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Chemical Vapor Deposition of Photocatalyst Nanoparticles on PVDF Membranes for Advanced Oxidation Processes

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Abstract: The chemical binding of photocatalytic materials, such as TiO₂ and ZnO nanoparticles, onto porous polymer membranes requires a series of chemical reactions and long purification processes, which often result in small amounts of trapped nanoparticles with reduced photocatalytic activity. In this work, a chemical vapor deposition technique was investigated in order to allow the nucleation and growth of ZnO and TiO₂ nanoparticles onto polyvinylidene difluoride (PVDF) porous membranes for application in advanced oxidation processes. The thickness of obtained surface coatings by sputtered nanoparticles was found to depend on process conditions. The photocatalytic efficiency of sputtered membranes was tested against both a model drug and a model organic pollutant in a small continuous flow reactor.

Keywords: chemical vapor deposition; polyvinylidene difluoride; titanium dioxide; zinc oxide; photocatalysis

1. Introduction

Thin films are material layers that have thicknesses varying from tens of nanometers to a few micrometers [1]. They are generally obtained by deposition processes on the surface of given substrates. The thin film growth is generally a two-step process where an initial random nucleation step is followed by an ordered growth. Nucleation and growth—and consequently the film structure—depend on the substrate chemistry (surface composition and structure), the method used [2], and deposition conditions [3].

Thin film deposition methods are classified in solid, liquid, and gas phase deposition methods according to the physical state of the deposited material. A further classification of gas deposition methods distinguishes chemical vapor deposition (CVD) and physical vapor deposition (PVD) processes. Both methods involve the deposition of atoms or molecules carried in their vapor phase

onto the substrate surface. In a CVD process, the deposited (target) material reacts chemically with the substrate, while in PVD processes, the deposited molecules and substrate are still distinct [4].

Separation, concentration, and purification processes present some challenges for chemical industries. Efficient separation and purification are also important for food and pharmaceutical plants in order to guarantee high-quality water after removal of toxic components from industrial wastewater. Pharmaceutical active compounds can be considered as hazardous substances because of their potential threat to health and the environment. They are also considered an emerging pollutant due to the failure of classical treatments (such as filtration, adsorption, bio-oxidation, sedimentation, coagulation, chlorination, and UV-irradiation) to effectively remove them [5–7]. In addition to common chemical pollutants, over 80 pharmaceutical active compounds have been detected in wastewater effluents and surface water across the world [8], with concentration ranging from few ng $\rm L^{-1}$ to several g $\rm L^{-1}$. The most important sources of pharmaceutical active compounds are incorrect disposal of unused drugs and effluents of wastewater treatment plants (including pharmaceutical industries, hospital wastewater, aqua-farming, and cattle-breeding) [9–11]. While the measured concentrations can result in water that is low or below drinking water guidelines and health criteria [12], their continuous accumulation in aquatic environment can also represent a real hazard.

More recently, polymer membranes have been used as innovative separation materials [13]. A membrane can be defined as "a barrier that separates and/or contacts two different regions and controls the exchange of matter and energy between the regions" [14]. Today, membranes are efficiently used for water desalinization, wastewater purification, recovery of valuable constituents from production waste, gas separation in petrochemical processes, concentration and purification in food and drug applications, artificial organs and therapeutic systems, energy conversion, and storage systems [15–17]. In addition to their technical simplicity and energy efficiency, membrane processes can be easily upscaled from batchwise treatment of small quantities to large-scale continuous operations.

An advanced oxidation process (AOP) is a simple technique that allows an efficient degradation of organic pollutants generally found in wastewater. In an AOP, organic pollutants are mineralized by the generation of highly reactive hydroxyl radicals [18]. Among the advanced treatment technologies, UV photocatalysis by nanoparticles (e.g., ZnO and TiO_2), has attracted great interest in recent years [19].

In a photooxidation process (Figure 1), electrons are promoted from valence band to conduction band—resulting in the formation of electron-hole pairs—when the catalyst nanoparticle (*CNp*) is irradiated by UV light with energy intensity larger than the characteristic band gap (3.37 eV for ZnO and 3.2 eV for TiO₂, respectively) [20,21].

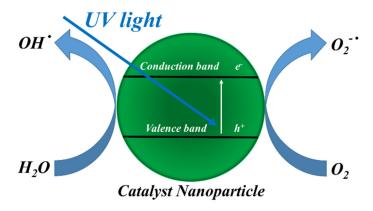


Figure 1. Schematization of the photoactivity of a catalyst nanoparticle (*CNp*).

Both electrons and holes can move to the semiconductor surface and produce radicals, which can oxidize organic compounds (*OC*), whereas electrons can reduce them, according to the reactions

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reported [19,22]. The degradation of organic compounds (*OC*) by photocatalyst nanoparticles (*CNp*) is shown below.

$$CNp + hv \rightarrow e_{cb}^{-} + h_{vb}^{+}$$

$$O_{2} + e_{cb}^{-} \rightarrow O_{2}^{-\bullet}$$

$$H_{2}O + h_{vb}^{+} \rightarrow OH^{\bullet} + H^{+}$$

$$O_{2}^{-\bullet} + H_{2}O \rightarrow H_{2}O_{2} \rightarrow 2OH^{\bullet}$$

$$OH^{\bullet} + OC \rightarrow OC_{ox}$$

$$OC + e_{cb}^{-} \rightarrow OC_{red}$$

ZnO and TiO₂ are the most commonly used photocatalysts due to their redox ability, chemical stability, reduced toxicity towards the environment and health, and low cost [23]. In addition to the mineralization of organic compounds, the reactive redox species—such as hydroxyl radicals (OH^{\bullet}) , superoxide anions $(O_2^{-\bullet})$, and hydrogen peroxide molecules (H_2O_2) generated by UV irradiation—can damage the cell membrane of microorganisms [24] and kill bacteria, viruses, fungi, and algae [25], thus conferring long-term antibacterial and antifungal properties [26–28] to photocatalysts.

