

Article

# Visible Light Photodegradation of Formaldehyde over TiO<sub>2</sub> Nanotubes Synthesized via Electrochemical Anodization of Titanium Foil

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**Abstract:** In this study, a series of TiO<sub>2</sub> nanotubes (NTs) were synthesized employing electrochemical anodization of titanium foil in an ionic liquid solution containing a mixture of glycerol and choline chloride, acting as electrolyte. The as-synthesized TiO<sub>2</sub> NTs were calcined at 350, 450, or 550 °C for a 2 h duration to investigate the influence of calcination temperature on NTs formation, morphology, surface properties, crystallinity, and subsequent photocatalytic activity for visible light photodegradation of gaseous formaldehyde (HCHO). Results showed that the calcination temperature has a significant effect on the structure and coverage of TiO<sub>2</sub> NTs on the surface. Freshly synthesized TiO<sub>2</sub> NTs showed better-ordered structure compared to calcined samples. There was significant pore rupture with increasing calcination temperature. The transformation from anatase to rutile phase appeared after calcination at 450 °C and the weight fraction of the rutile phase increased from 19% to 36% upon increasing the calcination temperature to 550 °C. The band gaps of the TiO<sub>2</sub> NTs were in the range from 2.80 to 2.74 eV, shifting the active region of the materials to visible light. The presence of mixed anatase–rutile TiO<sub>2</sub> phases in the sample calcined at 450 °C showed enhanced photoactivity, which was confirmed by the 21.56 mg·L<sup>-1</sup>·g<sup>-1</sup> removal of gaseous formaldehyde under 120 min of visible light irradiation and displayed enhanced quantum yield,  $\theta_{HCHO}$  of 17%.

Keywords: TiO<sub>2</sub> nanotubes; anodization; ionic liquid; glycerol; formaldehyde degradation

## 1. Introduction

Indoor air pollution is one of the major worldwide human health concerns related to volatile organic compounds (VOCs) since it can lead to sick building syndromes such as headache and fatigue. VOCs such as formaldehyde, toluene, and chloroform are the most commonly investigated airborne contaminants [1]. Formaldehyde (HCHO) is considered as one of the most hazardous VOCs because long-term exposure to it causes adverse effects on human health such as eye irritation, breathing difficulties, and skin irritation [2,3]. The World Health Organization (WHO) guidelines for indoor air HCHO concentration is 0.08 ppm. Nielsen et al. [4] reported that the maximum HCHO concentration in a house inhabited by asthmatics in Boston was 162  $\mu$ g/m<sup>3</sup> while for homes in Japan the maximum



concentration was 58  $\mu$ g/m<sup>3</sup>, although the value can increase to 220  $\mu$ g/m<sup>3</sup> in the summer. Hence, it is crucial to eliminate this chemical substance in order to improve indoor air quality and to comply with stringent environmental regulations.

Several conventional physicochemical methods have been investigated for the removal of HCHO from the air [5–7]. However, these methods have several major drawbacks including their incapability to remove HCHO completely from the air, long retention time, and the production of secondary pollutants [8]. Among the advanced oxidation processes (AOPs), photocatalytic technology provides an affordable, economical, and sustainable alternative for the degradation of gaseous formaldehyde [6]. For example, Wu has decomposed 85.6% of formaldehyde over 0.15 g of nitrogen-doped ZnO [9]. Similarly, Chang et al. [10] reported 93.2% formaldehyde degradation over platinum-tin oxide core-shell structured nanoparticles. The major advantage of photocatalytic oxidation is that it offers the possibility of using sunlight as renewable solar energy for photocatalytic degradation of HCHO, thereby making the process green and sustainable. Photocatalytic technology using titanium dioxide (TiO<sub>2</sub>) has been considered as a popular advanced oxidation process [11,12] because of its predominant photocatalytic activity, low toxicity, and high resistance toward corrosion, the low production cost of  $TiO_2$  [13,14], and its effectiveness in swift degradation of recalcitrant organic compounds and complete mineralization into  $CO_2$  and water [15–17]. Pure or modified  $TiO_2$  has been used to remove gaseous formaldehyde from airstreams. Zhu and Wu reported 98.3% formaldehyde degradation over Pt-modified TiO<sub>2</sub> [18]. However, high aggregation tendency and difficulty in separation and recovery of the  $TiO_2$  are some of the most important limitations of photocatalysis for practical application in air decontamination. To overcome these limitations, the immobilization of TiO<sub>2</sub> on support medium has become a preferable alternative. For instance, Xu et al. [19] coated TiO<sub>2</sub> on fabrics for the degradation of HCHO. Their results demonstrated that the  $TiO_2$  coated on fabrics showed better photocatalytic performance compared to non-coated TiO<sub>2</sub>. Nevertheless, the drawbacks associated with immobilization or coating such as concentrating the target contaminant around the TiO<sub>2</sub> nanoparticles are also difficult to address convincingly. The other bottleneck for TiO<sub>2</sub> to be utilized in visible light photocatalysis is its wide bandgap.

