

Article



The Water Content Drives the Susceptibility of the Lichen *Evernia prunastri* and the Moss *Brachythecium* sp. to High Ozone Concentrations

Andrea Vannini ¹, Giulia Canali ¹, Mario Pica ², Cristina Nali ³ and Stefano Loppi ^{1,*}

- ¹ Department of Life Sciences, University of Siena and Italy, 53100 Siena, Italy; andrea.vannini@unisi.it (A.V.); giulia.canali@student.unisi.it (G.C.)
- ² Bioredox, 00187 Rome, Italy; bioredoxsrl@gmail.com
- ³ Department of Agriculture, Food and Environment, University of Pisa, 56124 Pisa, Italy; cristina.nali@unipi.it
- * Correspondence: stefano.loppi@unisi.it

Received: 13 March 2020; Accepted: 22 April 2020; Published: 27 April 2020

Abstract: The aim of this study was to evaluate the tolerance of lichens (*Evernia prunastri*) and mosses (*Brachythecium* sp.) to short-term (1 h), acute (1 ppm) O₃ fumigation under different hydration states (dry, <10% water content, metabolism almost inactive; wet, >200% water content, metabolism fully active). We hypothesized that stronger damage would occur following exposure under wet conditions. In addition, we checked for the effect of recovery (1 week) after the exposure. Ozone fumigation negatively affected the content of chlorophyll only in wet samples, but in the moss, such a difference was no longer evident after one week of recovery. Photosynthetic efficiency was always impaired by O₃ exposure, irrespective of the dry or wet state, and also after one week of recovery, but the effect was much stronger in wet samples. The antioxidant power was increased in wet moss and in dry lichen, while a decrease was found for wet lichens after 1 week. Our results confirm that the tolerance to O₃ of lichens and mosses may be determined by their low water content, which is the case during the peaks of O₃ occurring during the Mediterranean summer. The role of antioxidant power as a mechanism of resistance to high O₃ concentrations needs to be further investigated.

Keywords: antioxidants; chlorophyll; cryptogams; hydration state; photosynthesis

1. Introduction

Ozone (O₃) is a strongly oxidizing pollutant occurring at ground level as a consequence of the interaction between solar irradiation and gases such as nitrogen oxides (NOx), volatile organic compounds (VOCs), and carbon monoxide (CO). Considering its known phytotoxicity [1] and the global tendency of increasing background concentrations of about 0.3 ppb/y as a consequence of the global increases in temperature and precursors [2], investigations of the impact of O₃ on vegetation have been strongly prompted during the last few years [3], especially because this compound may affect plants' productivity [4].

Lichens, despite their wide use as bioindicators of air quality [5], are known for being rather insensitive to O₃ pollution, with lab and field studies showing similar results: a (very) limited influence of O₃ on lichen physiology and biodiversity. In detail, field studies did not find any evidence of a correlation between lichen biodiversity and O₃ concentrations, as evaluated indirectly by the damage occurring in the O₃ supersensitive plant species *Nicotiana tabacum* Bel-W3 during summer exposures [6–8], while O₃ fumigations at ecologically relevant concentrations under controlled conditions failed to cause relevant injuries to the photosynthetic systems of several lichen species [9–11]. These results suggested a very high tolerance of these organisms to environmental O₃

concentrations, with a toxicity threshold higher than the natural concentrations to which they are commonly exposed to.

However, upon increasing the concentrations of O_3 up to levels less or not at all ecologically relevant, the results were different. Fumigations carried out at ca. 1 ppm O_3 indicated contrasting results: a decrease in net photosynthesis in *Parmelia sulcata* [12] and no change in the chlorophyll content of *Cladonia arbuscula*, the photobiont of *Cladonia stellaris* [13] or the photosynthesis of *Cladonia rangiformis* [14]. However, acute fumigations at 3 ppm O_3 induced strong physiological and ultrastructural damage in both the photobiont and the mycobiont of the pollution-tolerant species *Xanthoria parietina* [15]. Nevertheless, these authors suggested that the hydration state may play a major role in determining the severity of the damage, thus explaining the ecological insensitivity of lichens to the high environmental levels of O_3 occurring during dry Mediterranean summers.

Similarly to lichens, a limited susceptibility to O₃ of bryophytes has been reported. Fumigations with 150 ppb O₃ for 5 h induced a slight decrease in photosynthesis only in one out of four *Sphagnum* species [16], while fumigations with 70–80 ppb O₃ for 6–9 weeks induced only a modest photosynthetic injury in *Sphagnum recurvum* compared with in *Polytrichum commune* [17]. In addition, chronic fumigations of *Sphagnum* in open top chambers did not induce reductions in the chlorophyll contents after exposure to 50, 100 and 150 ppb O₃ [18]. Nevertheless, fumigations carried out for 10 weeks (6 h/d, 4 d/w) with 240 and 320 ppb were shown to cause reductions in the abundance of four mosses species [19].

Poikilohydric organisms like lichens and mosses are assumed to have higher resistance to gaseous pollutants during the dry state, having a metabolism strictly dependent on their water content [20], and since during the summer periods of higher O₃ concentrations the metabolism of these cryptogams is largely reduced, they are likely to be less prone to being injured [6]. Nevertheless, information about the resistance of mosses to O₃ exposure during different hydration states is lacking and in need of clarification.