More recently, submerged membranes have been integrated by semiconductor photocatalysts in order to photomineralize membrane fouling [29,30], thereby reducing cleaning and maintenance costs. In particular, Ho et al. obtained an enhancement in the filtration flux of a submerged membrane reactor by integration of photooxidation process and membrane filtration [31], while Mendez-Arriaga et al. [32] studied the TiO₂ photocatalytic degradation of pharmaceutical compounds such as diclofenac, naproxen, and ibuprofen.

The combination of membrane separation and advanced oxidation processes is an emerging technology for the complete removal of pollutants because each technique complements the advantages of the other. In particular, the AOP eliminates membrane fouling and allows the remediation of the concentrate while, at the same time, the membrane process filters the feed and concentrates pollutants to an optimal level for AOP [33–35].

Nevertheless, the chemical binding of photocatalysts onto porous polymer membranes requires a series of chemical reactions and long cleaning processes, which often result in small amounts of trapped nanoparticles with reduced photocatalytic activity. In addition, some polymers, such as polytetrafluoroethylene and polyvinylidene fluoride, are hardly functionalizable in order to trap photocatalyst molecules.

In this work, a chemical vapor deposition process was investigated in order to allow the nucleation and growth of ZnO and TiO₂ photocatalytic nanoparticles onto polyvinylidene difluoride (PVDF) porous membranes for applications in AOP.

The purpose of this work was the coupling of filtration properties of polymer membranes with the photocatalytic activity of ZnO and TiO₂ nanoparticles nucleated and grown on PVDF porous membranes using the CVD technique. The substrates used were membrane disks in PVDF, which is a thermoplastic material characterized by high strength and nontoxicity and, consequently, widely used in membrane processes and food applications. Moreover, PVDF is characterized by high chemical and UV stability, which renders this material particularly interesting for photocatalytic applications.

CVD is a well-known chemical process for the production of high-purity, high-performance solid thin films. In a typical CVD process, the substrate is exposed to one or more volatile precursors, which react on the substrate surface to produce the desired layer (Figure 2).

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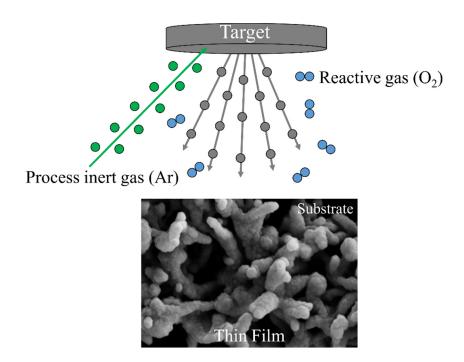


Figure 2. Schematization of the chemical vapor deposition (CVD) technique. The substrate is exposed to one or more volatile precursors, which react on the substrate surface to produce the desired thin film.

The photocatalytic efficiency of sputtered membranes was tested against a model drug (diclofenac sodium) and a model pollutant (methylene blue) in a small continuous flow reactor.

2. Materials and Methods

The substrates used were PVDF membrane disks with a diameter of 47 mm, a porosity of 70%, and a mean pore size of 0.10 μ m (Durapore ©, Merck KGaA, Darmstadt, Germany). Prior to use, membranes were washed in methanol (Sigma-Aldrich, Milan, Italy) by an ultrasonic bath (model M1800H-E, Bransonic, Danbury, CT, USA). The deposition of nanostructured photocatalysts on PVDF membranes was obtained by sputtering of appropriate targets by process inert gas ions (argon) in a Edwards AUTO-306 sputtering system (Edwards, Burgess Hill, UK).

ZnO was deposited on PVDF membranes by Argon (purity 99.999%) sputtering of a ZnO target (purity 99.99%, Goodfellow Cambridge Ltd., Huntingdon, England). The deposition of nanostructured TiO₂ was obtained by reactive sputtering using a Ti target (purity 99.999%, Goodfellow Cambridge Ltd., Huntingdon, England) in presence of a gaseous mixture of argon and oxygen (purity 99.999%, pressure of gas mixture: $p(Ar) = 2.8 \times 10^{-3}$ mbar and $p(O_2) = 1.2 \times 10^{-3}$ mbar, $p(Ar)/p(O_2) = 2.3$). The reactive gas mixture reacts with the substrate and sputtered atoms, forming a thin film of desired compound onto the substrate. The particular pressure ratio between Ar and O_2 was chosen in order to form the anatase polymorph of TiO₂ [36], which is more photoactive than rutile polymorph [37].

A microRaman spectrometer (Labram, Horiba Jobin Yvon) equipped with an Olympus microscope and interfaced to a color camera was used to confirm the presence of TiO_2 anatase thin layers. The Raman spectra were collected through a $100\times$ objective using a He–Ne laser (emission wavelength at 632.8 nm, power 5 mW).

