



Article

Bacterial Disinfection by CuFe_2O_4 Nanoparticles Enhanced by NH_2OH : A Mechanistic Study

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Abstract: Many disinfection technologies have emerged recently in water treatment industry, which are designed to inactivate water pathogens with extraordinary efficiency and minimum side effects and costs. Current disinfection processes, including chlorination, ozonation, UV irradiation, and so on, have their inherent drawbacks, and have been proven ineffective under certain scenarios. Bacterial inactivation by noble metals has been traditionally used, and copper is an ideal candidate as a bactericidal agent owing to its high abundance and low cost. Building on previous findings, we explored the bactericidal efficiency of Cu(I) and attempted to develop it into a novel water disinfection platform. Nanosized copper ferrite was synthesized, and it was reduced by hydroxylamine to form surface bound Cu(I) species. Our results showed that the generated Cu(I) on copper ferrite surface could inactivate *E. coli* at a much higher efficiency than Cu(II) species. Elevated reactive oxygen species' content inside the cell primarily accounted for the strong bactericidal role of Cu(I), which may eventually lead to enhanced oxidative stress towards cell membrane, DNA, and functional proteins. The developed platform in this study is promising to be integrated into current water treatment industry.

Keywords: copper ferrite; hydroxylamine; Cu(I); reactive oxygen species; water pathogen

1. Introduction

Water pathogens are a great concern that threaten the safety of public drinking water. It has been reported that outbreaks of mass diseases in the cities are most likely linked to the failed disinfection facilities [1–3]. So far, water disinfection has been a widely researched topic, but nonetheless, available water disinfection techniques are limited. Most of the currently used water disinfection methods are chlorination, ozonation, and UV irradiation, and so on, all of which have significant drawbacks [4–8]. For example, chlorination byproducts after reaction with organic compounds in water are reported to be carcinogenic and are not generally avoided around the world [9,10]. Ozonation is a rather clean and powerful method, however, the electrochemical production of ozone relies on special anode materials and high applied voltage to surpass the overpotential of ozone evolution. Moreover, toxic nitric oxide species might also be generated during the electrochemical production of ozone [11,12]. Compared with the above two methods, UV irradiation is relatively simple and less vigorous, because lights in the UV range could penetrate the cell membrane to impair DNA and cause gene breakdown. While demonstrated in clear water bodies such as potable water, UV irradiation is ineffective in dark water bodies as the lights' travel is blocked [8]. In addition, it has to be noted that water

pathogens are observed to develop resistance to those traditional disinfection technologies, including chlorination/chloramination [13–15], ozonation [16–18], and UV irradiation [19–21]. Therefore, the development of more potent and environmentally-friendly techniques is necessary.

Recently, multifarious nanomaterials have been developed for bacterial inactivation purposes. For instance, a recent study reported that g-C₃N₄ nanolayers under visible light could kill *Escherichia coli* with high efficiency by generating reactive oxygen species [22,23]. However, the mass production of high quality g-C₃N₄ nanolayers has not been achieved so far, impeding the application of this material in the water disinfection industry. Graphene-based nanoparticles as a disinfectant have also attracted numerous attention [24,25], but face the same issue of upscaled production. Another widely used antibacterial reagent belongs to the silver-based material family [26–28]. Silver inactivates bacteria mainly through binding to the thiol groups of functional proteins and destroying the protein native structure [27]. The issue that prohibits silver from large-scale application into water disinfection industry is its high cost.

Copper has been used as an antibacterial material of low cost and easy manufacturing for hundreds of years [29–31]. It has been revealed that copper inactivates bacteria through a mechanism similar to silver's bactericidal role, that is, deactivating functional proteins by chelating thiol groups [27]. Further studies showed that, among all copper valency (Cu(0), Cu(I), and Cu(II)), Cu(I) owns the highest antibacterial activity, mainly because Cu(I) has strong binding affinity, as well as its reduction capability towards functional proteins [7]. These two synergistic functions of Cu(I) lead to its bacteria inactivation performance with hundreds of times higher efficiency than Cu(II) and Cu(0). It has been previously reported that Cu(II) could be reduced by hydroxylamine (NH₂OH) to produce Cu(I) with a high efficiency [32,33], but Cu(II) ion should be prohibited from drinking water because of its strong toxicity to human beings [34,35]. Specifically, the U.S. Environmental Protection Agency (EPA)-permitted copper ion concentration in drinking water is 1.3 ppm [36]. Herein, we demonstrate that heterogeneous copper ferrite (CuFe₂O₄) nanoparticles with minimum leached toxic copper ions, after reduction by hydroxylamine, show significantly higher antibacterial activity than that without hydroxylamine addition. In this study, we chose *E. coli* as the target species, mainly because it has wide infectivity in various water bodies [37,38]. The bactericidal mechanism by CuFe₂O₄/NH₂OH was also revealed in this study with several molecular probes. Overall, this controllable manner of chemical addition and powerful bactericidal performance could attract the attention of the water disinfection industry.

2. Materials and Methods

2.1. Materials

CuSO₄ (Alfa Aesar, Ward Hill, MA, USA), Fe₂(SO₄)₃•5H₂O (ACROS Organics, Morris Plains, NJ, USA) and dodecyltrimethylammonium bromide (Sigma Aldrich, Natick, MA, USA) were used to synthesize CuFe₂O₄ nanoparticles. Hydroxylamine (NH₂OH) was purchased from Sigma Aldrich (Natick, MA, USA). LB medium and agar from BD Difco (Pittsburgh, PA, US) were used to culture *E. coli* cells. MOPS buffer (ACROS Organics, Morris Plains, NJ, USA) was used to maintain the physiological integrity of cell membrane. HPF probe (3'-p-(hydroxyphenyl) fluorescein, Thermo Fisher Scientific, Bedford, MA, USA) and EDTA (Millipore Sigma, Burlington, MA, USA) were used to scavenge ROS (reactive oxygen species) generation. Milli-Q water was used throughout the study, and all nutrients were autoclaved before being used to culture *E. coli* cells.

2.2. Synthesis and Characterization of Copper Ferrite

We followed a reported procedure to synthesize copper ferrite nanoparticles [39]. Briefly, 0.1 M dodecyltrimethylammonium bromide was used as capping agent. Then, 1.6 g CuSO₄ and 4.9 g Fe₂(SO₄)₃•5H₂O was added into the solution to achieve a Fe/Cu molar ratio of 2:1. The solution was stirred by a magnetic bar for 15 min to totally dissolve copper and iron salts. Then, solution pH was

adjusted to pH 12.5 with 5 M NaOH, and the solution was then stirred for 45 min to allow sufficient precipitation. Subsequently, the solution was transferred into an autoclave vessel, and kept at 120 °C for 1 h. After hydrothermal treatment, the obtained powders were then extensively washed with hexane. Eventually, the powders were sintered in a 100 °C oven overnight. The produced CuFe₂O₄ nanoparticle samples were collected for future use.

Copper and iron ions concentration in solution were quantified with ICP (inductively coupled plasma). Then, 2% HNO₃ solution was used to dissolve particles. Copper and iron were analyzed with emission wavelengths of 324.754 nm and 259.940 nm, respectively.