These challenges can be overcome by fabricating the nanosized  $TiO_2$  on some larger particulates and fabricating visible light active  $TiO_2$ . The advent of one-dimensional nanostructures such as  $TiO_2$ NTs and its fabrication by simple electrochemical anodization of titanium (Ti) foil has greatly eliminated the requirement of a filtration unit or catalyst recovery from the treated air stream. Since then, the decontamination of air with this high surface area and vertically aligned homogeneous  $TiO_2$  (NTs) has gained increasing research interest. Recently, more research activities were focused on the formation, growth, structural modifications, and applications of  $TiO_2$  NTs [20]. Several methods exist for the fabrication of TiO<sub>2</sub> NTs. However, electrochemical anodization offers a simple and more robust way to fabricate TiO<sub>2</sub> NTs at low cost. Furthermore, TiO<sub>2</sub> NTs can be grown on titanium substrate producing large surface area materials, hence improving the interaction between the gas and nanostructures while possessing an immobilized characteristic at the same time [15,21]. Although this technique has various advantages, some issues could occur in terms of the reproducibility of the prepared samples. It has been reported that the photocatalytic performance of TiO<sub>2</sub> NTs strongly depends on its surface morphology [22]. The surface morphology of  $TiO_2$  NTs are controlled by various experimental parameters such as anodization duration, anodization potential, and electrolyte composition [13]. Calcination temperature of the NTs is another important factor that determines the NTs properties and reaction performance of the photocatalyst [15].

There are many articles that previously reported on the formation of self-organized TiO<sub>2</sub> NTs by anodization of Ti in an electrolyte containing fluoride ions [12,15,22]. The NTs demonstrated interesting photocatalytic degradation of organic pollutants. For example, Sreekantan et al. [23] reported that uniform and well-aligned TiO<sub>2</sub> NTs were produced in fluorinated glycerol electrolytes at an anodization voltage of 20 V, which exhibited better photocatalytic activity for degradation of methyl orange. Liang et al. reported more than 90% degradation of 2,3-dicholorphenol over anodic TiO<sub>2</sub> NT

arrays [15]. However, a fluoride-free electrolyte is desired as discussed by Nguyen et al. [24] due to the ease in handling and synthesizing NTs without involving hazardous chemicals, thus allowing the NTs to be grown in a safer manner. There are a few research groups that have investigated the formation of  $TiO_2$  NTs by anodization process using a chloride-containing electrolyte as an alternative to fluoride-based electrolyte [24–27]. For instance, Hahn et al. [28] synthesized  $TiO_2$  and  $WO_3$  NTs in  $HCIO_4$  and  $NaCIO_4$  electrolytes. Nguyen and co-workers [24] also reported that NTs can be fabricated using NaCl dissolved in either water, ethylene glycol, or glycerol. Their results showed the formation of NTs with a relatively smaller diameter at a faster growth rate compared to those NTs fabricated in a fluoride-based electrolyte. Excellent reviews have been published on the fabrication, modification, and application of  $TiO_2$  NTs [29,30]. It is evident that  $TiO_2$  NTs have gained tremendous research attention and there has been an exponential growth in this field.

Heat treatment also plays a crucial role in producing NTs with enhanced photocatalytic activity since high thermal treatment can significantly affect the crystallographic structure of the TiO<sub>2</sub> phase [31]. Hurum et al. [32] reported that anatase TiO<sub>2</sub> exhibits lower recombination rates compared to the rutile phase and is regarded as a more photochemically active phase. Other pieces of literature also supported that mixed-phase TiO<sub>2</sub> exhibited higher photocatalytic activity compared to pure phases alone. Bickley et al. [33] proposed the hypothesis of good photocatalytic activity of mixed-phase TiO<sub>2</sub> is due to the transfer of electrons from anatase (3.2 eV) to lower energy rutile (3.0 eV) electrons trapping site which serves to reduce the recombination rate of anatase. However, there is no consensus on whether a pure anatase phase or mixed-phase (anatase and rutile) TiO<sub>2</sub> NTs are the most photochemically active catalyst.

The main aim of the current study is to fabricate visible-light active  $TiO_2$  NTs for photodegradation of gaseous formaldehyde.  $TiO_2$  NTs were fabricated by electrochemical anodization of Ti foil in an ionic liquid solution containing a mixture of glycerol and choline chloride (chloride-based electrolyte). The effects of calcination temperature on the properties and photocatalytic activity of the synthesized  $TiO_2$  NTs were investigated.

#### 2. Materials and Methods

#### 2.1. Synthesis of Ionic Liquid

Choline chloride (ChCl, 98%, Sigma Aldrich, Darmstadt, Germany), and glycerol ( $C_3H_8O_3$ , 98%, Fischer Scientific, NH, USA) were mixed in a molar ratio of 1:2 and heated to 80 °C for 30 min to form a colorless ionic liquid [34,35]. It was used as the electrolyte solution for the anodization process of Ti foil to form TiO<sub>2</sub> NTs. The chemical structure of the mixture of ChCl and glycerol is shown in Figure 1.



ChCl + Glycerol

Figure 1. Chemical structure of choline chloride and glycerol [36].