Membrane and nanoparticle morphology was investigated by scanning electron microscopy (SEM). Observations were performed on membrane top surfaces coated with a thin gold or graphite film by a Leica LEO 420 (Leica Microsystems, Cambridge, England) or a Quanta 200 (FEI/Philips, Eindhoven, Netherlands) scanning electron microscope equipped with a backscatter electron detector. Energy-dispersive X-ray (EDX) maps were obtained with a Phenom ProX SEM (Phenom-World, Eindhoven, The Netherlands). Transmission electron microscope (TEM) images were collected with a

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JEM 1400 Plus transmission electron microscope operating at 100 kV (Jeol, Akishima, Tokyo, Japan). The shape and size of nanoparticles was obtained by software analysis of TEM pictures. The number of measured nanoparticles—taken from different pictures of the same sample—was at least 100, and their size was evaluated with an image software (Motic Images Plus 2.0, MoticEurope S.L.U., Barcelona, Spain).

Static contact angles to pure water were measured with a CAM 200 contact angle meter (KSV Instruments LTD, Helsinki, Finland) at 25 $^{\circ}$ C. A drop (2 μ L) of water was put onto the sample surface by a microsyringe, and measurements were carried out by setting the tangents on both visible edges of the droplet on five different positions of each sample and calculating the average value of the measurements.

The permeation tests were carried out with distilled water using a tangential flow filtration cell having an active area of $14.5~\rm cm^2$. The feed solution (at $25\pm1~\rm ^{\circ}C$) was pumped parallel to the membrane surface by a gear pump at the transmembrane pressure of $0.4~\rm bar$. The feed flow rate was $250~\rm mL~min^{-1}$. Permeate samples were collected every $5~\rm min$ in order to determine the transmembrane flux, J, defined as:

$$J = \frac{V_p}{A \Delta t} \tag{1}$$

where V_p is the permeate volume passed through the membrane in the fixed time interval Δt and A is the effective membrane area. The photoactivity of ZnO and TiO₂ layers was tested in a small continuous plant where either a diclofenac sodium (9.3 \times 10⁻⁵ M, Sigma Aldrich, Milan, Italy) or methylene blue (1.3 \times 10⁻⁵ M, Sigma Aldrich, Milan, Italy) water solution was recirculated by a peristaltic system through a round cell, which was equipped with a quartz window to allow UV irradiation and divided into two compartments by membrane [38]. The sputtered side of membranes was exposed to the UV light from a medium-high pressure mercury vapor lamp (ZS lamp, Helios Italquartz, Italy) with an irradiance of 6 W m⁻². At the cell exit, the solution passed through a quartz flow cuvette placed inside a spectrophotometer able to read the absorbance value at either 275 nm or 665 nm, which are the wavelength of maximum absorption of diclofenac sodium and methylene blue, respectively. The absorbance was measured at regular intervals of 5 or 15 min.

3. Results and Discussion

The quality of the obtained thin film is strongly dependent on the process parameters. In particular, the sputtering yield (*Y*) is defined as the number of sputtered atoms per impinging ion. Consequently, a higher yield results in a higher sputtering deposition rate. The sputtering yield depends on several parameters [39], including the energy of the incident ions, the masses of the ions and target atoms, the binding energy of atoms in the solid, and the incident angle of ions. The sputtering yield can be expressed as:

$$Y = \alpha \frac{Mm E_m}{(M+m)^2 U_M} \tag{2}$$

where m and M are the mass of the bombing ion and target atom, respectively, E_m is the kinetic energy of bombing ion, and U_m is the bonding energy of target metal. α takes into account the incident angle of ions.

It is important to recall that magnetic field strength, CVD chamber volume, power density, gas composition and pressure are other factors that can affect yield values [40]. In addition, the film deposition rate decreases for increasing target-substrate distances. Therefore, under the chosen sputtering process parameters (obtained starting from values based on the previous works [41,42] using the same sputtering source), an optimal deposition rate for ZnO and TiO₂ of about 2 and 1 nm min⁻¹, respectively, was gained. The sputtering time used for the ZnO target was twice that of the Ti one in order to have similar layer thicknesses.

The different sputtering conditions (sputtering power and time, target distance, and gas pressure) were tested in order to find the optimal set of parameters able to give a homogeneous membrane

coverage with no polymer substrate damage and very small photocatalyst nanoparticles. This last condition ensures a high photoactivity, with catalysis being a surface process. Due to inadequate sputtering conditions, typical sample drawbacks were inhomogeneous coverage, pore occlusion, and presence of cracks (damage of thin film), as shown in Figure 3. The best results—in terms of both coverage quality and nanoparticle size—were obtained with the sputtering conditions (sputtering power, target distance, gas pressure, sputtering time) shown in Table 1.

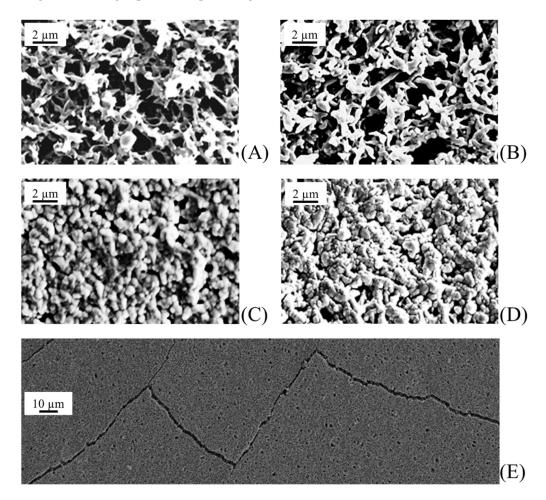


Figure 3. Morphology of virgin and TiO_2 sputtered polyvinylidene difluoride (PVDF) membranes after different sputtering times (t): (A) virgin PVDF membrane; t = 0, (B) t = 1 h; (C) t = 2 h; (D) t = 3 h; (E) t = 4 h.