The synthesized CuFe₂O₄ nanoparticles were then characterized by transmission electron microscopy (TEM, JEOL, Beijing, China). The crystal structure was analyzed by X-ray diffractometry using a Thermo Scientific ARL EQUINOX 1000 diffractometer, and X-ray photoelectron spectroscopy studies were performed utilizing a Thermo Scientific™ K-Alpha™ spectrometer to evaluate the electronic properties of elements on the surface of synthesized CuFe₂O₄ nanoparticles.

2.3. E. Coli Inactivation Assay

Exponential phase *E. coli* cells were used for the bacterial inactivation assays. A single *E. coli* K12 colony picked up from an LB-agar petri dish was cultured overnight in 5 mL LB medium at 37 °C. Then, 50 µL saturated *E. coli* cell solution was added into 5 mL fresh LB medium. The bacterial solution was then shaken at a speed of 250 rpm at 37 °C for around 2 h, until OD₆₀₀ reached ~0.7. Exponential phase *E. coli* cells were then collected by centrifugation at 5 g for 1 min, and washed extensively with 10 mM MOPS buffer (pH 7) to remove residual nutrients. Subsequently, cells were transferred in 5 mL of 10 mM MOPS buffer and stored at 4 °C. Bacterial solutions were used within the same day.

In a typical antibacterial assay, MOPS buffer was replaced by a solution containing 0.2 g/L CuFe₂O₄ and 2 mM NH₂OH with *E. coli* cells. The solutions were constantly shaken at 37 °C at a speed of 250 rpm. At each hour, 200 µL solution was withdrawn from the tube for analysis, and the bacteria survival rate was determined by a 10-fold serial dilution method in 96-well plates [7]. A volume of 10 µL from the six dilutions of each sample was dropped onto an LB-agar plate, and incubated at 37 °C overnight. To calculate the total survived cell number, CFUs (colony forming units) were counted.

In order to investigate whether the bactericidal role of CuFe₂O₄/NH₂OH was from the reaction between leached Cu²⁺ and NH₂OH, we performed a simulation experiment. It was determined that the leached Cu²⁺ concentration in solution was below 20 ppb. Therefore, 20 ppb Cu²⁺ was mixed with 2 mM NH₂OH, and bactericidal efficiency was then determined with the abovementioned assay.

2.4. ROS Quantification Assay

The ROS content in *E. coli* cells was quantified by a fluorescent HPF probe. In detail, bacterial cultures were incubated with 10 µM HPF at 37 °C. At 1, 2, and 3 h, and 200 µL was transferred into a 96-well plate for analysis. The fluorescence analysis was performed with excitation/emission maxima at 490/515 nm, respectively. The following positive and negative groups were researched [40]: (1). 0.2 g/L CuFe₂O₄, 2 mM NH₂OH, with *E. coli* cells; (2). 0.2 g/L CuFe₂O₄, with *E. coli* cells; (3). 2 mM NH₂OH, with *E. coli* cells; (4). 0.2 g/L CuFe₂O₄, 2 mM NH₂OH, without *E. coli* cells; and (5) *E. coli* cells. The (2)–(5) negative controls were used, to verify that the generated ROS signal in *E. coli* cells after CuFe₂O₄/NH₂OH reaction was indeed from the oxidative stress within bacterial cells. The results indicated that the (2)–(5) negative controls produced negligible fluorescence response (data not shown), suggesting that the used CuFe₂O₄ or NH₂OH chemical or *E. coli* cells have no influence on HPF fluorescence. Therefore, the HPF used in this method is valid for measuring ROS content in our study.

2.5. ROS Scavenging Assay

We conducted ROS scavenging assays, in order to verify that ROS played a vital role in inactivating *E. coli* cells. In detail, 2 mM EDTA or 10–100 mM DMSO was added into a bacterial solution containing 0.2 g/L CuFe₂O₄ nanoparticles and 2 mM NH₂OH. Reaction solutions were maintained at 37 °C,

and shaken at 250 rpm. The bacterial survival rate was determined by the abovementioned serial dilution method.

2.6. Recycling of Copper Ferrite

The reusability of heterogeneous CuFe_2O_4 nanoparticles was examined. After each round of aqueous reaction, CuFe_2O_4 nanoparticles were collected by centrifugation at 10 g for 5 min. The pellets were then transferred into an oven at 80 °C to heat for 2 h. Afterwards, the pellets were used for the subsequent round of bacterial inactivation assay.

2.7. Statistical Analysis

Bactericidal assays were performed with three independent replicates ($n = 3$), and statistical analysis was performed with *t*-test. Asterisks of *p*-values indicate the level of significance compared with *E. coli* control cells in MOPS buffer, that is, ** $p < 0.01$ and *** $p < 0.001$.

3. Results and Discussion

3.1. Characterization of Synthesized Copper Ferrite

The copper ferrite (CuFe_2O_4) used was synthesized via a hydrothermal method [39]. After synthesis, the powders were extensively washed to remove residual copper and iron salts from CuFe_2O_4 nanoparticles. The residual copper and iron ion concentrations in solution was below 30 ppb measured by ICP, which was primarily because of leaching. The synthesized CuFe_2O_4 nanoparticles were then subjected to analysis to confirm their identity. As shown in Figure 1a,b, the CuFe_2O_4 morphology was examined by TEM. The CuFe_2O_4 nanoparticle sizes were in the range of 20–80 nm, showing relatively homogeneous distribution. The pseudospherical shape of the nanoparticles was owed to the isotropic growth of the crystal from a core. Besides, it is also noted that these nanoparticles tended to aggregate, because of the iron magnetic interactions between particles. It was further determined that the atomic ratio between Cu and Fe is 1.8:1 (Figure 1c), close to theoretical 2:1 value. The detected carbon element was attributed to the used capping reagent and hexane cleaning agent. The crystal structure of CuFe_2O_4 was then probed by XRD (X-ray diffraction). It was shown in Figure 2 that the diffraction pattern of obtained sample matched well with the standard, indicating that the main phase of the powder was cuprospinel. In particular, the main peak at $2\theta = 35.64$ degree dominated both the standard and the obtained sample.

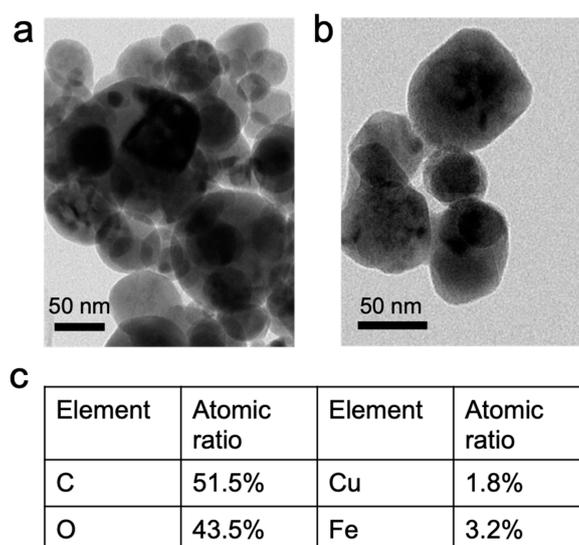


Figure 1. (a,b) Morphology of synthesized CuFe_2O_4 nanoparticles by transmission electron microscopy (TEM), and (c) elemental composition of CuFe_2O_4 nanoparticles.