#### 2.2. Synthesis of TiO<sub>2</sub> Nanotubes

Technical grade Ti foil (Titanium Ti Gr5/Tc4 Grade 5 ASTM B265 Thin Plate Sheet, Nanjing, Jiangsu, China) with 0.1 mm thickness was cut into 2 cm  $\times$  1 cm squares which were used as the substrates for

anodization process. The Ti substrates were ultrasonically cleaned in acetone for 10 min, followed by thorough rinsing with deionized (DI) water, and dried in air prior to anodization. Figure 2 shows the experimental set up for the fabrication of TiO<sub>2</sub> NTs. The anodization process was conducted in an electrochemical set up consisting of Ti substrate as the anode and platinum rod as the cathode. The distance between cathode and anode was fixed at 2.5 cm. The electrodes were submerged in 35 mL of ionic liquid electrolyte solution and the experiment was conducted for 1 h at a constant voltage of 20 V (DC power supply) and under ambient condition. After anodization, the as-synthesized samples were removed immediately from the electrolyte solution and rinsed with distilled water. The samples were air-dried in ambient atmosphere. The as-synthesized TiO<sub>2</sub> NTs (denoted as TiO<sub>2</sub>) were calcined at 350, 450, or 550 °C for 2 h with a heating rate of 5 °C/min in static air. The calcined TiO<sub>2</sub> NTs were labeled as T\_TiO<sub>2</sub>. NTs calcined at 350 °C.



Figure 2. Schematic diagram of the anodization setup.

#### 2.3. TiO<sub>2</sub> Nanotubes Characterization

The prepared TiO<sub>2</sub> NTs were characterized using Field emission scanning electron microscopy (FESEM) from Carl Zeiss instrument (SUPRA 55VP, Oberkochen, Germany) to investigate the surface morphology. The FESEM images were captured at 50 kX at an acceleration voltage of 20 kV. The crystalline structure and phase composition of TiO<sub>2</sub> NTs were determined using X-ray diffractometer (PANalytical X'Pert<sup>3</sup> Powder, AA Almelo, Almelo, The Netherlands) with Cu K $\alpha$  radiation (40 kV, 40 mA) at a 2 $\theta$  angle of 10° to 80° with the step size of 0.01°. The weight fractions of the anatase and rutile phases of the NTs, calcined at various temperatures, were estimated using Equation (1) [15,23] based on the relative intensities of the most dominant peaks for rutile (*I*<sub>R</sub>) and anatase (*I*<sub>A</sub>) at (110) and (101) planes, respectively:

$$f_r = \frac{1.26I_R}{I_A} + 1.26(I_R) \tag{1}$$

The average crystallite sizes of the TiO<sub>2</sub> NTs were determined from the characteristic diffraction peaks matching the (101) plane of anatase TiO<sub>2</sub> at  $2\theta = 25^{\circ}$  and the rutile crystallite was calculated from the rutile peak located at  $2\theta = 27^{\circ}$  using Scherrer's formula as shown in Equation (2):

$$D = \frac{K\lambda}{\beta Cos\theta}$$
(2)

where K is the Scherrer's constant (0.9),  $\lambda$  is the X-ray wavelength (0.15418 nm),  $\beta$  is the full width at half maximum of the selected diffraction peak (in radian), and  $\theta$  is the Bragg's angle.

The TiO<sub>2</sub> NTs were further analyzed using XPS from Thermo-Fischer (K-alpha, Madison, WI, USA). Al K $\alpha$  (1486.60 eV) was used as an X-ray excitation source with C1s correction at 285.73 eV

for calibration. The Ti2p and O1s XPS signals were deconvoluted using Gaussian curve fitting. Photoluminescence (PL) spectra of the  $TiO_2$  NTs were recorded using Horiba LabRam HR Evolution spectrometer (Minamiku Kyoto, Japan) at room temperature operated at 325 nm excitation light equipped with He-Cd laser. The PL spectra were recorded in the range from 350 nm to 700 nm. The bandgap values were estimated using Planck's expression as shown in Equation (3) below:

$$E = \frac{hc}{\lambda} \tag{3}$$

where *E* is photon energy (eV), h is Planck's constant (6.626 ×  $10^{-34}$  J.s), c is speed of light constant (3.00 ×  $10^8$  ms<sup>-1</sup>), and  $\lambda$  is wavelength (m).

#### 2.4. Photodegradation of Formaldehyde

The photocatalytic experiments for photodegradation of HCHO were carried out in batch mode. The photoreactor system for HCHO photodegradation under visible light irradiation is shown in Figure 3. The anodized Ti foil containing TiO<sub>2</sub> NTs was positioned inside a 250 mL quartz photoreactor with a holder. HCHO (37% in aqueous solution, Sigma Aldrich, Darmstadt, Germany) of 3.22 ppm was added inside a tightly sealed stainless-steel container (1). The quartz photoreactor (2) was filled with HCHO vapor by diffusion when V1 and V2 valves were opened. After 30 min equilibration in the dark, the quartz photoreactor was irradiated with a halogen lamp (150 W) as the light source, which mainly consists of a visible light region, and was positioned 7 cm beneath the photoreactor. The intensity and wavelength of the light was 1379.67 W/m<sup>2</sup> and 400–750 nm, respectively. The photodegradation of HCHO was monitored by sampling at 30 min interval for 150 min. A digital formaldehyde sensing meter (Hal Tech, Wetherill Parl, NSW Australia) (3) was connected to V3 to measure the HCHO vapor concentration. The working range of the sensor is up to 10 ppm. During the photoreaction, the temperature of the photoreactor was maintained at 25 ± 1 °C using a cooling fan. At the end of the reaction, the remaining gas flows through scrubber (4) containing H<sub>2</sub>O, which acts as the scrubber for HCHO.



Figure 3. Photoreactor system for formaldehyde (HCHO) photodegradation.