Table 1. Optimal sputtering parameters able to give a homogeneous coverage with no substrate damage and small nanoparticles.

Target	Photocatalyst Layer	Sputtering Power/W	Target Distance/10 ⁻² m	Pressure/10 ⁻⁶ bar	Sputtering Time/min	
ZnO	ZnO	35	8	P(Ar) = 4.5	30	
Ti	TiO_2	65	7	$P(Ar) = 2.8^{1}$	60	
1 P(O ₂) = 1.2 × 10 ⁻⁶ bar.						

Figure 4 shows the morphology of the top surface in virgin and sputtered PVDF membranes under the experimental conditions reported in Table 1. Both photocatalyst coatings (Figure 4B,C for ZnO and TiO₂, respectively) were homogeneous with no evident alteration/damage of the virgin PVDF membrane (Figure 4A). In addition, no occlusion of membrane pores was present. Coatings had

a cauliflower structure with aggregate diameters of around 100 nm and formed by agglomeration of smaller primary nanoparticles (see later).

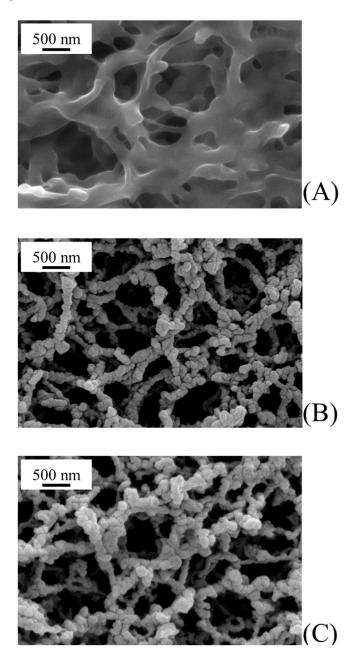


Figure 4. Morphology of virgin and sputtered PVDF membranes under the experimental conditions reported in Table 1: (**A**) virgin PVDF membrane; (**B**) ZnO sputtered PVDF membrane; and (**C**) TiO_2 sputtered PVDF membrane.

In order to further confirm that PVDF membranes were homogeneously covered with photocatalysts, SEM backscattering electron micrographs and spot EDX analysis on sputtered PVDF membranes were performed. Figure 5 shows SEM backscattering electron micrographs and EDX maps for ZnO and TiO_2 sputtered membranes.

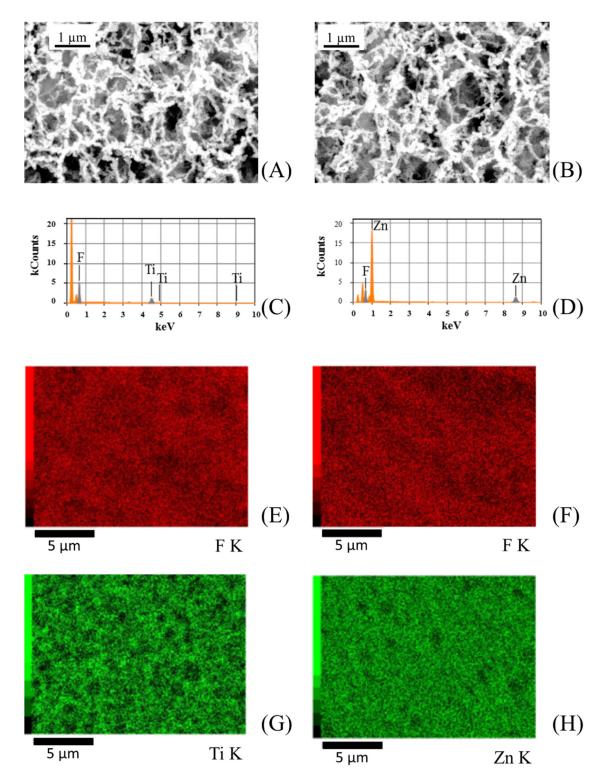


Figure 5. (**A,B**) Scanning electron microscopy (SEM) backscattering electron micrographs, (**C,D**) elemental mapping, and (**E–H**) energy-dispersive X-ray (EDX) analysis performed on TiO₂ (on the **left**) and ZnO (on the **right**) sputtered PVDF membranes.

At larger magnifications, TEM analysis allows characterizing the morphology of primary nanoparticles grown on PVDF membranes. As shown in Figure 6A,B, both ZnO and TiO₂ primary nanoparticles were rather spherical in shape with similar average diameters of 11.6 ± 4.2 and 12.1 ± 3.4 nm, respectively (Figure 7). Primary nanoparticles agglomerated into larger aggregates.