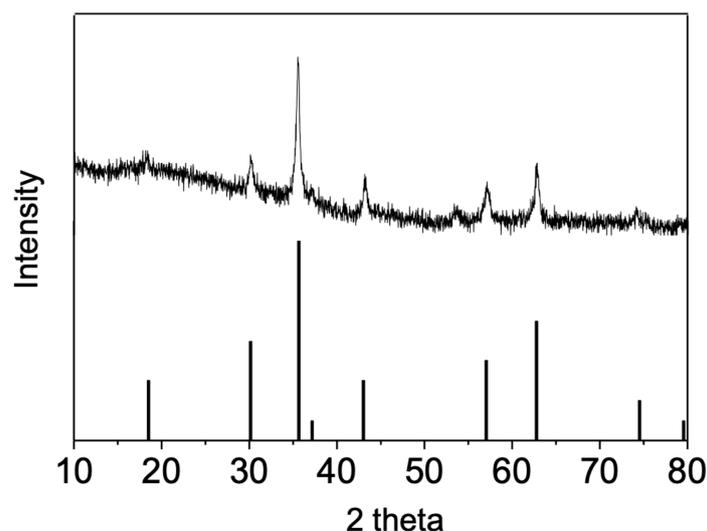


Figure 2. X-ray diffraction (XRD) characterization of synthesized CuFe_2O_4 nanoparticles.

3.2. Enhanced Bactericidal Performance of Copper Ferrite by Hydroxylamine Addition

We next investigated the bactericidal potential of the synthesized CuFe_2O_4 nanoparticles. The exponential phase *E. coli* cells were incubated with 0.2 g/L CuFe_2O_4 nanoparticles at 37 °C under shaking conditions, and bacterial viability was measured every hour. It was shown that CuFe_2O_4 nanoparticles only induce slight bacterial death during the 3 h incubation, indicating that CuFe_2O_4 nanoparticles are a weak antibacterial agent, primarily because of limited surface exposed Cu(II) species (Figure 3).

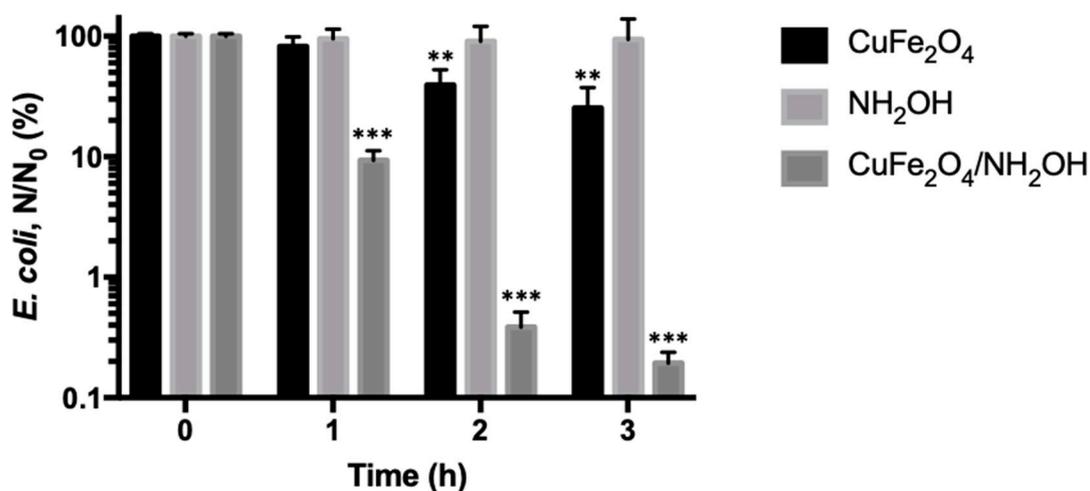


Figure 3. Inactivation of *E. coli* cells by CuFe_2O_4 nanoparticles, NH_2OH , and $\text{CuFe}_2\text{O}_4/\text{NH}_2\text{OH}$. During reactions, 0.2 g/L CuFe_2O_4 , 2 mM NH_2OH , and 10^8 CFU/mL *E. coli* cells in 10 mM MOPS buffer at pH 7 were used. ** $p < 0.01$, *** $p < 0.001$.

It is interesting to observe that, after addition of hydroxylamine (NH_2OH), the antibacterial potency of CuFe_2O_4 nanoparticles increased remarkably. For instance, the *E. coli* cell inactivation rate increased from 0.60-log by CuFe_2O_4 nanoparticles to 2.71-log by coupled $\text{CuFe}_2\text{O}_4/\text{NH}_2\text{OH}$ reaction after incubation for 3 h. In addition, it was observed that NH_2OH alone did not show detectable toxicity to *E. coli* cells. Specifically, after incubating exponential phase *E. coli* cells with 2 mM NH_2OH for 3 h, the cell inactivation rate was 0.03-log. The above results demonstrated that the bactericidal capacity by $\text{CuFe}_2\text{O}_4/\text{NH}_2\text{OH}$ reaction was in fact from a new generated species rather than either CuFe_2O_4 or NH_2OH alone. Besides, the bacterial inactivation action by $\text{CuFe}_2\text{O}_4/\text{NH}_2\text{OH}$ reaction

exhibited a time-dependent pattern, and the *E. coli* cells' inactivation rate at 1, 2, and 3 h was 1.03-, 2.41-, and 2.71-log, respectively (Figure 3). The progressive bacterial death thus indicated a persistent antibacterial mode by $\text{CuFe}_2\text{O}_4/\text{NH}_2\text{OH}$ reaction. It is worth noting that the bacterial amount in drinking water bodies is around 10^4 CFU per mL, and 99% removal efficiency is desired in most cases [41–43]. As the CuFe_2O_4 nanoparticles could inactivate 2.71-log of 10^8 CFU/mL *E. coli* cells after addition of NH_2OH , the disinfection process developed in this study holds great potential in the streamlined water treatment industry.

It should be noted that the observed strong antibacterial capability of $\text{CuFe}_2\text{O}_4/\text{NH}_2\text{OH}$ reaction could be attributed to leached copper ions into the solution. For this purpose, we used ICP to analyze dissolved copper ions in the solution. It was revealed that the detected copper ion concentration was below 20 ppb. To test if such an amount of copper ions could play a role in inactivating *E. coli* cells, we spiked 20 ppb Cu(II) ion into the bacterial solutions with or without 2 mM NH_2OH . The results showed that the homogeneous Cu(II)/ NH_2OH reaction had no effect on bacterial inactivation (Figure 4).

