

Review

Coral Reef Resilience in Taiwan: Lessons from Long-Term Ecological Research on the Coral Reefs of Kenting National Park (Taiwan)

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Abstract: Coral reefs in the Anthropocene are being subjected to unprecedented levels of stressors, including local disturbances—such as overfishing, habitat destruction, and pollution—and large-scale destruction related to the global impacts of climate change—such as typhoons and coral bleaching. Thus, the future of corals and coral reefs in any given community and coral-Symbiodiniaceae associations over time will depend on their level of resilience, from individual corals to entire ecosystems. Herein we review the environmental settings and long-term ecological research on coral reefs, based on both coral resilience and space, in Kenting National Park (KNP), Hengchun Peninsula, southern Taiwan, wherein fringing reefs have developed along the coast of both capes and a semi-closed bay, known as Nanwan, within the peninsula. These reefs are influenced by a branch of Kuroshio Current, the monsoon-induced South China Sea Surface Current, and a tide-induced upwelling that not only shapes coral communities, but also reduces the seawater temperature and creates fluctuating thermal environments which over time have favoured thermal-resistant corals, particularly those corals close to the thermal effluent of a nuclear power plant in the west Nanwan. Although living coral cover (LCC) has fluctuated through time in concordance with major typhoons and coral bleaching between 1986 and 2019, spatial heterogeneity in LCC recovery has been detected, suggesting that coral reef resilience is variable among subregions in KNP. In addition, corals exposed to progressively warmer and fluctuating thermal environments show not only a dominance of associated, thermally-tolerant *Durusdinium* spp. but also the ability to shuffle their symbiont communities in response to seasonal variations in seawater temperature without bleaching. We demonstrate that coral reefs in a small geographical range with unique environmental settings and ecological characteristics, such as the KNP reef, may be resilient to bleaching and deserve novel conservation efforts. Thus, this review calls for conservation efforts that use resilience-based management programs to reduce local stresses and meet the challenge of climate change.

Keywords: Taiwan; coral reef; marine national park; nuclear power plant; community dynamics; Symbiodiniaceae; long-term ecological data

1. Introduction

1.1. Coral Reef Ecosystems and the Impacts of Environmental Change

Coral reefs are one of the world's most diverse and productive marine ecosystems, providing essential goods and services for millions of people. A remarkably high diversity of species and interactions makes this ecosystem particularly sophisticated and sensitive to human-induced perturbations [1]. Overfishing, pollution and habitat destruction are among the most common local stressors altering coral reef ecosystem dynamics (reviewed in [1–4]). In addition, the increases in seawater temperature and ocean acidification are considered to be the two major global stressors responsible for the worldwide degradation of coral reef health (reviewed in [3]). Seawater temperatures 0.5–1.0 °C above summer average persisting for days to weeks could result in disruption of mutualistic relationship between coral hosts and their symbiotic algae (also known as zooxanthellae, family Symbiodiniaceae), resulting in “coral bleaching”; if this persists for months, it can result in mass coral mortality [5]. Several major severe bleaching episodes have been recorded since 1979. Among them, a 1998 event had the most devastating known effect, impacting over 75% of reefs worldwide, and wiping out nearly 16% of them [6]. In addition, between 2014 and 2017, “back-to-back” thermal anomalies occurred on the northeastern coast of Australia, causing massive coral bleaching in the north and mid sections of the Great Barrier Reef; this is arguably the worst-ever bleaching in the history of the GBR [7,8] and other parts of Australia [9,10]. Similar global-scale coral bleaching events (GCBE) that result in high coral mortality, the rapid decline of reef structures, and unprecedented environmental impacts have also been reported in the Indian [11,12], Pacific [13–15], and Atlantic Oceans [16,17]. Scientists have therefore concluded that the 2014–2017 GCBE represents the first multi-year, global-scale coral bleaching event to cause bleaching and mortality two or more times over the 3-year event [18].

Ocean acidification—the ongoing decrease in the pH of the ocean caused by the uptake of anthropogenic carbon dioxide (CO₂)—is, on the other hand, decreasing the calcification, reducing coral growth, and limiting reef development. There appears to be no chance of maintaining atmospheric CO₂ concentrations under 450 ppm [19] and limiting global temperature increase to less than 2 °C [20] by the end of the century; however, corals and coral reef communities at low and high latitudes might demonstrate exceptional acclimatization and adaptation capacities to survive the future environmental changes [21], although it has been suggested that most species will fail to develop mechanisms to survive future conditions, and a worldwide decline in coral reefs now seems inevitable [3]. Recent reports have shown that recurrent and prolonged thermal anomalies above the threshold limit corals' resistance to stress [22] and leads to increased bleaching, mass destruction of coral reefs, and the loss of many coral species [7,8]. The latest IPCC reports show an imminent threat to tropical coral reefs as soon as 2030 if carbon emissions continue to increase and subsequently increase average seawater temperatures. It was concluded at 1.5 °C global heating, the world will lose 70%–90% of its coral reefs, but at 2 °C, virtually all of the world's coral reefs will be lost [19].

In addition to driving more intensive mass coral bleaching, warming oceans will also enhance the destructive potential of tropical storms, including typhoons in the West Pacific, hurricanes in the Caribbean, and cyclones in the South Pacific and Indian Oceans, by either increasing their frequencies [23] or intensities [24], although other studies suggest that the global frequency of cyclones might remain stable or even decrease by up to 40% under greenhouse conditions by the end of this century [25–27]. Despite these projections, future impacts remain debatable. The future climate might, in turn, be influenced by some of the most severe ecological impacts on coral reefs through direct physical disturbances, turbidity, sedimentation, or salinity changes that would result from the destruction of reef structures [28]. As a consequence, the diversity and biomass of fish and other fauna that require corals for shelter or food will be dramatically reduced [29].

1.2. Coral Reef Resilience under the Impacts of Environmental Change

Resilience, a theory introduced to describe how ecosystems respond to disturbances [20], has recently been applied to coral reef ecosystems to examine how organisms respond to and interact as the result of local and global stresses. Many studies [30–35] including [36] have developed different definitions of resilience when applied to coral reef ecosystems including that the resilience is influenced by many stochastic factors [36]. In other words, stochastic resilience or ecosystem resilience is the capacity of an ecosystem to overcome disturbances and reorganize to maintain original fundamental state [36–38]. In case of coral reefs, the disturbances vary in terms of time and space with different intensities. As a result, depending on the location and local environmental factors, resilience of a given reef and its coral communities will vary. Resilience-based management has been proposed as a realistic model to predict which sustainability measures can be achieved in coral reefs in the face of ocean warming and acidification and various local disturbances, and this strategy will help us set achievable goals for regional and local-scale management programs [39]. Moreover, beyond the ecosystem level, resilience also covers the overall ability of individuals, populations, or communities to respond positively after disturbance and restore some part of their original state [4]. For example, individual corals can show physiological resilience via survival, sustained growth, reproduction, and/or by shuffling their symbionts towards more thermally-tolerant genera and species (see the review by Carballo-Bolaños et al. in this issue). Coral populations can re-populate by recruiting new individuals, and communities can re-organise ecosystem traits such as productivity, diversity, trophic linkages, and sustained biomass through shifts in species composition [reviewed in 4]. In this review, we adopted the broad definition of resilience [4] to examine the resilience of coral-Symbiodiniaceae associations and coral communities to the long-term disturbances in KNP.

1.3. Coral-Symbiodiniaceae Associations Play a Key Role in Coral Resilience to Thermal Stress

Coral-Symbiodiniaceae associations are the crux of coral health and functioning, even though corals are multi-bionts [40]. One of the most important factors of this association is its resilience to seawater temperature anomalies. Coral-Symbiodiniaceae resistance mechanisms to temperature stress depend on the combination of mechanisms involving the coral host and/or its Symbiodiniaceae partners and whether the relationship that the coral host has with Symbiodiniaceae is specific or flexible [41–43]. Corals are known to associate with a wide range of Symbiodiniaceae genera. There are nine genera in Symbiodiniaceae, and each genus has its own characteristics that help corals survive in a wide range of environmental niches [44]. Studies have shown that symbiosis between coral hosts and different Symbiodiniaceae genera contribute to the divergence in their thermal tolerance under different environmental conditions [42,45].

For example, stress-tolerant *Durusdinium trenchii* has proven to be heat-tolerant, whereas *Cladocopium* species are the most sensitive to stress [41,46–48]. *Durusdinium*-associated corals are also known to inhabit reef environments that experience large surface seawater temperature fluctuations [42,49,50] and possess better survival chances under heat-treatment experiments [51]. Symbiodiniaceae diversity thus provides a mechanism for corals to adapt and/or acclimatize to changing environmental conditions. Since environmental factors vary both spatially and temporally, even at a micro-geographic scale, differences in Symbiodiniaceae diversity within a single host or among multiple species are important and contribute to coral resilience.

1.4. Coral Reefs in Taiwan, with a Focus on Kenting Coral Reefs

Taiwan, a continental island with several offshore islets, is located at the centre or junction of the Philippine-Japan island arc; the Tropic of Cancer, runs through the middle of Taiwan (Figure 1). Scleractinian coral occurrence and distribution in Taiwan is influenced by sea surface currents and seawater temperatures; it is found in patchy non-reefal coral communities, similar to a high-latitude environment, and occupies the Penghu Archipelago and northern, northeastern, and rocky eastern

coasts of Taiwan [52–54]. The marine environmental conditions in southern Taiwan—facing the junction of the Pacific Ocean, Bashi Channel, and South China Sea—are influenced by a branch of the Kuroshio Current (KC) and the South China Sea Surface Current (SCSSC) (Figure 1). Thus, tropical fringing reefs developed along the coast of the Hengchun Peninsula, with about 300 coral species described [52–54].

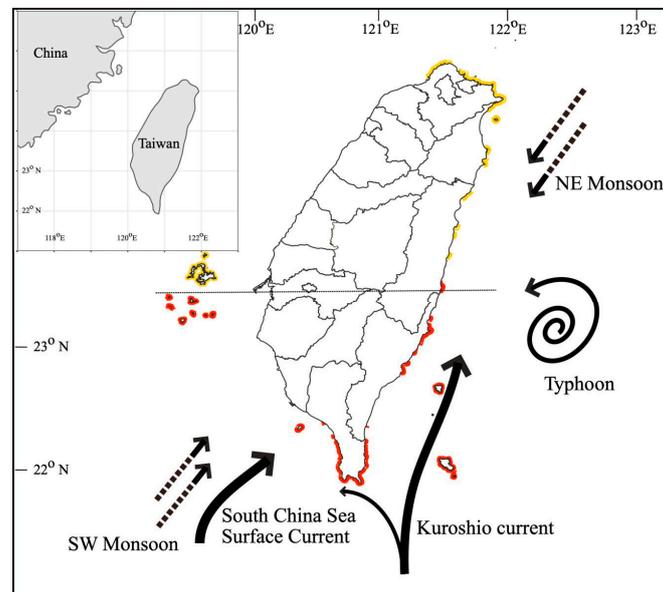


Figure 1. The distribution of corals in Taiwan, including non-reefal communities (yellow) and tropical coral reefs (red) divided roughly by the Tropic of Cancer, and the main climate factors, including the Kuroshio Current, typhoons, and northeast and southwest monsoons. The two types of coral communities are divided by the Tropic of Cancer. The information in the map was adapted from (Dai, C.-F. 2018). Taiwan map was downloaded from free to use Taiwan Map Store (<https://whgis.nlsc.gov.tw/English/0-1Introduction.aspx>).

In 1982, Kenting National Park (KNP), the first national park in Taiwan, was established to manage and conserve the uplifted reef landscape as well as the modern coral reef system (Figure 2). In addition, a nuclear power plant (NPP) that started operating in 1985 discharges heated seawater into Nanwan, KNP (Figure 2). Biological surveys and environmental monitoring have been conducted in KNP since the late 1970s to collect baseline data to manage KNP and monitor the environmental impact of NPP during construction and follow-up operations, particularly the thermal stress on the reef adjacent to the heated-water outlet (OL); data collection was interrupted from time to time due to shortages in funding or termination of monitoring projects (Table 1). Nevertheless, this long-term research effort (> 30 years) collected enough data for us to assess spatial and temporal changes in Taiwan’s coral reef ecosystems (Figure 2). However, similar to reefs from most of the tropical waters [55–58], the reefs in KNP have been subjected to human disturbances—including overfishing, habitat destruction, and sewage discharge—due to a growing population and poor management [59–62]. Combined with the synergistic effects from climate change, there has been a declining trend in KNP’s living coral cover [60,63].