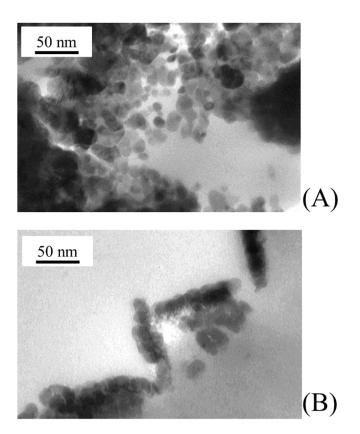


Figure 6. Morphology of catalyst nanoparticles grown on PVDF membranes under the experimental conditions reported in Table 1: (A) ZnO sputtered PVDF membrane; and (B) TiO_2 sputtered PVDF membrane.

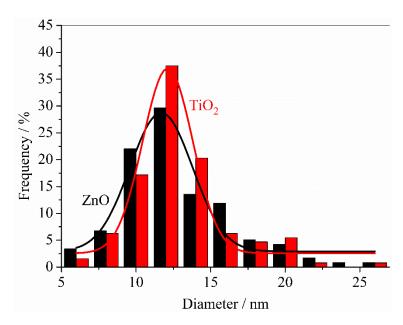


Figure 7. Distribution of diameters shown by primary nanoparticles present on ZnO and TiO_2 sputtered PVDF membranes.

MicroRaman spectrum of PVDF membrane obtained by reactive sputtering using a Ti target is reported in Figure 8.

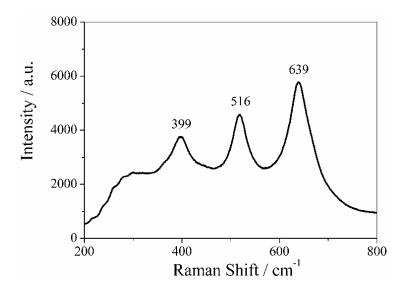


Figure 8. Raman spectrum of TiO_2 nanoparticles sputtered on PVDF membrane. The peaks at 399, 516, and 639 cm⁻¹ are associated to the Raman active modes B1g, A1g, and Eg, respectively, confirming the anatase structure of TiO_2 .

The Raman spectrum of TiO_2 nanoparticles consisted of three peaks with strong intensities at 399, 516, and 639 cm⁻¹, which can be associated to the Raman active modes B1g, A1g, and Eg of anatase structure of TiO_2 thin layers. These values are in good agreement with the Raman bands reported in literature [43]. The fourth active Raman mode of anatase structure of TiO_2 , which is generally placed at 196 cm⁻¹ (Eg mode), was out of the instrument range. It is expected that the deposition of ZnO and TiO_2 thin films could change the hydrophilicity of virgin PVDF membranes.

Table 2 shows the static contact angles to pure water measured for virgin and sputtered membranes. Both ZnO and TiO_2 thin films drastically reduced the contact angle, i.e., increased the hydrophilicity of virgin PVDF membrane from 61° to 27° and 26°, respectively, with a consequent decrease in fouling. It is important to recall that the photoactivity of ZnO and TiO_2 layers is able to further overcome this drawback. Moreover, the increase in hydrophilicity is expected to have a positive effect in the membrane permeation properties. Table 2 also shows the transmembrane fluxes of virgin and sputtered membranes. The performant effect of catalyst thin film increased the transmembrane flux of virgin PVDF from a value of 200 to 760 and 710 L m⁻² h⁻¹ for ZnO and TiO_2 thin films, respectively, as a result of the increased hydrophilicity. These results also confirm that the sputtered thin layer did not occlude the membrane pores.

Table 2. Contact angle and transmembrane flux of virgin and sputtered PVDF membranes under the experimental conditions reported in Table 1.

Photocatalyst	Contact Angle/deg	Transmembrane Flux/L $\mathrm{m}^{-2}~\mathrm{h}^{-1}$
Virgin PVDF	61 ± 1	200 ± 15
ZnO	27 ± 2	760 ± 15
TiO ₂	26 ± 2	710 ± 15

In order to test their photocatalytic activity against pharmaceutical active compounds and organic pollutants, sputtered membranes were placed in a small continuous flow reactor where a water solution of either diclofenac sodium—a well-known anti-inflammatory drug—or methylene blue—a well-known organic dye—was circulated.

Figure 9 shows the activity of ZnO and TiO₂ sputtered membranes in the diclofenac sodium salt photodegradation. The drug photodegradation by PVDF membranes with sputtered ZnO and

 TiO_2 nanoparticles followed a first order kinetics with similar rate constants of 6.8×10^{-3} min⁻¹ and 8.3×10^{-3} min⁻¹, respectively. An almost complete photodegradation of diclofenac sodium salt was obtained within 6 h in both cases.

Similarly, Figure 10 shows the activity of ZnO and TiO_2 sputtered membranes in the methylene blue photodegradation. Also in this case, the photodegradation kinetics of the organic pollutant by PVDF membranes with sputtered ZnO and TiO_2 nanoparticles was a first order kinetics but with larger rate constants of 2.2×10^{-2} min⁻¹ and 2.8×10^{-2} min⁻¹, respectively. Nevertheless, the methylene blue photodegradation stopped after 4 h with a plateau of 33% and 8% for PVDF membranes with sputtered ZnO and TiO_2 nanoparticles, respectively.

The behavior of virgin PVDF membrane reported in Figures 9 and 10 takes into account the UV photolysis of diclofenac sodium and methylene blue, respectively.