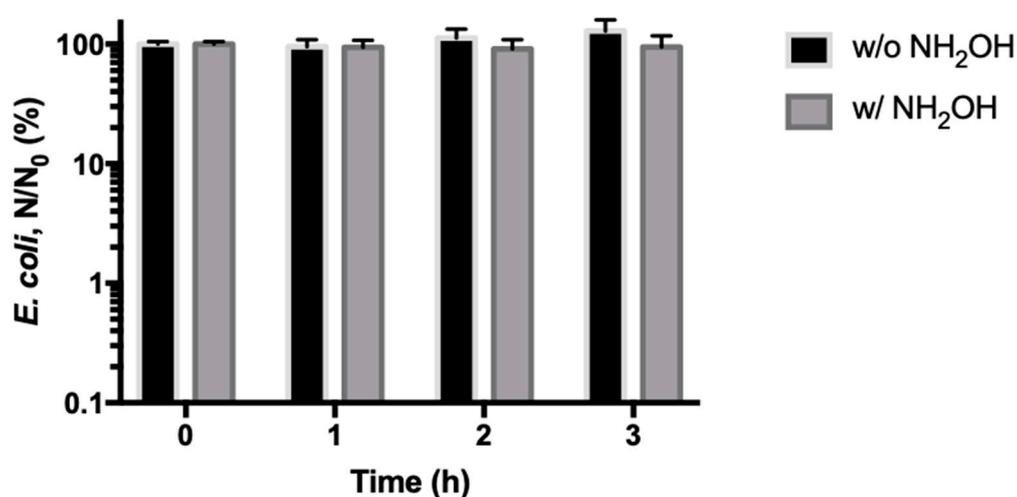


Figure 4. Inactivation of *E. coli* cells by homogeneous Cu(II)/ NH_2OH reaction with or without NH_2OH . During reactions, 20 ppb Cu(II) ions, 2 mM NH_2OH , and 10^8 CFU/mL *E. coli* cells in 10 mM MOPS buffer at pH 7 were used.

The effect of CuFe_2O_4 nanoparticles or NH_2OH concentrations was further investigated. At first, the CuFe_2O_4 nanoparticle concentration varied between 0.1 and 1 g/L, while NH_2OH concentration was fixed at 2 mM. The results showed that, at a concentration of 0.1, 0.2, 0.4, and 1 g/L CuFe_2O_4 nanoparticles, the *E. coli* inactivation rate was 0.27-, 2.71-, 3.55-, and 4.74-log, respectively (Figure 5a), suggesting a dose-dependent CuFe_2O_4 nanoparticle-induced reduction of bacterial viability. This is presumably because more exposed Cu(I) species mediated by NH_2OH reduction acted as a highly potent antibacterial agent [32,33]. We subsequently evaluated the effect of NH_2OH concentration. A total of 1 to 10 mM NH_2OH was used to mix with 0.2 g/L CuFe_2O_4 nanoparticle for bactericidal assays. It was shown that at 1, 2, 4, and 10 mM NH_2OH , 1.27-, 2.71-, 3.07-, and 3.41-log *E. coli* inactivation rate was obtained (Figure 5b). The plateaued enhancement of cell inactivation by increased NH_2OH doses was perhaps because of the fact that the majority of copper species on 0.2 g/L CuFe_2O_4 were reduced by 4 mM NH_2OH , and the further increase in NH_2OH concentration did not improve copper reduction. Such a trend has also been observed in other studies [7].

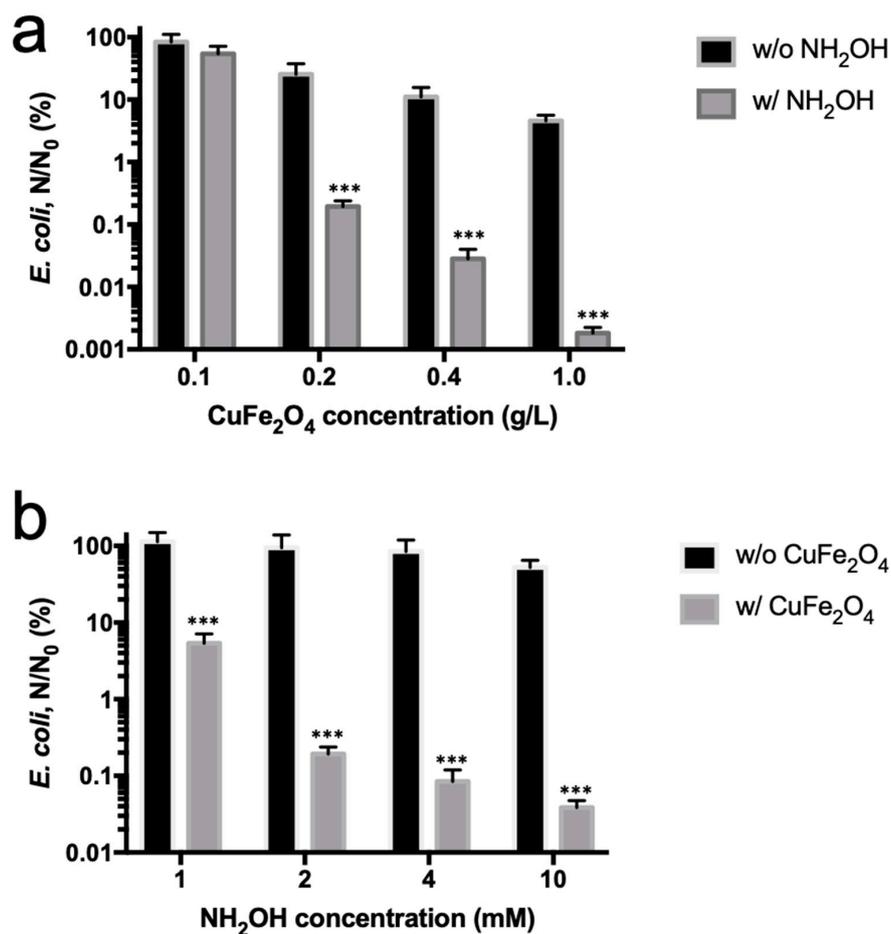


Figure 5. Effect of (a) CuFe₂O₄ and (b) NH₂OH concentration on *E. coli* inactivation by CuFe₂O₄/NH₂OH reaction. During reactions, (a) 0.1–1 g/L CuFe₂O₄, 2 mM NH₂OH, and 10⁸ CFU/mL *E. coli* cells in 10 mM MOPS buffer at pH 7 were used; (b) 0.2 g/L CuFe₂O₄, 1–10 mM NH₂OH, and 10⁸ CFU/mL *E. coli* cells in 10 mM MOPS buffer at pH 7 were used. *** $p < 0.001$.

3.3. Reduction of Surface Cu(II) into Cu(I) by Hydroxylamine

It has been previously reported that NH₂OH was able to efficiently transform Cu(II) ion into Cu(I) ion, which shows approximately 100–1000-fold enhancement in terms of antibacterial activity [7,33]. Because CuFe₂O₄ also owns Cu(II) species exposed onto the nanoparticle surface, we thus explored if the highly bactericidal Cu(I) species was formed by NH₂OH reduction. XPS (X-ray photoelectron spectroscopy) was utilized to detect the electronic properties of elements on the surface of CuFe₂O₄ nanoparticles. The electronic properties of CuFe₂O₄ nanoparticles were tested before and after reduction by NH₂OH.