The photodegradation performance, X was monitored and calculated using Equation (4):

$$X(\%) = \frac{C_0 - C_t}{C_0} \times 100\%$$
(4)

where X% denotes the percentage of HCHO removal,  $C_0$  is the initial concentration of HCHO, and  $C_t$  represents the concentration of formaldehyde at sampling time, t.

### 3. Results and Discussion

#### 3.1. Characterization of the TiO<sub>2</sub> Nanotubes

#### 3.1.1. Surface Morphology

Figure 4 shows the effect of calcination temperature on the morphology of the synthesized  $TiO_2$  NTs. Figure 4a of as-synthesized  $TiO_2$  NTs showed discrete and irregular shaped  $TiO_2$  NTs formed covering the surface of the Ti foil when no heat treatment was introduced. It can be observed that almost the whole surface is covered by  $TiO_2$  NTs and is denser than other samples that were calcined at different temperatures. As for the  $TiO_2$  NTs calcined at 350, 450, and 550 °C, obvious changes can be seen from Figure 4b–d. The tubular structure was distorted, and a more severe pore rupture and disintegration of the surface was observed upon increasing calcination temperature to 550 °C. It is quite evident that the NTs after being calcined apparently had a poorly ordered structure, which can be attributed to the increase in internal stress with the shrinkage of thin-film during the calcination process [37] that destroys the NTs structure. This may also be ascribed to destruction and coalescence at the top of the NTs walls [38]. Furthermore, with increasing calcination temperature, a large area of the surface was covered by isolated NTs indicating that the better-ordered structure was destroyed to a great extent, which can be due to the phase transformation from anatase to rutile [37]. In fact, the phase transition from anatase to rutile was confirmed by X-ray diffraction analysis as discussed in Section 3.1.4.



**Figure 4.** Field emission scanning electron microscopy (FESEM) images of anodized Ti foil for (**a**) uncalcined and after calcination at (**b**) 350 °C, (**c**) 450 °C, and (**d**) 550 °C at 50 kX magnification.

It is important to note that chloride-containing electrolyte (ChCl) was employed for the fabrication of TiO<sub>2</sub> NTs in the present study as opposed to previous works where fluoride-based electrolytes were

predominantly used for the fabrication of NTs. Previous reports by Hahn et al. [28] and Ng et al. [39] provides evidence that chloride ion in an organic electrolyte can be used to grow  $TiO_2$  NTs efficiently. In addition, the viscosity of organic-based electrolytes can influence the diffusion of ionic species, resulting in the altering of the morphology of NTs. Bervian et al. [40] found that  $TiO_2$  NTs anodized in a glycerol-based electrolyte displayed higher mobility of ionic species, thus resulted in increased growth rate.

The mechanism of TiO<sub>2</sub> NTs formation on the surface of Ti foil can be explained using Equations (4)–(7). The growth of oxide on the surface of Ti foil is from the interaction between cations (Ti<sup>4+</sup>) and O<sup>2-</sup> species formed in the organic electrolyte [41]. During the early stage of the anodization process, a TiO<sub>2</sub> layer was grown on the surface of the Ti foil after a fixed potential (20 V) was applied (Equation (6)). According to reported principle, TiO<sub>2</sub> growth was based on three key processes [42]. On the surface of Ti substrate, there was the formation of a TiO<sub>2</sub> layer, which can be expressed by Equation (8) [42–44]:

Process 1: Field-assisted oxidation at metal/oxide interface

$$\mathrm{Ti} \to \mathrm{Ti}^{4+} + 4\mathrm{e}^{-} \tag{5}$$

$$\mathrm{Ti}^{4+} + 2\mathrm{O}^{2-} \to \mathrm{Ti}\mathrm{O}_2 \tag{6}$$

$$Ti(OH)_4 \to TiO_2 + 2H_2O \tag{7}$$

$$2\text{Ti} + 2\text{H}_2\text{O} \rightarrow 2\text{TiO}_2 + 4\text{e}^- + 4\text{H}^+$$
 (8)

After that,  $Cl^-$  ions randomly attacked  $TiO_2$  and  $TiO_2$  started dissolving due to electrochemical etching leading to the formation of pits and a thick layer of  $TiO_2$  due to migration of oxide ions. NTs were then grown and elongated inside of the pits and the oxide layer releases periodically until the complete transformation of Ti metal to  $TiO_2$ . The negatively charged ions within the electrolyte, particularly  $OH^-$  and  $Cl^-$ , moved toward Ti. The  $OH^-$  would be responsible for the formation of  $TiO_2$ . In our case, the  $OH^-$  is most probably coming from the water content that is present in the mixture of choline chloride and glycerol. In fact, the water content in the mixture was confirmed from water content analysis using Karl Fischer assay and was found to be 0.5%. A sample of the mixture of ChCl and glycerol along with the reagent (CombiTitrant) were introduced into a titration cell and dissolved. The reagent was released by the induction of an electrical current and the amount of current required to convert the water is the determinant of the amount of moisture present in the sample. Karl Fischer assay was used because it has the capability to measure the moisture or water content as low as 200 µg. Meanwhile, randomly attacked  $Cl^-$  ions etched the Ti metal, which cause dissolution of Ti as  $[TiCl_6]^{2-}$  for nanotubes formation [43,45–47]. The formation of  $TiO_2$  nanotubes under the influence of  $Cl^-$  is given in Equations (9) and (10) [48]:

Process 2: Field-assisted dissolution at oxide/electrolyte interface at tube bottom

$$\mathrm{Ti}^{4+} + 6\mathrm{Cl}^- \to [\mathrm{Ti}\mathrm{Cl}_6]^{2-} \tag{9}$$

Process 3: Chemical dissolution/etching of the fabricated tubes at the tubes top

$$\text{TiO}_2 + 4\text{H}^+ + 6\text{Cl}^- \rightarrow [\text{TiCl}_6]^{2-} + 2\text{H}_2\text{O}$$
 (10)

To further understand the mechanism of nanotubes formation, the anodization reaction could be represented by the simplified diagram shown in Figure 5 [39]:



Figure 5. Illustration of the formation TiO<sub>2</sub> nanotubes adapted from Reference [39].

#### 3.1.2. XPS Analysis

XPS analysis was performed in order to obtain better insight into the surface properties and nature of the chemical bonding in the synthesized TiO<sub>2</sub> NTs. Figure 6 shows the high-resolution deconvoluted Ti2p and O1s spectra of the TiO<sub>2</sub> and TiO<sub>2</sub> NTs calcined at different temperatures. For the TiO<sub>2</sub> sample, the binding energies (BE) of Ti2p<sub>3/2</sub> and Ti2p<sub>1/2</sub> were located at 459.4 and 465.2 eV, respectively, which are the characteristic features of anatase. The calculated difference in BE of Ti2p<sub>3/2</sub> and Ti2p<sub>1/2</sub> ( $\Delta$ BE = BE Ti2p<sub>3/2</sub> – Ti2p<sub>1/2</sub>) is equaled to 5.8 eV, which are characteristic to Ti<sup>4+</sup>–O bonds in TiO<sub>2</sub> [45,46]. Similar Ti2p<sub>3/2</sub> and Ti2p<sub>1/2</sub> peak positions were observed for the 350\_TiO<sub>2</sub> NTs. However, the Ti2p<sub>3/2</sub> and Ti2p<sub>1/2</sub> peaks of the sample calcined at 450 °C exhibit a negative shift of 0.1 eV and the peaks are now located at 459.3 eV and 464.8 eV. Compared to the TiO<sub>2</sub> sample, the 550\_TiO<sub>2</sub> NTs shows a negative shift 0.6 eV of Ti2p<sub>3/2</sub> and Ti2p<sub>1/2</sub> and the peaks are now centered at 458.8 and 464.5 eV, respectively.

The shift in the position of Ti2p<sub>3/2</sub> and Ti2p<sub>1/2</sub> peaks indicated the influence of calcination temperature on the electronic state of the Ti element; most probably, the Ti<sup>4+</sup> is reduced partially due to the loss of oxygen as the heat treatment temperature was increased from 350 to 550 °C [47]. According to previous reports on TiO<sub>2</sub> NTs, a negative shift in BE is associated with the additional screening of extra electrons in the crystal field with an increase in calcination temperature [49]. However, the  $\Delta$ BE value between Tip<sub>1/2</sub> and Tip<sub>3/2</sub> for samples 350\_TiO<sub>2</sub>, 450\_TiO<sub>2</sub>, and 550\_TiO<sub>2</sub> were 5.7, 5.6, and 5.7 eV, respectively, which are typical of the existence of Ti<sup>4+</sup> on the surface of TiO<sub>2</sub> lattice [45].

The major peaks of O1s with BE regions of 530.8–530.0 eV refer to  $Ti^{4+}$ –O in  $TiO_2$  [37,50]. The rest of the two peaks at BE region 532.1–531.2 eV and 533.7–532.2 eV were associated with adsorbed oxygen and OH group, respectively [51,52]. It was reported that the OH group gradually decreased by increasing the calcination temperature, which was suggested due to the chemical reaction that took place on the surface of  $TiO_2$  during the heat treatment process, as shown in Equation (11) [50]:



**Figure 6.** XPS spectra of (**a**–**d**) Ti2p and (**e**–**h**) O1s of TiO<sub>2</sub> nanotubes (NTs) calcined at various temperatures.

## 3.1.3. Optical Properties

The photocatalytic activity of  $TiO_2$  depends on the duration of the separation between the excited electrons and holes on its surface. Therefore, the photoluminescence (PL) spectra of the  $TiO_2$  NTs were measured to investigate the effect of calcination temperature on the optical properties and behavior of electron-hole recombination. The emission spectra of the as-synthesized and calcined  $TiO_2$  NTs in the wavelength range of 350–700 nm, with excitation at 325 nm, are shown in Figure 7a. The main

emission peaks of all the  $TiO_2$  NTs appeared at 386 nm (3.21 eV), 445 nm (2.80 eV), 555 nm (2.23 eV), and 596 nm (2.08 eV). The peak at 386 nm was attributed to the bandgap transition, corresponding to the bandgap energy of anatase. The samples  $0_TiO_2$ ,  $350_TiO_2$ ,  $450_TiO_2$ , and  $550_TiO_2$  showed the emission peaks at 442 nm, 445 nm, 448 nm, and 452 nm, respectively, which can be attributed to the band-edge free excitation. These wavelengths (nm) were converted to energy (eV) according to Pishkar et al. [53] and the values are given in Table 1. From the inspection of Table 1, it appears that the bandgap gap energy decreased with increasing calcination temperature and the results are consistent with the results reported by Mioduska et al. [54] that the calcination temperature can influence the energy bandgap. It is suggested that the increase in temperature can cause an incremental increase in the absorption coefficient due to the increase of defects sites. Electron-hole pairs are produced through photon absorption, generating a field that could change the optical attributes and electronic structure of the product [55].



**Figure 7.** The effect of calcination temperatures on (**a**) the PL spectra and (**b**) the valance band positions of TiO<sub>2</sub> NTs.

**Table 1.** Crystallite size, inner tube diameter, and bandgap of the  $TiO_2$  photocatalyst from different characterization methods.

Photocatalyst	Average Crystallite Size (nm)		Band Gap Calculated	Valance Band
	Anatase	Rutile	from PL (eV) Position	Positions (eV)
TiO <sub>2</sub>	30.31	-	2.80	2.77
350_TiO <sub>2</sub>	60.50	-	2.76	2.80
450_TiO <sub>2</sub>	44.36	13.93	2.79	2.43
550_TiO <sub>2</sub>	33.27	27.85	2.74	1.84

The PL peak intensities of the as-synthesized TiO<sub>2</sub> NTs showed a significant decrease compared to those calcined at 350, 450, and 550 °C. Furthermore, the intensity variations suggest that the PL spectrum of TiO<sub>2</sub> NTs with larger crystallite size is dominated by a radiative recombination of electrons via intrinsic defects states. The blue emission is visible due to the available surface energy that induces quantum confinement effect and also due to the presence of oxygen vacancies defects on the surface of TiO<sub>2</sub> [56].

The bandgap values obtained from the PL spectrum are shown in Table 1. No significant effect of calcination temperature on the bandgap values of the NTs was observed in the present study. However, a variation in calcination temperatures showed a significant effect on the valence band positions of the  $TiO_2$  NTs, as shown by the valence band (VB) XPS spectra in Figure 7b. The valence band of the as-synthesized NTs was located at 2.77 eV, while the valence band positions

were shifted to 2.80, 2.43, and 1.84 eV with the increase in calcination temperature to 350, 450, and 550 °C, respectively. Pishkar et al. [53] and Ghows and Entezari [57] reported in their findings that bandgap values calculated from the PL spectrum are in good agreement with their diffuse reflectance spectra (DRS) results. It can be observed that the PL peak intensity gradually increases with increasing calcination temperature. Yu et al. and Sang et al. [58,59] reported that low PL intensity suggests a low recombination rate of excited charges on the surface of the photocatalyst. As expected, the high crystallinity and low number of defects sites results in a lower PL intensity, which is consistent with XRD patterns. However, it should be noted that the sample calcined at 350 °C is highly crystalline as compared to the uncalcined sample. but the slight increase in PL intensity of the 350\_TiO<sub>2</sub> can be due to the presence of random cracks on the surface of Ti foil [58]. The bandgap energy decreases with increasing calcination temperature, which reflects anatase to rutile phase alteration as inferred from the XRD analysis. A previous study also reported that, usually, PL intensity is proportional to the amount of produced hydroxyl radicals that are needed to enhance photocatalytic performance [38].

#### 3.1.4. XRD Analysis

The mechanism of phase transformation is important in order to control the nanostructure and material properties. The crystallite size and crystallinity, as well as phase structure and composition, play an important role in the photocatalytic activity of TiO<sub>2</sub>. Figure 8a displays the XRD patterns of the synthesized TiO<sub>2</sub> NTs. TiO<sub>2</sub>, 350\_TiO<sub>2</sub>, and 450\_TiO<sub>2</sub> patterns congruously show characteristic diffraction peaks at  $2\theta = 25.2^{\circ}$ ,  $48.1^{\circ}$ , and  $54.1^{\circ}$  corresponding to 101, 200, and 105 planes of the anatase phase, respectively. It can be observed for the peak at  $2\theta = 25.2^{\circ}$  that its intensity increased at 350 °C calcination temperature but decreased when higher heat treatment was introduced. Compared to the other samples, TiO<sub>2</sub> NTs calcined at 450 °C exhibited the rutile phase with a peak at  $2\theta = 27.5^{\circ}$ , as shown in Figure 8b. The weight fraction of the rutile peak was 19%. The sample calcined at 550 °C (550\_TiO<sub>2</sub> NTs) exhibited one additional peak of  $2\theta = 69.9^{\circ}$  corresponding to 220 planes of rutile. With this additional rutile peak, the weight fraction was increased to 36%. The results suggest that the anatase to rutile phase transformation occurs at a higher calcination temperature. It has been reported that the anatase crystal phase can be induced by thermal treatment starting from 350 °C and the shift of the diffraction peaks to a higher angle could be attributed to the difference in particle size, impurities, and synthesis methods [50].



**Figure 8.** XRD patterns for TiO<sub>2</sub> calcined at different calcination temperature (**a**,**b**) magnified XRD patterns from 20° to 30° (2 $\theta$ ).

The average crystallite sizes of the TiO<sub>2</sub> NTs calcined at different temperatures are listed in Table 1. The TiO<sub>2</sub> NTs calcined at 350 °C showed the highest crystallite size (60.50 nm) compared to other samples. It should be noted that the calculated crystallite size was larger than the inner tube diameter

of  $350_{\text{TiO}_2}$  sample, which can be due to the inhomogeneous distribution of TiO<sub>2</sub> NTs [60]. There is a decreasing trend of the inner tube diameter in the temperature range 450-550 °C for the average size of the anatase grains (from 48.84 to 25.05 nm). This suggests that in this range of temperature, some parts of anatase grains started to transform into rutile TiO<sub>2</sub> [61]. Anatase to rutile transformation happens by coarsening, so rutile crystallite size is expected to be bigger than anatase. In the present study, the crystallite size (13.93 nm) of rutile was smaller than the crystallite size of anatase (44.36 nm) at a calcination temperature of 450 °C. One of the possible reasons for this is that the values of crystallite sizes reported here are only relative because the contribution of the strain on peak broadening has not been taken into consideration in the calculation using the Scherrer Equation. The other reason could be the appearance of rutile at a lower temperature (450 °C) in our prepared samples as compared to previous reports, where they reported the transformation of anatase to rutile to have occurred at 600 °C.

However, the crystallite size of rutile rose up to 27.85 nm when the calcination temperature was increased to 550 °C [61]. Generally, the rutile phase appears at higher temperature ca. 500 °C. For instance, Low et al. studied the crystalline behavior of pure and chromium-doped TiO<sub>2</sub> nanotubes, where they reported the transformation of anatase to rutile to have occurred at 600 °C [62]. On inspection of Table 1, it is evident that the crystallite size of pure TiO<sub>2</sub> NTs was smaller compared to the calcined TiO<sub>2</sub> NTs samples. Initially, the crystallite size of anatase rose up to 60.50 nm when the calcination temperature was increased to 350 °C. However, the anatase crystallites size began to decrease when the calcination temperature was raised to 450 and 550 °C. This could be due to the fact greater anatase phase crystallites first turn into the rutile phase during the phase transition, which results in the decrease of average crystallite size of the anatase.

The dominant peak at  $2\theta = 25.2^{\circ}$  showed the highest intensity when calcined at 350 °C. However, a higher calcination temperature led to a decreased intensity of the anatase peak. This implies that the crystallinity of the TiO<sub>2</sub> NTs decreased with increasing calcination temperature.

#### 3.2. Photocatalytic Activity

The photocatalytic activities of the fabricated  $TiO_2$  NTs were examined for the removal of gaseous HCHO. Figure 9a shows the removal of HCHO using four different type TiO<sub>2</sub> NTs: as-synthesized NTs, and NTs calcined at 350 °C, 450 °C, and 550 °C under visible light irradiation. Without light (adsorption equilibrium), almost 9% of HCHO was removed by TiO<sub>2</sub> NTs. A control experiment has been done for blank where no reaction has been observed. A noticeable improvement of HCHO removal was observed in the presence of light suggesting significant photocatalytic activity of the fabricated TiO<sub>2</sub> NTs. The higher HCHO removal efficiency (62%) was obtained by TiO<sub>2</sub> NTs calcined at 450 °C. The higher efficiency of 450\_TiO<sub>2</sub> NTs can be attributed to its mixed crystal composition of anatase and rutile, where the rutile phase is dispersed into the anatase phase, resulting in an improvement of electronic interactions between the anatase and rutile phases. In other words, it is believed that the lower bandgap rutile phase absorbs photons and formed electron-hole pairs while the anatase phase traps the electrons. Thus, reducing the recombination of electron and allowing the hole to transfer to the surface to react. It has been reported that the number of photons being in contact with the surface of the photocatalyst actually controls the degradation reaction [63]. The latter is an indication that the reaction takes place only in the adsorbed phase of the photocatalyst. Although uncalcined  $TiO_2$  NTs exhibits anatase phase, its photocatalytic activity was lower (46%) than the calcined samples. Nevertheless, the HCHO removal efficiency in the current study is higher as compared to a previous study, where they only removed 40% of HCHO from air using  $TiO_2$  immobilized on a low melting point polymer (TiO<sub>2</sub>@LMPET) [51].



**Figure 9.** (a) Effect of calcination temperatures on the photodegradation of formaldehyde. Experimental conditions: 3.22 ppm initial formaldehyde concentration, 25 °C, 150 W, and (b) recycling results of 450\_TiO<sub>2</sub> HCHO removal.

The synthesized TiO<sub>2</sub> NTs showed dissimilar HCHO removal concentration. This can be due to the structural dissimilarity and differently sized inner tube diameter. The uncalcined sample and the NTs samples calcined at 350 and 550 showed 16.17, 17.25, and 19.01 mg·L<sup>-1</sup>·g<sup>-1</sup> HCHO removal concentration, respectively. The HCHO removal concentration of  $450_{-}$ TiO<sub>2</sub> NTs was 21.56 mg·L<sup>-1</sup>·g<sup>-1</sup>, which is highest among all the samples. The results demonstrate that the HCHO removal concentration shows a significant increase when increasing the calcination temperature to  $450 \,^{\circ}$ C. However, above 450 °C, the photocatalytic removal concentration of HCHO declines. TiO<sub>2</sub> NTs sample calcined at 450 °C achieved the highest degree of HCHO degradation may be ascribed due to the optimum crystallinity of anatase and rutile phase developed at this temperature [15]. It is known that Degussa P25-TiO<sub>2</sub> is often used as the benchmark photocatalyst due to its superior photocatalytic activity, which contains 80% anatase and 20% rutile phases [64]. This implies that the presence of optimum anatase to rutile mass ratio is also beneficial to increase the photocatalytic activity [65]. Furthermore, Razali et al. stated that anatase fraction can provide a high amount of oxygen vacancy