Several parameters can affect the degradation efficiency of photocatalysts including the particular dye/drug, the pH of the solution, the presence of oxygen, the addition of hydrogen peroxide, the nanoparticle average size, and the amount and type of catalyst [44]. In particular, several experimental investigations have found that TiO_2 nanoparticles show a photocatalytic efficiency higher than ZnO nanoparticles due to their band gap values [45,46]. Indeed, ZnO samples have a larger band gap, which leads to the production of less radicals and, consequently, to a lower dye photodegradation. On the contrary, TiO_2 has a higher quantum yield engendered by a relatively slower electron-hole pair recombination, faster electron-hole pair migration to the surface, fewer defects, and exciton traps in the crystal lattice [47].

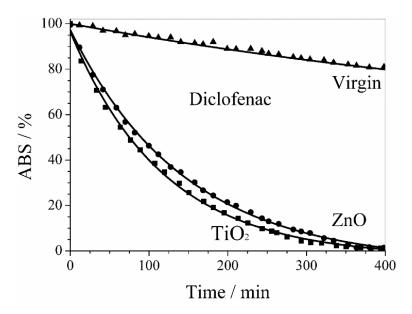


Figure 9. Photodegradation of diclofenac sodium salt by PVDF membranes with sputtered ZnO and TiO_2 nanoparticles. The behavior of virgin PVDF membrane takes into account the UV photolysis of diclofenac.

Even if kinetic rate constants are strongly dependent on membrane composition, investigated pollutants and the UV lamp power used [48], it is important to note that the obtained values of kinetic rate constants are the same order of magnitude as those found in other literature works where the photoactive nanomaterial was either directly synthesized or immobilized on polymer substrates [49–51]. These rate values make the CVD of photocatalyst nanoparticles on porous polymer membranes a suitable technique for applications in the field of advanced oxidation processes. In fact, the CVD of photocatalysts is a fast process that avoids expensive and time-consuming syntheses and cleaning post-treatments, as it is possible to directly sputter photocatalysts onto commercially available membranes while keeping good photocatalytic activities at the same time. Recycling of a catalyst

is a very important property in practical applications. In order to assess the recycling properties of photocatalysts, ten photodegradation cycles were performed using the same film and fresh methylene blue. Both ZnO and TiO₂ sputtered membranes were reused in successive runs without performing any cleaning procedure and gave percentages of degraded methylene blue similar to those obtained after the first run. These results demonstrated the ability of the ZnO and TiO₂ sputtered membranes to fully preserve/restore their initial photocatalytic efficiency. In addition, the long-term stability of photocatalyst deposit onto polymer substrates was checked after ten cycles of successive photocatalysis processes. No evident damage was revealed in the nanoparticles layer morphology, confirming the main advantage of an easy reuse of membranes with sputtered catalysts over the homogeneous and heterogeneous catalysis processes, which suffer the drawbacks of catalyst recovery and damage of polymers used to functionalize membranes or bind catalysts [52]. Further work is in progress in order to control the primary particle size in a finer way.

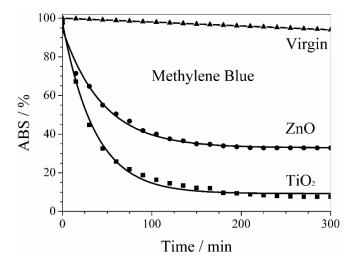


Figure 10. Photodegradation of methylene blue by PVDF membranes with sputtered ZnO and TiO_2 nanoparticles. The behavior of virgin PVDF membrane takes into account the UV photolysis of methylene blue.

4. Conclusions

In this work, the results of a CVD functionalization of polymer porous membranes with photocatalytic nanoparticles, and an application of their use in AOP were shown. The overall sputtering process took less than 1 h, which is lower than conventional times employed in organic synthesis processes of similar materials. The thickness of obtained surface coating by sputtered nanoparticles was found to depend on process conditions. The membranes functionalized with ZnO and TiO₂ nanoparticles were characterized by contact angles lower than that shown by virgin membrane, making these composite membranes suitable for the filtration of aqueous solutions. The deposition of a thin layer of nanoparticles increased the transmembrane fluxes as hydrophilicity increased, and no pore occlusion occurred. In addition to its long-term stability and solvent-free features, the proposed process of membrane functionalization can be easily upscaled to manufacture membrane modules for the efficient degradation of organic pollutants generally found in wastewater.

Author Contributions: E.P., K.A., P.F., T.P., E.F. and C.M. performed the experiments and the data analysis. G.D.F., G.D.P., A.I.M. and F.P.N. provided expert input into the experiments and discussions. G.D.F. and F.P.N. supervised the study and provided scientific discussion. All the authors contributed to the writing and revision of the article.

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Conflicts of Interest: The authors declare no conflict of interest.

References

1. Chidichimo, G.; Cupelli, D.; De Filpo, G.; Formoso, P.; Nicoletta, F.P. Nanoparticles as a smart technology for remediation. In *Sustainable Development in Chemical Engineering, Innovative Technologies;* Piemonte, V., De Falco, M., Basile, A., Eds.; John Wiley & Sons, Ltd.: Chichester, UK, 2013; pp. 297–329. ISBN 9781119953524.