XPS results are shown in Figure 6. It was observed that Fe 2p_{3/2} peaks almost did not show any detectable change in either octahedral or tetrahedral site Fe(III) species (Figure 6b), suggesting that iron species may not participate in the redox evolution of CuFe₂O₄/NH₂OH reaction. Besides, it was shown in Figure 6c that, after addition of NH₂OH, the surface adsorbed H₂O molecules were diminished, primarily because NH₂OH repulsed H₂O molecules away to approach the CuFe₂O₄ surface. Interestingly, the addition of NH₂OH drastically changed the speciation of copper (i.e., Cu(I) and Cu(II)) on CuFe₂O₄ nanoparticle surfaces based on Cu 2p_{3/2} deconvolution results (Figure 6a). Specifically, the fraction of Cu(I) before and after CuFe₂O₄/NH₂OH reaction was 27.4% and 75.2%, respectively (Table 1). The remarkable increase of Cu(I) species was primarily because of the reductive action of NH₂OH. Besides, CuFe₂O₄/NH₂OH reaction also mediated a significant change in O 1s electronic property. For example, O 1s of CuFe₂O₄ was mainly composed by lattice O (28.6%), surface

OH (66%), and adsorbed H₂O (47.4%). However, after reaction with NH₂OH, the components of O 1s on CuFe₂O₄ nanoparticle surface became lattice O (32.5%) and surface OH (67.5%), whereas the surface adsorbed H₂O molecules disappeared (Table 1). This is probably because NH₂OH might need to repel surface adsorbed H₂O molecule before accessing the reactive center on CuFe₂O₄ nanoparticle. Overall, the results indicated that Cu(I) fraction was successfully increased on the surface of CuFe₂O₄ after NH₂OH reduction, and the transformed nascent Cu(I) species is supposed to play a major role in *E. coli* inactivation.

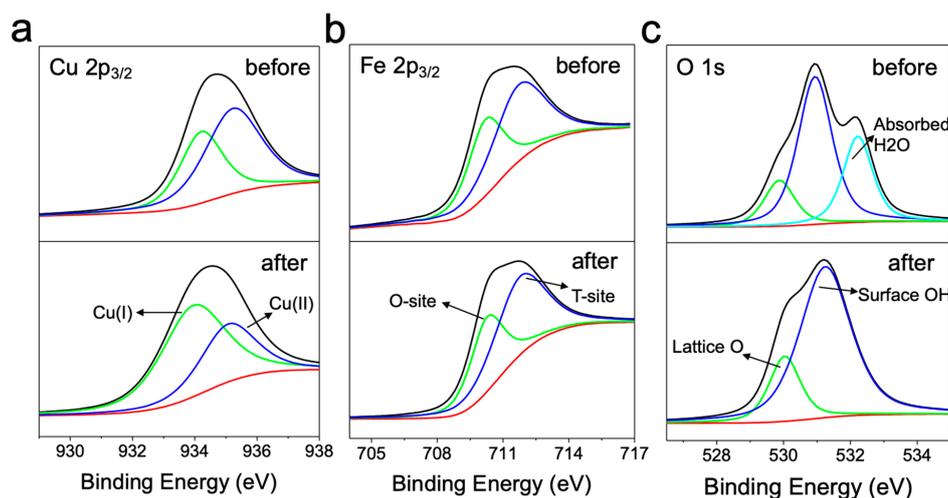


Figure 6. X-ray photoelectron spectroscopy (XPS) characterization of synthesized CuFe₂O₄ nanoparticles before (upper panel) and after (lower panel) reduction by NH₂OH. (a) Cu 2p_{3/2}, (b) Fe 2p_{3/2}, and (c) O 1s peaks were deconvoluted.

Table 1. Species distribution of Cu 2p, Fe 2p, and O 1s on CuFe₂O₄ nanoparticle surface before and after reduction by NH₂OH.

	Cu 2p		Fe 2p		O 1s		
	Cu(II)	Cu(I)	O-Site Fe(III)	T-Site Fe(III)	Lattice O	Surface OH	Absorbed H ₂ O
Before	72.6%	27.4%	45.5%	54.5%	28.6%	66%	47.4%
After	24.8%	75.2%	42.4%	57.6%	32.5%	67.5%	n/a

3.4. Bactericidal Action by CuFe₂O₄/NH₂OH Reaction

We were interested in understanding the molecular biology mechanism associated with the bactericidal action of CuFe₂O₄/NH₂OH reaction. It has been reported that Cu(I) is a strong complexing and denaturing agent for functional proteins in particular membrane proteins [44–48]. In fact, the antibacterial potency of Cu(I), which is called contact-killing [47,48], is significantly higher than other well-established heavy metals such as silver. Although the exact bactericidal actions of Cu(I) are unclear, it is widely accepted that it increases the oxidative stress inside the cell [44,47,48]. We thus attempted to evaluate the ROS content—A direct oxidative stress indicator—with a fluorescent probe [49,50]. The results suggested that the incubation of CuFe₂O₄ and CuFe₂O₄/NH₂OH with *E. coli* cells could increase the ROS content by comparison with the control (Figure 7). Specifically, after incubation for 3 h, the fluorescence change for the control, CuFe₂O₄, NH₂OH, and CuFe₂O₄/NH₂OH was 7.48%, 15.34%, 8.49%, and 25.69%, respectively. In addition, the CuFe₂O₄/NH₂OH treatment mediated a more significant increase in ROS content than the CuFe₂O₄ treatment, indicating that the generated Cu(I) species is more powerful in producing oxidative stress, in agreement with previous literatures [44,47,48].

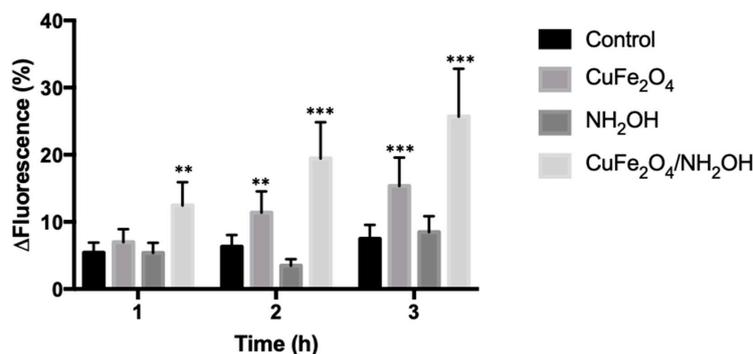


Figure 7. HPF (3'-p-(hydroxyphenyl) molecular probe fluorescence change of *E. coli* cells by CuFe₂O₄/NH₂OH and related controls. Excitation/emission at 490/515 nm was used for fluorescence determination. Control: No addition of CuFe₂O₄ or NH₂OH. During reactions, 10 μM HPF probe, 0.2 g/L CuFe₂O₄, 2 mM NH₂OH, and 10⁸ CFU/mL *E. coli* cells in 10 mM MOPS buffer at pH 7 were used. ** $p < 0.01$, *** $p < 0.001$.

The interaction between Cu(I) species and *E. coli* cell was further explored. At first, we added 2 mM EDTA as a complexing reagent to block the effective binding of copper species to membrane proteins, and found that the bacterial inactivation was negligible (0.07-log) (Figure 8). The above results verified the bactericidal role of copper species in CuFe₂O₄. We further investigated if the bacterial inactivation could be alleviated by adding a ROS scavenger. DMSO was used as a ROS scavenger [51–53], and a different DMSO concentration (10–100 mM) was used. It was found that the addition of DMSO could suppress the bactericidal potency of CuFe₂O₄/NH₂OH reaction. In detail, after the addition of 10, 20, 40, and 100 mM DMSO, *E. coli* inactivation efficiency by CuFe₂O₄/NH₂OH reaction was 1.60-, 0.98-, 0.40-, and 0.31-log, respectively. The results clearly indicated that ROS generation was the major reason accounting for bacterial inactivation in our system, which is in good accordance with other reports [7,15]. A detailed mechanistic illustration is shown in Figure 9.