The aim of this study was to evaluate the tolerance of lichens and mosses to short-term acute O₃ fumigation under different hydration states. We hypothesized that stronger damage would occur following exposure under wet conditions. In addition, we checked for the effect of recovery after exposure.

2. Materials and Methods

2.1. Experimental

Samples of the lichen *Evernia prunastri* and the moss *Brachythecium* sp. were collected at the end of September 2019 in a remote area of the Siena province (Tuscany, Central Italy), far from any local source of pollution. These species were chosen considering their wide use as biological indicators as well as their ease of collection and handling. In the laboratory, samples were cleaned from extraneous material under a stereoscopic microscope using plastic tweezers. After cleaning, one batch of 18 samples was air-dried (residual water <10%) overnight in a climatic chamber at 16 °C and 55% relative humidity (RH), while another batch of 18 samples was fully hydrated (water content up to 250% of dry weight for the lichen, and up to 600% for the moss) overnight in a climatic chamber at 16 °C and 90% RH. Two thirds of the air-dried and fully hydrated samples were fumigated for 1 h at an ozone concentration of 1 ppm using an ozone generator (GPC2000, Ozonosoluzioni, Italy); the remaining third were fumigated for 1 h with O₃-free air (control samples). To evaluate the possible recovery of the samples, 50% of the fumigated and control samples were left for 1 week under environmental conditions.

2.2. Physiological Parameters

2.2.1. Chlorophyll Content

The total chlorophyll content of the lichen and moss samples was measured using a chlorophyllcontent meter (CCM-300, Opti-Science, Hudson, NY, USA), which indicates the chlorophyll content on a surface basis (mg/m²). The effectiveness of this non-destructive method for the analysis of the chlorophyll content has already been tested [21]. Ten replicates were measured for each experimental unit.

2.2.2. Chlorophyll Fluorescence Analysis

The analysis of the chlorophyll a fluorescence of the photosystem II and the analysis of the chlorophyll fluorescence transient (OJIP) are known reliable techniques to assess the performance of photosynthetic organisms following O₃ exposures [22]. In this study, the chlorophyll a fluorescence was investigated using its most common parameter Fv/FM, which indicates the maximum quantum efficiency of PS II photochemistry, an indicator of photosynthetic efficiency. The analysis of the chlorophyll a fluorescent transient (OJIP test) was used as a supplementary photosynthetic indicator. Prior to analysis, samples were hydrated in a climatic chamber at 16 °C, 90% RH and 40 µmol/m2/s photosynthetically active radiation (PAR), to obtain their maximal photosynthetic activation. The analysis was run, lighting the samples for 1 sec with a saturating 3000 µmol/m²/s red light pulse, using a Plant Efficiency Analyzer (Handy PEA, Hansatech Ltd, Norfolk, UK). Fifteen replicates were measured for each experimental unit.

2.2.3. Total Antioxidant Power

The DPPH assay is a simple and functional method to evaluate the response of the total antioxidant activity after O₃ exposure [23]. Samples of ca. 50 mg were homogenized in 1 mL of a solution of ethanol/water (80:20; v/v). Of the homogenate, 100 μ L were added to 1 mL of a 100 μ M DPPH solution prepared by dissolving 3.9 mg of this compound in 100 mL of methanol/water (80:20; v/v). After the reaction, which occurred in 1 h, samples were read at 517 nm and the results were expressed as % Antiradical Activity (ARA%) according to the formula:

$$ARA\% = 100 \times [1- (control absorbance/sample absorbance)]$$
(1)

where control absorbance = the absorbance of the reagents only. Five replicates were measured for each experimental unit.

2.3. Statistical Analysis

Results were expressed as ratios to control samples. Owing to the limited data set, nonparametric statistics were used. Outliers were sought for using the Tukey test. Differences between fumigated and control samples were checked with the Mann-Withney U test. Differences between dry and wet samples, both immediately after fumigation and after recovery, were checked using the non-parametric Kruskal–Wallis ANOVA. The Dunn's test was used for post-hoc comparisons, except for temporal ones, for which the Wilcoxon rank sum test was used. In both cases, a correction for multiple testing was applied according to [24]. All calculations were run using the free software R [25].

3. Results

The physiological parameters in *Evernia* sp. and *Brachythecium* sp. immediately after O_3 fumigation, as well as after one week from the latter, are summarized in Table 1.

Table 1. Chlorophyll content (Chl), photosynthetic efficiency (Fv/FM) and antioxidant power % (ARA) in samples of the lichen (L) *Evernia prunastri* and the moss (M) *Brachythecium* sp. exposed dry (D) or wet (W), immediately after (E) 1 h fumigation with 1 ppm O₃ or after 1 week of recovery (R). Values are expressed as median \pm median absolute deviation of ratios to control values. Values in bold indicate significant (p < 0.05) differences with control values, capital letters indicate significant (p < 0.05) differences between D and W, and lowercase letters indicate significant (p < 0.05) differences between E and R. * values were almost zero and were disregarded for the statistical analysis but were considered significantly different from controls and dry samples.