- 2. Kudryashov, M.A.; Mashin, A.I.; Logunov, A.A.; Chidichimo, G.; De Filpo, G. Frequency dependence of the electrical conductivity in Ag/PAN nanocomposites. *Tech. Phys.* **2012**, *57*, 965–970. [CrossRef]
- 3. Löbl, P.; Huppertz, M.; Mergel, D. Nucleation and growth in TiO₂ films prepared by sputtering and evaporation. *Thin Solid Films* **1994**, *251*, 72–79. [CrossRef]
- 4. Carter, C.B.; Norton, M.G. Thin films and vapor deposition. In *Ceramic Materials*; Carter, C.B., Norton, M.G., Eds.; Springer: New York, NY, USA, 2007; ISBN 978-0-387-46270-7.
- 5. Drioli, E.; Ali, A.; Macedonio, F. Membrane distillation: Recent developments and perspectives. *Desalination* **2015**, *356*, 56–84. [CrossRef]
- 6. Bhattacharyya, D. Functionalized membranes and environmental applications. *Clean Technol. Environ. Policy* **2007**, *9*, 81–83. [CrossRef]
- 7. Di Profio, G.; Polino, M.; Nicoletta, F.P.; Belviso, B.D.; Caliandro, R.; Fontananova, E.; De Filpo, G.; Curcio, E.; Drioli, E. Tailored hydrogel membranes for efficient protein crystallization. *Adv. Funct. Mater.* **2014**, 24, 1582–1590. [CrossRef]
- 8. Heberer, T. Occurrence, fate, and removal of pharmaceutical residues in the aquatic environment: A review of recent research data. *Toxicol. Lett.* **2002**, *131*, 5–17. [CrossRef]
- 9. Jones, O.A.H.; Voulvoulis, N.; Lester, J.N. Human pharmaceuticals in the aquatic environment a review. *Environ. Technol.* **2001**, 22, 1383–1394. [CrossRef] [PubMed]
- Santos, J.L.; Aparicio, I.; Alonso, E. Occurrence and risk assessment of pharmaceutically active compounds in wastewater treatment plants. A case study: Seville city (Spain). *Environ. Int.* 2007, 33, 596–601. [CrossRef] [PubMed]
- 11. Ziylan, A.; Ince, N.H. The occurrence and fate of anti-inflammatory and analgesic pharmaceuticals in sewage and fresh water: Treatability by conventional and non-conventional processes. *J. Hazard. Mater.* **2011**, *187*, 24–36. [CrossRef] [PubMed]
- 12. Kolpin, D.W.; Furlong, E.T.; Meyer, M.T.; Thurman, E.M.; Zaugg, S.D.; Barber, L.B.; Buxton, H.T. Pharmaceuticals, hormones, and other organic wastewater contaminants in U.S. Streams, 1999–2000: A national reconnaissance. *Environ. Sci. Technol.* **2002**, *36*, 1202–1211. [CrossRef] [PubMed]
- 13. Salehi, S.M.; Di Profio, G.; Fontananova, E.; Nicoletta, F.P.; Curcio, E.; De Filpo, G. Membrane distillation by novel hydrogel composite membranes. *J. Membr. Sci.* **2016**, *504*, 220–229. [CrossRef]
- 14. Strathmann, H. *Introduction to Membrane Science and Technology;* Wiley-VCH Verlag GmbH & Co. KGaA: Weinheim, Germany, 2011; ISBN 978-3-527-32451-4.
- 15. Drioli, E.; Giorno, L. *Encyclopedia of Membranes*; Springer: Berlin/Heidelberg, Germany, 2016; ISBN 978-3-662-44325-5.
- 16. Nicoletta, F.P.; Cupelli, D.; Formoso, P.; De Filpo, G.; Colella, V.; Gugliuzza, A. Light responsive polymer membranes: A review. *Membranes* **2012**, *134*, 134–197. [CrossRef] [PubMed]
- 17. Formoso, P.; Pantuso, E.; De Filpo, G.; Nicoletta, F.P. Electro-conductive membranes for permeation enhancement and fouling mitigation: A short review. *Membranes* **2017**, 7, 39. [CrossRef] [PubMed]
- 18. Pelizzetti, E.; Serpone, N. *Photo-Catalysis Fundamentals and Applications*; John Wiley & Sons, Inc.: New York, NY, USA, 1989.
- 19. Arana, J.; Melian, J.A.H.; Rodriguez, J.M.D.; Pena, P.P. TiO₂-photocatalysis as a tertiary treatment of naturally treated wastewater. *Catal. Today* **2002**, *76*, 279–289. [CrossRef]
- 20. Hoffmann, M.R.; Martin, S.T.; Choi, W.Y.; Bahnemann, D.W. Environmental applications of semiconductor photo-catalysis. *Chem. Rev.* **1995**, *95*, 69–96. [CrossRef]
- 21. Minero, C. A rigorous kinetic approach to model primary oxidative steps of photo-catalytic degradations. *Sol. Energy Mater. Sol. Cells* **1995**, *38*, 421–430. [CrossRef]
- 22. Mahmoodi, N.M.; Arami, M.; Limaee, N.Y.; Tabrizi, N.S. Kinetics of heterogeneous photo-catalytic degradation of reactive dyes in an immobilized TiO₂ photo-catalytic reactor. *J. Colloid Interface Sci.* **2006**, 295, 159–164. [CrossRef] [PubMed]

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23. Bhatkhande, D.S.; Pangarkar, V.G.; Beenackers, A. Photocatalytic degradation for environmental applications: A review. *J. Chem. Technol. Biotechnol.* **2002**, 77, 102–116. [CrossRef]