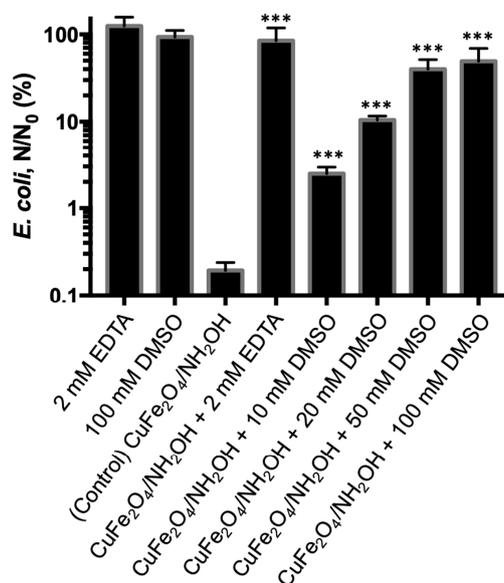


Figure 8. *E. coli* inactivation by CuFe₂O₄/NH₂OH after addition of EDTA and DMSO scavengers and related controls. During reactions, 0.2 g/L CuFe₂O₄, 2 mM NH₂OH, and 10⁸ CFU/mL *E. coli* cells in 10 mM MOPS buffer at pH 7 were used. EDTA of 2 mM and DMSO of 10–100 mM were added in the above bacterial solution. *** $p < 0.001$.

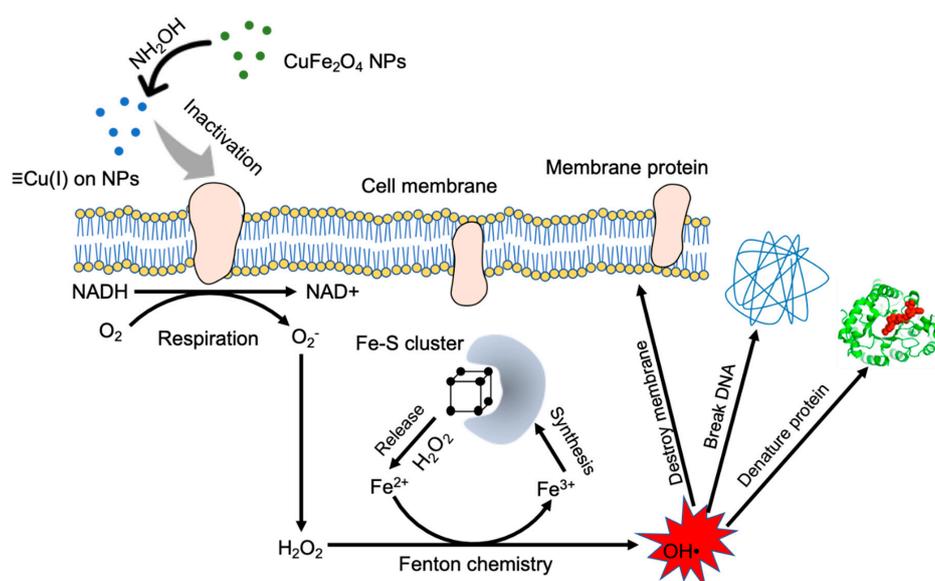


Figure 9. Proposed mechanism of $\text{CuFe}_2\text{O}_4/\text{NH}_2\text{OH}$ reaction inactivating *E. coli* cells. Generated Cu(I) species on nanoparticle surface after NH_2OH reduced CuFe_2O_4 inactivates membrane proteins, leading to generation of H_2O_2 . The disruption of the Fe–S cluster by H_2O_2 releases free Fe^{2+} ions, which catalyzes the Fenton chemistry to convert H_2O_2 into $\text{HO}\bullet$ (a reactive oxygen species). The $\text{HO}\bullet$ inactivates *E. coli* cell through destroying the membrane structure, breaking DNA, and denaturing functional proteins. NPs, nanoparticles.

3.5. Recycling Assay of Copper Ferrite Nanoparticle

An attractive advantage associated with heterogeneous water disinfection system is that the antibacterial agents could be reused for multiple rounds. To test if CuFe_2O_4 nanoparticle could be reused in our developed water disinfection platform, after each round of bactericidal assay, the nanoparticles were centrifuged and collected for a subsequent round of analysis. Then, 10^4 CFU/mL exponential phase *E. coli* cells were used to simulate real water bodies. Figure 10 shows that, during the 10 rounds of tests, the reused CuFe_2O_4 nanoparticles showed a steady antibacterial efficiency, from 96.1% to 99.9%, indicating that the proposed $\text{CuFe}_2\text{O}_4/\text{NH}_2\text{OH}$ antibacterial platform could be used for water treatment industry with a low cost. Besides, it is worth to mention that CuFe_2O_4 nanoparticles (NPs) have a strong magnetic property and could be collected by magnetic attraction. This property further simplifies the reuse procedure in industry because magnetic enrichment and collection has been well established.

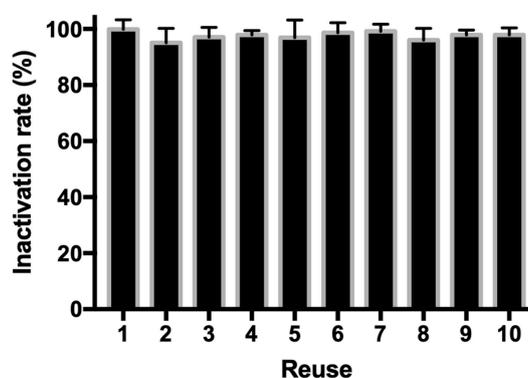


Figure 10. Inactivation of *E. coli* by reused CuFe_2O_4 nanoparticles and addition of NH_2OH . During reactions, 0.2 g/L fresh or reused CuFe_2O_4 , 2 mM NH_2OH , and 10^4 CFU/mL *E. coli* cells in 10 mM MOPS buffer at pH 7 were used.

4. Conclusions

In this study, we showed that the antibacterial capability of CuFe₂O₄ nanomaterial could be significantly enhanced after addition of hydroxylamine. This was because surface Cu(II) species was successfully reduced to Cu(I), as evidenced by XPS. Cu(I) has a much stronger binding affinity and reduction capability to functional proteins on bacterial cell membrane than Cu(II) species, leading to a contact-killing phenomenon. It is worth noting that the bacterial death caused by CuFe₂O₄/NH₂OH reaction was mainly because of the Cu(I) species on the nanoparticle surface, rather than that dissolved in solution. This implies that the minimum leaching of CuFe₂O₄ nanoparticle guarantees its safe application in the water disinfection industry. Besides, NH₂OH has also been widely used in water treatment, and meets the criteria of public drinking water safety.

Further, the bactericidal mechanism of CuFe₂O₄/NH₂OH reaction towards *E. coli* was revealed with multiple molecular approaches. The results indicated that ROS content is elevated inside the cell, which might impair vital cellular components and cause leakage, presumably accounting for the death of *E. coli* cells. In addition, CuFe₂O₄ nanoparticles were reused for several rounds in our study, delivering uncompromised *E. coli* inactivation performance. In conclusion, considering the low cost of the chemicals and negligible secondary contamination concern, these results demonstrated that the generation of Cu(I) species immobilized on CuFe₂O₄ nanoparticles after reduction by NH₂OH is a viable option for water pathogens' disinfection.

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References

1. Tien, J.H.; Earn, D.J. Multiple transmission pathways and disease dynamics in a waterborne pathogen model. *Bull. Math. Biol.* **2010**, *72*, 1506–1533. [[CrossRef](#)]
2. Fouz, B.; Toranzo, A.E.; Milan, M.; Amaro, C. Evidence that water transmits the disease caused by the fish pathogen *Photobacterium damsela* subsp. *damsela*. *J. Appl. Microbiol.* **2000**, *88*, 531–535. [[CrossRef](#)]
3. Ashbolt, N.J. Microbial contamination of drinking water and disease outcomes in developing regions. *Toxicology* **2004**, *198*, 229–238. [[CrossRef](#)] [[PubMed](#)]
4. Boorman, G.A. Drinking water disinfection byproducts: Review and approach to toxicity evaluation. *Environ. Health Perspect.* **1999**, *107*, 207–217. [[PubMed](#)]
5. McGuigan, K.G.; Conroy, R.M.; Mosler, H.J.; du Preez, M.; Ubomba-Jaswa, E.; Fernandez-Ibanez, P. Solar water disinfection (SODIS): A review from bench-top to roof-top. *J. Hazard. Mater.* **2012**, *235*, 29–46. [[CrossRef](#)] [[PubMed](#)]
6. Song, K.; Mohseni, M.; Taghipour, F. Application of ultraviolet light-emitting diodes (UV-LEDs) for water disinfection: A review. *Water Res.* **2016**, *94*, 341–349. [[CrossRef](#)]
7. Chen, L.; Tang, M.; Chen, C.; Chen, M.; Luo, K.; Xu, J.; Zhou, D.; Wu, F. Efficient bacterial inactivation by transition metal catalyzed auto-oxidation of sulfite. *Environ. Sci. Technol.* **2017**, *51*, 12663–12671. [[CrossRef](#)]
8. Chen, L.; Pinto, A.; Alshwabkeh, A.N. Activated Carbon as a Cathode for Water Disinfection through the Electro-Fenton Process. *Catalysts* **2019**, *9*, 601. [[CrossRef](#)]
9. Morris, R.D.; Audet, A.M.; Angelillo, I.F.; Chalmers, T.C.; Mosteller, F. Chlorination, chlorination by-products, and cancer: A meta-analysis. *Am. J. Public Health* **1992**, *82*, 955–963. [[CrossRef](#)]
10. Kim, J.; Chung, Y.; Shin, D.; Kim, M.; Lee, Y.; Lim, Y.; Lee, D. Chlorination by-products in surface water treatment process. *Desalination* **2003**, *151*, 1–9. [[CrossRef](#)]
11. Jyoti, K.K.; Pandit, A.B. Ozone and cavitation for water disinfection. *Biochem. Eng. J.* **2004**, *18*, 9–19. [[CrossRef](#)]

12. Kerwick, M.I.; Reddy, S.M.; Chamberlain, A.H.L.; Holt, D.M. Electrochemical disinfection, an environmentally acceptable method of drinking water disinfection? *Electrochim. Acta* **2005**, *50*, 5270–5277. [[CrossRef](#)]
13. Thomas, V.; Bouchez, T.; Nicolas, V.; Robert, S.; Loret, J.F.; Levi, Y. Amoebae in domestic water systems: Resistance to disinfection treatments and implication in Legionella persistence. *J. Appl. Microbiol.* **2004**, *97*, 950–963. [[CrossRef](#)]
14. Chiao, T.H.; Clancy, T.M.; Pinto, A.; Xi, C.; Raskin, L. Differential resistance of drinking water bacterial populations to monochloramine disinfection. *Environ. Sci. Technol.* **2014**, *48*, 4038–4047. [[CrossRef](#)]
15. Berry, D.; Xi, C.; Raskin, L. Microbial ecology of drinking water distribution systems. *Curr. Opin. Biotechnol.* **2006**, *17*, 297–302. [[CrossRef](#)]
16. Colberg, P.J.; Lingg, A.J. Effect of ozonation on microbial fish pathogens, ammonia, nitrate, nitrite, and BOD in simulated reuse hatchery water. *J. Fish. Board Can.* **1978**, *35*, 1290–1296. [[CrossRef](#)]
17. Wedemeyer, G.A.; Nelson, N.C. Survival of two bacterial fish pathogens (*Aeromonas salmonicida* and the enteric redmouth bacterium) in ozonated, chlorinated, and untreated waters. *J. Fish. Board Can.* **1977**, *34*, 429–432. [[CrossRef](#)]
18. Liltved, H.; Hektoen, H.; Efraimsson, H. Inactivation of bacterial and viral fish pathogens by ozonation or UV irradiation in water of different salinity. *Aquacult. Eng.* **1995**, *14*, 107–122. [[CrossRef](#)]
19. Zyara, A.; Torvinen, E.; Veijalainen, A.M.; Heinonen-Tanski, H. The effect of UV and combined chlorine/UV treatment on coliphages in drinking water disinfection. *Water* **2016**, *8*, 130. [[CrossRef](#)]
20. Guerrero-Latorre, L.; Gonzales-Gustavson, E.; Hundesa, A.; Sommer, R.; Rosina, G. UV disinfection and flocculation-chlorination sachets to reduce hepatitis E virus in drinking water. *Int. J. Hyg. Environ. Health* **2016**, *219*, 405–411. [[CrossRef](#)]
21. Hull, N.M.; Isola, M.R.; Petri, B.; Chan, P.S.; Linden, K.G. Algal DNA repair kinetics support culture-based enumeration for validation of ultraviolet disinfection ballast water treatment systems. *Environ. Sci. Technol. Lett.* **2017**, *4*, 192–196. [[CrossRef](#)]
22. Ma, S.; Zhan, S.; Jia, Y.; Shi, Q.; Zhou, Q. Enhanced disinfection application of Ag-modified g-C₃N₄ composite under visible light. *Appl. Catal. B* **2016**, *186*, 77–87. [[CrossRef](#)]
23. Teng, Z.; Yang, N.; Lv, H.; Wang, S.; Hu, M.; Wang, C.; Wang, D.; Wang, G. Edge-Functionalized g-C₃N₄ Nanosheets as a Highly Efficient Metal-free Photocatalyst for Safe Drinking Water. *Chem* **2019**, *5*, 664–680. [[CrossRef](#)]
24. Chen, M.; Li, Z.; Chen, L. Highly antibacterial rGO/Cu₂O nanocomposite from a biomass precursor: Synthesis, performance, and mechanism. *Nano Mater. Sci.* **2019**. [[CrossRef](#)]
25. Chen, L.; Li, Z.; Chen, M. Facile production of silver-reduced graphene oxide nanocomposite with highly effective antibacterial performance. *J. Environ. Chem. Eng.* **2019**, *7*, 103160. [[CrossRef](#)]
26. Jin, X.; Li, M.; Wang, J.; Marambio-Jones, C.; Peng, F.; Huang, X.; Damoiseaux, R.; Hoek, E.M. High-throughput screening of silver nanoparticle stability and bacterial inactivation in aquatic media: Influence of specific ions. *Environ. Sci. Technol.* **2010**, *44*, 7321–7328. [[CrossRef](#)]
27. Thurman, R.B.; Gerba, C.P.; Bitton, G. The molecular mechanisms of copper and silver ion disinfection of bacteria and viruses. *Crit. Rev. Environ. Sci. Technol.* **1989**, *18*, 295–315. [[CrossRef](#)]
28. Wang, L.; Luo, J.; Shan, S.; Crew, E.; Yin, J.; Zhong, C.J.; Wallek, B.; Wong, S.S. Bacterial inactivation using silver-coated magnetic nanoparticles as functional antimicrobial agents. *Anal. Chem.* **2011**, *83*, 8688–8695. [[CrossRef](#)]
29. Mikolay, A.; Huggett, S.; Tikana, L.; Grass, G.; Braun, J.; Nies, D.H. Survival of bacteria on metallic copper surfaces in a hospital trial. *Appl. Microbiol. Biotechnol.* **2010**, *87*, 1875–1879. [[CrossRef](#)]
30. Santo, C.E.; Lam, E.W.; Elowsky, C.G.; Quaranta, D.; Domaille, D.W.; Chang, C.J.; Grass, G. Bacterial killing by dry metallic copper surfaces. *Appl. Environ. Microbiol.* **2011**, *77*, 794–802. [[CrossRef](#)]
31. Borkow, G.; Gabbay, J. Putting copper into action: Copper-impregnated products with potent biocidal activities. *FASEB J.* **2004**, *18*, 1728–1730. [[CrossRef](#)] [[PubMed](#)]
32. Lee, H.J.; Kim, H.E.; Lee, C. Combination of cupric ion with hydroxylamine and hydrogen peroxide for the control of bacterial biofilms on RO membranes. *Water Res.* **2017**, *110*, 83–90. [[CrossRef](#)] [[PubMed](#)]
33. Kim, H.E.; Nguyen, T.T.; Lee, H.; Lee, C. Enhanced inactivation of *Escherichia coli* and MS2 coliphage by cupric ion in the presence of hydroxylamine: Dual microbicidal effects. *Environ. Sci. Technol.* **2015**, *49*, 14416–14423. [[CrossRef](#)]

34. Morgada, M.N.; Abriata, L.A.; Cefaro, C.; Gajda, K.; Banci, L.; Vila, A.J. Loop recognition and copper-mediated disulfide reduction underpin metal site assembly of CuA in human cytochrome oxidase. *Proc. Natl. Acad. Sci. USA* **2015**, *112*, 11771–11776. [CrossRef] [PubMed]
35. Brewer, G.J. Risks of copper and iron toxicity during aging in humans. *Chem. Res. Toxicol.* **2009**, *23*, 319–326. [CrossRef] [PubMed]
36. US EPA Standard for Copper in Drinking Water. Available online: <https://www.epa.gov/dwreginfo/lead-and-copper-rule> (accessed on 30 November 2019).
37. Brenner, K.P.; Rankin, C.C.; Roybal, Y.R.; Stelma, G.N.; Scarpino, P.V.; Dufour, A.P. New medium for the simultaneous detection of total coliforms and *Escherichia coli* in water. *Appl. Environ. Microbiol.* **1993**, *59*, 3534–3544.
38. Heijnen, L.; Medema, G. Quantitative detection of *E. coli*, *E. coli* O157 and other shiga toxin producing *E. coli* in water samples using a culture method combined with real-time PCR. *J. Water Health* **2006**, *4*, 487–498. [CrossRef]
39. Atia, T.A.; Altimari, P.; Moscardini, E.; Pettiti, I.; Toro, L.; Pagnanelli, F. Synthesis and characterization of copper ferrite magnetic nanoparticles by hydrothermal route. *Chem. Eng. Trans.* **2016**, *47*, 151–156.
40. Petersen, E.J.; Hirsch, C.; Elliott, J.T.; Krug, H.F.; Aengenheister, L.; Arif, A.T.; Bogni, A.; Kinsner-Ovaskainen, A.; May, S.; Walser, T.; et al. Cause-and-Effect Analysis as a Tool To Improve the Reproducibility of Nanobioassays: Four Case Studies. *Chem. Res. Toxicol.* **2019**. [CrossRef] [PubMed]
41. Liu, Y.; Gilchrist, A.; Zhang, J.; Li, X.F. Detection of viable but nonculturable *Escherichia coli* O157: H7 bacteria in drinking water and river water. *Appl. Environ. Microbiol.* **2008**, *74*, 1502–1507. [CrossRef]
42. Joyce, T.M.; McGuigan, K.G.; Elmore-Meegan, M.; Conroy, R.M. Inactivation of fecal bacteria in drinking water by solar heating. *Appl. Environ. Microbiol.* **1996**, *62*, 399–402.
43. Wolfe, R.L.; Ward, N.R.; Olson, B.H. Inactivation of heterotrophic bacterial populations in finished drinking water by chlorine and chloramines. *Water Res.* **1985**, *19*, 1393–1403. [CrossRef]
44. Santo, C.E.; Quaranta, D.; Grass, G. Antimicrobial metallic copper surfaces kill *Staphylococcus haemolyticus* via membrane damage. *Microbiologyopen* **2012**, *1*, 46–52. [CrossRef]
45. Silver, S. Bacterial resistances to toxic metal ions—A review. *Gene* **1996**, *179*, 9–19. [CrossRef]
46. Bondarczuk, K.; Piotrowska-Seget, Z. Molecular basis of active copper resistance mechanisms in Gram-negative bacteria. *Cell Biol. Toxicol.* **2013**, *29*, 397–405. [CrossRef]
47. Saphier, M.; Silberstein, E.; Shotland, Y.; Popov, S.; Saphier, O. Prevalence of Monovalent Copper Over Divalent in Killing *Escherichia coli* and *Staphylococcus Aureus*. *Curr. Microbiol.* **2018**, *75*, 426–430. [CrossRef]
48. Borkow, G. Using copper to improve the well-being of the skin. *Curr. Chem. Biol.* **2014**, *8*, 89–102. [CrossRef]
49. Rowe, L.A.; Degtyareva, N.; Doetsch, P.W. DNA damage-induced reactive oxygen species (ROS) stress response in *Saccharomyces cerevisiae*. *Free Radic. Biol. Med.* **2008**, *45*, 1167–1177. [CrossRef]
50. Tomizawa, S.; Imai, H.; Tsukada, S.; Simizu, T.; Honda, F.; Nakamura, M.; Nagano, T.; Urano, Y.; Matsuoka, Y.; Fukasaku, N.; et al. The detection and quantification of highly reactive oxygen species using the novel HPF fluorescence probe in a rat model of focal cerebral ischemia. *J. Neurosci. Res.* **2005**, *53*, 304–313. [CrossRef]
51. Veltwisch, D.; Janata, E.; Asmus, K.D. Primary processes in the reaction of OH⁻radicals with sulphoxides. *J. Chem. Soc. Perkin Trans. I* **1980**, 146–153. [CrossRef]
52. Reuvers, A.P.; Greenstock, C.L.; Borsa, J.; Chapman, J.D. Studies on the mechanism of chemical radioprotection by dimethyl sulphoxide. *Int. J. Radiat. Biol. Relat. Stud. Phys. Chem. Med.* **1973**, *24*, 533–536. [CrossRef]
53. Mi, H.; Wang, D.; Xue, Y.; Zhang, Z.; Niu, J.; Hong, Y.; Drlica, K.; Zhao, X. Dimethyl sulfoxide protects *Escherichia coli* from rapid antimicrobial-mediated killing. *Antimicrob. Agents Chemother.* **2016**, *60*, 5054–5058. [CrossRef]