Parameter	LED	LEW	LRD	LRW
Chl	$0.85 \pm 0.19 \; \text{A}$	0.00B*	$0.89\pm0.47~\mathrm{A}$	0.00B*
Fv/Fм	$0.21 \pm 0.09 \; \mathrm{A}$	0.00 B*	$0.19\pm0.14~\mathrm{A}$	0.00B*
ARA	1.51 ± 0.09 Aa	1.11 ± 0.08 Ba	1.19 ± 0.02 Ab	0.83 ± 0.01 Bb
	MED	MEW	MRD	MRW
Chl	0.96 ± 0.13 A	0.48 ± 0.17 Ba	0.73 ± 0.27	0.73 ± 0.18 b
Fv/Fм	0.35 ± 0.11 Aa	0.00B*	0.70 ± 0.14 Ab	0.00B*
ARA	1.16 ± 0.20 a	$1.30 \pm 0.11a$	$1.04 \pm 0.05 \text{ Ab}$	1.17 ± 0.01 Bb

Compared with control samples, 1 h of fumigation with 1 ppm O₃ negatively affected the content of chlorophyll only in wet samples, but in moss samples, such a difference was no longer evident after one week of recovery. Photosynthetic efficiency was always impaired by O₃ exposure, irrespective of dry or wet conditions, and also after 1 week of recovery, but the effect was much stronger for wet samples. The antioxidant power was increased in wet moss and in dry lichen, while a decrease was found for wet lichens after 1 week.

Differences between dry and wet samples emerged for all the investigated parameters, with wet samples generally showing lower values.

After one week of recovery, moss wet samples showed an almost complete restoration of their chlorophyll content. The photosynthetic efficiency of dry samples almost doubled in moss but did not change in lichens. The antioxidant power decreased in both lichens and mosses.

The fluorescence transient curves (Figure 1) confirm that remarkable injury occurred to the photobiont of both organisms following O₃ fumigations, which persisted after 1 week of recovery. The transient curves flattened out, losing their typical sequence of the OJIP steps as indicated for healthy (control) samples. Fluorescence emission was characterized by a marked reduction in F_M values in dry fumigated samples of both organisms, while negligible differences between F₀ and F_M values were observed for wet fumigated samples, which showed almost flat transients.



Figure 1. OJIP fluorescence transients of the lichen *Evernia prunastri* (up) and the moss *Brachythecium sp.* (down) after the O₃ fumigation (left) and the recovery time (right). Legend: (\circ) control samples (mean of dry and wet), (Δ) dry fumigated samples, (\diamond) wet fumigated samples.

4. Discussion

Short-term acute fumigation with 1 ppm O₃ impaired the photosynthetic efficiency of the lichen, with effects being much more evident in the hydrated state and with no recovery after the damage. However, a decreased chlorophyll content was found only in wet samples, suggesting that the mechanisms of action of O₃ causing these alterations are not identical. The antioxidant power always being higher in dry samples suggests that a counteraction of the strong oxidizing effect of O₃ may be a key factor to protect the integrity of the chlorophyll. Nevertheless, there is clear evidence that O₃ sensitivity is dependent upon the hydration state, since — also in terms of photosynthetic efficiency — the damage was much more striking in wet samples, of which the capacity to convert light was almost abolished. The response was similar for the moss, which was, however, less sensitive to O₃ stress and showed clear recovery after one week from the fumigation, even if full recovery was achieved only for the chlorophyll content. Unlike the lichen, the moss showed higher values of antioxidant power in wet samples, thus suggesting a role of this latter parameter in protecting the photosynthetic system.

The higher impact of O₃ on wet samples may be determined by the combination of two main factors, the great solubility of this pollutant in water and the fully activated metabolism of both organisms during the hydrated state. Ozone is a very soluble molecule that exhibits a water solubility higher than that of oxygen and a lower affinity for the cellular hydrophobic layers [26]. After its dissolution in water, O₃ starts its decomposition, leading to the formation of reactive oxygen species (ROS), responsible for all of the oxidizing reactions that further occurred [27]. Despite ROS being (at low concentrations) natural components of cell metabolism, produced as a result of aerobic respiration and used in several cell processes, exposure to O₃ tends to increase their concentration, generating cellular perturbations, changes in cellular homeostasis and further physiological injuries, following a process known as "oxidative burst" [28]. The increased metabolism of these cryptogamic species during the hydrated state may have increased their susceptibility to O₃. In fact, being poikilohydric organisms, with metabolic activity largely dependent on the hydration state, lichens and mosses tend to increase their net photosynthetic rate (NPR) under increasing hydration conditions [29–32]. However, the combination of these two factors—the O₃ dissolution in water and the activated metabolism during the hydrated state—are probably the main reasons explaining the differential toxicity of lichens and mosses to this pollutant. This behavior, dependent on the hydration state, is known also for other toxic gaseous pollutants such as SO2, which shows a higher toxicity to lichens and mosses when they are completely wet [33,34], in spite of their activated metabolisms and the high affinity of SO₂ for water [35].

The mechanism of action of O₃ against the chlorophyll pool is complex [26], but it is generally assumed to be related either to the direct action of ROS on chlorophylls, with subsequent degradation (chlorosis), or to a sort of protection mechanism of the photosynthetic system [36]. The reduction in the chlorophyll content of wet samples of *E. prunastri* after O₃ fumigation is at variance with the results of [15], which showed a reduction in dry samples of the lichen *Xanthoria parietina*, probably determined by the absence of water, which allowed easy O₃ intrusion into the lichen thallus, as noted for CO₂, for which water has an active role in limiting its diffusion inside the lichen cortex [29,37]. Lichens may show remarkable differences in O₃ uptake during the dry state [38], probably depending on differences in their morphology [39] or tissue structure—as also suggested for vascular plants [26]— or on other (still unclear) physiological characteristics that also drive their resistance to others pollutants.

The ecologies of *Evernia* and *Xanthoria* is quite different, the former being hygrophytic and the latter, xerophytic [40]; thus, they are naturally less hydrated, with a consequent increase in their tolerance to pollutants such as SO₂ [34,41]. In spite of their differential ecophysiologies, *Evernia* is known to be more hydrophobic than *Xanthoria* [42,43], the latter having the capacity to hydrate up to 300% of its dry weight vs. the 200% for the former. In addition, once fully hydrated, to achieve full evaporation, *Evernia* requires up to 60 min, while *Xanthoria* may prolong the hydrated state for a much longer time, up to 180 min [43]. These marked differences may well explain the different responses of these two species to O₃.

Biology 2020, 9, 90

6 of 10

The limited chlorophyll reduction in the wet samples of *Brachythecium* is consistent with chronic fumigations (4–6 weeks) carried out on *Sphagnum* in open top chambers, during the dry (Summer, RH = 52%–80%) and the wet (Autumn, RH = 71%–92%) seasons, for which significant reductions in the chlorophyll contents were not found after exposure to a wide range (50–150 ppb) of O₃ concentrations [18]. These results suggest a limited effect of O₃ on chlorophyll when samples are not under fully hydrated conditions. However, reduction in the chlorophyll content as a consequence of the toxic action of O₃ has been reported for several vascular plants when fumigated under different concentrations, e.g., for soybean (*Glycine max*) at 50–130 ppb O₃ [44], wheat (*Triticum aestivum*) at 25–35 ppb O₃ [45], bean (*Phaseolus vulgaris*) at 50–90 ppb O₃ [46], and *Tilia americana* at 120 ppb O₃ [47].

The acute O₃ fumigation severely impaired the photosynthetic efficiency of both the lichen and the moss, with reductions of ca. 100% in wet samples and ca. 65%-80% in dry samples. Similar results were observed for the lichen X. parietina, for which wet and dry samples showed reductions of ca. 85% and 69%, respectively [15]. Slight reductions in the photosynthetic efficiency of *Evernia* were observed after 6 weeks of exposure to 90 ppb O₃ in a partially continuous hydration state [38], but were not observed when fumigated for 14 days with 300 ppb of O₃ (4 h/d) under high levels of humidity (RH = 97%) [48]. The fumigations of four species of Sphangnum at 150 ppb O₃ for 5 h induced slightly decreased photosynthetic efficiency in one species only [16], while fumigations of Sphagnum recurvum and Polytrichum commune with 70-80 ppb O3 for 6-9 weeks induced the occurrence of modest photosynthetic injuries only in the former [17]. The effect of O₃ fumigations on the photosynthetic efficiency of higher plants is well documented, and significant effects were reported in ca. 50% of the studies [49], e.g., in the summer squash (Cucurbita pepo), which showed an impairment of this parameter when fumigated for 5 h/d for 5 days with 150 ppb O₃ [36]. Furthermore, reductions were also observed in *Tilia americana* after 28–42 days of exposure to 120 ppb O₃ at 5 h/day [47], in sugar beet (Beta vulgaris) and spring rape (Brassica napus) at 35 ppb O₃ [50], and wheat (Triticum aestivum) at 50 ppb O3 at different growth stages [51].

After 1 week following fumigation, the dry samples of *Brachythecium* showed an almost complete recovery of the chlorophyll content and an increase in photosynthetic efficiency. The occurrence of photosynthetic recovery after O₃ exposure was also observed in some vascular plants after restoration in O₃-free air. In detail, *Solanum tuberosum* showed a recovery of its net photosynthesis after 4 days of fumigation at 60–80 ppb O₃ [52], while *Nicotiana tabacum* Bel B (ozone tolerant) did so after 17 h following fumigation with 300 ppb [53]. In addition, *Quercus ilex* and *Q. pubescens* showed recoveries in their photosynthesis after 72 h following 4 days of exposure to 9 h/d 300 ppb O₃ [54].

The analysis of the chlorophyll fluorescence transient confirmed the strong negative effect of O_3 on wet samples and the lower effect on dry samples. Both the lichen and the moss, after the wet exposure, showed a total flattening of the typical OJIP steps, indicating that reduction of plastoquinone A by electrons did not occur, while dry samples showed an intermediate decrease, indicating that the reduction of plastoquinone B by electrons did not occur, confirming the lower susceptibility of the photosynthetic system under the dry state. This behavior was probably determined by the reduced presence of water in these samples (<10%), which may have limited the diffusion of O₃, since its direct effect on the PSII has been excluded [55]. In fact, when these cryptogamic organisms are dry, most of them put into action strategies to protect their photosynthetic system from photoinactivation [56], which make them less sensitive to pollutants. In addition, the fast disappearance of O₃, due to its short half-life that ranges from minutes to hours [27], may play an important role.

The response of the antioxidant activity showed slight increases in the dry fumigated samples of *E. prunastri* (+150%) and in the wet ones of *Brachythecium* (+128%). The increases in the antioxidant power after the exposure to O₃ may be related to the necessity for the organism to maintain a stable balance between ROS production and elimination to preserve its cellular and metabolic integrity. ROS can work as signals to stimulate cellular defenses but, following the exposure to oxidizing compounds (O₃), their extracellular concentration increases, generating cellular damage. The effects of seasonal O₃ variations on the antioxidant levels of *Picea rubens*, clones of *Populus* and *Triticum*

Biology 2020, 9, 90

aestivum were reported [57–59], and similarly, fumigations with very low O₃ concentrations caused increases in all of the indicators of antioxidant activity during the first 20 min of exposure of *Carica papaya* [60]. In addition, prolonged exposure to 50 ppb O₃ increased the total phenolic content and peroxidase activity of two different wheat cultivars [51], and fumigations with 32 ppb O₃ for 4 weeks increased the phenolic contents of *Trifolium pratense* cv. Bjursele [61]. The increase in the antioxidant power in dry samples of *E. prunastri* may be, and is probably, related to the ability to counteract the (limited) ROS production, irrespective of an inefficient metabolism. The slight increase in the antioxidant power in the wet samples of *Brachythecium* was probably due to a first step of the metabolism to counteract the deleterious effects of ROS on the photosynthetic machinery, as suggested by the reductions in photosynthetic efficiency.

After one week of recovery from the O₃ fumigation, dry samples of *E. prunastri* still showed a higher antioxidant power than controls, while wet samples showed a decrease; wet samples of *Brachythecium* showed a higher antioxidant power compared with control samples. The low antioxidant power of wet *E. prunastri* samples may be, and is probably, due to a sort of metabolic stress caused by ROS, as has been observed for higher plants after prolonged exposure to O₃ [60]. Nevertheless, we may also speculate that the degradation of photobiont cells could have resulted in the formation of molecules with antioxidant properties.

The higher antioxidant expression of wet moss may be due to the already turned on antioxidant metabolism, stimulated to maintain a sort of protection mechanism for the photosynthetic system during its complete recovery [58] or as a consequence of a physiological stress-memory process that protects the organism from subsequent oxidative exposures, as can occur in higher plants following ecological changes [62].

5. Conclusions

The fumigation of the lichen *Evernia prunastri* and the moss *Brachythecium* sp. with 1 ppm O₃ for 1 h highlighted the higher susceptibility of these cryptogamic organisms to this oxidizing pollutant under the hydrated state. An impairment of the photosynthetic efficiency and a reduction in the chlorophyll content were evident. The role of antioxidant power as mechanism of resistance to high O₃ levels needs to be better clarified with further investigations. Our results confirm that the tolerance to O₃ of lichens and mosses may be determined by their low water content, which is the case during the peaks of O₃ occurring during the Mediterranean summer.

Author Contributions: S.L. conceived and designed the experiments; A.V. and G.C. performed the experiments; S.L. and A.V. analyzed the data; M.P. contributed analysis tools; A.V. and S.L. wrote the paper, C.N. supervised the text. All authors have read and agreed to the published version of the manuscript.

Acknowledgments: Thanks to Dr. Ilaria Bonini for moss identification.

Funding: This research received no external funding.

Conflicts of Interest: The authors declare no conflict of interest.

References

- Wilkinson, S.; Mills, G.; Illidge, R.; Davies, W.J. How is ozone pollution reducing our food supply? J. Exp. Bot. 2011, 63, 527–536, doi:10.1093/jxb/err317.
- Santurtún, A.; Gonzalez-Hidalgo, J.C.; Sanchez-Lorenzo, A.; Zarrabeitia, M.T. Surface ozone concentration trends and its relationship with weather types in Spain (2001–2010). *Atmospheric Environ.* 2015, 101, 10–22, doi:10.1016/j.atmosenv.2014.11.005.
- Fusaro, L.; Palma, A.; Salvatori, E.; Basile, A.; Maresca, V.; Karam, E.A.; Manes, F. Functional indicators of response mechanisms to nitrogen deposition, ozone, and their interaction in two Mediterranean tree species. *PLOS ONE* 2017, *12*, e0185836, doi:10.1371/journal.pone.0185836.
- Ainsworth, E.A.; Lemonnier, P.; Wedow, J.M. The influence of rising tropospheric carbon dioxide and ozone on plant productivity. *Plant Boil.* 2019, 22, 5–11, doi:10.1111/plb.12973.
- Nimis, P.L.; Scheidegger, C.; Wolseley, P.A. Monitoring with Lichens–Monitoring Lichens; Kluwer Academic Publ.: Dordrecht, Netherlands, 2002

- Gombert, S.; Asta, J.; Seaward, M. Lichens and tobacco plants as complementary biomonitors of air pollution in the Grenoble area (Isère, southeast France). *Ecol. Indic.* 2006, *6*, 429–443, doi:10.1016/j.ecolind.2005.06.001.
- 8. Nali, C.; Balducci, E.; Frati, L.; Paoli, L.; Loppi, S.; Lorenzini, G. Integrated biomonitoring of air quality with plants and lichens: A case study on ambient ozone from central Italy. *Chemosphere* **2007**, *67*, 2169–2176, doi:10.1016/j.chemosphere.2006.12.036.
- 9. Bertuzzi, S.; Davies, L.; Power, S.A.; Tretiach, M. Why lichens are bad biomonitors of ozone pollution? *Ecol. Indic.* **2013**, *34*, 391–397, doi:10.1016/j.ecolind.2013.05.023.
- 10. Riddell, J.; Padgett, P.; Nash, T. Physiological responses of lichens to factorial fumigations with nitric acid and ozone. *Environ. Pollut.* **2012**, *170*, 202–210, doi:10.1016/j.envpol.2012.06.014.
- Pellegrini, E.; Bertuzzi, S.; Carniel, F.C.; Lorenzini, G.; Nali, C.; Tretiach, M. Ozone tolerance in lichens: A possible explanation from biochemical to physiological level using Flavoparmelia caperata as test organism. *J. Plant Physiol.* 2014, *171*, 1514–1523, doi:10.1016/j.jplph.2014.06.017.
- Iii, T.H.N.; Sigal, L.L. Gross Photosynthetic Response of Lichens to Short-Term Ozone Fumigations. *Bryol.* 1979, 82, 280, doi:10.2307/3242087.
- Rosentreter, R.; Ahmadjian, V. Effect of Ozone on the Lichen *Cladonia arbuscula* and the *Trebouxia Phycobiont* of Cladina stellaris. *Bryol.* 1977, 80, 600, doi:10.2307/3242416.
- 14. Brown, D.H.; Smirnofff, N. Observations on the Effect of Ozone on *Cladonia Rangiformis*. *Lichenol*. **1978**, *10*, 91–94, doi:10.1017/s0024282978000109.
- Vannini, A.; Paoli, L.; Ceccarelli, S.; Sorbo, S.; Basile, A.; Carginale, V.; Nali, C.; Lorenzini, G.; Pica, M.; Loppi, S. Physiological and ultrastructural effects of acute ozone fumigation in the lichen Xanthoria parietina: the role of parietin and hydration state. *Environ. Sci. Pollut. Res.* 2017, 25, 8104–8112, doi:10.1007/s11356-017-9545-9.
- 16. Potter, L.; Foot, J.P.; Caporn, S.; Lee, J.A. Responses of four Sphagnum species to acute ozone fumigation. *J. Bryol.* **1996**, *19*, 19–32, doi:10.1179/jbr.1996.19.1.19.
- Potter, L.; Foot, J.P.; Caporn, S.J.M.; Lee, J.A. The effects of long-term elevated ozone concentrations on the growth and photosynthesis of Sphagnum recurvum and Polytrichum commune. *New Phytol.* 1996, 134, 649–656, doi:10.1111/j.1469-8137.1996.tb04930.x.
- Niemi, R.; Martikainen, P.J.; Silvola, J.; Holopainen, T. Ozone effects on Sphagnum mosses, carbon dioxide exchange and methane emission in boreal peatland microcosms. *Sci. Total. Environ.* 2002, 289, 1–12, doi:10.1016/s0048-9697(01)01012-9.
- 19. Stanosz, G.; Smith, V.; Bruck, R. Effect of ozone on growth of mosses on disturbed forest soil. *Environ. Pollut.* **1990**, *63*, 319–327, doi:10.1016/0269-7491(90)90138-3.
- 20. Nash III, T.H.; Egan, R.S. *The biology of lichens and bryophytes. In: Lichens, Briophytes and Air Quality*; Bibl. Lichenol.; Berlin, Stuttgart, Germany, 1988, Volume 30, pp. 11–22.
- Vannini, A.; Paoli, L.; Russo, A.; Loppi, S. Contribution of submicronic (PM1) and coarse (PM>1) particulate matter deposition to the heavy metal load of lichens transplanted along a busy road. *Chemosphere* 2019, 231, 121–125, doi:10.1016/j.chemosphere.2019.05.085.
- 22. Bussotti, F.; Strasser, R.J.; Schaub, M. Photosynthetic behavior of woody species under high ozone exposure probed with the JIP-test: A review. *Environ. Pollut.* **2007**, *147*, 430–437, doi:10.1016/j.envpol.2006.08.036.
- 23. Zhang, Z.; Diao, E.; Shen, X.; Ma, W.; Ji, N.; Dong, H. Ozone-Induced Changes in Phenols and Antioxidant Capacities of Peanut Skins. J. Food Process. Eng. 2014, 37, 506–514, doi:10.1111/jfpe.12108.
- 24. Benjamini, Y.; Hochberg, Y. Controlling the False Discovery Rate: A Practical and Powerful Approach to Multiple Testing. J. R. Stat. Soc. Ser. B (Statistical Methodol.) **1995**, 57, 289–300, doi:10.1111/j.2517-6161.1995.tb02031.x.
- Anonymous. The R Project for Statistical Computing. Available online: http://www.r-project.org/ (accessed on 13 February 2012).
- 26. Heath, R.L. Possible mechanisms for the inhibition of photosynthesis by ozone. *Photosynth. Res.* **1994**, *39*, 439–451, doi:10.1007/bf00014597.
- 27. Von Gunten, U. Ozonation of drinking water: Part I. Oxidation kinetics and product formation. *Water Res.* **2003**, *37*, 1443–1467, doi:10.1016/s0043-1354(02)00457-8.

- Schraudner, M.; Wiese, C.; Van Camp, W.; Moeder, W.; Inzé, D.; Langebartels, C.; Jr, H.S. Ozone-induced oxidative burst in the ozone biomonitor plant, tobacco Bel W3. *Plant J.* 1998, *16*, 235–245, doi:10.1046/j.1365-313x.1998.00294.x.
- Green, T.G.A.; Kulle, D.; Pannewitz, S.; Sancho, L.; Schroeter, B. UV-A protection in mosses growing in continental Antarctica. *Polar Boil.* 2005, 28, 822–827, doi:10.1007/s00300-005-0011-7.
- Titus, J.E.; Wagner, D.J.; Stephens, M.D. Contrasting Water Relations of Photosynthesis for Two Sphagnum Mosses. *Ecol.* 1983, 64, 1109–1115, doi:10.2307/1937821.
- Lange, O.L.; Green, T.G.A. Lichens show that fungi can acclimate their respiration to seasonal changes in temperature. *Oecologia* 2004, 142, 11–19, doi:10.1007/s00442-004-1697-x.
- Pannewitz, S.; Green, T.A.; Maysek, K.; Schlensog, M.; Seppelt, R.; Sancho, L.; Türk, R.; Schroeter, B. Photosynthetic responses of three common mosses from continental Antarctica. *Antarct. Sci.* 2005, 17, 341– 352, doi:10.1017/s0954102005002774.
- Winner, W.E.; Bewley, J.D. Photosynthesis and respiration of feather mosses fumigated at different hydration levels with SO2. *Can. J. Bot.* 1983, 61, 1456–1461, doi:10.1139/b83-158.
- Rürk, R.; Wirth, V.; Lange, O.L. CO2-Gaswechsel-Untersuchungen zur SO2-Resistenz von Flechten. Oecologia 1974, 15, 33–64, doi:10.1007/bf00345227.
- Nieboer, E.; Tomassini, F.D.; Puckett, K.J.; Richardson, D.H.S. A model for the relationship between gaseous and aqueous concentrations of sulphur dioxide in lichen exposure studies. *New Phytol.* 1977, 79, 157–162, doi:10.1111/j.1469-8137.1977.tb02191.x.
- Castagna, A.; Nali, C.; Ciompi, S.; Lorenzini, G.; Soldatini, G.F.; Ranieri, A. Ozone exposure affects photosynthesis of pumpkin (*Cucurbita pepo*) plants. *New Phytol.* 2001, 152, 223–229, doi:10.1046/j.0028-646X.2001.00253.x.
- Snelgar, W.P.; Green, T.G.A.; Wilkins, A.L. Carbon dioxide exchange in lichens: resistances to CO2 uptake at different thallus water contents. *New Phytol.* 1981, *88*, 353–361, doi:10.1111/j.1469-8137.1981.tb01730.x.
- Scheidegger, C.; Schroeter, B. Effects of ozone fumigation on epiphytic macrolichens: Ultrastructure, CO2 gas exchange and chlorophyll fluorescence. *Environ. Pollut.* 1995, *88*, 345–354, doi:10.1016/0269-7491(95)93449-a.
- Ross, L.J.; Nash, T.H. Effect of ozone on gross photosynthesis of lichens. *Environ. Exp. Bot.* 1983, 23, 71–77, doi:10.1016/0098-8472(83)90022-9.
- 40. Nimis, P.L. The Lichens of Italy. A Second Annotated Catalogue. EUT: Trieste, Italy, 2016.
- Grace, B.; Gillespie, T.J.; Puckett, K.J. Uptake of gaseous sulphur dioxide by the lichen Cladina rangiferina. *Can. J. Bot.* 1985, 63, 797–805, doi:10.1139/b85-102.
- Hauck, M.; Jürgens, S.-R.; Brinkmann, M.; Herminghaus, S. Surface Hydrophobicity Causes SO2 Tolerance in Lichens. *Ann. Bot.* 2007, 101, 531–539, doi:10.1093/aob/mcm306.
- Paoli, L.; Pisani, T.; Munzi, S.; Gaggi, C.; Loppi, S. Influence of sun irradiance and water availability on lichen photosynthetic pigments during a Mediterranean summer. *Boil.* 2010, 65, 776–783, doi:10.2478/s11756-010-0087-6.
- Reich, P.B.; Schoettle, A.W.; Raba, R.M.; Amundson, R.G. Response of Soybean to Low Concentrations of Ozone: I. Reductions in Leaf and Whole Plant Net Photosynthesis and Leaf Chlorophyll Content. J. Environ. Qual. 1986, 15, 31–36, doi:10.2134/jeq1986.00472425001500010008x.
- Sandelius, A.S.; Näslund, K.; Carlsson, A.S.; Pleijel, H.; Selldén, G. Exposure of spring wheat (*Triticum aestivum*) to ozone in open-top chambers. Effects on acyl lipid composition and chlorophyll content of flag leaves. *New Phytol.* 1995, 131, 231–239, doi:10.1111/j.1469-8137.1995.tb05724.x.
- Knudson, L.L.; Tibbitts, T.W.; Edwards, G.E. Measurement of Ozone Injury by Determination of Leaf Chlorophyll Concentration. *Plant Physiol.* 1977, 60, 606–608, doi:10.1104/pp.60.4.606.
- Pellegrini, E. PSII photochemistry is the primary target of oxidative stress imposed by ozone in Tilia americana. Urban For. Urban Green. 2014, 13, 94–102, doi:10.1016/j.ufug.2013.10.006.
- Calatayud, A.; Temple, P.; Barreno, E. Chlorophyll a Fluorescence Emission, Xanthophyll Cycle Activity, and Net Photosynthetic Rate Responses to Ozone in Some Foliose and Fruticose Lichen Species. *Photosynth.* 2000, *38*, 281–286, doi:10.1023/a:1007214915785.
- Bussotti, F.; Desotgiu, R.; Cascio, C.; Pollastrini, M.; Gravano, E.; Gerosa, G.; Marzuoli, R.; Nali, C.; Lorenzini, G.; Salvatori, E.; et al. Ozone stress in woody plants assessed with chlorophyll a fluorescence. A critical reassessment of existing data. *Environ. Exp. Bot.* 2011, 73, 19–30, doi:10.1016/j.envexpbot.2010.10.022.