- 24. Cho, M.; Chung, H.; Choi, W.; Yoon, J. Linear correlation between inactivation of *E. coli* and OH radical concentration in TiO₂ photocatalytic disinfection. *Water Res.* **2004**, *38*, 1069–1077. [CrossRef] [PubMed]
- 25. Huang, Z.; Maness, P.C.; Blake, D.; William, A.J. Bactericidal mode of titanium dioxide photocatalysis. *J. Photochem. Photobiol. A* **2000**, *130*, 163–170. [CrossRef]
- 26. De Filpo, G.; Palermo, A.M.; Tolmino, R.; Formoso, P.; Nicoletta, F.P. Gellan gum hybrid hydrogels for the cleaning of paper artworks contaminated with Aspergillus versicolor. *Cellulose* **2016**, 23, 3265–3279. [CrossRef]
- 27. De Filpo, G.; Palermo, A.M.; Rachiele, F.; Nicoletta, F.P. Preventing fungal growth in wood by titanium dioxide nanoparticles. *Int. Biodeterior. Biodegrad.* **2013**, *85*, 217–222. [CrossRef]
- 28. De Filpo, G.; Palermo, A.M.; Munno, R.; Molinaro, L.; Formoso, P.; Nicoletta, F.P. Gellan gum/titanium dioxide nanoparticle hybrid hydrogels for the cleaning and disinfection of parchment. *Int. Biodeterior. Biodegrad.* **2015**, *103*, 51–58. [CrossRef]
- 29. Shon, H.K.; Vigneswaran, S.; Ngo, H.H.; Kim, J.H. Chemical coupling of photocatalysis with flocculation and adsorption in the removal of organic matter. *Water Res.* **2005**, *39*, 2549–2558. [CrossRef] [PubMed]
- 30. Mozia, S.; Morawski, A.W. Hybridization of photocatalysis and membrane distillation for purification of wastewater. *Catal. Today* **2006**, *118*, 181–188. [CrossRef]
- 31. Ho, D.P.; Vigneswaran, S.; Ngo, H.H. Integration of photocatalysis and microfiltration in removing effluent organic matter from treated sewage effluent. *Sep. Sci. Technol.* **2010**, *45*, 155–162. [CrossRef]
- 32. Mendez-Arriaga, F.; Esplugas, S.; Gimenez, J. Photocatalytic degradation of nonsteroidal anti-inflammatory drugs with TiO₂ and simulated solar irradiation. *Water Res.* **2008**, 42, 585–594. [CrossRef] [PubMed]
- 33. Ganiyu, S.O.; Van Hullebusch, E.D.; Cretin, M.; Esposito, G. Coupling of membrane filtration and advanced oxidation processes for removal of pharmaceutical residues: A critical review. *Sep. Purif. Technol.* **2015**, *156*, 891–914. [CrossRef]
- 34. Molinari, R.; Caruso, A.; Argurio, P.; Poerio, T. Degradation of the drugs Gemfibrozil and Tamoxifen in R. pressurized and de-pressurized membrane photoreactors using suspended polycrystalline TiO₂ as catalyst. *J. Membr. Sci.* **2008**, *319*, 54–63. [CrossRef]
- 35. Sarasidis, V.C.; Plakas, K.V.; Patsios, S.I.; Karabelas, A.J. Investigation of Diclofenac degradation in a continuous photo-catalytic membrane reactor: Influence of operating parameters. *Chem. Eng. J.* **2014**, 239, 299–311. [CrossRef]
- 36. Zeman, P.; Takabayashi, S. Effect of total and oxygen partial pressures on structure of photocatalytic TiO₂ films sputtered on unheated substrate. *Surf. Coat. Technol.* **2002**, *153*, 93–99. [CrossRef]
- 37. Luttrell, T.; Halpegamage, S.; Tao, J.; Kramer, A.; Sutter, E.; Batzill, M. Why is anatase a better photocatalyst than rutile?—Model studies on epitaxial TiO₂ films. *Sci. Rep.* **2014**, *4*, 4043. [CrossRef] [PubMed]
- 38. Mastropietro, T.F.; Meringolo, C.; Poerio, T.; Scarpelli, F.; Godbert, N.; Di Profio, G.; Fontananova, E. Multistimuli activation of TiO₂/α-Alumina membranes for degradation of methylene blue. *Ind. Eng. Chem. Res.* **2017**, *56*, 11049–11057. [CrossRef]
- 39. Plummer, J.D.; Deal, M.; Griffin, P.D. *Silicon VLSI Technology*; Prentice Hall Press: Upper Saddle River, NJ, USA, 2008.
- 40. Tavakolmoghadam, M.; Mohammadi, T.; Hemmati, M.; Naeimpour, F. Surface modification of PVDF membranes by sputtered TiO₂: Fouling reduction potential in membrane bioreactors. *Desalin. Water Treat.* **2014**, *57*, 3328–3338. [CrossRef]
- 41. Marques, S.M.; Tavares, C.J.; Lanceros-Mendez, S.; Denchev, Z. X-ray scattering experiments on sputtered titanium dioxide coating onto PVDF polymers for self-cleaning application. *J. Appl. Polym. Sci.* **2011**, *119*, 726–731. [CrossRef]
- 42. Tavares, C.J.; Marques, S.M.; Lanceros-Mendez, S.; Robouta, L.; Alves, E.; Barradas, N.P.; Munnik, F.; