



Article Impact of Thermal Stress on Abrasive Dust from a Carbon Fiber-Reinforced Concrete Composite

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Abstract: Recently, a novel corrosion-resistant construction material, Carbon Concrete Composite (C³), consisting of coated carbon fibers embedded in a concrete matrix, was introduced. However, thermal exposure during domestic fires may impact the release of organic pollutants and fibers during abrasive processing and/or demolition. Consequently, the objective of this study was to explore the emission characteristics of toxic compounds and harmful fibers during the dry-cutting after exposure to 25–600 °C (3 h, air). These parameters mimic the abrasive machining and dismantling after a domestic fire event. Mass spectrometry and chromatography served as analytical methodologies, and no organic pollutants for exposure temperatures ≥ 400 °C were found. In contrast, significant amounts of pyrolysis products from the organic fiber coating were released at lower temperatures. Studying the morphology of the released fibers by electron microscopy revealed a decrease in fiber diameter for temperatures exceeding 450 °C. At \geq 550 °C, harmful fibers, according to the World Health Organization (WHO) definition, occurred (28–41 \times 10³ WHO fibers/m³ at 550–600 °C). This leads to the conclusion that there is a demand for restraining and protection measures, such as the use of wet cutting processes, suction devices, particle filtering masks and protective clothing, to handle thermally stressed C³.

Keywords: carbon fibers; concrete; WHO fibers; thermal exposure; scanning electron microscopy; evolved gas analysis; mass spectrometry

1. Introduction

At present, concrete is the mainstay in the construction industry, and will remain so in the near future. Since the most commonly used steel bar-reinforced concrete is susceptible to corrosion, buildings have a limited lifetime of only about 40–80 years [1]. Indeed, corrosion-resistant alternatives are needed. Extensive research was conducted on identifying resilient and durable reinforcement materials such as asbestos [2,3], glass [4,5], basalt [5,6], or carbon fibers [7,8]. Although glass and basalt fibers have poor resistance against alkaline matrices such as the cementitious matrix [9,10], asbestos fibers were found to cause severe health impairments such as asbestosis or lung cancer [11]. Therefore, carbon fibers can be a promising reinforcement material for concrete composites.



Citation: Koch, A.; Friederici, L.; Fiala, P.; Springer, A.; Di Bucchianico, S.; Stintz, M.; Frank, M.; Rüger, C.P.; Streibel, T.; Zimmermann, R. Impact of Thermal Stress on Abrasive Dust from a Carbon Fiber-Reinforced Concrete Composite. *Fibers* **2022**, *10*, 39. https://doi.org/10.3390/ fib10050039

Academic Editor: Doo-Yeol Yoo

Received: 19 February 2022 Accepted: 24 April 2022 Published: 26 April 2022

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A new carbon fiber-reinforced construction material called Carbon Concrete Composite (C^3) has been developed since the 1990s by a cooperation of the Technical University (TU) Dresden and Rheinisch-Westfälische Technische Hochschule (RWTH) Aachen [12,13]. C^3 consists of a grid- or rod-shaped reinforcement material made from continuous, unidirectional carbon fibers embedded in a concrete matrix [1,12,14]. The carbon fibers are produced from various organic precursor fibers such as polyacrylonitrile (PAN), pitch or cellulose, by (1) stabilizing them at 200–400 °C [1,15–17] under air atmosphere, (2) carbonizing them at 400–1600 °C [1,15–17] under inert nitrogen or argon [18] atmosphere and, if high tensile strength is required, (3) graphitizing at temperatures of up to 3000 °C [15–17]. As a result, carbon fibers with a carbon fraction of >98% [15] and \sim 7 µm diameter [12] are formed due to oxidation, cyclization and dehydrogenation of the precursor fibers [19,20]. A roving is formed with up to 48,000 carbon fibers combined and coated by an organic polymer matrix, such as a styrene-butadiene resin (SBR) or epoxy resin. The organic coating improves the grip between the reinforcement material and the concrete matrix to increase the stability of the resulting C^3 material. Subsequently, rovings are combined in a weaving process to produce the reinforcement material. The resulting C^3 material yields a five-fold higher tensile strength compared to steel bar-reinforced concrete and is expected to have higher longevity due to its corrosion resistance [12].

Construction materials are exposed to various stress situations during their lifecycle, such as weathering, abrasive processing, thermal recycling or domestic fires. However, only a limited number of studies address the release of chemicals or fibers, as observed for asbestos concrete [21], and its linked health impacts in these scenarios. Weiler and Vollpracht (2019) [22] studied the release of heavy metals and trace elements leached from C^3 during wet weathering, but found concentrations considered that were not environmentally harmful. Bienkowski et al. (2018) [14], Hillemann et al. (2018) [12], and Koch et al. (2021) [23] showed that no harmful fibers were formed during crushing and cutting processes of C^3 . According to the World Health Organization (WHO), biopersistent, inorganic fibers with length >5 μ m, diameter <3 μ m and a length-to-diameter ratio >3 are considered to be harmful to human health [24]. However, carbon fibers with $\sim 7 \,\mu m$ diameter in the total dust fraction were found, and a considerable amount of inhalable particulate matter (PM) fraction PM₁₀ (aerodynamic diameter $\leq 10 \,\mu$ m) was released during dry-cutting of C^{3} [23]. Furthermore, Hillemann et al. (2018) [12] and Koch et al. (2021) [23] found organic compounds in the gas phase and in the inhalable PM fractions $PM_{2.5}$ (aerodynamic diameter $\leq 2.5 \,\mu$ m) and PM₁₀, posing health risks associated with exposure. These compounds resulted from pyrolysis of the organic matrix of the reinforcement material at the saw blade edge during dry-cutting.

In contrast, the impact of thermal stress during domestic fires or thermal recycling on the abrasive dust from cutting C^3 materials has not been investigated yet. However, the thermal behavior of carbon fiber-reinforced plastics (CFRPs) with a comparable composition as C^3 reinforcement materials is well studied. No changes in the surface morphology of fibers or mass loss due to fiber degradation were found in scanning electron microscopy (SEM), and thermal analysis (TA) experiments at temperatures below 800 °C under N₂ atmosphere [25–27]. However, these fibers are coated by a pyrolytic carbon layer resulting from the degradation of the polymer matrix. This layer can impair the bonding properties of recycled fibers with their new matrix [25]. Oxygen-containing atmospheres can remove the pyrolytic carbon from the fiber surface due to oxidation, potentially altering the fiber morphology and structure [25–31]. Degradation effects, such as localized fiber oxidation and pitting processes [28,30,31] or total mass loss [25,27], were observed at a broad temperature range between >540 °C (Yatim et al. (2020) [26]) and >700 °C (Tranchard et al. (2017) [27]).

The aim of this study was to investigate the impact of thermal stress on the release of toxic organic pollutants and harmful fibers during abrasive processing of a C³ material. This experimental design primarily mimics the building demolition scenario after a domestic fire event. Moreover, this study may also act as a proxy for abrasive processing after

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recycling under oxidative conditions, e.g., to remove the pyrolytic carbon layer. For this purpose, chemical speciation techniques, such as mass spectrometric detection of volatile organic products, and electron microscopy, to study fiber morphology, were deployed. Furthermore, this study was motivated by giving advice regarding the personal protective equipment and the general health risk of construction workers involved in the demolition of C^3 -reinforced buildings after domestic fires. It is expected that elevated temperatures will affect the organic composition and fiber morphology of the C^3 reinforcement materials. This may have an impact on the toxic profile of PM formed during subsequent abrasive processing. Particularly, morphological investigations were of interest, as a study of Černý et al. (2007) [32] on cement reinforced with short, disordered pitch-based carbon fibers treated at 600 and 1000 °C (air atmosphere, 1 h), and preliminary studies on thermally stressed C^3 , showed a decrease in fiber diameter and, thus, an increased health concern.

2. Materials and Methods

2.1. Samples and Equipment

A C³ material reinforced with an epoxy resin-coated carbon grid of the type "solidian GRID Q95/95 CCE-38" manufactured by Solidian GmbH (Albstadt, Germany), and its concrete reference, were studied. The carbon fibers of the reinforcement material were made from PAN precursor fibers. Concrete plates were manufactured in a pouring process. The concrete was poured into a 175 mm \times 295 mm \times 15 mm cuboid form. One layer of the reinforcement grid was centered in the concrete and the plates were stored in a dry environment. It was always ensured that the reinforcement material was completely covered by concrete. To expose the concrete plates to thermal stress, a muffle furnace of the type "L9/S27" from Nabertherm GmbH (Lilienthal, Germany) was used. The reinforcement material, and a C³ sample before and after thermal exposure, are shown in the Supplementary Material (Figure S1b,c).

For homogenization of thermally stressed reinforcement materials, required for chemical analysis, a cryomill of the type "6750 Freezer/Mill" from SPEX Certiprep Inc. (Metuchen, NJ, USA) was used. Cutting of thermally stressed C³ was performed using an "Abrasive Dust Generator" from Vitrocell Systems GmbH (Waldkirch, Germany) equipped with a diamond saw blade. To measure particle size distributions of the PM formed during the dry-cutting process, a laser aerosol particle sizer "LAP322" from Topas GmbH (Dresden, Germany) was used (Figure S1a). Filter samples were collected using a filter control device "LVS3.1" from Comde-Derenda GmbH (Stahnsdorf, Germany). Quartz fiber (47 mm, grade T293) and methylcellulose filters (47 mm, 1.2 μ m) were purchased from Munktell and GE Healthcare Life Sciences, respectively. Quartz fiber filters were heated at 550 °C for 12 h prior to sampling. Gases were purchased with a purity of >99.9% and used without further purification. Samples were stored at -20 °C until analysis.

2.2. Sample Preparation

2.2.1. Thermal Stress

Concrete plates were heated in the muffle furnace under air atmosphere with a temperature ramp of 1.5 K/min and held at the final temperature for 3 h. Directly thereafter, the concrete plates were removed from the muffle furnace and allowed to cool under air atmosphere (cooling period of 24 h) until room temperature was reached. These conditions were sufficient to prevent cracking of the concrete and, thus, an increase in oxygen transportation rate from the concrete surface towards the reinforcement material, achieving high repeatability. Three C³ plates were heated to 25, 100, 200, 250, 300, 312, 325, 338, 350, 400, 450, 500, 550, and 600 °C (14 temperatures, n = 3). Furthermore, three concrete reference plates were heated to 25, 325, and 600 °C (3 temperatures, n = 3). C³ reinforcement materials were mechanically separated from thermally stressed C³ and homogenized using the cryomill (Supplementary Material, Section S2).

2.2.2. Dry-Cutting Simulation

Thermally stressed concrete and C^3 plates were cut in a dry-cutting simulation using the Abrasive Dust Generator as described earlier by Koch et al. (2021) [23]. An illustration of the dry-cutting simulation arrangement is provided in the Supplementary Material (Figure S1a). Dry-cutting was performed with a sawblade rotation of 50 Hz, cutting depth of 10 mm and cutting speed of 0.1 mm/s for 15 min. It was always ensured to cut along and through the rovings at their full depth to produce concentrations of organic compounds and fibers comparable to realistic demolition scenarios. To obtain reproducible aerosols, the sampling process was started when the deepest point of the saw blade entered and stopped before the foremost point of the saw blade left the concrete plate. The inhalable PM fractions PM_{2.5} and PM₁₀ were collected with a flow rate of 1.0 m³/h on quartz fiber filters for chemical analysis and methylcellulose filters for morphological analysis using PM_{2.5} and PM₁₀ impactors. One quartz fiber filter and one methylcellulose filter per PM fraction and concrete plate were collected to gather information about the repeatability of the whole sampling process, including the thermal stress period.

2.3. Analytical Methods

2.3.1. Online Particle Size Determination

Particles produced during dry-cutting of thermally stressed concrete plates were determined online by the laser aerosol particle sizer in order to obtain information about the particle size distribution and concentration, as described earlier by Koch et al. (2021) [23]. Particle size distributions and concentrations/cm³ × minute were recorded and averaged over the whole sampling period. Scattered light-equivalent diameters of particles measured by this method are limited to a range from 0.2 to 40 μ m and concentrations below 10⁴ particles/cm³. Calibration was performed using 13 monodisperse particle size standards containing spherical polystyrene latex (ISO 21501).

2.3.2. Pyrolysis Gas Chromatography—Mass Spectrometry

Pyrolysis gas chromatography (Py-GC) was applied as a standard technique to investigate the organic composition of the homogenized reinforcement materials mechanically separated from the thermally stressed C³. Therefore, roughly 0.2–0.9 mg of the homogenized reinforcement material was injected into the pyrolyzer (model PY-2020iD, Frontier Laboratories) connected to a HP 6890 gas chromatograph mass spectrometer. The sample was pyrolyzed at 500 °C for 1 min. Evolved gases were separated by an SGE-BPX5 column (30 m × 250 µm I.D. × 0.30 µm film) using the following temperature program: hold for 5 min at 40 °C, ramp to 330 °C with 10 K/min, and hold for 10 min. Mass spectrometric detection of the effluent gases from the column was performed by 70 eV electron ionization (EI) and analysis by a quadrupole (Q) mass analyzer. Mass spectra were recorded in scan mode from m/z 40 to 500 [33–35] (details in Supplementary Material, Section S3.1).

2.3.3. Thermal Analysis—Mass Spectrometry

A thermobalance (TG 209, Netzsch Gerätebau, Selb, GER) was hyphenated to a modified Bruker GC-APCI II source. Atmospheric pressure photoionization (APPI) performed by a Kr vacuum ultraviolet lamp preserving the molecular pattern was used as a soft ionization technique [36]. Thermal analysis was carried out with a constant nitrogen flow of 200 mL/min holding the temperature isothermal for 2 min at 20 °C, raising to 600 °C with 10 K/min, and holding for 10 min at 600 °C. Evolved gases were transferred into the ion source of the mass spectrometer via a 280 °C interface and 280 °C transfer line. Mass spectrometric detection was carried out with a Bruker Apex II ultra FTMS equipped with a 7 T superconducting magnet recording from m/z 100 to 1000 with a two-second transient length (4M datapoints), achieving a resolving power of roughly 300,000 at m/z400. Time-resolved broadband spectra, each with ten summed micro scans, were recorded. Time-resolved processing and sum formula calculation was performed using Bruker Data-Analysis 5.1 for m/z precalibration of the summed mass spectra and a self-written tool CERES based on in-house MATLAB scripting (R2018b). Elemental composition attribution for MS spectra was performed with the following restrictions: #C 2–100, #H 4–200, #N 0–2, #O 0–6, #S 0, H/C ratio 0.4–2.4, ring and double bond equivalent (DBE) 0–30, and a maximal sum formula error of 1 ppm. A detailed description of the setup and methodology can be found elsewhere [35,37] (details in Supplementary Material, Section S3.2).

2.3.4. Thermo-Optical Carbon Analyzer—Mass Spectrometry

Filter punches with an 8 mm diameter were taken from the quartz fiber filter samples obtained during the dry-cutting simulation of thermally stressed concrete and measured according to the Improve A protocol [38] with a thermo-optical carbon analyzer (TOCA) of the type "Model 2001 A" from Desert Research Institute (Reno, USA). This technique yields information about the carbon composition of the inhalable PM by thermally releasing carbon-containing compounds from the PM, oxidizing them to carbon dioxide and subsequently reducing the CO_2 to methane, which is measured by a flame ionization detector (FID). A detailed description of the setup and analysis principle can be found elsewhere [23,39–41]. Organic carbon (OC) was released during four thermal fractions (OC1, OC2, OC3, and OC4) in helium atmosphere at 140, 280, 480, and 580 °C, respectively. By adding 2 v-% oxygen, elemental carbon (EC) was oxidized to CO_2 and evolved in three thermal fractions (EC1, EC2, and EC3) at 580, 740, and 840 °C, respectively. Optical correction of pyrolyzed OC was performed by measuring the filter reflectance using a laser with a wavelength of 635 nm. A decrease in laser reflectance after OC2 for C^3 samples exposed to temperatures \leq 350 °C indicated enhanced pyrolysis of organic matrix molecules during analysis of these samples. Hence, OC1 and OC2 were combined to form the fraction thermal desorption OC. Since the fractions OC3 and OC4 most likely contain pyrolysis compounds formed during the analysis procedure, those fractions were neglected. To ensure a sufficient comparability of all samples, C³ samples exposed to temperatures >350 °C prior to abrasive treatment were treated equally during data evaluation, although no decrease in laser reflectance was observed.

To gather further information about the organic composition of the inhalable PM fractions formed during dry-cutting of thermally stressed C^3 , the TOCA was hyphenated to a photo ionization mass spectrometer (PIMS) [39–41]. Resonance-enhanced multiphoton ionization (REMPI) [42–44] using a wavelength of 248 nm is an efficient technique to obtain information about the aromatic constituents in abrasive PM from C^3 [23]. Thus, this wavelength produced by a krypton fluoride excimer laser of the type PhotonEx (Photonion GmbH, Schwerin, GER) and a repetition rate of 500 Hz was used in this study. The formed ions were analyzed by an orthogonal accelerated time-of-flight (oa-Tof) analyzer with an extraction rate of 20 kHz into the mass analyzer. The TOCA-PIMS coupling was manufactured by Photonion GmbH (Schwerin, Germany).

2.3.5. Scanning Electron Microscopy

 C^3 reinforcement materials were mechanically separated from thermally stressed C^3 and cropped to ~1 cm long segments. Methylcellulose filters were prepared for SEM measurements by carefully cutting out a circular segment of about 4 cm². The samples were mounted on Al-SEM-carriers with adhesive conductive carbon tape (PLANO GmbH, Wetzlar, Germany). To minimize the induction of a charging process, the reinforcement materials were coated under vacuum with a thin carbon layer of about 5 nm, and the filter samples were sputtered with a 1–2 nm platinum layer using a CCU 010 HV-coating device (Safematic GmbH, Zizers, Switzerland) and a SCD-040 sputtering device (Bal-Tec, Pfäffikon, Switzerland), respectively. SEM measurements were performed with field emission devices of the type MERLIN[®] VP Compact (Zeiss, Oberkochen, Germany, acceleration voltage 5 kV, operating distance 11 mm) and GeminiSEM 300 (Zeiss, Oberkochen, Germany, acceleration voltage 2 kV, operating distance 5 mm), respectively.

Quantification of harmful WHO fibers was conducted in accordance with the VDI guideline 3492 [45]. To eliminate fluctuation in the WHO fiber concentration when cutting

several thermally stressed C³ plates, 60–70 fibers on each filter were counted and sized, analyzing 33% of the filter area using a $2000 \times$ magnification. Hence, in total, at least 180 fibers per exposure temperature were considered by performing three cuts through C³ plates that were thermally stressed independently from each other.

2.3.6. Statistical Evaluation

Released particle concentrations were divided into four sizing groups according to their scattered light-equivalent diameter described in Section 3.1. Particle concentrations of the different sizing groups were compared for all samples. Particle concentrations for each sizing group, and fiber dimensions including length, diameter and length-to-diameter (L/D) aspect ratio, were additionally compared between all samples. Statistical evaluation was performed by a two-sided Student's t-test. Significance was accepted with $p \leq 0.05$.

3. Results and Discussion

3.1. Particle Size Distribution

Concentrations of particles formed during abrasive machining of C³ and concrete are shown in Figure 1. The particles were categorized by their scattered light-equivalent diameter into the following groups: 0.2–0.5 μ m, 0.5–1.0 μ m, 1.0–2.5 μ m, and 2.5–10.0 μ m. Particles with diameters >10 μ m are negligible. Furthermore, the ratio between fine particles with 0.2–1.0 μ m diameter and larger particles with 1.0–10.0 μ m diameter is indicated. According to the literature [46,47], PM with aerodynamic diameters of \leq 2.5 μ m (PM_{2.5}) and \leq 10 μ m (PM₁₀) can penetrate into the alveolar and tracheobronchial regions of the mammal respiratory system, respectively. Thus, those are considered to have the largest impact on the toxicity of PM. Concentrations of PM with equivalent scattered light-equivalent diameters as PM_{2.5} (0.2–2.5 μ m), PM₁₀ (0.2–10.0 μ m) and total PM (0.2–40 μ m) are summarized in Table 1.



□ 0.2–0.5 μm □ 0.5–1.0 μm □ 1.0–2.5 μm □ 2.5–10.0 μm • ratio 0.2–1.0 μm/1.0–10.0 μm

Figure 1. Particle concentration per cm³ and minute divided into four groups by their scattered light-equivalent diameter formed during abrasive processing of thermally stressed C³ materials and concrete (25–600 $^{\circ}$ C, air atmosphere, 3 h). Error bars indicate the standard deviation of the distinct sizing group.

	0.2–2.5 μm [particles/cm ³ * min]	0.2–10 μm [particles/cm ³ * min]	0.2–40 μm [particles/cm ³ * min]
C ³			
25 °C	7600 ± 200	8600 ± 300	8600 ± 300
100 °C	6400 ± 200	7300 ± 200	7400 ± 200
200 °C	6700 ± 200	7600 ± 200	7600 ± 200
250 °C	6200 ± 200	7000 ± 200	7000 ± 200
300 °C	8000 ± 300	8800 ± 300	8800 ± 300
312 °C	7800 ± 200	8500 ± 300	8500 ± 300
325 °C	7500 ± 200	8200 ± 100	8300 ± 100
338 °C	7900 ± 200	8800 ± 200	8800 ± 200
350 °C	7700 ± 200	8500 ± 200	8600 ± 200
400 °C	8300 ± 300	8900 ± 200	8900 ± 200
450 °C	7100 ± 300	7600 ± 200	7700 ± 300
500 °C	7100 ± 200	7500 ± 200	7500 ± 200
550 °C	4700 ± 200	4900 ± 200	5000 ± 200
600 °C	4500 ± 200	4800 ± 200	4900 ± 200
Concrete			
25 °C	6900 ± 100	7800 ± 200	7900 ± 200
325 °C	4500 ± 200	5100 ± 200	5100 ± 200
600 °C	3800 ± 100	4000 ± 100	4000 ± 100

Table 1. Concentrations of inhalable PM fractions with scattered light equivalent diameters of 0.2–2.5 μ m, 0.2–10 μ m and 0.2–40 μ m in particles per cm³ and minute formed during dry-cutting of thermally stressed C³ (25–600 °C, air atmosphere, 3 h). These size groups imitate the inhalable PM fractions PM_{2.5}, PM₁₀ and total PM, respectively, categorized by the aerodynamic diameter.

During abrasive processing of C³ exposed to 25 °C, the concentration of fine particles (0.2–1.0 µm in diameter) is $1.5 \times$ higher compared to larger particles (1.0–10.0 µm in diameter). At 400 °C, the concentration of fine particles increases slightly (p = 0.0005, reference: 350 °C), while the number of larger particles decreases (p = 0.0002, reference: 350 °C). Exposure to temperatures \geq 450 °C leads to a slight decrease in the concentration of fine particles (p = 0.002, reference: 400 °C) in the 0.5–1.0 µm diameter group. In contrast, the concentration of larger particles decreases more severely (p = 0.002, reference: 400 °C) in the 1.0–2.5 µm and the 2.5–10.0 µm diameter groups. These effects lead to an overall increase in the ratio between fine particles and larger particles from 1.5 at 25 °C to 4.1 at 550 °C. However, the concentration of particles in all size groups released during dry-cutting of C³ decreases when C³ is exposed to \geq 450 °C prior to abrasive processing.

Concrete without reinforcement materials exposed to temperatures of 600 °C also shows a significant ($p = 1 \times 10^{-15}$) increase in the ratio between fine particles and larger particles from 1.6 at 25 °C to 3.5 at 600 °C. This is due to a larger decrease in the concentration of larger particles compared to the 0.2–1.0 µm diameter fraction. Hence, those effects cannot be solely assigned to C³ materials. Nonetheless, the released concentration of total PM during dry-cutting of concrete is considerably smaller ($p = 1 \times 10^{-11}$) compared to C³, when the material is exposed to 325 °C prior to abrasive machining. At 300–350 °C, the organic coating of the reinforcement material pyrolyzed (Section 3.2). Thus, the reduced particle concentration released at 325 °C may be affected by the pyrolysis of the organic coating of the reinforcement material. However, the data gathered in this study are not sufficient to fully investigate this effect.

It must be acknowledged that the measurement of scattered light-equivalent diameters is dependent on the particle shape, the absorption potential, and the angle between light and particle [48]. The released aerosols formed during dry-cutting contained a mixture of rather spherical and non-spherical concrete particles, and fibrous carbon particles. Hence, artifacts due to the measurement method can occur and an underestimation of particle diameters cannot be excluded.

3.2. Organic Composition

The thermally treated samples were analyzed using various analytical techniques to determine the differences in the organic composition before and after the cutting process. A simple way to determine chemical compounds on the carbon fibers is thermal analysis. It becomes clear that after treatment of \geq 400 °C for an extended time period of several hours, no organic molecules can be detected to evaporate from the fibers (Figures S2 and S3). Below this temperature, the fraction of evaporable compounds increases (Table 2). Therefore, the fibers treated with 350 °C and lower temperatures are within the focus of chemical analysis. It is noticeable that while the evaporable fraction decreases with increasing exposure temperature, the number of sum formulae increases. During the thermal exposure period, partial pyrolysis of the organic coating of the reinforcement material results in the evaporation of small pyrolysis products and, thus, a reduction in the ratio between the organic coating and carbon fibers. Furthermore, non-evaporable pyrolysis products of the organic coating are formed during the thermal exposure period. In subsequent thermal analysis, those pyrolysis products and the remaining organic coating material are pyrolyzed, leading to a more complex pyrolytic degradation pathway. As a result, the number of sum formulae increases although the evaporable fraction decreases.

Table 2. Percentage of evaporable organic compounds, residue and the number of attributed sum formulae in fibers from thermally stressed C³ (25, 300, 325 and 350 °C, air atmosphere, 3 h) measured by TA-APPI-MS (10 K/min, 600 °C, hold for 10 min). Compounds evaporating during the temperature program (25–600 °C) are combined as the evaporable fraction. The remaining fraction is defined as residue.

Exposure Temperature [°C]	Residue [wt%]	Evaporable Fraction [wt%]	Number of Sum Formulae
25	71.2	28.8	1897
300	76.5	23.5	2951
325	84.9	15.1	2274
350	93.2	6.8	2584

3.2.1. Organic Composition of Reinforcement before Cutting

Evolved gas analysis [49] is a useful tool for the investigation of fiber-reinforced plastics (FRPs) [50,51]. In this process, the polymer matrix is thermally degraded into smaller components, which are subsequently analyzed using a spectrometric/spectroscopic method. Herein, Py-GC-EI-QMS was used for this purpose. The corresponding total ion count (TIC) chromatogram is shown in Figure 2. Two regions of interest (10–20 min/26–32 min) can be seen for the different samples. In the first region (Figure S5), almost exclusively small compounds are observed for the thermally treated materials, which do not occur in the room temperature sample. The authors hypothesize that these smaller constituents are thermal fragments, primarily from the coating material. Structural assignment of these released chemicals was carried out by the hard electron ionization fragment pattern and database matching (Table S1). The various phenols originate from the thermal degradation and release of bisphenol A (BPA), which is used as a building block in the polymer matrix (BPA as dominant peak: 26.74 min, Figure 2). Presumably, rearrangements occur at about 300 °C under the influence of oxygen, which favors the formation of smaller molecules during pyrolysis at elevated temperatures. Since the fiber diameter does not change at these temperatures (Section 3.3.1), a significant influence on the carbon fibers can be excluded.



Figure 2. Total ion count (TIC) chromatogram of reinforcement material mechanically separated from thermally stressed C³ (RT (25 °C), 300, 325, 350 °C) measured by Py-GC-EI-QMS. Exposure to room temperature leads to characteristic larger pyrolysis fragments, e.g., BPA derivatives, at higher retention times. C³ exposed to higher temperatures shows a variety of additional smaller compounds at 300 and 325 °C, e.g., phenol derivatives, homologous of benzene, and other methoxyphenols, and a drastic decrease in detected compounds at 350 °C.

The second region (26–32 min, Figure S6) primarily belongs to the resin matrix coating components. The main component BPA can be identified as the base peak. In addition, several other intense compounds are present and elute after the BPA signal. Based on manual investigation of the fragmentation pattern, these signals presumably have similar structural motives but are not available in the database for automated attribution (Table S1). Hence, to obtain in-depth information and verify these findings, a soft ionization approach, thermal analysis (thermogravimetry) with APPI, was applied. APPI preserved the molecular information, and with the help of high-resolution mass spectrometry, the exact masses of the released compounds can be determined with sub-ppm deviation. From the exact mass, the sum formulae can directly be obtained as chemical information. Using this technique and comparing the sum formulae to previously published structures assigned to pyrolysis of epoxy resin-coated C³ reinforcement materials [23], compounds with the sum formulae C₁₅H₁₆O₂ and C₁₈H₁₈O₂ occurring in the second region could be confirmed as BPA benzofuran derivatives (Figure S6 and Table S1). The same structural motives could be depicted from the APPI mass spectra, thus verifying the outlined Py-GC-EI-QMS results.

Based on the number of attributed sum formulae (Table 2), it is noticeable that the complexity increases significantly with increasing thermal treatment temperature. With 40.2% (1573 sum formulae), most compounds are found in all samples exposed to 300, 325, and 350 °C (Figure S7). Although the samples exposed to 300 and 350 °C form a large pool of unique compounds, of 24.3% (948 sum formulae) and 15.5% (606 sum formulae), respectively, the sample exposed to 325 °C only contains 1.4% (53 sum formulae) of unique compounds. Hence, a direct shift in the compound pattern from 300 to 350 °C takes place during the thermal exposure period.

For further investigation of the complex data, the molecular attributions can be grouped into compound classes. To ensure comparability with the TOCA-PIMS data, data evaluation was restricted to the thermal desorption fraction (20–280 °C). This approach revealed large differences between the samples (Figure S8). With increasing temperature from 300 to 350 °C, the amount of oxygen- and nitrogen-containing compounds in the evaporated mixture decreases significantly due to evaporation of these compounds during the

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thermal treatment prior to thermal analysis. In contrast, the quantity of these compounds is significantly lower in the sample exposed to room temperature. Nitrogen compounds, such as isophorone diamine or the triglycidylether of para-aminophenol [52], are introduced into the matrix as hardeners, as cross-linker in the coating. Oxygen-containing groups mainly belong to benzofuran derivatives of BPA and phenolic thermal degradation products of BPA (Table S1). Nitrogen is most likely introduced into these compounds by thermal degradation of the hardener in larger epoxy resin molecules during the thermal exposure period. In subsequent abrasive machining steps, those compounds may be released.

3.2.2. Organic Composition of Abrasive PM

The inhalable PM fractions PM_{2.5} and PM₁₀ released during dry-cutting of thermally stressed C³ materials were analyzed by TOCA-PIMS. At exposure temperatures \leq 250 °C, only small concentrations of thermal desorption OC (20–280 °C) were found in PM₁₀ with significantly higher concentrations in C³ compared to concrete at room temperature exposure (Figure S13). At exposure temperatures of 300–312 °C, three-fold higher thermal desorption OC concentrations could be detected. Higher exposure temperatures of \geq 325 °C lead to a decrease of the thermal desorption OC content.

For further information about the thermal desorption OC composition, the samples were investigated with REMPI mass spectrometry preserving the molecular pattern (Figure S3). Compounds detected in the abrasive PM formed during dry-cutting of C^3 exposed to elevated temperatures are summarized in Table 3. Averaged REMPI spectra of PM_{2.5} and PM₁₀ from C^3 exposed to 300 °C are illustrated in Figure 3. The corresponding REMPI spectra for exposure temperatures of 312, 325, 338, and 350 °C are provided in the Supplementary Material (Figures S14–S17).

Table 3. Organic compounds detected in the inhalable PM fractions $PM_{2.5}$ and PM_{10} formed during dry-cutting of thermally stressed C³ (25–600 °C, air atmosphere, 3 h). Differences between $PM_{2.5}$ and PM_{10} occurred only in the signal intensities. The compounds were observed by TOCA-REMPI-MS and evaluated by comparison with the Py-GC-EI-QMS and TA-APPI-MS data.

#	Identified Substance	Compound Name	m/z	25–250 °C	300–312 °C	325–350 °C	400–600 °C
1	ОН	Phenol	94	Х	Х	-	-
2		4-isopropyl phenol	136	-	Х	-	-
3 4	C ₁₁ H ₁₂ O C ₁₂ H ₁₄ O	-	160 174	-	X X	X X	-
5		Phenanthrene	178	х	Х	Х	-
6		Anthracene	178	Х	Х	Х	-
7		Fluoranthene	202	-	Х	-	-
8		Pyrene	202	-	Х	-	-
9	ССССОН	4-(2-phenyl-2-propanyl) phenol	212	х	Х	-	-
10	Me	n-methyl-5-[1-methyl-1- (n-methylbenzofuran-5-yl) ethyl] benzofuran	304	-	Х	Х	-



Figure 3. Averaged REMPI spectra (248 nm, 500 Hz) of thermal desorption OC fraction (20–280 °C, OC1 + OC2) of PM_{2.5} (**a**) and PM₁₀ (**b**) formed during dry-cutting of C³ exposed to 300 °C (air atmosphere, 3 h).

The release of organic compounds can be divided into three major exposure temperature regions: \leq 250, 300–350, and \geq 400 °C. The same behavior for PM_{2.5} and PM₁₀, only differing in their intensities, can be found. At temperatures ≤ 250 °C, phenol (*m*/*z* 94), dehydroxylated BPA (m/z 212) and the PAHs phenanthrene/anthracene (m/z 178) were detected. Those compounds can be traced back to pyrolysis of the epoxy resin coating at the saw blade edge during the dry-cutting process, as described earlier by Koch et al. (2021) [23]. Higher exposure temperatures of 300–350 °C lead to the release of a different molecular pattern. At 300-312 °C (Figures 3 and S14), a benzofuran derivative of BPA (m/z 304), three phenolic BPA degradation products (m/z 136, 160, and 174), and the PAHs fluoranthene/pyrene (m/z 202), were detected additionally. The complexity of the molecular pattern decreases with increasing exposure temperatures up to 350 °C in this temperature region. At high exposure temperatures ≥ 400 °C, no organic compounds were detected. Those molecular patterns are caused by a slow pyrolysis process of the organic coating material during thermal exposure leading to partial fragmentation of the epoxy resin (300–350 °C) and complete evaporation of the organic coating (\geq 400 °C). These results largely confirm the results obtained from Py-GC-EI-QMS and TA-APPI-MS.

Fully developed fires reach temperatures up to 1050 °C [53,54]. Hence, most likely, no organic compounds will be released during abrasive demolition of C^3 buildings after fully developed domestic fire events. However, when fires are extinguished by firefighters at an early stage, temperatures of 300–350 °C can occur. This can lead to the release of BPA derivatives and phenolic degradation products, in addition to the PAHs phenan-threne/anthracene and fluoranthene/pyrene. BPA is a well-known endocrine disruptor that can impair the reproductive system, in addition to liver and kidney function [55]. No information about the toxicity of the identified BPA derivatives is available yet. In addition, PAHs are carcinogenic [56]. Thus, we recommend that the pyrolysis products of the organic coating should be considered during the manufacturing process of C^3 reinforcement materials. Furthermore, we recommend using organic coating materials not based on BPA-based epoxy resin to minimize the risk of toxic effects.

3.3. Fiber Morphology

Carbon fibers were investigated with respect to their morphology before and after drycutting of thermally stressed C^3 . Therefore, reinforcement material mechanically separated from thermally stressed C^3 and fibers in the inhalable PM fraction PM_{10} were studied. Furthermore, the impact of thermal stress on the occurrence and the concentration of harmful WHO fibers released during abrasive machining of thermally stressed C^3 were evaluated, allowing the definition of personal protection measures for handling thermally stressed C^3 materials [57]. Pictures of the reinforcement materials mechanically separated from thermally stressed C^3 are illustrated in the Supplementary Material (Figure S18). It is noticeable that the reinforcement structure is partially degraded at 312 °C due to the pyrolysis of the organic coating material. Exposure to 450 °C leads to a complete loss of the reinforcement structure and the formation of a fluffy bundle of isolated fibers.

3.3.1. Fiber Morphology before Cutting

Electron microscopy is a well-established technique to study fiber morphology [25,28,29]. SEM micrographs of fibers exposed to various temperatures (air atmosphere, 3 h) while embedded in concrete are illustrated in Figure 4a–f. Fibers exposed to low temperatures \leq 325 °C are all well-coated with the epoxy resin matrix. Each fiber is covered individually by a smooth epoxy resin film. Furthermore, the epoxy resin leads to the fibers adhering to each other. Degradation effects of the epoxy resin were not observed. Hence, a high stability of the reinforcement grid can be expected due to improved binding properties between the fibers and between the fibers and concrete particles. However, higher exposure temperatures \geq 450 °C lead to a total loss of the epoxy resin matrix due to degradation. As a result, the binding properties of the fibers are weakened and isolated fibers can be obtained. At high exposure temperatures \geq 550 °C, degradation effects of the carbon fibers, such as pitting processes, can be observed with small notches in the fiber surface and heavy weakening of the fibrous structure occurring at 550 and 600 °C (Figure 4e–f), respectively. These effects are caused by oxidation of the carbon fiber surfaces.

Figure 4g shows the median fiber diameters for exposure temperatures between 25 and 600 °C measured with SEM. At exposure temperatures \leq 500 °C, the carbon fibers have a median diameter of 7.1 \pm 1.5 µm (n = 147), which corresponds to the fiber diameter obtained due to the manufacturing process. Although the epoxy resin matrix degrades between 300 and 350 °C, the fiber diameter is retained. Thus, the degradation of the epoxy resin matrix has no influence on the fiber diameter. However, at high exposure temperatures \geq 550 °C, the fiber diameter is substantially decreased by a factor of 1.3 and 2.3 at 550 and 600 °C, respectively. The reduction in the fiber diameters is caused by the oxidation processes at the fiber surface occurring at high exposure temperatures \geq 550 °C. This leads to the formation of fibers with diameters <3 µm at exposure temperatures of 600 °C. Thus, harmful fibers, according to the WHO definition [24], with length >5 µm, diameter <3 µm and a length-to-diameter ratio >3, may occur in subsequent processing steps such as abrasive machining.



Figure 4. SEM micrographs (1000× magnification). Morphology of epoxy resin-coated carbon fibers extracted from thermally stressed C³ (25–600 °C, air atmosphere, 3 h, (**a**–**f**)). Pitting processes occurring at \geq 550 °C lead to a decrease in the median fiber diameter (**g**). WHO fiber diameters <3 µm are indicated by a red line and arrow.

3.3.2. Fiber Morphology/Occurrence in Abrasive PM₁₀

The inhalable PM fraction PM_{10} was investigated with respect to the fiber dimensions and the occurrence of harmful fibers. SEM micrographs of fibers in PM_{10} from a dry-cutting simulation of C³ exposed to 500 and 600 °C are shown in Figure 5a,b. Although, in PM_{10} from C³ exposed to lower temperatures \leq 450 °C (Figure S19) only isolated fibers with 7 ± 1 µm in diameter were observed, as already shown by Koch et al. (2021) [23] for nonthermally stressed C³, higher exposure temperatures >450 °C lead to a 4.5-fold increase (reference: 450 °C) at 500–550 °C and a 15-fold increase (reference: 450 °C) at 600 °C in the fiber number on the filters due to the isolation of fibers decomposed in the thermally stressed concrete. Although fibers in PM_{10} of C³ exposed to 500 °C show only small notches on their surface due to localized oxidation processes, exposure temperatures of 600 °C lead to a heavy weakening in the fibrous structure of the carbon fibers. Thus, the formation of harmful WHO fibers becomes more likely with increasing temperatures above 500 °C as the fiber diameter may fall within the speciation prior to the dry-cutting.



Figure 5. SEM micrographs (5000× magnification). Morphology of fibers in PM₁₀ formed during dry-cutting of C³ exposed to 500 (**a**) and 600 °C ((**b**), air atmosphere, 3 h). Pitting processes start to occur at 500 °C, leading to decreasing fiber diameters (**c**), while fiber lengths (**d**) show no clear trend. The speciation for harmful fibers according to the WHO definition (length > 5 µm, diameter < 3 µm) is indicated by a red line and arrow.

The impact of thermal stress on the dimensions of fibers released during dry-cutting of thermally stressed C³ exposed to temperatures ≥ 450 °C (air atmosphere, 3 h) were further investigated by measuring the length and diameter of fibers in the inhalable PM fraction PM₁₀. The corresponding lengths and diameters are shown in Figure 5c,d. It is observed that the median length of the fibers does not change. However, the median fiber diameter decreases with exposure temperatures above 450 °C. Although at 450 °C only isolated fibers with diameters ~3 µm are observed, exposure temperatures ≥ 550 °C lead to the formation of a larger number of fibers with diameters <3 µm and, thus, harmful WHO fibers. The corresponding L/D ratios are predominantly within the speciation of the WHO definition with a value >3 (Figure S20). Table 4 summarizes the measured length and

diameter spans of the carbon fibers compared to asbestos fibers. In comparison to asbestos fibers, the carbon fibers are longer and have a larger diameter.

Table 4. Spans of fiber length and diameters in abrasive PM_{10} formed during dry-cutting of thermally stressed C³ (450, 500, 550, and 600 °C, air atmosphere, 3 h) in comparison to asbestos fibers.

Exposure Temperature [°C]/material	Fiber Length Span [µm]	Fiber Diameter Span [µm]
$450/C^{3}$	10.3–769.2	2.6-8.1
$500/C^{3}$	22.3–1573	3.8–7.4
$550/C^{3}$	1.6–3447	0.1–7.0
$600/C^3$	5.4-1514	0.1–6.0
room temperature/asbestos	0.2–38.0 [58]	0.01–2.9 [58]

Since the highest risk of exposure to harmful WHO fibers occurs during dry-cutting of C^3 exposed to high temperatures of 550 and 600 °C, the concentrations of these two scenarios were measured in accordance to VDI guideline 3492 [45]. Dry-cutting of C³ exposed to 550 and 600 °C leads to a release of 27,940 WHO fibers/m³ (32% of all fibers, 2643 WHO fibers/cm³ cut volume, cut volume: 10.57 cm³/m³ analyzed volume) and 40,500 WHO fibers/m³ (46% of all fibers, 3709 WHO fibers/cm³ cut volume, cut volume: 10.92 cm³/m³ analyzed volume), respectively. Exposure concentrations for the worker may be smaller due to a dilution of the aerosol stream between the spot of fiber release at the saw blade edge and the inhalation area of the worker. Data about the dilution behavior of released carbon fibers with respect to the distance from the cutting device are not available yet. For asbestos fibers, exposure concentrations of <10,000 WHO fibers/m³ [59] and 10,000-100,000 WHO fibers/m³ [59] are considered to cause a small or medium risk of cancer development, respectively, when a daily exposure on workdays for a time span of 40 years can be assumed [59]. Thus, to minimize health risks working with C^3 , the fiber concentrations should be decreased by using wet cutting processes and suction devices during cutting C^3 that is thermally stressed by high temperatures \geq 550 °C. Furthermore, construction workers working with C³ exposed to high temperatures should wear particle protection masks and protective clothes [57]. However, in fire conditions, the oxygen concentration may be reduced due to oxygen consumption. This may reduce the risk of fiber damage and lead to the release of lower WHO concentrations. Furthermore, shorter exposure times to high temperatures \geq 550 °C during domestic fire events are also expected to reduce the risk of fiber damage. In contrast, a higher risk of WHO fiber release can be expected when the C³ material is exposed to temperatures >600 $^{\circ}$ C (air atmosphere, 3 h) due to an increased risk of fiber damage. Typical temperatures expected in domestic fire events can reach up to 1050 °C [53,54].

In comparison to asbestos cement, the released WHO fiber concentrations are two orders of magnitude lower during cutting C^3 materials. Abelmann et al. (2018) [21] measured the exposure concentration to asbestos fibers during dry-cutting of asbestos cement pipes (two cuts, outer diameter: 10.2 cm, inner diameter: 7.6 cm, cut width: 0.3 cm, cut volume: 21.8 cm³/m³ analyzed volume) and found concentrations of 5,200,000 WHO fibers/m³ (238,533 WHO fibers/cm³ cut volume) to 12,400,000 WHO fibers/m³ (568,808 WHO fibers/cm³ cut volume). Hence, the herein observed WHO fiber concentrations indicate that the health risk for construction workers involved in the demolition of C³ buildings after domestic fire events is much lower compared to that involved in the demolition of asbestos cement buildings.

4. Conclusions

In this study, the release of toxic organic compounds and fibers during the dry-cutting of a Carbon Concrete Composite (C^3) was investigated. Prior, the material was thermally stressed at air for 3 h at various temperatures (25–600 °C). These conditions mimic building demolition after domestic fire events and abrasive processing after thermal recycling.

A release of a variety of epoxy resin degradation products, such as phenols, BPA benzofuran derivatives, and PAHs was observed. These pollutants are exposed due to the partial decomposition of the organic fiber coating. Consequently, a substantial environmental and health risk in abrasive building demolition scenarios after domestic fire events can be proposed. Hence, potential degradation products of the organic coating material should be considered in the manufacturing process.

In contrast, the release of WHO fibers observed due to a decrease in fiber diameter at elevated temperatures above 550 °C poses a medium health risk. A medium health risk is assumed when cancer development is sufficiently likely, but can be tolerated if further safety precautions are applied [59]. Hence, the exposure to WHO fibers during abrasive processing of thermally stressed C³ in domestic fire events or thermal recycling should be minimized by applying wet cutting processes, work protection equipment, and a suction device [57]. Since domestic fire conditions vary, future research should address the thermal impact on C³ materials using varying dwell times and atmosphere conditions. Moreover, the distribution behavior of the released carbon WHO fibers at the workplace should be defined by performing exposure measurements at various distances from the saw blade. This data can be used to establish an exposure model for worker and coworker exposure [60]. In addition, various C³ reinforcement materials can behave differently during thermal exposure. Therefore, studies on the decomposition behavior of different C³ reinforcement materials focusing on the fiber and organic coating degradation are currently being performed.

This study focused on the physicochemical description, and toxicity assessment was based on literature guidelines. Thus, future work will investigate the direct toxicity of dust from abrasive machining of thermally stressed C³. However, it is expected that the described organic pollutants will mainly drive the toxicity for low temperature scenarios (\leq 350 °C). In contrast, released WHO fibers will most likely be the main driving force when the C³ is exposed to high temperatures \geq 550 °C. In vitro studies addressing these questions are currently being carried out.

Supplementary Materials: The following Supplementary Materials are available online at: https:// www.mdpi.com/article/10.3390/fib10050039/s1. Figure S1: Illustration of the dry-cutting simulation and pictures of the C³ reinforcement material and a C³ material before and after thermal exposure to 400 °C. Figure S2: Organic composition of reinforcement materials mechanically separated from C³ exposed to 300-400 °C measured by TA-APPI-MS. Figure S3: Averaged REMPI spectra of thermal desorption OC in PM_{2.5} and PM₁₀ from dry-cutting of C³ exposed to 25–600 °C. Figures S4–S6: Total ion count chromatogram (TIC) and magnified TIC chromatograms (12-20 min/25-32 min) of reinforcement material mechanically separated from C³ exposed to 25, 300, 325, and 350 °C. Figure S7: Venn diagram of the sum formulae assignment for reinforcement materials mechanically separated from C³ exposed to 300, 325, and 350 °C. Figure S8: Absolute intensities of the compound classes grouped for the reinforcement materials mechanically separated from C³ exposed to 25, 300, 325, and 350 °C. Figures S9-S12: In-depth chemical characterization of the emitted gas mixture for the thermal analysis of reinforcement materials mechanically separated from C³ exposed to 25, 300, 325, and 350 °C by APPI-FT-MS. Figure S13: Concentrations of thermal desorption OC in PM2.5 and PM10 released from dry-cutting of C³ exposed to 25–600 °C. Figures S14–S17: Averaged REMPI spectra of $PM_{2.5}$ and PM_{10} released from dry-cutting of C³ exposed to 312, 325, 338, and 350 °C. Figure S18: Pictures of C³ reinforcement materials mechanically separated from C³ exposed to 25–600 °C. Figure S19: SEM micrographs of fibers in PM₁₀ from dry-cutting of C³ exposed to 25–450 °C. Figure S20: Length-to diameter (L/D) aspect ratio of carbon fibers in PM_{10} from dry-cutting of C³ exposed to 550, and 600 °C. Table S1: Organic constituents in reinforcement materials mechanically separated from C³ exposed to 300–350 °C. Furthermore, a description of the homogenization of reinforcement materials mechanically separated from thermally stressed C^3 is given in Section S2. Details of the Py-GC-EI-QMS and TA-APPI-MS measurements are given in Section S3.

Author Contributions: Conceptualization, A.K., L.F., C.P.R., T.S. and R.Z.; formal analysis, A.K. and L.F.; funding acquisition, R.Z.; investigation, A.K., L.F., P.F. and A.S.; methodology, A.K., L.F., P.F. and A.S.; project administration, T.S.; resources, A.K., L.F., P.F. and A.S.; supervision, R.Z.; validation, A.K., L.F. and C.P.R.; visualization, A.K. and L.F.; writing—original draft, A.K.; writing—review and editing, L.F., P.F., A.S., S.D.B., M.S., M.F., C.P.R., T.S. and R.Z. All authors have read and agreed to the published version of the manuscript.

Funding: This research is part of the C³—Carbon Concrete Composites project funded by the German Federal Ministry of Education and Research, grant number 03ZZ0332A, and Research in the loan program "Zwanzig20-Partnerschaft für Innovation. We acknowledge financial support by German Research Foundation (DFG) and University of Rostock/University Medical Center Rostock within the funding program Open Access Publishing, grant number 325496636.

Data Availability Statement: The data presented in this study are available in the article and the Supplementary Material.

Acknowledgments: The authors thank their colleagues from the Institute of Construction Management at the TU Dresden for providing the C³ samples. Funding by the Horizon 2020 program for the EU FT-ICR MS project (European Network of Fourier-Transform Ion-Cyclotron-Resonance Mass Spectrometry Centers, Grant agreement ID: 731077) is gratefully acknowledged. The authors thank the German Research Foundation (DFG) for funding of the Bruker FT-ICR MS (INST 264/56).

Conflicts of Interest: The authors declare no conflict of interest. The funders had no role in the design of the study, in the collection, analyses, or interpretation of data, in the writing of the manuscript, or in the decision to publish the results.

Abbreviations

APPI—atmospheric pressure photoionization; BPA—bisphenol A; C³—Carbon Concrete Composite; CFRP—carbon fiber-reinforced plastic; DBE—double bond equivalent; EC—elemental carbon; EI—electron ionization; FID—flame ionization detector; FRP—fiber-reinforced plastic; GC—gas chromatography; L/D—length-to-diameter aspect ratio; MS—mass spectrometry; oa-Tof—orthogonal accelerated time-of-flight; OC—organic carbon; PAN—polyacrylonitrile; PIMS—photoionization mass spectrometry; PM—particulate matter; Py—pyrolysis; QMS—quadrupole mass spectrometry; REMPI—resonance-enhanced multiphoton ionization; SBR—styrene-butadiene resin; SEM—scanning electron microscopy; TA—thermal analysis; TIC—total ion count; TOCA—thermo-optical carbon analyzer; WHO—World Health Organization.

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