final value of the RRF index showed the expected value of community richness according to the sample size of each PA [21]. This correction allows for the direct comparison of the richness of two samples that were originally of different sizes.

Hurlbert (1971) [31] proposed the following algorithm to determine the RRF:

$$Sn = \sum_{i=1}^{S} (1 - qi)$$
 (2)

where $qi = \frac{\left(\frac{n-2i}{n}\right)}{\left(\frac{N}{n}\right)}$ would represent the probability that community *i* does not appear in a sample of size *n*, *z*_i is the number of *i* communities, and (*N*/*n*) is the binomial coefficient or the number of ways in which n out of N can be chosen.

This algorithm was calculated by applying an analytical solution known as "Mao's tau", in which standard deviations are converted to 95% confidence intervals [37].

4.5. Priority Conservation Areas (PCA)

The method proposed by Vane-Wright et al. (1993) [38] was used to determine the value of the PCAs. In a first cycle, the PAs with the most communities were selected. Subsequently, the communities within the PA were eliminated from the BDM. This procedure was repeated in successive cycles that included the remaining communities that had not been included in the PAs already selected. The values obtained in each cycle refer to the total number of communities in the study area.

With the values of these three variables (STD, RRF and ACP), a Principal Component Analysis (using the variance–covariance matrix method) was implemented to check the influence of each one of them on the different protected areas, based on their vegetation communities, as well as the one that presented a higher level of incidence in the way these areas are grouped. The calculations of these three indices were carried out using PAST v. 4.11 software [39,40].

4.6. Calculation of the DRA Index

The proposed DRA index was established as a tool for assessing the suitability of the assignment of protection categories to natural areas. For this reason, it was calculated by means of the product of the values obtained and subsequently standardized from the DST, RRF and PCA indices, all of which were independent. In this way, the range of variation of the values obtained was adjusted to a similar range to that obtained for the LPI (Table S1).

$$DRA = DST_{st} * RRF_{st} * PCA_{st}$$
(3)

The values obtained from the calculation of the three indices and DRA are correlated (Spearman's non-parametric correlation) with the LPI, which is the reference index for the current protection status.

DRA values were compared with those of the LPI (Table S1) to check how they are distributed along the trend lines of both. This was done with excel V. 2204. With the residues obtained, divergences between the DRA and the LPI in the different protected areas were checked using the "confidence tunnel" method [28].

Finally, in order to test the functionality of the new DRA index, a clustering of the PAs studied was carried out on the basis of the three components of the index and the current protection categories (LPI), taking into account the IUCN categories for the protection of areas and that of the Andalusian Network of Protected Spaces, and drawing a Fenon Line at 80% similarity. This was calculated with the PAST 4.1.1 software using the Euclidean distance and the WPGMA algorithm (co-phenetic correlation of 0.87).

Supplementary Materials: The following supporting information can be downloaded at: https: //www.mdpi.com/article/10.3390/plants12020406/s1, Table S1: Values of all calculated indices; Table S2: Distribution of communities by protected areas.

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Appendix A. Syntaxonomic Scheme of the Protected Areas [25]. Abbreviations of Plant Communities

Cakiletea maritimae Tüxen & Preising ex Br.-Bl. & Tüxen 1952 Cakmar
 Cakiletalia integrifoliae Tüxen ex Oberdorfer 1949 corr. Rivas-Martínez, Costa & Loidi 1992 Caklia
 Cakilion maritimae Pignatti 1953 Cakion
 Hypochoerido radicatae-Glaucietum flavi Rivas Goday & Rivas-Martínez 1958 HypGla
 Salsolo kali-Cakiletum maritimae Costa & Mansanet 1981 SalCak
 Cisto-Lavanduletea stoechadis Br.-Bl. in Br.-Bl., Molinier & Wagner 1940 CisLav
 Stauracantho genistoidis-Halimietalia calycini Rivas-Martínez, Lousã, T.E. Díaz,
 Fernández-González & J.C. Costa 1990 StaHal
 Coremation albi Rothmaler 1943 Coralb
 Cisto salviifolii-Ulicetum australis A.V. Pérez, Nieto & Cabezudo 1993 CisSal

12 of 17

Fumano juniperinae-Cistetum crispi Sánchez & Galán 1996 FumCis

Halimio halimifolii-Stauracanthetum genistoidis Rivas-Martínez 1979 HalSta Thymo albicantis-Stauracanthetum genistoidis Galán, I. Sánchez & Vicente 1997 ThyAlb Crithmo maritimi-Limonietea Br.-Bl. in Br.-Bl., Roussine & Nègre 1952 Cretea Crithmo maritimi-Limonietalia Molinier 1934 Cralia Crithmo maritimi-Daucion halophili Rivas-Martínez, Lousã, T. E. Díaz, Fernández-González & J. C. Costa 1990 Cdnion Limonietum emarginati Asensi 1984 Limemar Crithmo maritimi-Limonion pseudominuti Molinier 1934Clnion Crithmo maritimi-Limonietum malacitani Diez Garretas 1977 corr. Diez Garretas 1981 CriLim Limonio cossoniani-Lycietum intricate Esteve 1976 corr. Alcaraz, P. Sánchez, De la Torre, Ríos & J. Alvarez 1991 LimLyc Cytisetea-Scopario striati Rivas-Martínez 1974 CytSco Cytisetalia scopario-striati Rivas-Martínez 1974 Cyalia Retamion monospermae Rivas-Martínez & Cantó 2002 Retmon Pycnocomono rutaefolii-Retametum monospermae Pérez Chiscano 1983 PycRet Euphorbio paraliae-Ammophiletea australis Géhu & Rivas-Martínez 2011 EupAmm Ammophiletalia australis Br.-Bl. 1933 Amalia Ammophilion australis Br.-Bl. 1921 Amlion Loto cretici-Ammophiletum australis Rivas-Martínez 1965 corr. Rivas-Martínez, T.E. Díaz, Fernández-González, Izco, Loidi, Lousã & Penas 2002 LotAmm Honckenyo peploidis-Elytrigion boreoatlanticae Tüxen in Br.-Bl. & Tüxen 1952 HonEly Cypero mucronati-Elytrigietum junceae Br.-Bl. 1933 CypEly Euphorbio paraliae-Elytrigietum boreoatlanticae Tüxen in Br.-Bl. & Tüxen 1952 EupEly Sporobolion arenarii (Géhu & Géhu-Franck ex Géhu & Biondi 1994) Rivas-Martínez & Cantó 2002 Spoare Eryngio maritimi-Sporoboletum arenarii (Arenes ex Géhu & Biondi 1994) Rivas-Martínez & Cantó 2002 ErySpo Sporoboletum arenarii Rothmaler 1943 Spoari Crucianelletalia maritimae Sissing 1974 Crulia Crucianellion maritimae Rivas Goday & Rivas-Martínez 1959 Cruion Loto cretici-Crucianelletum maritimae Alcaraz, T.E. Díaz, Rivas-Martínez & P. Sánchez 1989 LotCru Helichrysion picardii (Rivas-Martínez, Costa & Izco in Rivas-Martínez, Lousã, T.E. Díaz, Fernández-González & J.C. Costa 1990) Rivas-Martínez, Fernández-González & Loidi 1999 Hesion Artemisio crithmifoliae-Armerietum pungentis Rivas Goday & Rivas-Martínez 1959 ArtArm Halodulo wrightii-Thalassietea testudinum Den Hartog ex Rivas-Martínez, Fernández Gonzalez & Loidi 1999 HalTha Thalassio-Syringodietalia filiformis Borhidi, Muñiz & Del Risco in Borhidi 1996 ThaSyr Syringodio-Thalassion testudinum Borhidi 1996 SyrTha Cymodoceetum nodosae Feldmann 1937 Cymnod Juncetea maritimi Br.-Bl. in Br.-Bl., Roussine & Nègre 1952 Juetea Juncetalia maritimi Br.-Bl. in Br.-Bl., Roussine & Nègre 1952 Jualia Juncion maritimi Br.-Bl. in Br.-Bl., Roussine & Nègre 1952 Jucion Elymo elongati-Juncetum maritimi Alcaraz, Garre, Peinado & Martínez Parras 1986 ElyJun Polygono equisetiformis-Juncetum maritimi J. C. Costa in J. C. Costa, Lousã & Espírito Santo 1997 PolJun

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Magnocarici elatae-Phragmitetea australis Klika in Klika & V. Novák 1941 MagPhr

Bolboschoenetalia compacti Dahl & Hadac 1941 corr. Rivas-Martínez, Costa, Castroviejo & E. Valdés 1980 **Boalia**

Bolboschoenion compacti Dahl & Hadac 1941 corr. Rivas-Martínez, Costa, Castroviejo & E. Valdés 1980 Bolion

Bolboschoeno maritimi-Schoenoplectetum litoralis Br.-Bl. in Br.-Bl., Roussine & Nègre 1952 corr. Rivas-Martínez, Costa, Castroviejo & E. Valdés 1980 BolJun

Nerio-Tamaricetea Br.-Bl. & O. Bolòs 1958 NerTam

Tamaricetalia Br.-Bl. & O. Bolós 1958 Tamlia

Tamaricion africanae Br.-Bl. & O. Bolós 1958 Tamafr

Polygono equisetiformis-Tamaricetum africanae Rivas-Martínez & Costa in Rivas- Martínez, Costa, Castroviejo & E. Valdés 1980 **PolTam**

<i>Tamaricion boveano-canariensis</i> Izco, Fernández-González & A. Molina 1984 Tambov Inulo crithmoidis-Tamaricetum boveanae Izco, Fernández-González & A. Molina 1984 InuTam
Parietarietea judaicae Rivas-Martínez in Rivas Goday 1964 Paetea
Parietarietalia judaicae (Rivas-Martínez 1960) Rivas Goday 1964 Paalia
Lavaterion maritimae Rivas-Martínez & Cantó 2002 Lavmar
Rosmarinetum tomentosi F. Casas & M. López in F. Casas 1972 Rostom
Pegano harmalae-Salsoletea vermiculatae BrBl. & O. Bolos 1958. PegSal
Salsolo vermiculatae-Peganetalia harmalae BrBl. & O. Bolos 1954 SalPeg
Suisoio oppositijoitue-Suueuton motiis Rigual 1972 Salsua
1997 FraSal
Withanio frutescentis-Lucietum intricate Alcaraz P Sánchez De la Torre Ríos & I Alvárez
1991 WhiLvc
Posidonietea Den Hartog 1976 Poetea
Posidonietalia Den Hartog 1976 Poalia
Posidonion BrBl., Roussine & Nègre 1952 Ponion
Posidonietum oceanicae Funk 1927 Posoce
Quercetea ilicis BrBl. ex A. & O. Bolòs 1950 Quetea
Pistacio lentisci-Rhamnetalia alaterni Rivas-Martínez 1975 PisRha
Asparago albi-Rhamnion oleoidis Rivas Goday ex Rivas-Martínez 1975
AspRha
Cneoro tricocci-Buxetum balearicae Rivas Goday & Rivas-Martínez 1969 CneBux
Juniperion turbinatae Rivas-Martínez 1975 corr. Rivas-Martínez 1987 Juntur
Osyrio quadripartitae-Juniperetum turbinatae Rivas-Martinez ex Rivas-Martinez, Lousa, I.E.
Diaz, Fernandez-Gonzalez & J.C. Costa 1990 OsyJun
Chamaeropo numilis-Juniperetum navicularis Sanchez Garcia, Sanchez Guilon, Linares Perea &
Rhamno anoustifoliae. Junineretum turbinatae Rivas-Martínoz ex Freitae 1971 RanJun
Rhamno alegidis-Juniperetum macrocarnae Rivas-Martínez 1965 Rollun
Perinlocion angustifoliae Rivas-Martínez 1975 Perang
Zizinhetum loti Rivas Goday & Bellot 1944 Zizlot
Rubio longifoliae-Coremation albi Rivas-Martínez in Rivas-Martínez, Costa, Castroviejo & E.
Valdés 1980 Rution
Rubio longifoliae-Corematetum albi Rivas-Martínez in Rivas-Martínez, Costa, Castroviejo & E.
Valdés 1980 RubCor
Querco rotundifoliae-Oleion sylvestris Barbéro, Quézel & Rivas-Martínez in Rivas-Martínez,
Costa & Izco 1986 QueOle
Aro neglecti-Quercetum suberis Rivas-Martínez & Díez-Garretas 2011 AroQue
Rosmarinetea officinalis Rivas-Martínez, T.E. Díaz, F. Prieto, Loidi & Penas 2002 Rosoff
Anthyllidetalia terniflorae Rivas Goday, Rigual, Esteve, Borja & Rivas-Martínez in Rivas Goday &
Borja 1961 Antter
Inymo moroderi-Sideritiaion leucanthae O. Bolos 1957 corr. Alcaraz, I.E. Diaz, Rivas-Martinez &
P. Sanchez 1989 InySia
Cruz 1985 Tou Hol
Saginetea maririmae Westhoff Van Leeuwen & Adriani 1962 Sagmar
<i>Frankenietalia nulverulentae</i> Rivas-Martínez ex Castrovieio & Porta 1976 Fralia
Frankenion vulverulentae Rivas-Martínez ex Castroviejo & Porta 1976 Fraion
Parapholido incurvae-Frankenietum pulverulentae Rivas-Martínez ex Castroviejo & Porta 1976.
ParFra
Hordeion marini Ladero, F. Navarro, C. Valle, Marcos, Ruiz & M. T. Santos 1984 Hormar
Hainardio cylindricae-Lophochloetum hispidae Rivas-Martínez, Costa, Castroviejo & E. Valdés
1980 HaiLop
Sarlicornietea fruticosae BrBl. & Tüxen ex A. & O. Bolós 1950 Saetea
Limonietalia BrBl. & O. Bolós 1958 Lialia
Limoniastrion monopetali Pignatti 1953 Limmon
Polygono equisetiformis-Limoniastretum monopetali Rivas-Martínez & Costa in Rivas-Martínez,
Costa, Castroviejo & E. Valdes 1980 PolLim
Limonion confusi (drdl. 1955) Kivas-Martinez & Costa 1984 Limcon

Limonietum ferulacei Rothmaler 1943 Limfer Lygeo sparti-Limonion furfuracei Rigual 1972 LygLim Limonietum angustebracteato-delicatuli Rivas-Martínez & Alcaraz in Alcaraz 1984 Limang Arthrocnemo macrostachyi-Suaedetalia braun-blanquetii Rufo, Fuente & Sánchez Mata 2016 ArtSua Arthrocnemion glauci (Rivas-Martínez in Rivas-Martínez & al. 1980) Rivas-Martínez & Costa 1984 Artgla Arthrocnemo macrostachyi-Sarcocornietum hispanicae Fuente, Rufo, Teijeiro & Sánchez-Mata 2013 ArtSar Inulo crithmoidis-Arthrocnemetum macrostachyi Fontes ex Géhu & Géhu Franck 1977 InuArt Salicornietalia fruticosae Br.-Bl. 1933 Sarlia Sarlicornion fruticosae Br.-Bl. 1933 Sarion Limonio cossoniani-Sarcocornietum lagascae M.A. Alonso & De la Torre 2002 corr. Rufo et al. 2016 LimSar Sarcocornio pruinosae-Halimionetalia portulacoidis Rufo, Fuente & Sánchez Mata 2016 SarHal Halimionion portulacoidis Géhu 1976 Halpor Cistancho phelypaeae-Sarcocornietum pruinosae Géhu ex Géhu & Géhu-Franck 1977 corr. Rufo et al. 2016 CisSar Sarcocornio perennis-Puccinellietum ibericae J. C. Costa in J. C. Costa, Lousã & Espírito Santo 1997 corr. Rivas-Martínez, Fernández-González, Loidi, Lousã, Penas & Izco 2002 SarPuc Sarcocornion alpini (Rivas-Martínez et al. 1990) Brullo et al. 2002 Sarcon Halimiono portulacoidis-Sarcocornietum alpini Rivas-Martínez & Costa 1984 HalSar Sarcocornietum alpini Br.-Bl. 1933 corr. Rivas-Martínez, Lousã, T. E. Díaz, Fernández González & J. C. Costa 1990 Saralp Suaedion verae (Rivas-Martínez, Lousã, T. E. Díaz, Fernández González & J. C. Costa 1990) Rivas-Martínez, Fernández González & Loidi 1999 Suaver Cistancho phelypaeae-Suaedetum verae Géhu & Géhu-Franck 1977 CisSua Frankenio corymbosae-Suaedetum verae Alonso & De la Torre 2002 FraSua Spartinetea maritimae Tüxen in Beeftink 1962 Spetea Spartinetalia maritimae Conard 1935 Spalia Spartinion maritimae Beeftink & Géhu 1973 Spaion Spartinetum densiflorae Rivas-Martínez, Costa, Castroviejo & E. Valdés 1980 Spaden Spartinetum maritimae Béguinot ex Corillion 1953 Spamar Thero-Salicornietea Tüxen in Tüxen & Oberdorfer ex Géhu & Géhu-Frank 1984 Thetea Thero-Salicornietalia Tüxen in Tüxen & Oberdorfer ex Géhu & Géhu-Frank 1984 Tsalia Salicornion patulae Géhu & Géhu-Frank 1984 Salpat Suaedo spicatae-Salicornietum patulae Brullo & Furnari ex Géhu & Géhu-Franck 1984 corr. Alcaraz, Ríos, De la Torre, Delgado & Inocencio 1998 Suaspi Suaedo splendentis-Salicornietum patulae Rivas-Martínez, Costa, Castroviejo & E. Valdés 1980 corr. Rivas-Martínez 1991 Suasal Thero-Suaedetalia Br.-Bl. & O. Bolòs 1958 Sualia Thero-Suaedion Br.-Bl in Br.-Bl, Roussine & Nègre 1952 Theion Suaedo splendentis-Salsoletum sodae Br.-Bl. 1933 Salsod Tuberarietea guttatae (Br.-Bl. in Br.-Bl., Roussine & Nègre 1952) Rivas Goday & Rivas-Martínez 1963 em Rivas-Martínez 1978 Tubgut Cutandietalia maritimae Rivas-Martínez, Díez Garretas & Asensi 2002 Cutmar Alkanno-Maresion nanae Rivas Goday ex Rivas Goday & Rivas-Martínez 1963 corr. Diez-Garretas, Asensi & Rivas-Martínez 2001 AlkMar Wahlenbergio nutabundae-Loeflingietum pentandrae Alcaraz, Díez Garretas & Asensi in Ferre, Díez-Garretas & Asensi 1985 WalLoe Linarion pedunculatae Díez-Garretas, Asensi & Esteve in Díez-Garretas 1984 Linped Ononido variegatae-Linarietum pedunculatae Diez-Garretas, Asensi & Esteve ex Izco, P. & J. Guitián 1988 OnoLin Triplachno nitentis-Silenetum ramosissimae Peinado, Martínez Parras, Alcaraz, Garre & de La Cruz 1985 TriSil Malcolmietalia Rivas Goday 1958 Maalia Anthyllido hamosae-Malcolmion lacerae Rivas Goday 1958 em. Rivas-Martínez 1978 AntMal Linario donyanae-Loeflingietum baeticae Rivas-Martínez, Costa, Castroviejo & E. Valdés 1980 LinLoe

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