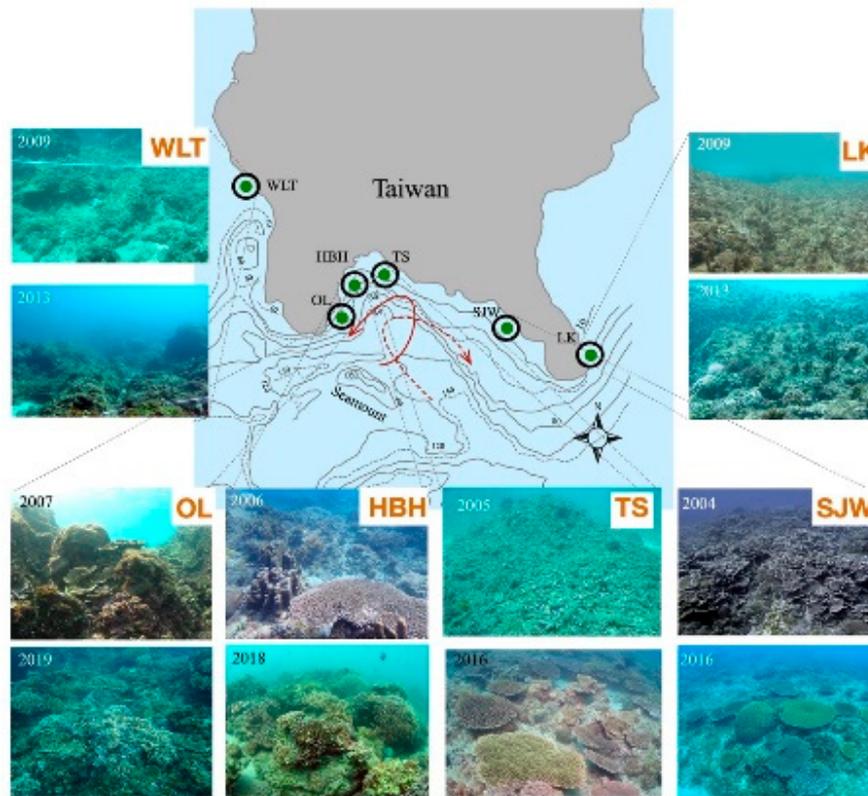


Figure 2. Depth contours in Nanwan, Kenting National Park, and coral communities at six sites and two time points, showing the status of coral communities through time. The sites include Wanlitung (WLT) on the west coast of Hengchun Peninsula, NPP Outlet (OL) and Houbihu (HBH) on the west coast of Nanwan, Tiaoshi (TS) and Siangjiaowan (SJW) on the east coast of Nanwan, and Longkeng (LK) on the east coast of Hengchun Peninsula. The solid and dotted arrow line indicates the direction of tidal flow patterns in flows and eddies. Depth contour figure was adapted from Lee, H.-J. etc. 1999a.

In this paper, we review the long-term studies of (1) environmental settings of KNP coral reefs, focusing on the cooling effect of tide-induced upwelling; (2) spatial and temporal dynamics of benthic communities and how they respond to large-scale disturbances, such as coral bleaching and typhoons; and (3) Symbiodiniaceae diversity and the responses of KNP corals to thermal-induced bleaching. By synthesizing the long-term ecological research data on coral reefs in KNP, we show that, by introducing resilience-based management programs to reduce local stresses, coral distribution along a small geographic range with unique environmental settings and Symbiodiniaceae diversity can be potentially resilient to climate change.

This section can be summarized as; (1) Coral reefs are facing increasing pressure from natural and anthropogenic disturbances and (2) Resilience of corals to such disturbances will depend on local-scale management programs, physiological resilience of individual species, recruitment potential, etc.

Table 1. Summary of the historical dataset conducting National Park-wide qualitative survey since 1986.

Year	Sites	Survey Method	Survey Area per Transect	Number of Replicates at Each Site	Survey Depth	Methods	Identification Level	Reference
1986–1987, 1998–1999	Wanlitung (WLT) Hungchai (HC) Leidashih (LDS) Houbihu (HBH) Tiaoshi (TS) Siangjiaowan (SJW)	Line intercept transect	10 m	25	3–23	A transect tape was placed perpendicular to the coast and extended seaward from 3 m depth to the reef edge at 25 m depth. A 10 m metal chain was placed parallel to the transect at 15 m intervals.	Species for corals Total algae	Dai, C.-F. (1988); Dai, C.-F. etc. (1998, 1999)
2003–2005 2008–2014 2016, 2018	Howan (HW) Wanlitung (WLT) Hungchai (HC) Leidashih (LDS) Houbihu (HBH) NPP Inlet (IL) Tiaoshi (TS) Siangjiaowan (SJW) Longkeng (LK) Jialeshuei (JLS)	30 m × 0.25 m belt transects except 20 m × 0.25 m in TS	7.5 m ² and 5 m ² in TS	3, except 9 in TS	5–10	Three or nine (at Tiaoshi) permanent belt transects were established along depth contours between 5 and 10 m depth at each site. Benthic organisms were quantified using 25 × 25 cm ² photo-quadrats. The percent cover of the benthic categories was determined using Coral Count with Excel Extensions software (Kohler, K.E. 2016), with 30 random points per quadrat. Surveys were conducted between March and May each year.	Species for corals (2003–2005, 2011, 2014); Genus for corals (2008–2010, 2012–2013); Morphology for scleractinian corals (2016); Genus and morphology for scleractinian corals (2018) Macroalgae Turf algae	Kuo, C.-Y. etc. (2012); Shao, K.-T. etc. (2002); Fang, L.-S. etc. (2003–2006); Wang, W.-H. etc. (2007); Ho, M.-J. etc. (2016); Ho, P.-H. etc. (2008–2011); Chen, C.A. etc. (2012–2014, 2016, 2018); Kuo, C.-Y. etc. (2007)

2. Environmental Settings of Coral Reef in KNP with Focus on Tide-Induced Upwelling

KNP is located on Hengchun Peninsula in southernmost Taiwan and receives seasonal influence from a branch of KC and SCSSC; the park can be divided into three geographic sub-regions, namely the west coast of Hengchun Peninsula facing the Taiwan Strait; the east coast of Hengchun Peninsula adjacent to the Pacific Ocean; and Nanwan, a semi-enclosed bay between two capes facing the Bashi Channel that connects the Pacific Ocean and Taiwan Strait (Figure 1). Distance between the east cape, known as Eluanbi, protruding farther south and the west cape, Maobitou, is about 14 km. The semi-enclosed basin of Nanwan is characterized by a zonally elongated seamount partially blocking the bay mouth. The seamount reaches up to 50 m below the sea surface (Figure 2). Toward the west side of Nanwan, there is almost no continental shelf. On the contrary, the shallow continental shelf shoreward of 80 m isobath is about 4 km wide, with isobaths running more or less parallel to the coastline at the east side of Nanwan. With this unique geomorphologic setting, the deeper portion of Nanwan forms an arc-shaped channel open at both ends between the seamount to the south and landmasses to the north [61–63].

By applying acoustic Doppler current profiler surveys, moored measurements, and numerical modelling, it has been shown that the headlands on either side of the Nanwan generate strong tidally-induced upwelling (TIU) within the bay during each phase of the tide [64,65]. Considerable difference in size between the flood and ebb eddies were observed following the geometry of the region. The entire Nanwan basin is filled by the flood eddy, while the western and central regions are filled by the ebb eddy. Eventually, the upwelling occurs within each eddy, causing two temperature drops per tidal cycle in western and central Nanwan, and only one drop in the eastern part (Figure 3) [64,65].

The significant temperature drops caused by TIU (Figure 3) might not only play an important role in shaping the coral community structure in different sub-regions of KNP, but also have a crucial cooling effect on corals in the Nanwan that resists the rising seawater temperature, both locally and globally. Analysis of in situ hourly temperature data showed a spatial heterogeneity in response to TIU at different sub-regions in KNP (Figure 3). No TIU was detected on the west coast of Hengchun Peninsula; thus, seawater temperature is constantly stable, with small fluctuations through the year (Figure 3a,b). In contrast, within the Nanwan and the east coast of Hengchun Peninsula, seawater temperature responds in concordance, but with different amplitudes, to the TIU, following the moon phases (Figure 3c–j). In west Nanwan, where the OL is located, large amplitudes were observed around new moons, with a maximum temperature fluctuation of 4.08 °C and 4.89 °C at 2 m and 7 m deep in the summer, respectively (Figure 3c–f), providing a significant cooling effect to remove thermal stress on coral reefs near the OL (Table 2). Towards east Nanwan and the corner of east Hengchun Peninsula, the effect of temperature drop by TIU is reduced in the summer, as the model predicted [64,65], but remains relatively strong in the winter (Figure 3g–j).

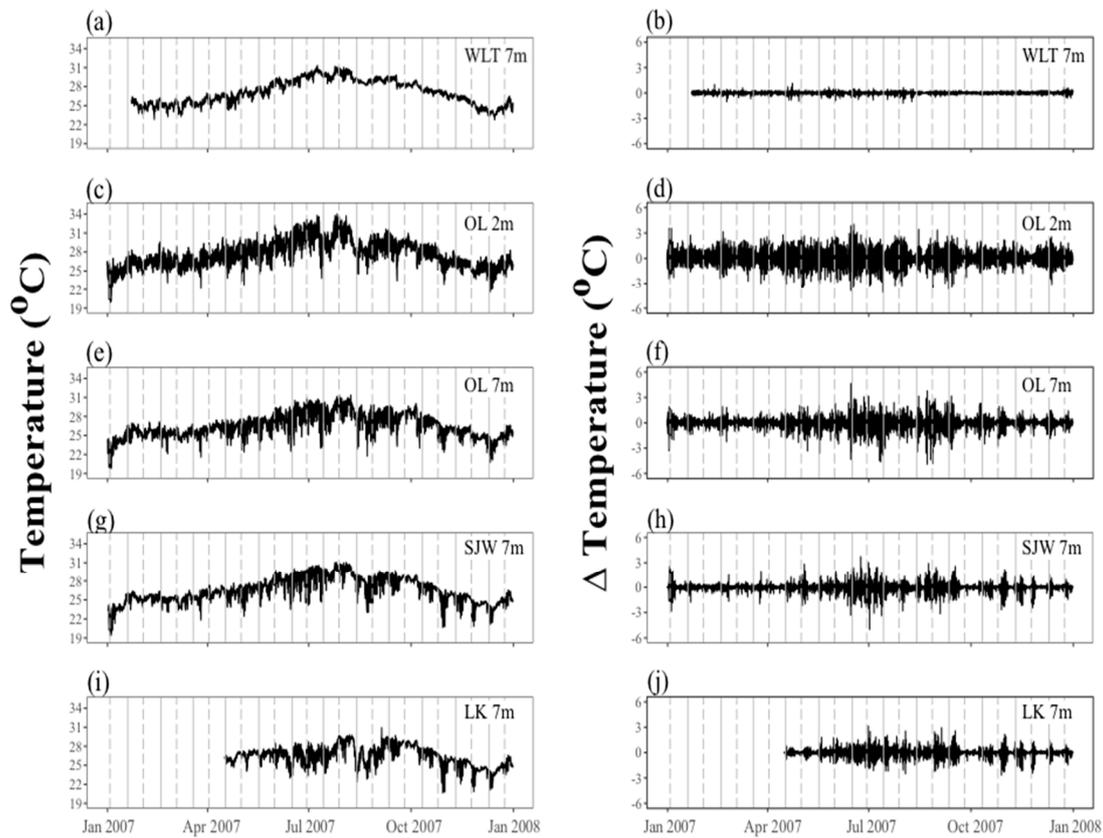


Figure 3. The in situ hourly average seawater temperatures and hourly average temperature differences among Wanlitung (WLT) (7 m deep: a,b), Outlet (OL) (2 m: c,d; 7 m: e,f), Siangjiaowan (SJW) (7 m: g,h), and Longkeng (LK) (i,j) from 1 January 2007 to 1 January 2008. The vertical dotted and solid lines indicate full moon and new moon, respectively. The seawater temperature was recorded using in situ underwater data loggers (Hobo Pendant[®] Temperature data logger, Onset Computer Corporation, Bourne, MA, USA). We calculated the hourly average temperature first because temperature was measured at different time interval, such as every 15 min or every 30 min, among sites. Hourly average temperature differences were defined as the difference in hourly average temperatures between any given hour and the hour next. We used hourly instead of daily average temperature difference in order to reveal the temperature difference caused by TIU instead of day cycle.

Table 2. Characteristics of the in-situ temperature record at the five sites from 1 January 2007 to 1 January 2008.

Subregions	Site	Depth (m)	Yearly Mean Temperature	Max SST	Min SST	Max. Variability within 2 Hours
West						
Hengchun Peninsula	Wanlitung	7	27.090	31.33 (July)	22.77 (Dec.)	3.15 (July)
West Nawan	Outlet	2	27.580	34.02 (July)	19.91 (Jan.)	4.08 (July)
West Nawan	Outlet	7	26.620	31.36 (Aug.)	18.70 (Jan.)	4.90 (Aug.)
East Nawan	Siangjiaowan	7	26.580	31.09 (July)	17.82 (Jan.)	4.99 (July)
East						
Hengchun Peninsula	Longkeng	7	26.460	30.93 (Sep.)	20.63 (Oct.)	3.13 (June)

This section can be summarized as; (1) There is a presence of strong tidally induced upwelling in Nanwan and (2) Significant temperature drops caused by upwelling could possibly play a major role in shaping coral community structure and long-term resilience in different sub-regions of KNP

2.1. Spatial and Temporal Variability of Coral Communities in Responding to Large-Scale Disturbances

The coral reefs in KNP have been explored since the latter half of Japanese rule, when efforts began to describe coral species and produce a checklist [66–73]. Efforts to describe reef geomorphology and coral communities began in the 1970s, when KNP was established and the third nuclear power plant was constructed [73–80]. These studies focused on describing the reef morphology, coral and fish diversity, and ecology of planktons and fishes. The coral reefs in KNP are separated by sand channels of various widths and vary in their coral fauna among different locations summarised in [79]. There is no significant correlation between species diversity (Shannon-Wiener’s index, H') per 10 m line intercept transect and depth, except at the northern tip of Nanwan, where the species diversity decreases with depth [79,80]. It was concluded that KNP contains two types of coral communities, one mainly composed of scleractinian corals with a few alcyonaceans (less than 6% in total coverage) in the wave-protected areas, and the other dominated by alcyonaceans in wave-exposed areas (total coverage > 50%) [80].

Local tourism significantly increased in the 1990s, and long-term ecological research (LTER) was introduced to monitor the impact of local human disturbances—such as overfishing, habitat destruction, and pollution—on KNP coral reefs [59,60,62]. Observations have shown that variations in large-scale physical disturbances, such as typhoons, in different sub-regions had more of an impact on environmental and biological processes than human disturbance, and resulted in spatial heterogeneity of coral communities in KNP [78]. Analysis showed that typhoons in Taiwan took four types of routes between 1911 and 2018: 26.3% traveled north-westward (Type I), 12.3% traveled northward along the east coast (Type II), 10.26% traveled south-eastward (Type III), and 6.84% traveled northward along the west coast (Type IV) (Figure 4). Interestingly, historic records suggest that Type I typhoons are the major contributor of large-scale disturbances that shape coral communities in KNP (Table 3).

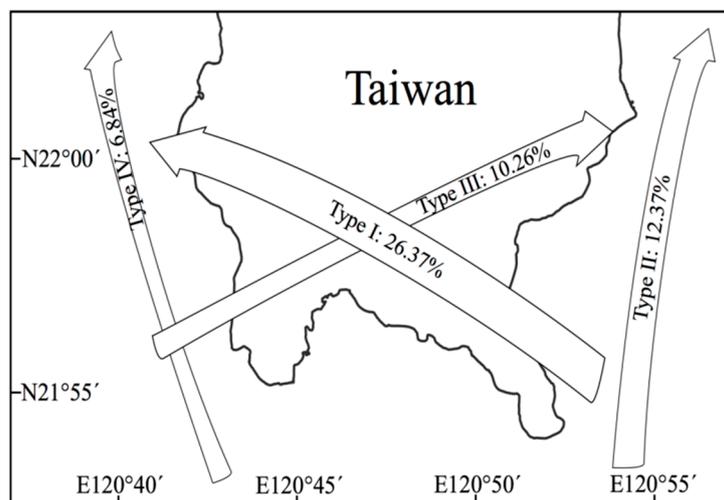


Figure 4. Common trajectories that typhoons took through southern Taiwan between 1911 and 2018. Type I and III indicate a typhoon running north-westwards and north-eastwards, respectively. The other two indicate a typhoon running northwards along the east (type II) and west coasts (type IV). The data were collected from the Typhoon Database, Central Weather Bureau, Taiwan.

Table 3. Summary of the historical disturbances, typhoons, coral bleaching events, and biological outbreaks in Kenting National Park since 1986. Type I and IV of typhoons referred the direction of typhoon used in Figure 4. Typhoon categories are inside the brackets after typhoon names.

Category	Name	Year	Note
Typhoon	Peggy (5)	1986	Type I
Typhoon	Gerald (4)	1987	Type I
Typhoon	Lynn (5)	1987	Type I
Typhoon	Herb (5)	1996	Type I
Typhoon	Chanchu (4)	2006	Type IV
Typhoon	Morakot (2)	2009	Type I
Typhoon	Nanmadol (5)	2011	Type I
Typhoon	Usagi (4)	2013	Type I
Typhoon	Soudelor (5)	2015	Type I
Typhoon	Meranti (5)	2016	Type I
Typhoon	Megi (4)	2016	Type I
Temperature anomaly	Bleaching	1998	Nearly all the colonies on the reefs shallower than 5m in depth in Outlet were bleached.
Temperature anomaly	Bleaching	2002	Minor, very small scale and local bleaching event were recorded in Wanlitung, Houbihu, and Siangjiaowan
Temperature anomaly	Bleaching	2007	50% In Outlet and up to 25% on the West coast of Hengchun Peninsula and Nanwan
Temperature anomaly	Bleaching	2010	Minor scale on the hallow reef of the NPP OL
Temperature anomaly	Bleaching	2014	There were around 30% of the corals bleaching in KNP except in the Outlet that 50% and 20% of the corals were bleached on the shallow (shallower than 5m in depth) and deep (10m in depth) reefs.
Temperature anomaly	Bleaching	2016	Minor scale from Outlet to Nanwan beach
Temperature anomaly	Bleaching	2017	Minor scale on the West coast of Hengchun Peninsula
Ship grounding	Amorgos	2001	Limited on the East coast of Hengchun Penunsula, in particular Longkeng
Ship grounding	Colombo Queen	2009	East coast of Hengchun Peninsula
Ship grounding	WO-BUDMO	2009	West coast of Hengchun Peninsula
Biological outbreak	Sea anemone <i>Condylactis</i> sp.	Late 1996–2008	Limited in the shallow area of Tiaoshi
Biological outbreak	Green alga <i>Codium edule</i>	Late 1996–2002	Limited in the shallow area of Tiaoshi with significantly seasonal variation

LTER data of 5 to 10 m depths showed a declining trend in mean living coral cover (LCC) in KNP, from 48.56% in 1986 to 29.33% in 2018 (Figure 5a), mainly due to the synergistic effects from multiple disturbances (Table 3). Three time-intervals—1986–2000, 2000–2006, 2006–present—shed light on coral community dynamics. First, although no data is available between 1987 and 1997, typhoons in 1986, 1987, and 1996 and the 1998 mass bleaching event were proposed to account for LCC decreasing to 36% by 1999 [81,82], and as a result of local pollution and habitat destruction, was followed by outbreaks of the macroalgae *Codium* spp. [83] and sea anemone *Condylactis* sp. in Taioshi [84] and see references in Table 1 [62,79–82,85–101]. Second, between 2000 and 2006, the LCC returned to a level similar to that of 1986 (> 45%) in 2003, 2004, and 2005 (Figure 5a) due to a lack of major typhoons or bleaching (except a minor one in 2002) between 1999 and 2005 [62,101].

Interestingly, dominant species that were recovered and found to contribute to the increasing LCC during this period were significantly different from those in 1986; for example, in Wanlitung, a LTER site on the west coast of Hengchun Peninsula, the LCC was composed of *Montipora*, *Heliopora*, and *Poritidae* during this period, whereas *Acropora* was the dominant coral genus in 1986 [62,101]. Third, intense disturbances by typhoons (Table 3) caused LCC in KNP to decline in 2006 and stay low until 2016 (Figure 5a). For example, Typhoon Morakot in 2009, the deadliest one in the recorded history (although recorded as category 2), stayed on top of Taiwan for 2 days, causing flooding and big waves that brought the LCC down to 21.07% in 2010 [102,103].

Although the LCC recovered by 2016 to 43.86%, typhoons Meranti and Megi (ranked category 5 and 4, respectively, Table 3), directly hit KNP in September 2016 and combined with minor coral bleaching in 2017, again caused a dramatic decline of LCC down to 29.33% in 2018 [103].

Spatial variability in the response of coral reefs to disturbances was observed among sub-regions in KNP (Figure 5b–e) by analysing the coral cover using the method described in [104]. The coral reef at the west coast of Hengchun Peninsula (Figure 5b) and east coast of Nanwan (Figure 5e) showed a similar trend of LCC dynamics of the overall KNP, but the latter showed no LCC decline between 1986 and 2000. The east coast of Hengchun Peninsula showed an increasing trend at the time interval between 2000 and 2006, declined due to the impact of typhoon Morakot in 2009, and reached its highest LCC of 44.77% in 2012 before declining again between 2012 and 2016 and, finally recovering by 2018 (Figure 5d). The west coast of Nanwan has maintained a higher LCC (> 40%) than other subregions in the last 36 years (Figure 5c) due to the lack of direct impact from typhoons.

Although typhoons have played a notable role of causing the decline in LCC in KNP, it seems they also have benefited the coral community in KNP. The positive effect of typhoons in terms of cooling of the sea surface temperature and breaking down accumulated heat stress in summer by mixing the heated surface water with cooler water from deeper areas [105,106] has resulted in reduced or no coral bleaching. The best example of this in the KNP is NPP Outlet (OL) located at the west coast of Nanwan. This site is most protected from storm surges and at the same time most exposed to the thermal stress caused by the nuclear power plant discharge. In September 2009, while the storm surges of typhoon Morakot did not cause a significant damage on the LCC of the reefs in NPP Outlet (OL) (Figure 5c), compared to the reefs on the east coast of Nanwan (Figure 5e), it did reduce the temperature of the constantly heated sea surface water at the shallow part of the OL by > 6 °C (Figure 6). The cooling effect created an environmental condition of temperature equal to winter and lasted for 3 days after the typhoon passed.

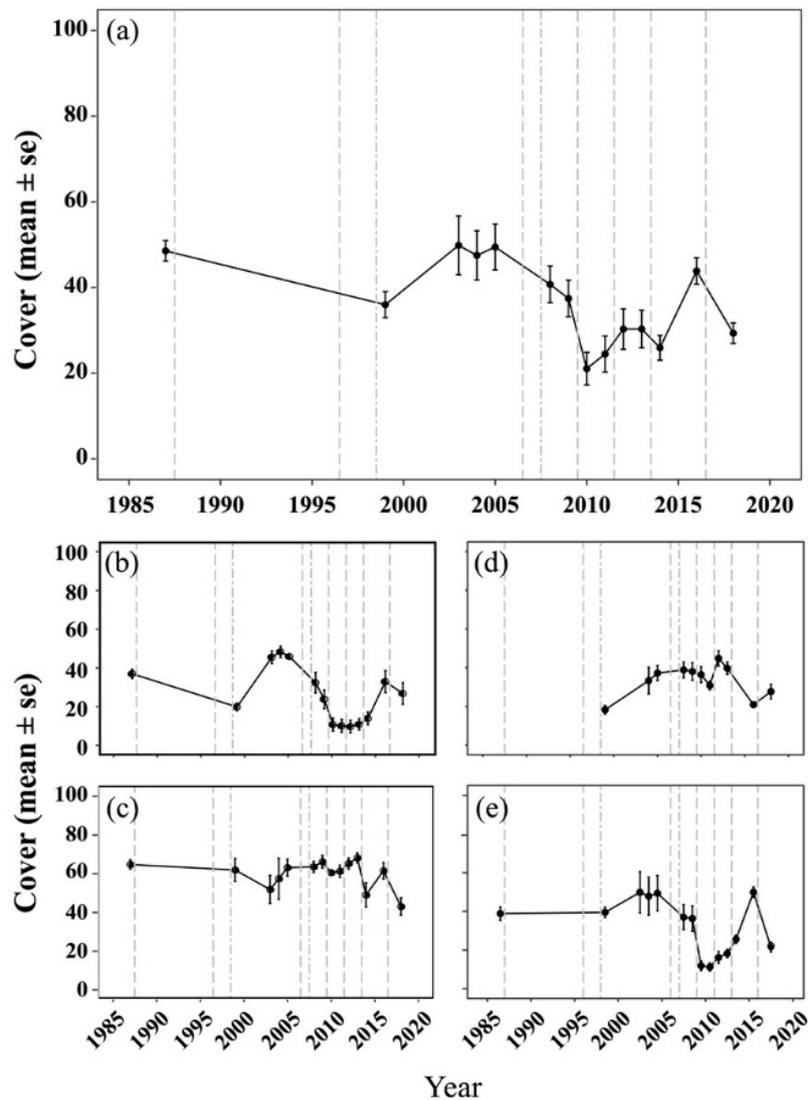


Figure 5. Spatial and temporal long-term trend of average living total coral cover (LCC, per transect \pm standard error) in KNP from 1985 to the present. (a) the average LCC in the Kenting National Park scale, (b) the west, and (d) the east coast of Hengchun Peninsula; (c) the west, and (e) the east coast of Nanwan. The wide and narrow dashed lines indicate typhoons and bleaching events in KNP. The data were collected from the dataset listed in Table 1. The detailed survey methods of the data used in this figure are listed in Table 1. In order to compare the LCC at similar depth, only the transects laid out between 5 and 10m in depth at each site from 1986 to 1999 (Dai, C.-F. 1988; Dai, C.-F. etc. 1998, 1999) were used to generate the figure, combined with the data collected after 2003, including data from a PhD thesis (Dai, C.-F. 1988), one article published in the proceeding of the 6th International Coral Reef Symposium Dai, C.-F. (1988) and two local journal articles written in Mandarin with English abstract (Dai, C.-F. etc. 1998, 1999). The data of LCC after 2003 were collected by the co-first author C.-Y.K. and the corresponding author C.A.C. Different parts of this dataset have been published as a Master’s thesis (Kuo, C.-Y. 2007), in local project reports (Kuo, C.-Y. etc. 2012; Shao, K.-T. etc. 2002; Fang, L.-S. etc. 2003–2006; Wang, W.-H. etc. 2007; Ho, M.-J. etc. 2016; Ho, P.-H. etc. 2008–2011; Chen, C.A. etc. 2012–2014, 2016, 2018; Kuo, C.-Y. etc. 2007 and journal articles (Kuo, C.-Y. etc. 2011, 2012).

The coral communities in the reef adjacent to the nuclear power plant outlet (OL) are protected by storm surges and cooled down by TIU and typhoons, however, the local rising and variable sea surface temperatures [107–110] has resulted in shifting dominant coral species (Figure 7) due to multiple coral bleaching events over time (Table 3). Also at NPP OL, the warm discharged water is trapped in the shallow waters (up to 4 m deep) and flows southwestward in Nanwan because of a near-shore current and tides [111], resulting in a 2.0–3.0 °C higher summer average seawater temperature than at other coral reef sites in Kenting [107,112,113]. Comparing the living coral assemblage in 1986, 1995, 2010, and 2019 in shallow water (3 m) at OL, there was a sharp change in coral genus composition in 1995 (Figure 7a). While *Acropora* dominated in 1986 (31.58% of relative LCC) and 1995 (59.42%), *Galaxea* replaced it and became dominant in 2010 (31.64%) and 2019 (21.11%). *Montipora* remained relatively constant throughout the monitoring period, and *Seriatopora* and *Stylophora* were completely absent at 3 m at OL after 2010 [107]. In addition, coral genus composition remained similar in 1986, 1995, 2010, and 2019 in deep water (7 m), although their relative abundance fluctuated through time (Figure 7b).

However, the response of coral communities to typhoons are varied and has resulted in spatial variation in long term changes of LCC among subregions in KNP (Figure 5). In the nuclear power plant outlet (OL), the reef most protected from typhoons, local, small scale variation of temperature has caused multiple bleaching events and resulted in the loss of temperature sensitive taxa.

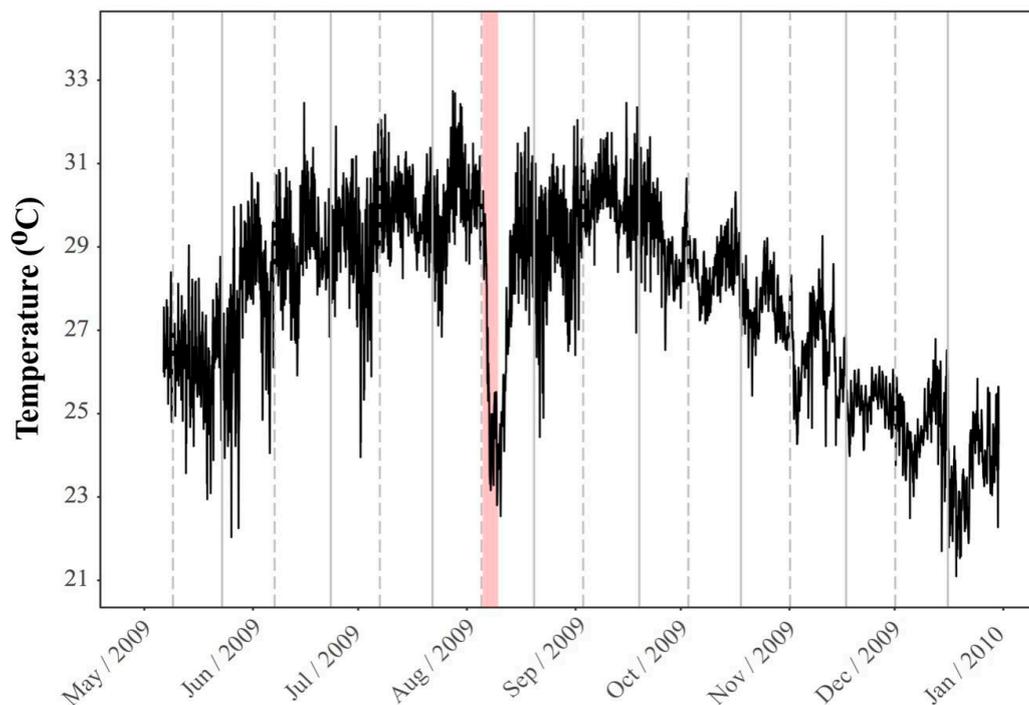


Figure 6. The in situ hourly average seawater temperatures at 2 m depth in Outlet (OL) from 6 June 2009 to 31 December 2009. The vertical dotted and solid lines indicate full moon and new moon, respectively. The pink area indicates the period, from 5th August, 20:30 to 10th August 5:30 2009, the sea warning for typhoon Morakot, the deadliest one in the history of Taiwan, was issued by the Central Weather Bureau.

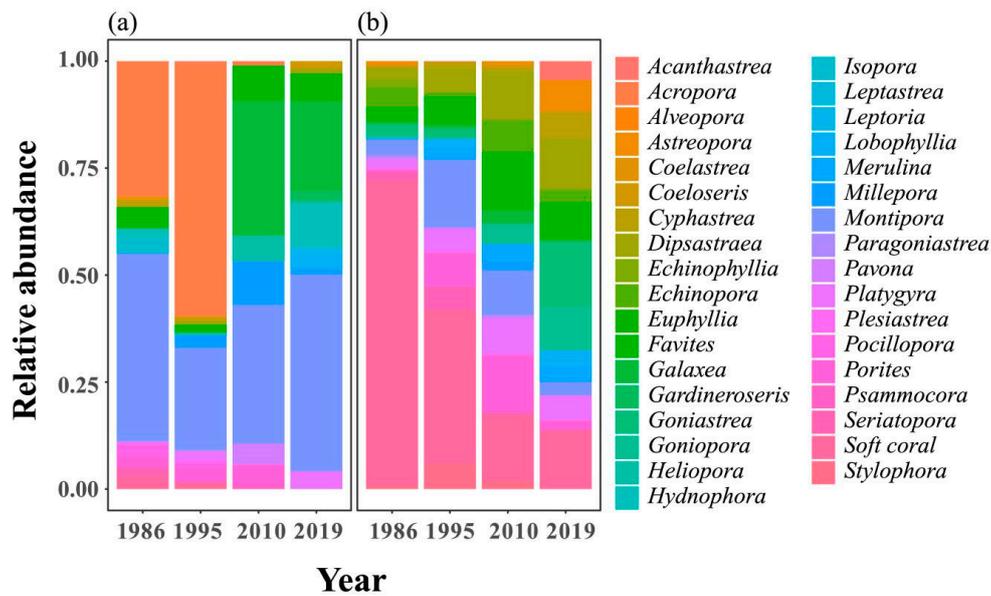


Figure 7. The relative abundance of coral genera at (a) 3 m (shallow) and (b) 7 m (deep) in NPP OL in 1986, 1995, 2010, and 2019. This figure was redrawn using data in (Keshavmurthy, S. etc. 2014) and combined with a survey conducted in 2019.

This section can be summarized as; (1) The effect of typhoons and bleaching has made notable impact on environmental and biological processes resulting in spatial heterogeneity of coral communities and (2) Response of coral communities to typhoons are varied and has resulted in spatial variation in long term changes of LCC among subregions in KNP. In the nuclear power plant outlet (OL), the reef most protected from typhoons, local, small scale variation of temperature has caused multiple bleaching events and resulted in the loss of temperature sensitive taxa.

2.2. Symbiont Community Dynamics over Space and Time in KNP

Coral-associated Symbiodiniaceae from different locations in KNP have been analysed since 1997 [107–110,114–116]. Our studies on Symbiodiniaceae diversity in KNP over the past 24 years demonstrate fine-scale, micro-geographic, temporal, and species-specific associations among genera in addition to/other than coral-associated *Cladocopium* spp. In particular, the ability of corals to associate with multiple Symbiodiniaceae genera and change (i.e., shuffle) between stress-resistant and stress-sensitive genera/species depending on environmental conditions [46,47,117–122] is a critical requirement of their resilience towards stress. At the community level, *Cladocopium* spp. were the dominant Symbiodiniaceae associated with corals in KNP (previously Clade C), as they are elsewhere in the Pacific and South China Sea [123–125]. No matter the molecular technique applied to phylotype Symbiodiniaceae, the results consistently shows that *Cladocopium* sp. is dominant and co-occurs with *Durusdinium* sp. (previously Clade D).

Fine-scale techniques such as ITS2-DGGE, used from 2009 onwards, revealed a fine-scale variation in associated species within each genus; for example, *Cladocopium* C1, C3 and *Durusdinium trenchii* (previously D1a) have all been shown to be dominant. However, recent use of up-to-date NGS amplicon sequencing has revealed more diversity within genera *Cladocopium* and *Durusdinium* (Table 4). In addition, a study published in 2014 [107] showed depth and species-related differences in the coral-associated Symbiodiniaceae in KNP. Samples collected from 16 genera from eight locations and two depths in KNP revealed some interesting trends (Figure 8). *Cladocopium* spp. were more dominant in deeper than shallow water, especially in the corals occurring near OL and in Nanwan.

Table 4. Coral-Symbiodiniaceae associations in Kenting National Park through time analyzed using more sensitive over time. Until 2010 all data are shown by old taxonomy used for Symbiodiniaceae. C = *Cladocopium* spp. (previously clade C), D = *Durusdinium* spp. (previously clade D).

Year	Host Species (Family/Genus)	Symbiodiniaceae Clade/Type/Genera/Species	Study Sites in KNP	Genetic Method for ID	Reference
1997–2001	<i>Acropora</i>	C3, C1, D1, D2,		srDNA- RFLP	Chen, C.A. etc. 2005
	<i>Montipora</i>	C1			
	Pocilloporidae	C1, C2			
	Euphyllidae	C1, D1			
	Poritidae	C1			
	Siderastreidae	C1			
	Agariciidae	C1			
	Oculinidae	C3			
	Merulinidae	C1			
Faviidae	C1				
2000–2001	<i>Isopora palifera</i>	C, D	Tantzei Bay	srDNA- RFLP	Hsu, C.-M. etc. 2012
2006–2009	<i>Isopora palifera</i>	C3, D1a	Tantzei Bay, Maobitou, Siatanzai, NPP-OL, Shiaowan, Siangjiaowan Longkeng	srDNA- RFLP ITS2-DGGE	
2009	<i>Platygyra verweyi</i>	C3, D1a	Leidashih, Siatanzai, Maobitou, NPP-OL, Wanlitung, Hungchai, NPP-IL, Tiaoshi, Tantzei Bay, Longkeng	ITS2-DGGE ITS1-qPCR	Keshavmurthy, S. etc. 2012
2009–2010	<i>Acanthastrea</i>	C1, D1a	Houbihu, NPP-OL	srDNA- RFLP	Keshavmurthy, S. etc. 2014
	<i>Acropora</i>	C21a, C3, D1a	Siangjiaowan, NPP-IL	ITS2-DGGE	
	<i>Cyphastrea</i>	C3, D1a	Wanlitung, Tiaoshi	ITS1-qPCR	
	<i>Favia</i>	C3, D1a	Tantzei Bay, Longkeng		
	<i>Favites</i>	C3, D1a			
	<i>Galaxea</i>	C1, D1a			
	<i>Goniastrea</i>	C1, D1a			
	<i>Isopora</i>	C3, D1a			
	<i>Leptastrea</i>	D1a			
	<i>Leptoria</i>	C1, D1a			
	<i>Montastrea</i>	C1, C3, D1a			
	<i>Montipora</i>	C15, D1a			
	<i>Pavona</i>	C1, D1a			
	<i>Platygyra</i>	C3, D1a			
	<i>Pocillopora</i>	C3, D1a			
	<i>Porites</i>	C15, D1a			
<i>Seriatopora</i>	C1				
<i>Stylophora</i>	C1				

Table 4. Cont.

Year	Host Species (Family/Genus)	Symbiodiniaceae Clade/Type/Genera/Species	Study Sites in KNP	Genetic Method for ID	Reference
2016–2017	<i>Leptoria phrygia</i>	<i>Durusdinium glynii</i>	Wanlitung	ITS2-DGGE	Carballo-Bolaños, R. etc. 2019
		<i>Durusdinium trenchii</i>	NPP-OL	ITS1-qPCR	
		<i>Cladocopium</i> C3w			
		<i>Cladocopium</i> C21a			
		<i>Cladocopium</i> sp.			
2019	<i>Leptoria phrygia</i>	<i>Durusdinium glynii</i>	Wanlitung	ITS2 Amplicon	Huang, Y-Y. etc. 2019
		<i>Durusdinium trenchii</i>	NPP-OL		
		<i>Durusdinium</i> D1.6, D17, D2, D5, D6			
		<i>Cladocopium</i> C116, C15.7, C21a, C2r, C3.1			
		C3.8, C3b, C3d, C3e, C3s, C50			

RFLP = Restriction Fragment Length Polymorphism, DGGE = Denaturing Gradient Gel Electrophoresis, qPCR = Quantitative Real Time Polymerase Chain Reaction.

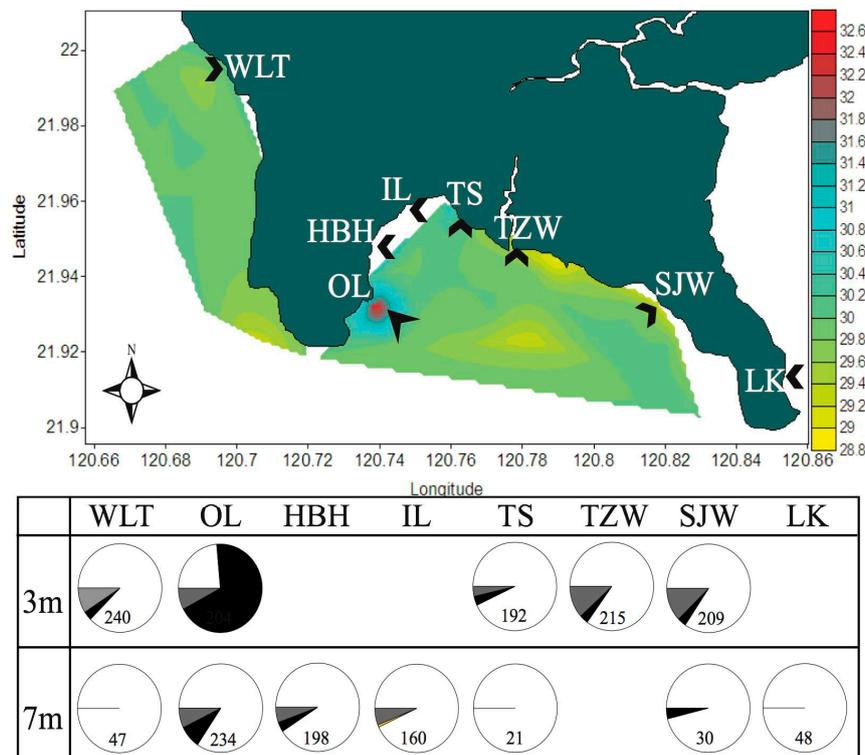


Figure 8. Symbiont diversity associated with corals at different locations—Wanlitung (WLT), NPP outlet (OL), Houbihu (HBH), NPP Inlet (IL), Taioshi (TS), Taziwan (TZW), Siangjiaowan (SJW), and Longkeng (LK)—and from the reef near the nuclear power plant outlet (OL) in Kenting National Park. Samples were collected in 2009 and 2010 from 3 m and 7 m depths. Pie-Charts were drawn using data from previous publications (Keshavmurthy, S. etc. 2014). Analysis of samples collected was done using srDNA-RFLP, ITS2-DGGE and ITS1-qPCR (see Table 4). White = *Cladocopium* sp. (previously clade C), Black = *Durusdinium* sp. (previously clade D), and Grey = co-occurrence of *Cladocopium* sp. and *Durusdinium* sp. The values inside the pie-charts are total sample numbers for each location.

For example, *Isopora palifera* samples in the Taziwan (TZW) population were dominated by *Durusdinium* spp. in 2001 (three years after the 1998 mass coral bleaching event), *Cladocopium* spp. in

2005, co-dominant by two Symbiodiniaceae genera in 2009, and returned to *Durusdinium* dominance in 2015 (Figure 9). This is concordant with previous studies that show that the occurrence of multiple Symbiodiniaceae genera at low concentrations densities might lead to either shuffling or switching to beneficial Symbiodiniaceae genera over time [121,122,126]; in some cases, the coral host may revert back to its original composition of either a single dominant Symbiodiniaceae species or multiple dominant species/genera [108,122]. In addition to temporal changes, our study also indicated a very efficient spatial difference in Symbiodiniaceae associated with *I. palifera*. Samples collected from different locations in KNP showed [108] that corals at a site next to the OL were exclusively associated with or dominated by *D. trenchii/glynnii*, and those at sites away from the OL were associated with *Cladocopium* C3 (Figure 9). Such a micro-geographic difference in Symbiodiniaceae composition is due to the presence of the OL, which has released hot water onto to the coral reefs for over 35 years. Hence, corals near the OL have adapted/acclimatized to associate with *Durusdinium* spp. that are generally stress- as well as temperature-tolerant. Similar spatial differences in the association with Symbiodiniaceae was found in the coral *Platygyra verweyi* and *Leptoria phygia* [108,110]. Nonetheless, there are also cases, irrespective of environmental perturbations, in which the host maintains stable symbiosis with a particular Symbiodiniaceae genus [123,127,128].

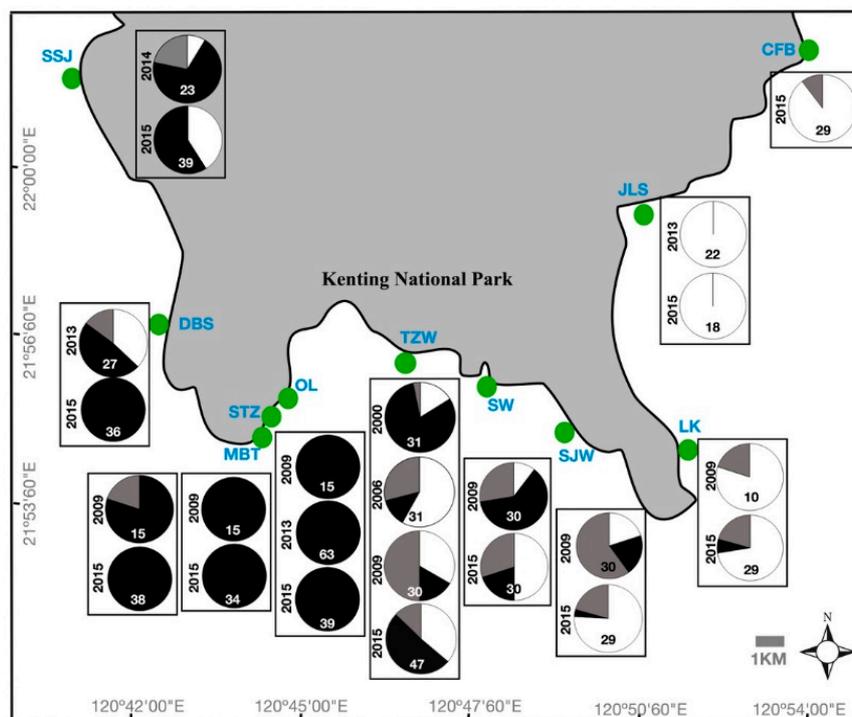


Figure 9. Spatial and temporal variation in *Cladocopium* sp. and *Durusdinium* sp. data in the coral *Isopora palifera* from different locations—Siashuijue (SSJ), Dingbaisha (DBS), Maobitou (MBT), Siatanzai (STZ), nuclear power plant outlet (OL), Taniwan (TZW), Shiaowan (SW), Siangjiaowan (SJW), Longkeng (LK), Jialeshuei (JLS), and Chufengbi (CFB) in Kenting National Park. The data were collected from 2000 to 2015 are adjacent to the pie-charts. Pie-Charts were drawn using data from previous publications (Hsu, C.-M etc. 2012; Fong, W.-L. 2016). Analysis of samples collected in 2000 was done using srDNA-RFLP and all the samples obtained between 2006–2015 were analysed by srDNA-RFLP and ITS2-DGGE (see Table 3) White = *Cladocopium* sp. (previously clade C), Black = *Durusdinium* sp. (previously clade D), and Grey = co-occurrence of *Cladocopium* sp. and *Durusdinium* sp. The values inside the pie-charts are total sample numbers for every location.

This section can be summarized as; (1) Dominant Symbiodiniaceae associated with corals in the KNP is *Cladocopium* sp. However, majority of coral species in the shallows of NPP OL are associated with *Durusdinium* sp. Shuffling of Symbiodiniaceae in corals is almost non-existent.

3. Discussion and Conclusions

There is a serious concern that coral reefs will almost entirely disappear by 2050 if average ocean temperatures increase by 2 °C, with just 10–30% of existing reefs surviving if ocean temperatures increase by 1.5 °C [19]. The “mission-impossible” goal is to drastically reduce CO₂ emissions to net zero and maintain an only 1.5 °C temperature increase; this might lead to 10% of current reefs surviving after 2050. Some global efforts, such as the 50 Reefs Initiative, have used Modern Portfolio Theory (MPT) to identify coral reef locations that represent imperative conservation investments and to ensure their survival; the goal of these efforts is to prepare these areas for repopulation once the climate has been stabilized [129]. If 10%–30% of existing reefs will indeed survive if ocean temperatures increase by only 1.5 °C, an important question is what to do if a reef does not pass the MPT criteria, but does possess certain localized environmental settings, reef topologies, and coral species that do or could resist the impacts of climate change. Herein we argue that the coral reef in KNP could have this great potential to resist the impacts of climate change and deserves novel conservation efforts to ensure its contribution to the coral reef resilience not only in Taiwan but also in the West Pacific.

First, the KNP coral reefs probably receive a constant supply of coral larvae from the south to replenish the reefs after disturbances, and also serve as a “stepping stone” to connect to the reefs or coral communities at high latitudes with currents such as the Kuroshio Current (KC) and the South China Sea Surface Current (SCSSC). Southern Taiwan is bordered by Luzon Island (the Philippines), the north boundary of the “Coral Triangle” (CT). It is argued that KNP has a relatively high scleractinian and reef-associated species diversity because of its connection to the CT [52–54]. Preliminary studies on the genetic connectivity of coral reef fishes in the West Pacific and South China Sea support this scenario [130,131]. Further research on the genetic connectivity of scleractinian corals using high-resolution molecular markers could help elucidate the resilience role of the KNP reef in the region.

Second, many typhoons in the north-western Pacific pass through Taiwan; thus, typhoons might play both negative and positive roles in shaping the coral community structure at different sub-regions in KNP. Of the four types of typhoons recorded (Figure 4), type I (Table 3)—created by a southeast-northwest vortex—contributes most to the temporal dynamics and spatial heterogeneity of coral communities in different sub-regions of KNP (Figure 5). While there is always some mechanical damage, KNP coral reefs can also benefit from typhoons during the warm summer months. As ocean surface waters become warmer during the summer, corals often experience thermal stress. Typhoons can relieve thermal stress by (1) absorbing energy from surface waters through the transfer of latent heat; (2) inducing local upwelling, bringing deeper, cooler water to the surface; and (3) creating clouds of typhoons to shade the ocean surface from solar heating, allowing the water to cool and reducing light stress. Although the projected impacts of climate change on typhoons remain debatable, monitoring the physical damages caused by typhoons and their joint effect with thermal-induced coral bleaching will be crucial for us to develop a management plan to improve coral reef resilience in KNP.

Third, upwelling has been proposed to be a cooling mechanism to protect coral reefs against bleaching by reducing seawater temperatures or creating fluctuating thermal environments that induce corals to build thermal resistance over time [132–135]. However, it has also been suggested that upwelling areas do not always guarantee refuge for coral reefs in a warming ocean unless the thermal threat and upwelling coincide [136]. Some large-scale seasonal upwelling with cold, nutrient-rich, and naturally acidic subsurface water—such as the upwelling in Gulf of Panama and Papagayo upwelling of Costa Rica, in the tropical eastern Pacific—indeed hinders the development of coral reefs [137–139]. In KNP, upwelling is small-scale, localized, and induced by tides flowing from east to west Nanwan that create a temperature difference within the bay of Nanwan and two sides of Hengchun Peninsula, where the water is cooler in the east and warmer in the west (Table 2).

In addition, the upwelling helps reduce the thermal stress, particularly in the reef adjacent to the nuclear power plant (OL), by creating significant temperature drops during spring tides in the summer (Table 2) and fluctuating thermal environments that induce corals to build thermal resistance (Figures 8 and 9). These positive effects ensure that KNP remains a refuge for coral reefs to survive in a warming

ocean. Monitoring whether the tide-induced upwelling will be enhanced or hindered by the rising background seawater temperature in the region should be considered as a research priority in KNP.

Fourth, Symbiodiniaceae play a crucial role in bleaching tolerance. Many species or genera of Symbiodiniaceae have been identified [44], and different genera display varying thermal, and therefore bleaching, resistances. It has been suggested that, by associating with or shuffling the symbiont community towards making thermal-tolerant Symbiodiniaceae, such as *Durusdinium* spp, dominant, corals can increase their thermal tolerance by 1.0 °C–1.5 °C [118,119]. Our long-term monitoring of symbiont community diversity shows that corals constantly exposed to warming and fluctuating thermal environments (OL) or constantly higher seawater temperatures (west coast of Hengchun Peninsula) have a dominance of *Durusdinium* spp, whereas the same species located on the cooler east coast of Hengchun Peninsula are increasingly associated with *Cladocopium* spp. (Figure 8). This subregional difference in symbiont community is concordant with the influence of TIU that pumps cooler water from Eluanbi (east cape), protruding towards Maobitou (east cape), creating two temperature drops per tidal cycle in western and central Nanwan and one drop in the eastern part, but no having impact on the west coast of the peninsula. The TIU affects the seawater temperature in KNP and not only drives sub-regional variability in symbiont communities, but also provides the signal for corals to shuffle their symbionts in response to seasonally fluctuating seawater temperature; this is not, however, true for corals in shallow water (< 3 m) of the reef adjacent to the nuclear power plant OL, which are associated dominantly with *Durusdinium* spp. and do not show sign of shuffling [107–109]. In reciprocal transplantation experiments (RTE), corals from WLT to OL did not survive under a prolonged seawater temperature anomaly, even though they showed signs of shuffling from *Cladocopium* to *Durusdinium* dominance [22]. These results imply that corals in the shallow water of the OL reef already live at the ceiling of thermal tolerance, and future climate change trends might be untenable for those corals [110].

Despite these environmental, ecological, and biological characteristics, adaptive management strategies such as implementing sewage treatment systems, banning the serving of herbivorous fishes in restaurants, and promoting eco-friendly tourism and public awareness in recent years has aided in the resilience of coral reefs in KNP. Present and future adaptive management in accordance with the framework of resilience-based management [39] might help sustain coral reef resilience in Kenting National under the impacts of climate change.

This section can be summarized as; Coral reefs in a small geographical range with unique environmental settings and ecological characteristics, such as the KNP reef, are resilient to bleaching and deserve novel conservation efforts. Thus, conservation efforts that use resilience-based management programs to reduce local stresses and meet the challenge of climate change is urgently needed.

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References

1. Atweberhan, M.; Feary, D.A.; Keshavmurthy, S.; Chen, C.A.; Schleyer, M.H.; Sheppard, R.C. Climate change impacts on coral reefs: Synergies with local effects, possibilities for acclimation, and management implications. *Mar. Pollut. Bull.* **2013**, *74*, 526–539. [[CrossRef](#)] [[PubMed](#)]
2. Hughes, T.P.; Baird, A.H.; Bellwood, D.R.; Card, M.; Connolly, S.R.; Folke, C.; Grosberg, R.; Hoegh-Guldberg, O.; Jackson, J.B.; Kleypas, J.; et al. Climate Change, Human Impacts, and the Resilience of Coral Reefs. *Science* **2013**, *301*, 929–933. [[CrossRef](#)] [[PubMed](#)]
3. Hoegh-Guldberg, O.; Mumby, P.J.; Hooten, A.J.; Steneck, R.S.; Greenfield, P.; Gomez, E.; Harvell, C.D.; Sale, P.F.; Edwards, A.J.; Caldeira, K.; et al. Coral reefs under rapid climate change and ocean acidification. *Science* **2007**, *318*, 1737–1742. [[CrossRef](#)] [[PubMed](#)]
4. National Academies of Sciences, Engineering, and Medicine. *A Research Review of Interventions to Increase the Persistence and Resilience of Coral Reefs*; The National Academies Press: Washington, DC, USA, 2018; 258p.
5. Brown, B.E. Coral bleaching: Causes and consequences. *Coral Reefs* **1997**, *16*, S129–S138. [[CrossRef](#)]
6. Hoegh-Guldberg, O. Climate change, coral bleaching and the future of the world's coral reefs. *Mar. Freshw. Res.* **1999**, *50*, 839–866. [[CrossRef](#)]
7. Hughes, T.P.; Kerry, J.T.; Álvarez-Noriega, M.; Álvarez-Romero, J.G.; Anderson, K.D.; Baird, A.H.; Babcock, R.C.; Beger, M.; Bellwood, D.R.; Berkelmans, R.; et al. Global warming and recurrent mass bleaching of corals. *Nature* **2017**, *543*, 373–377. [[CrossRef](#)]
8. Hughes, T.P.; Anderson, K.D.; Connolly, S.R.; Heron, S.F.; Kerry, J.T.; Lough, J.M.; Baird, A.H.; Baum, J.K.; Berumen, M.L.; Bridge, T.C.; et al. Spatial and temporal patterns of mass bleaching of corals in the Anthropocene. *Science* **2018**, *359*, 80–83. [[CrossRef](#)]
9. Gilmour, J.P.; Cook, K.L.; Ryan, N.M.; Puotinen, M.L.; Green, R.H.; Shedrawi, G.; Hobbs, J.-P.A.; Thomson, D.P.; Babcock, R.C.; Buckee, J.; et al. The state of Western Australia's coral Reefs. *Coral Reefs* **2019**, *38*, 651–667. [[CrossRef](#)]
10. Goyen, S.; Camp, E.F.; Fujise, L.; Lloyd, A.; Nitschke, M.R.; LaJeunesse, T.; Kahlke, T.; Ralph, P.J.; Suggett, D. Mass coral bleaching of *P. verispora* in Sydney Harbour driven by the 2015–2016 heatwave. *Coral Reefs* **2019**, *38*, 815–830. [[CrossRef](#)]
11. Head, C.E.I.; Bayley, D.T.I.; Rowlands, G.; Roche, R.C.; Tickler, D.M.; Rogers, A.D.; Koldewey, H.; Turner, J.R.; Andradi-Brown, D.A. Coral bleaching impacts from back-to-back 2015–2016 thermal anomalies in the remote central Indian Ocean. *Coral Reefs* **2019**, *38*, 605–618. [[CrossRef](#)]
12. Montefalcone, M.; Morri, C.; Bianchi, C.N. Long-term change in bioconstruction potential of Maldivian coral reefs following extreme climatic anomalies. *Glob. Chang. Biol.* **2018**, *24*, 5629–5641. [[CrossRef](#)] [[PubMed](#)]
13. Fox, M.D.; Carter, A.L.; Edwards, C.B.; Takeshita, Y.; Johnson, M.D.; Petrovic, V.; Amir, C.G.; Sala, E.; Sandin, S.A.; Smith, J.E. Limited coral mortality following acute thermal stress and widespread bleaching on Palmyra Atoll, central Pacific. *Coral Reefs* **2019**, *38*, 701–712. [[CrossRef](#)]
14. Raymundo, L.J.; Burdick, D.; Hoot, W.C.; Miller, R.M.; Brown, V.; Reynolds, T.; Gault, J.; Idechong, J.; Fifer, J.; Williams, A. Successive bleaching events cause mass coral mortality in Guam, Micronesia. *Coral Reefs* **2019**, *38*, 677–700. [[CrossRef](#)]
15. Vargas-Ángel, B.; Huntington, B.; Brainard, R.E.; Venegas, R.; Oliver, T.; Barkley, H.; Cohen, A. El Niño-associated catastrophic coral mortality at Jarvis Island, central Equatorial Pacific. *Coral Reefs* **2019**, *38*, 731–741. [[CrossRef](#)]
16. Smith, K.M.; Payton, T.G.; Sims, R.J.; Stroud, C.S.; Jeanes, R.C.; Hyatt, T.B.; Childress, M.J. Impacts of consecutive bleaching events and local algal abundance on transplanted coral colonies in the Florida Keys. *Coral Reefs* **2019**, *38*, 851–861. [[CrossRef](#)]
17. Teixeira, C.D.; Leitão, R.L.L.; Ribeiro, F.V.; Moraes, F.C.; Neves, L.M.; Bastos, A.C.; Pereira-Filho, G.H.; Kampel, M.; Salomon, P.S.; Sá, J.A.; et al. Sustained mass coral bleaching (2016–2017) in Brazilian turbid-zone reefs: Taxonomic, cross-shelf and habitat-related trends. *Coral Reefs* **2019**, *38*, 801–813. [[CrossRef](#)]
18. Eakin, M.C.; Sweatman, H.P.A.; Brainard, R.E. The 2014–2017 global-scale coral bleaching event: Insights and impacts. *Coral Reefs* **2019**, *38*, 539–545. [[CrossRef](#)]

19. Masson-Delmotte, V.; Zhai, P.; Pörtner, H.O.; Roberts, D.; Skea, J.; Shukla, P.R.; Pirani, A.; Moufouma-Okia, W.; Péan, C.; Pidcock, R.; et al. (Eds.) IPCC, 2018: Summary for Policymakers. In *Global Warming of 1.5 °C. An IPCC Special Report on the Impacts of Global Warming of 1.5 °C above Pre-Industrial Levels and Related Global Greenhouse Gas Emission Pathways, in the Context of Strengthening the Global Response to the Threat of Climate Change, Sustainable Development, and Efforts to Eradicate Poverty*; World Meteorological Organization: Geneva, Switzerland, 2018; 32p.
20. Riegl, B.M.; Purkis, S.J.; Al-Cibahy, A.S.; Abdel-Moati, M.A.; Hoegh-Guldberg, O. Present limits to heat-adaptability in corals and population-level responses to climate extremes. *PLoS ONE* **2011**, *6*, e24802. [[CrossRef](#)]
21. Pandolfi, J.M.; Connolly, S.R.; Marshall, D.J.; Cohen, A.L. Projecting Coral Reef Futures Under Global Warming and Ocean Acidification. *Science* **2011**, *333*, 418–422. [[CrossRef](#)]
22. Kao, K.-W.; Keshavmurthy, S.; Tsao, C.-H.; Wang, J.-T.; Chen, C.A. Repeated and Prolonged Temperature Anomalies Negate Symbiodiniaceae Genera Shuffling in the Coral *Platygyra verweyi* (Scleractinia; Merulinidae). *Zool. Stud.* **2018**, *57*. [[CrossRef](#)]
23. Emanuel, K.A. Downscaling CMIP5 climate models shows increased tropical cyclone activity over the 21st century. *Proc. Natl. Acad. Sci. USA* **2013**, *110*, 12219–12224. [[CrossRef](#)] [[PubMed](#)]
24. Tsuboki, K.; Yoshioka, M.K.; Shinoda, T.; Kato, M.; Kanada, S.; Kitoh, A. Future increase of super typhoon intensity acclimated with climate change. *Geophys. Res. Lett.* **2015**, *42*, 646–652. [[CrossRef](#)]
25. Knutson, T.R.; McBride, J.L.; Chan, J.; Emanuel, K.; Holland, G.; Landsea, C.; Held, I.; Kossin, J.P.; Srivastava, A.K.; Sugi, M. Tropical cyclones and climate change. *Nat. Geosci.* **2010**, *3*, 157–163. [[CrossRef](#)]
26. Christensen, J.H.; Kanikicharla, K.K.; Marshall, G.; Turner, J. Climate phenomena and their relevance for future regional climate change. In *Climate Change 2013: The Physical Science Basis*; Stocker, T.F., Qin, D., Plattner, G.-K., Tignor, M., Allen, S.K., Boschung, J., Nauels, A., Xia, Y., Bex, V., Midgley, P.M., Eds.; Cambridge University Press: Cambridge, UK, 2013; pp. 1217–1308.
27. Walsh, K.J.; McBride, J.L.; Klotzbach, P.J.; Balachandran, S.; Camargo, S.J.; Holland, G.; Knutson, T.R.; Kossin, J.P.; Lee, T.; Sobel, A.; et al. Tropical cyclones and climate change. *WIREs Clim. Chang.* **2016**, *7*, 65–89. [[CrossRef](#)]
28. Dollar, S.J. Wave stress and coral community structure in Hawaii. *Coral Reefs* **1982**, *1*, 71–81. [[CrossRef](#)]
29. Cheal, A.J.; MacNeil, M.A.; Emslie, M.J.; Sweatman, H. The threat to coral reefs from more intense cyclones under climate change. *Glob. Chang. Biol.* **2017**, *23*, 1511–1524. [[CrossRef](#)]
30. Holling, C.S. Resilience and stability of ecological systems. *Annl. Rev. Ecol. Syst.* **1973**, *4*, 1–23. [[CrossRef](#)]
31. Nyström, M.; Folke, C. Spatial resilience of coral reefs. *Ecosystems* **2001**, *4*, 406–417. [[CrossRef](#)]
32. Nystrom, M. Redundancy and response diversity of functional groups: Implications for the resilience of coral reefs. *AMBIO* **2006**, *35*, 30–35. [[CrossRef](#)]
33. Norström, A.; Nyström, M.; Lokrantz, J.; Folke, C. Alternative states of coral reefs: Beyond coral-macroalgal phase shifts. *Mar. Ecol. Prog. Ser.* **2009**, *376*, 295–306. [[CrossRef](#)]
34. Nyström, M.; Norström, A.; Blenckner, T.; de la Torre-Castro, M.; Eklöf, J.S.; Folke, C.; Österblom, H.; Steneck, R.S.; Thyresson, M.; Troell, M. Confronting Feedbacks of Degraded Marine Ecosystems. *Ecosystems* **2012**, *15*, 695–710. [[CrossRef](#)]
35. Hughes, T.P.; Graham, N.A.J.; Jackson, J.B.C.; Mumby, P.J.; Steneck, R.S. Rising to the challenge of sustaining coral reef resilience. *Trends Ecol. Evolut.* **2010**, *25*, 633–642. [[CrossRef](#)] [[PubMed](#)]
36. Montefalcone, M.; Parravicini, V.; Bianchi, C.N. Quantification of coastal ecosystem resilience. In *Treatise on Estuarine and Coastal Science*; Elsevier: Amsterdam, The Netherlands, 2011; Volume 10, pp. 49–70.
37. Folke, C. Resilience: The emergence of a perspective for social–ecological systems analyses. *Glob. Environ. Chang.* **2006**, *16*, 253–267. [[CrossRef](#)]
38. Nystrom, M.; Graham, A.J.; Lokrantz, J.; Norstorm, A.V. Capturing the cornerstones of coral reef resilience: Linking theory to practice. *Coral Reefs* **2008**, *27*, 795–809. [[CrossRef](#)]
39. Anthony, K.R.N. Coral Reefs Under Climate Change and Ocean Acidification: Challenges and Opportunities for Management and Policy. *Annu. Rev. Environ. Res.* **2016**, *41*, 59–81. [[CrossRef](#)]

40. Barneah, O.; Brickner, I. Multibiont Symbioses in the Coral Reef Ecosystem. In *Symbioses and Stress. Cellular Origin, Life in Extreme Habitats and Astrobiology*; Seckbach, J., Grube, M., Eds.; Springer: Dordrecht, The Netherlands, 2010; Volume 17.
41. Baker, A.C. Flexibility and specificity in coral-algal symbiosis: Diversity, ecology, and biogeography of *Symbiodinium*. *Annu. Rev. Ecol. Evol. Syst.* **2003**, *34*, 661–689. [[CrossRef](#)]
42. Lajeunesse, T.C.; Pettay, D.T.; Sampayo, E.E.; Phongsuwan, N.; Brown, B.E.; Obura, D.O.; Hoegh-Guldberg, O.; Fitt, W.K. Long-standing environmental conditions, geographic isolation and host-symbiont specificity influence the relative ecological dominance and genetic diversification of coral endosymbionts in the genus *Symbiodinium*. *J. Biogeogr.* **2010**, *37*, 785–800. [[CrossRef](#)]
43. Silverstein, R.N.; Correa, A.M.S.; Baker, A.C. Specificity is rarely absolute in coral-algal symbiosis: Implications for coral response to climate change. *Proc. R. Soc. B Biol. Sci.* **2012**, *279*, 2609–2618. [[CrossRef](#)]
44. Lajeunesse, T.C.; Parkinson, J.E.; Gabrielson, P.W.; Jeong, H.J.; Reimer, J.D.; Voolstra, C.R.; Santos, S.R. Systematic Revision of Symbiodiniaceae Highlights the Antiquity and Diversity of Coral Endosymbionts. *Curr. Biol.* **2018**, *28*, 2570–2580. [[CrossRef](#)]
45. Weber, M.; Medina, M. The role of microalgal symbionts (*Symbiodinium*) in holobiont physiology. *Adv. Bot. Res.* **2012**, *64*, 119–140.
46. Sampayo, E.E.; Ridgway, T.; Bongaerts, P.; Hoegh-Guldberg, O. Bleaching susceptibility and mortality of corals are determined by fine-scale differences in symbiont type. *Proc. Natl. Acad. Sci. USA* **2008**, *105*, 10444–10449. [[CrossRef](#)]
47. Jones, A.M.; Berkelmans, R.; Van Oppen, M.J.H.; Mieog, J.C.; Sinclair, W. A community change in the algal endosymbionts of a scleractinian coral following a natural bleaching event: Field evidence of acclimatization. *Proc. R. Soc. B Biol. Sci.* **2008**, *275*, 1359–1365. [[CrossRef](#)] [[PubMed](#)]
48. Ulstrup, K.E.; Van Oppen, M. Geographic and habitat partitioning of genetically distinct zooxanthellae (*Symbiodinium*) in *Acropora* corals on the Great Barrier Reef. *Mol. Ecol.* **2003**, *12*, 3477–3484. [[CrossRef](#)]
49. Lien, Y.-T.; Nakano, Y.; Plathong, S.; Fukami, H.; Wang, J.-T.; Chen, C.A. Occurrence of the putatively heat-tolerant *Symbiodinium* phylotype D in high-latitude outlying coral communities. *Coral Reefs* **2007**, *26*, 35–44. [[CrossRef](#)]
50. Ghavam Mostafavi, P.; Fatemi, S.M.R.; Shahhosseiny, M.H.; Hoegh-Guldberg, O.; Loh, W.K.W. Predominance of clade D *Symbiodinium* in shallow-water reef-building corals off Kish and Larak Islands (Persian Gulf, Iran). *Mar. Biol.* **2007**, *153*, 25–34. [[CrossRef](#)]
51. Oliver, T.A.; Palumbi, S.R. Many corals host thermally resistant symbionts in high-temperature habitat. *Coral Reefs* **2011**, *30*, 241–250. [[CrossRef](#)]
52. Chen, C.A. Analysis of Scleractinian Distribution in Taiwan Indicating a Pattern Congruent with Sea Surface Temperatures and Currents: Examples from *Acropora* and Faviidae Corals. *Zool. Stud.* **1999**, *38*, 119–129.
53. Chen, C.A.; Keshavmurthy, S. Taiwan as a connective stepping-stone in the Kuroshio Triangle and the conservation of coral ecosystems under the impacts of climate change. *Kuroshio Sci.* **2009**, *3*, 15–22.
54. Dai, C.F. Coastal and shallower water ecosystem. In *Chapter 7. Regional Oceanography of Taiwan Version II*; Jan, S., Ed.; National Taiwan University Press: Taipei, Taiwan, 2018; pp. 263–299.
55. Dai, C.F. Assessment of the present health of coral reefs in Taiwan. In *Status of Coral Reefs in the Pacific*; UNIH Sea Grant CP-98-01; Grigg, R.W., Birkeland, C., Eds.; Sea Grant Program, University of Hawaii: Honolulu, HI, USA, 1997; pp. 123–131.
56. Bruno, J.F.; Selig, E.R. Regional Decline of Coral Cover in the Indo-Pacific: Timing, Extent, and Subregional Comparisons. *PLoS ONE* **2007**, *2*, e711. [[CrossRef](#)]
57. Jackson, J.B.C.; Donovan, M.K.; Cramer, K.L.; Lam, V.V. (Eds.) Status and Trends of Caribbean Coral Reefs: 1970–2012. In *Global Coral Reef Monitoring Network*; IUCN: Gland, Switzerland, 2014; 298p.
58. Meng, P.-J.; Chen, J.-P.; Chung, J.-N.; Liu, M.-C.; Fan, T.-Y.; Chang, C.-M.; Tian, W.-M.; Chang, Y.-C.; Lin, H.-J.; Fang, L.-S.; et al. Long-term ecological studies in Kenting National Park neighboring marine areas, on monitoring the impact factors from anthropogenic activities to the marine ecosystem and a preliminary database of its marine ecosystem. *J. Natl. Park* **2004**, *14*, 43–69, (In Chinese with English abstract).
59. Meng, P.-J.; Lee, H.-J.; Wang, J.-T.; Chen, C.-C.; Lin, H.-J.; Tew, K.S.; Hsieh, W.-J. A long-term survey on anthropogenic impacts to the water quality of coral reefs, southern Taiwan. *Environ. Pollut.* **2008**, *156*, 67–75. [[CrossRef](#)] [[PubMed](#)]

60. Liu, P.-J.; Meng, P.-J.; Liu, L.-L.; Wang, J.-T.; Leu, M.-Y. Impacts of human activities on coral reef ecosystems of southern Taiwan: A long-term study. *Mar. Pollut. Bull.* **2012**, *64*, 1129–1135. [[CrossRef](#)] [[PubMed](#)]
61. Chen, C.A. Status of Coral Reefs in Taiwan. In *Status of Coral Reefs in East Asian Seas Region*; Kimura, T., Tun, K., Chou, L.M., Eds.; Ministry of Environment: Tokyo, Japan, 2014; pp. 69–78.
62. Kuo, C.-Y.; Yuen, Y.-S.; Meng, P.-J.; Ho, P.-H.; Wang, J.T.; Liu, P.J.; Chang, Y.C.; Dai, C.F.; Fan, T.Y.; Lin, H.J.; et al. Recurrent disturbances and the degradation of hard coral communities in Taiwan. *PLoS ONE* **2012**, *7*, e44364. [[CrossRef](#)] [[PubMed](#)]
63. Ribas-Deulofeu, L.; Denis, V.; De Palmas, S.; Kuo, C.-Y.; Hsieh, H.J.; Chen, C.A. Structure of Benthic Communities along the Taiwan Latitudinal Gradient. *PLoS ONE* **2016**, *11*, e0160601. [[CrossRef](#)] [[PubMed](#)]
64. Lee, H.-J.; Chao, S.-Y.; Fan, K.-L.; Kuo, T.-Y. Tide-Induced Eddies and Upwelling in a Semi-enclosed Basin: Nan Wan. *Estuar. Coast. Shelf Sci.* **1999**, *49*, 775–787. [[CrossRef](#)]
65. Lee, H.-J.; Chao, S.-Y.; Fan, K.-L. Flood-ebb disparity of tidally induced recirculation Eddies in a semi-enclosed basin: Nan Wan Bay. *Cont. Shelf Res.* **1999**, *19*, 871–890. [[CrossRef](#)]
66. Sugiyama, T. Reef building corals of Japanese coast. *Pap. Inst. Geol. Paleontol. Tohoku Imp. Univ.* **1937**, *26*, 1–60.
67. Kawaguti, S. Coral fauna of the Taiwan waters. *Kagaku No Taiwan* **1942**, *11*, 1–6.
68. Kawaguti, S. Coral fauna of Garampi. *Trans. Formos. Nat. Hist. Soc.* **1943**, *33*, 258–259.
69. Kawaquti, S. Coral fauna of the island of Botel Tobago, Formosa with a list of corals from the Formosan waters. *Biol. J. Okoyama Univ.* **1953**, *1*, 185–198.
70. Ma, T.Y.H. The effect of warm and cold currents in the southern-western Pacific on the growth rate of reef corals. *Oceanogr. Sin.* **1957**, *5*, 1–34.
71. MA, T.Y.H. The relation of growth rate of reef corals to surface temperature of sea water as basis for study of causes of diastrophisms instigating evolution of life. In *Research on the Past Climate and Continental Drift*; World Book Co.: Taipei, Taiwan, 1958; Volume 14.
72. Ma, T.Y.H. Effect of water temperature on growth rate of corals. *Oceanogr. Sin. Specif.* **1959**, *1*, 116.
73. Jones, O.A.; Randall, R.H.; Cheng, Y.M.; Kami, H.T.; Mak, S.M. *A Marine Biological Survey of Southern Taiwan with Emphasis on Corals and Fishes*; Institute of Oceanography, National Taiwan University: Taipei, Taiwan, 1972; 93p.
74. Yang, R.T.; Huang, C.C.; Wang, C.H.; Yeh, S.Z.; Jan, Y.F.; Liu, S.L.; Chen, C.H.; Chen, S.J.; Chang, L.F.; Sun, C.L.; et al. *A Marine Biological Data Acquisition Program Pertaining to the Construction of a Power Plant in the Nan-Wan Bay Area. Phase, I. A Preliminary Reconnaissance Survey*; Institute of Oceanography, National Taiwan University: Taipei, Taiwan, 1976; 134p.
75. Yang, R.T.; Huang, C.C.; Wang, C.H.; Yeh, S.Z.; Jan, Y.F.; Liu, S.L.; Chen, C.H.; Chen, S.J.; Chang, L.F.; Sun, C.L.; et al. *A Marine Biological Data Acquisition Program Pertaining to the Construction of a Power Plant in the Nan-Wan Bay Area. Phasw 11, Biological Data Acquisition*; Institute of Oceanography, National Taiwan University: Taipei, Taiwan, 1977; pp. 1–34.
76. Randall, R.H.; Cheng, Y.M. Recent corals of Taiwan. Part, I. Description of reefs and coral environment. *Acta Geol. Taiwanica* **1977**, *19*, 79–102.
77. Randall, R.H.; Cheng, Y.M. Recent corals of Taiwan. Part II. Description of reefs and coral environment. *Acta Geol. Taiwanica* **1979**, *20*, 1–32.
78. Yang, R.T. Coral communities in Nan-wan Bay (Taiwan). In *Proceedings of the Fifth International Coral Reef Congress, Tahiti, French Polynesia, 27 May–1 June 1985*; Volume 6, pp. 273–278.
79. Dai, C.-F. Community Ecology of Corals on the Fringing Reefs of Southern Taiwan. Ph.D. Thesis, Yale University, New Haven, CT, USA, 1988; p. 315.
80. Dai, C.-F. Coral communities of southern Taiwan. In *Proceedings of the 6th International Coral Reef Symposium, Townsville, Australia, 8–12 August 1988*; Volume 2, pp. 647–652.
81. Dai, C.-F.; Chen, Y.-T.; Kuo, K.-M.; Chuang, C.-H. Changes of coral communities in Nanwan Bay, Kenting National Park: 1987–1997. *J. Natl. Park* **1998**, *8*, 79–99, (In Chinese with English abstract).
82. Dai, C.-F.; Kuo, K.-M.; Chen, Y.-T.; Chuang, C.-H. Changes of coral communities on the east and west coast of the Kenting National Park. *J. Natl. Park* **1999**, *9*, 112–130, (In Chinese with English abstract).
83. Chang, J.-S.; Dai, C.-F.; Chang, J. The gametangium-like structure as propagation buds in *Codium edule* Silva (Bryopsidales: Chlorophyta). *Bot. Mar.* **2003**, *46*, 431–437. [[CrossRef](#)]

84. Chen, C.A.; Dai, C.-F. Local phase shift from *Acropora*-dominant to *Condylactis*-dominant community in the Tiao-Shi Reef, Kenting National Park, southern Taiwan. *Coral Reefs* **2004**, *23*, 508. [[CrossRef](#)]
85. Shao, K.-T.; Fang, L.-S.; Liang, N.-K.; Meng, P.-J.; Chung, K.-N.; Li, J.-J.; Han, C.-C. *Long-Term Ecological Research (LTER) Program of the Coral Reefs in Kenting National Park, Taiwan*; Kenting National Park Headquarters, Construction and Planning Agency, Minister of the Interior: Taipei, Taiwan, 2002; p. 176. (In Chinese)
86. Fang, L.-S.; Shao, K.-T.; Tian, W.-M.; Chang, Y.-C.; Lin, H.-J.; Fan, T.-Y.; Chung, K.-N.; Chen, J.-P.; Liu, M.-Y.; Meng, P.-J. *Long-Term Ecological Research (LTER) Program of the Coral Reefs in Kenting National Park, Taiwan*; Kenting National Park Headquarters, Construction and Planning Agency, Minister of the Interior: Taipei, Taiwan, 2003; p. 249. (In Chinese)
87. Fang, L.-S.; Shao, K.-T.; Tian, W.-M.; Chang, Y.-C.; Lin, H.-J.; Fan, T.-Y.; Chung, K.-N.; Chen, J.-P.; Chen, M.-H.; Liu, M.-Y.; et al. *Long-Term Ecological Research (LTER) Program of the Coral Reefs in Kenting National Park, Taiwan*; Kenting National Park Headquarters, Construction and Planning Agency, Minister of the Interior: Taipei, Taiwan, 2004; p. 262. (In Chinese)
88. Fang, L.-S.; Shao, K.-T.; Meng, P.-J.; Chung, K.-N.; Chen, J.-P.; Chen, M.-H.; Liu, M.-Y.; Chang, Y.-C.; Fan, T.-Y.; Lin, H.-J. *Long-Term Ecological Research (LTER) Program of the Coral Reefs in the Kenting National Park*; Kenting National Park Headquarters, Construction and Planning Agency, Minister of the Interior: Taipei, Taiwan, 2005; p. 295. (In Chinese)
89. Fang, L.-S.; Shao, K.-T.; Meng, P.-J.; Chung, K.-N.; Chen, J.-P.; Chen, M.-H.; Liu, M.-Y.; Chang, Y.-C.; Fan, T.-Y.; Lin, H.-J. *Long-Term Ecological Research (LTER) Program of the Coral Reefs in Kenting National Park, Taiwan*; Kenting National Park Headquarters, Construction and Planning Agency, Minister of the Interior: Taipei, Taiwan, 2006; p. 233. (In Chinese)
90. Wang, W.-H.; Shao, K.-T.; Meng, P.-J.; Fan, T.-Y.; Chen, J.-P.; Chen, M.-H.; Liu, M.-Y.; Chang, Y.-C.; Lin, H.-J.; Ho, P.-H. *Long-Term Ecological Research (LTER) Program of the Coral Reefs in Kenting National Park, Taiwan*; Kenting National Park Headquarters, Construction and Planning Agency, Minister of the Interior: Taipei, Taiwan, 2007; p. 226. (In Chinese)
91. Ho, M.-J.; Cheng, N.-Y.; Chen, Y.-C.; Kuo, C.-Y.; Wen, C.K.-C.; Cherh, K.-L.; Chen, C.A. Investigating current status of benthic ecology in the Kenting National Park marine zones, Taiwan. *J. Natl. Park* **2016**, *26*, 46–56, (In Chinese with English abstract).
92. Ho, P.-H.; Chen, C.A.; Chen, H.-Y.; Chen, J.-P.; Chiu, Y.-W.; Ho, P.-H.; Lin, H.-J.; Chang, Y.-C. *Long-Term Ecological Research (LTER) Program of the Coral Reefs in Kenting National Park, Taiwan*; Kenting National Park Headquarters, Construction and Planning Agency, Minister of the Interior: Taipei, Taiwan, 2008; p. 270. (In Chinese)
93. Ho, P.-H.; Chen, C.A.; Meng, P.-J.; Chen, J.-P.; Chiu, Y.-W.; Lin, H.-J.; Chang, Y.-C. *Long-Term Ecological Research (LTER) Program of the Coral Reefs in Kenting National Park, Taiwan*; Kenting National Park Headquarters, Construction and Planning Agency, Minister of the Interior: Taipei, Taiwan, 2009; p. 276. (In Chinese)
94. Ho, P.-H.; Chen, C.A.; Meng, P.-J.; Chen, J.-P.; Chiu, Y.-W.; Lin, H.-J.; Chang, Y.-C.; Liu, P.-J. *Long-Term Ecological Research (LTER) Program of the Coral Reefs in Kenting National Park, Taiwan*; Kenting National Park Headquarters, Construction and Planning Agency, Minister of the Interior: Taipei, Taiwan, 2010; p. 300. (In Chinese)
95. Ho, P.-H.; Chen, C.A.; Meng, P.-J.; Chen, J.-P.; Chiu, Y.-W.; Lin, H.-J.; Chang, Y.-C.; Liu, P.-J. *Long-Term Ecological Research (LTER) Program of the Coral Reefs in Kenting National Park, Taiwan*; Kenting National Park Headquarters, Construction and Planning Agency, Minister of the Interior: Taipei, Taiwan, 2011; p. 330. (In Chinese)
96. Chen, C.A. *Long-Term Ecological Research (LTER) Program of the Coral Reefs in Kenting National Park, Taiwan*; Kenting National Park Headquarters, Construction and Planning Agency, Minister of the Interior: Taipei, Taiwan, 2012; p. 73. (In Chinese)
97. Chen, C.A.; Meng, P.-J.; Chiu, Y.-W. *Annual Monitoring of Benthic Community, Water Quality and Shellfish in Kenting National Park*; Kenting National Park Headquarters, Construction and Planning Agency, Minister of the Interior: Taipei, Taiwan, 2013; p. 119. (In Chinese)
98. Chen, C.A. *Evaluating and Re-Design the Zoning System in Kenting National Park, Taiwan*; Kenting National Park Headquarters, Construction and Planning Agency, Minister of the Interior: Taipei, Taiwan, 2014; p. 133. (In Chinese)

99. Chen, C.A. *Annual Survey of the Biodiversity of Coral Reef Ecosystem in Kenting National Park, Taiwan*; Kenting National Park Headquarters, Construction and Planning Agency, Minister of the Interior: Taipei, Taiwan, 2016; p. 82. (In Chinese)
100. Chen, C.A. *Annual Survey of the Biodiversity of Coral Reef Ecosystem in Kenting National Park, Taiwan*; Kenting National Park Headquarters, Construction and Planning Agency, Minister of the Interior: Taipei, Taiwan, 2018; p. 100. (In Chinese)
101. Kuo, C.-Y. The Structure and Variation of Benthic Communities in Coral Reefs of Southern Taiwan. Master's Thesis, National Sun Yat-sen University, Kaohsiung, Taiwan, 2007; p. 71. (In Chinese with English abstract).
102. Kuo, C.-Y.; Meng, P.-J.; Ho, P.-H.; Wang, J.-T.; Chen, J.-P.; Chiu, Y.-W.; Lin, H.-J.; Chang, Y.-C.; Fan, T.-Y.; Chen, C.M. Damage to the reefs of Siangjiao Bay marine protected area in Kenting National Park, Taiwan during Typhoon Morakot. *Zool. Stud.* **2011**, *50*, 85.
103. Kuo, C.-Y.; Ho, M.-J.; Chen, Y.-C.; Yang, S.-Y.; Huang, Y.-Y.; Hsieh, H.-J.; Jeng, M.-S.; Chen, C.A. *Status of Coral Reefs in East Asian Seas Region: Taiwan*; Ministry of the Environment of Japan; Japan Wildlife Research Center: Tokyo, Japan, 2018; pp. 33–40.
104. Kohler, K.E.; Gill, S.M. Coral point count with Excel extensions (CPCe): A visual basic program for the determination of coral and substrate coverage using random point count methodology. *Comput. Geosci.* **2006**, *32*, 1259–1269. [[CrossRef](#)]
105. Price, J.F. Upper ocean response to a hurricane. *J. Phys. Oceanogr.* **1981**, *11*, 153–175. [[CrossRef](#)]
106. Wada, A.; Uehara, T.; Ishizaki, S. Typhoon-induced sea surface cooling during the 2011 and 2012 typhoon seasons: Observational evidence and numerical investigations of the sea surface cooling effect using typhoon simulations. *Prog. Earth Planet. Sci.* **2014**, *64*, 3562–3578. [[CrossRef](#)]
107. Keshavmurthy, S.; Meng, P.-J.; Wang, J.-T.; Kuo, C.-Y.; Yang, S.-Y.; Hsu, C.-M.; Gan, C.-H.; Dai, C.-F.; Chen, C.A. Can resistant coral-*Symbiodinium* associations enable coral communities to survive climate change? A study of a site exposed to long-term hot water input. *PeerJ* **2014**, *2*, e327. [[CrossRef](#)] [[PubMed](#)]
108. Hsu, C.-M.; Keshavmurthy, S.; Denis, V.; Kuo, C.-Y.; Wang, J.-T.; Meng, P.-J.; Chen, C.A. Temporal and Spatial Variations in Symbiont Communities of Catch Bowl Coral *Isopora palifera* (Scleractinia: Acroporidae) on Reefs in Kenting National Park, Taiwan. *Zool. Stud.* **2012**, *51*, 1343–1353.
109. Keshavmurthy, S.; Hsu, C.-M.; Kuo, C.-Y.; Meng, P.-J.; Wang, J.-T.; Chen, C.A. Symbiont communities and host genetic structure of the brain coral *Platygyra verweyi*, at the outlet of a nuclear power plant and adjacent areas. *Mol. Ecol.* **2012**, *21*, 4393–4407. [[CrossRef](#)]
110. Carballo-Bolaños, R.; Denis, V.; Huang, Y.-Y.; Keshavmurthy, S.; Chen, C.A. Temporal variation and photochemical efficiency of species in Symbiodinaceae associated with coral *Leptoria phrygia* (Scleractinia; Merulinidae) exposed to contrasting temperature regimes. *PLoS ONE* **2019**, *14*, e0218801. [[CrossRef](#)]
111. Chiou, W.-D.; Cheng, L.-Z.; Ou, H.-C. Relationship between the dispersion of thermal effluent and the tidal current in the waters near the outlet of the third nuclear power plant in southern Taiwan. *J. Fish. Soc. Taiwan* **1993**, *20*, 207–220.
112. Fan, K.-L. The thermal effluent problems of three nuclear power plants in Taiwan. In *Oceanography of Asian Marginal Seas*, 54; Takano, K., Ed.; Elsevier: Amsterdam, The Netherlands, 1991; pp. 393–403.
113. Pier, J.-J. Power uprate effect on thermal effluent of nuclear power plants in Taiwan. In *Nuclear Power—Operation, Safety and Environment*; Tsvetkov, P., Ed.; InTech Publication: Rijeka, Croatia, 2011; 287p.
114. Chen, C.A.; Yang, Y.W.; Wei, N.V.; Tsai, W.S.; Fang, L.S. Symbiont diversity in scleractinian corals from tropical reefs and subtropical non-reef communities in Taiwan. *Coral Reefs* **2005**, *24*, 11–22. [[CrossRef](#)]
115. Huang, Y.-Y.; Carballo-Bolaños, R.; Kuo, C.-Y.; Keshavmurthy, S.; Chen, C.A. Seasonal dynamics of the Symbiodiniaceae in *Leptoria phrygia* (Scleractinia; Merulinidae). (submitted, under review).
116. Fong, W.-L. Spatial and Temporal Variations of Host Genetic Structure and Symbiont Community in the Catch Bowl Coral, *Isopora Palifera* (Scleractinia; Acroporidae), in Kenting National Park, Taiwan. Master's Thesis, Institute of Oceanography, National Taiwan University, Taipei, Taiwan, 2016.
117. Baker, A.C.; Starger, C.J.; McClanahan, T.R.; Glynn, P.W. Coral reefs: Corals' adaptive response to climate change. *Nature* **2004**, *430*, 741. [[CrossRef](#)]
118. Berkelmans, R.; Van Oppen, M. The role of zooxanthellae in the thermal tolerance of corals: A 'nugget of hope' for coral reefs in an era of climate change. *Proc. R. Soc. B Biol. Sci.* **2006**, *273*, 2305–2312. [[CrossRef](#)] [[PubMed](#)]

119. Silverstein, R.N.; Cunning, R.; Baker, A.C. Change in algal symbiont communities after bleaching, not prior heat exposure, increases heat tolerance of reef corals. *Glob. Chang. Biol.* **2015**, *21*, 236–249. [[CrossRef](#)] [[PubMed](#)]
120. Cunning, R.; Silverstein, R.N.; Baker, A.C. Investigating the causes and consequences of symbiont shuffling in a multi-partner reef coral symbiosis under environmental change. *Proc. R. Soc. B Biol. Sci.* **2015**, *282*, 20141725. [[CrossRef](#)]
121. Boulotte, N.M.; Dalton, S.J.; Carroll, A.G.; Harrison, P.L.; Putnam, H.M.; Peplow, L.M.; van Oppen, M.J. Exploring the *Symbiodinium* rare biosphere provides evidence for symbiont switching in reef-building corals. *ISME J.* **2016**. [[CrossRef](#)] [[PubMed](#)]
122. Thornhill, D.J.; Lajeunesse, T.C.; Kemp, D.W.; Fitt, W.K.; Schmidt, G.W. Multi-year, seasonal genotypic surveys of coral-algal symbioses reveal prevalent stability or post-bleaching reversion. *Mar. Biol.* **2006**, *148*, 711–722. [[CrossRef](#)]
123. Keshavmurthy, S.; Tang, K.-H.; Hsu, C.-M.; Gan, C.-H.; Kuo, C.-Y.; Soong, K.; Chou, H.-N.; Chen, C.A. *Symbiodinium* spp. associated with scleractinian corals from Dongsha Atoll (Pratas), Taiwan, in the South China Sea. *PeerJ* **2017**, *5*, e2871. [[CrossRef](#)]
124. Tong, H.; Cai, L.; Zhou, G.; Yuan, T.; Zhang, W.; Tian, R.; Huang, H.; Qian, P.-Y. Temperature shapes coral-algal symbiosis in the South China Sea. *Sci. Rep.* **2017**, *7*, 40118. [[CrossRef](#)] [[PubMed](#)]
125. Gong, S.; Chai, G.; Xiao, Y.; Xu, L.; Yu, K.; Li, J.; Liu, F.; Cheng, H.; Zhang, F.; Liao, B.; et al. Flexible symbiotic associations of *Symbiodinium* with five typical coral species in tropical and subtropical reef regions of the northern South China Sea. *Front. Microbiol.* **2018**, *9*, 2485. [[CrossRef](#)] [[PubMed](#)]
126. Chen, C.A.; Wang, J.-T.; Fang, L.-S.; Yang, Y.-W. Fluctuating algal symbiont communities in *Acropora palifera* (Scleractinia: Acroporidae) from Taiwan. *Mar. Ecol. Prog. Ser.* **2005**, *295*, 113–121. [[CrossRef](#)]
127. Thornhill, D.J.; Fitt, W.K.; Schmidt, G.W. Highly stable symbioses among western Atlantic brooding corals. *Coral Reefs* **2006**, *25*, 515–519. [[CrossRef](#)]
128. Tonk, L.; Sampayo, E.E.; Lajeunesse, T.C.; Schrammeyer, V.; Hoegh-Guldberg, O. *Symbiodinium* (Dinophyceae) diversity in reef-invertebrates along an offshore to inshore reef gradient near Lizard Island, Great Barrier Reef. *J. Phycol.* **2014**, *50*, 552–563. [[CrossRef](#)] [[PubMed](#)]
129. Beyer, H.L.; Kenndey, E.V.; Beger, M.; Chen, C.A.; Cinner, J.E.; Darling, E.S.; Eakin, C.M.; Gates, R.D.; Heron, S.F.; Knowlton, N.; et al. Risk-sensitive planning for conserving coral reefs under rapid climate change. *Conserv. Lett.* **2018**, *11*, e12587. [[CrossRef](#)]
130. Ablan, M.C.A.; McManus, J.W.; Chen, C.A.; Shao, K.T.; Bell, J.; Cabanban, A.S.; Tuan, V.S.; Arthana, I.W. Meso-scale transboundary units for the management of coral reefs in the South China Sea area. *NAGA Worldfish Cent. Q.* **2002**, *25*, 4–9.
131. Chen, C.A.; Ablan, M.C.A.; McManus, J.W.; Bell, J.D.; Tuan, V.S.; Cabanban, A.S.; Shao, K.-T. Population structure and genetic variability of six-bar wrasse (*Thalassoma hardwicki*) in northern South China Sea revealed by mitochondrial control region sequences. *Mar. Biotechnol.* **2004**, *6*, 312–326. [[CrossRef](#)]
132. Chou, L.M. Southeast Asian reefs—a status update: Cambodia, Indonesia, Malaysia, Philippines, Singapore, Thailand and Vietnam. In *Status of the Coral Reefs of the World: 2000*; Wilkinson, C., Ed.; Australian Institute of Marine Science: Townsville, Australia, 2000; pp. 117–129.
133. West, J.M.; Slam, R.V. Resistance and resilience to coral bleaching: Implications for coral reef conservation and management. *Conserv. Biol.* **2003**, *4*, 956–967. [[CrossRef](#)]
134. Riegl, B.; Piller, W.E. Possible refugia for reefs in times of environmental stress. *Int. J. Earth Sci.* **2003**, *92*, 520. [[CrossRef](#)]
135. Grimsditch, G.D.; Salm, R.V. *Coral Reef Resilience and Resistance to Bleaching*; IUCN: Gland, Switzerland, 2005; 50p.
136. Chollett, I.; Mumby, P.J.; Cortes, J. Upwelling areas do not guarantee refuge for corals in a warming ocean. *Mar. Ecol. Prog. Ser.* **2010**, *416*, 47–56. [[CrossRef](#)]
137. Glynn, P.W.; D’croz, L. Experimental evidence for high temperature stress as the cause of El Nino-coincident coral mortality. *Coral Reefs* **1990**, *8*, 181–191. [[CrossRef](#)]

138. Sánchez-Noguera, C.; Stuhldreier, I.; Cortés, J.; Jiménez, C.; Morales, Á.; Wild, C.; Rixen, T. Natural ocean acidification at Papagayo upwelling system (North Pacific Costa Rica): Implications for reef development. *Biogeosciences* **2018**, *15*, 2349–2360. [[CrossRef](#)]
139. Wizemann, A.; Nandini, S.D.; Stuhldreier, I.; Sánchez-Noguera, C.; Wisshak, M.; Westphal, H.; Rixen, T.; Wild, C.; Reymond, C.E. Rapid bioerosion in a tropical upwelling coral reef. *PLoS ONE* **2018**, *13*, e0202887. [[CrossRef](#)]



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