sites that could contribute to the production of active radicals. There is a direct correlation between the adsorption of organic pollutants and surface coverage of the TiO<sub>2</sub> photocatalyst [66]. Meanwhile, the quantum yield of HCHO degradation,  $\theta_{\text{HCHO}}$  for TiO<sub>2</sub> and 350\_TiO<sub>2</sub>, was 0.95 and 0.98, respectively. Most importantly, it was found that it further increased the calcination temperature to 450 and 550 °C led to the enhancement of  $\theta_{\text{HCHO}}$  by almost 17% ( $\theta_{\text{HCHO}} = 1.14$  and 1.28, respectively). Quantum yield is defined as the light efficiency of processes, the ratio of the reaction rate to the absorption rate of photons [67,68]. Significantly, the fabrication of TiO<sub>2</sub> NTs and application for the removal of gaseous HCHO from air have some other implications as well. The fabrication of TiO<sub>2</sub> NTs display a promising future in virtue of the mild operation conditions required for their use and suitability for large-scale fabrication, which could be employed in indoor pollution control technology.

The stability of the photocatalyst is an important consideration for industrial-scale applications in indoor air pollution abatement. To demonstrate the stability of the synthesized TiO<sub>2</sub> NTs photocatalyst, based on its high performance, 450\_TiO<sub>2</sub> was reused for the recycling tests for three consecutive runs due to its higher photocatalytic performance as. The results of the recyclability tests are shown in Figure 9b. The photocatalytic activity reduced from 20.00 mg·L<sup>-1</sup>·g<sup>-1</sup> to 17.50 mg·L<sup>-1</sup>·g<sup>-1</sup> after the three consecutive runs, and the photocatalytic activity was well retained. The results indicate that synthesized TiO<sub>2</sub> NTs are very stable under the employed reaction conditions.

Based on previously published reports, HCHO was adsorbed onto the surface of TiO<sub>2</sub> NTs [7,69,70]. In the below reaction pathway, it is possible that each oxidation step may proceed on the photocatalyst surface either by direct reaction of the adsorbed organic molecules with valence band holes (h<sub>VB</sub>) or by an indirect reaction path through hydroxyl radical attack [30]. For pure anatase  $TiO_2$ , when irradiated with ultraviolet light having photon energy higher than its bandgap, the electron in the conduction band can be photo-excited and transferred to the molecular oxygen provided that the gas inside the photoreactor was air-producing reactive oxygen radicals  $O_2^{-}$  (Equation (13)), where  $O_2$  acts as an oxidizing agent that can degrade HCHO molecules. In •OH, the radical-mediated path implies, first of all, the electrons were excited from the valance band to the conduction band of  $TiO_2$  NTs and leaving holes behind ( $h_{VB}$ ) in Equation (12) [30]. The  $h_{VB}$  reacts with adsorbed water, producing protons and reactive hydroxyl radicals in Equation (14). Photocatalytic oxidation of HCHO results from Equation (15), i.e., from the interaction of •OH radicals produced by the reaction of water with excited charge carrier ( $h_{VB}$ ), accumulated on the TiO<sub>2</sub> NTs under visible light irradiation. It is important to note that Equation (15) cannot proceed via a direct hole-mediated path, but can only proceed through the attack of hydroxyl radicals because an extra oxygen atom is required to transform HCHO into HCOOH, which can only be provided indirectly by water through •OH radicals.

$$TiO_2 + h_{VB} \rightarrow e^- + h^+ \tag{12}$$

$$e^- + O_2 \to \bullet O_2^- \tag{13}$$

$$h^+ + H_2O \to \bullet OH + H^+ \tag{14}$$

$$\bullet OH + HCHO \rightarrow H_2O + \bullet CHO \tag{15}$$

$$\bullet CHO + \bullet OH \to HCOOH \tag{16}$$

$$HCOOH + \bullet OH \to \bullet COOH + H^+$$
(17)

$$\bullet \text{COOH} + \bullet \text{OH} \to \text{CO}_2 + \text{H}_2\text{O} \tag{18}$$

#### 4. Conclusions

Mixed anatase–rutile phase  $TiO_2$  nanotubes were successfully fabricated via electrochemical anodization of titanium foil in this study. The  $TiO_2$  nanotubes with different crystallinities, lower bandgap, and varying crystal composition including pure anatase and mixed-phase anatase–rutile

were grown on Ti foil using a mixed electrolyte containing glycerol and choline chloride at a constant voltage of 20 V for 1 h. The effect of calcination temperature on the properties of the TiO<sub>2</sub> nanotubes was studied. The calcination temperature greatly affects the crystallization, crystalline structure, phase composition, and surface morphology. At a low temperature, pure anatase phase TiO<sub>2</sub> nanotubes were obtained. The phase transformation from anatase to rutile occurs at 450 °C to 550 °C. High calcination temperature led to the distortion of the TiO<sub>2</sub> nanotubes. The bandgap of the fabricated TiO<sub>2</sub> nanotubes was reduced to 2.74 eV from 3.20 eV of standard anatase TiO<sub>2</sub>. Interestingly, calcination at 450 °C led to enhanced photocatalytic activity of the TiO<sub>2</sub> nanotube as well as  $\emptyset_{\text{HCHO}}$  by 17%, which may be due to the synergistic effect of the mixed anatase–rutile phase. Almost 21.56 mg·L<sup>-1</sup>·g<sup>-1</sup> of gaseous formaldehyde was removed within 120 min of visible light reaction.

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