- Rai, R.; Agrawal, M. Assessment of competitive ability of two Indian wheat cultivars under ambient O3 at different developmental stages. *Environ. Sci. Pollut. Res.* 2013, *21*, 1039–1053, doi:10.1007/s11356-013-1981-6.
- 52. Dann, M.S.; Pell, E.J. Decline of Activity and Quantity of Ribulose Bisphosphate Carboxylase/Oxygenase and Net Photosynthesis in Ozone-Treated Potato Foliage. *Plant Physiol.* **1989**, *91*, 427–432, doi:10.1104/pp.91.1.427.
- Pasqualini, S.; Antonielli, M.; Ederli, L.; Piccioni, C.; Loreto, F. Ozone uptake and its effect on photosynthetic parameters of two tobacco cultivars with contrasting ozone sensitivity. *Plant Physiol. Biochem.* 2002, 40, 599–603, doi:10.1016/s0981-9428(02)01426-2.
- 54. Velikova, V.; Pinelli, P.; Pasqualini, S.; Reale, L.; Ferranti, F.; Loreto, F. Isoprene decreases the concentration of nitric oxide in leaves exposed to elevated ozone. *New Phytol.* **2005**, *166*, 419–426, doi:10.1111/j.1469-8137.2005.01409.x.
- 55. Schreiber, U.; Vidaver, W.; Runeckles, V.C.; Rosén, P. Chlorophyll Fluorescence Assay for Ozone Injury in Intact Plants. *Plant Physiol.* **1978**, *61*, 80–84, doi:10.1104/pp.61.1.80.
- 56. Heber, U.; Azarkovich, M.; Shuvalov, V. Activation of mechanisms of photoprotection by desiccation and by light: poikilohydric photoautotrophs*. *J. Exp. Bot.* **2007**, *58*, 2745–2759, doi:10.1093/jxb/erm139.
- 57. Hausladen, A.; Madamanchi, N.R.; Fellows, S.; Alscher, R.G.; Amundson, R.G. Seasonal changes in antioxidants in red spruce as affected by ozone. *New Phytol.* **1990**, *115*, 447–458, doi:10.1111/j.1469-8137.1990.tb00470.x.
- Gupta, A.S.; Alscher, R.G.; McCune, D. Response of Photosynthesis and Cellular Antioxidants to Ozone in Populus Leaves. *Plant Physiol.* 1991, *96*, 650–655, doi:10.1104/pp.96.2.650.
- 59. Bender, J.; Weigel, H.; Wegner, U.; Jäger, H. Response of cellular antioxidants to ozone in wheat flag leaves at different stages of plant development. *Environ. Pollut.* **1994**, *84*, 15–21, doi:10.1016/0269-7491(94)90065-5.
- Yeoh, W.K.; Ali, A.; Forney, C. Effects of ozone on major antioxidants and microbial populations of freshcut papaya. *Postharvest Boil. Technol.* 2014, 89, 56–58, doi:10.1016/j.postharvbio.2013.11.006.
- 61. Saviranta, N.M.; Julkunen-Tiitto, R.; Oksanen, E.; Karjalainen, R.O. Leaf phenolic compounds in red clover (*Trifolium pratense L.*) induced by exposure to moderately elevated ozone. *Environ. Pollut.* **2010**, *158*, 440–446, doi:10.1016/j.envpol.2009.08.029.
- 62. Walter, J.; Jentsch, A.; Beierkuhnlein, C.; Kreyling, J. Ecological stress memory and cross stress tolerance in plants in the face of climate extremes. *Environ. Exp. Bot.* **2013**, *94*, 3–8, doi:10.1016/j.envexpbot.2012.02.009.



© 2020 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (http://creativecommons.org/licenses/by/4.0/).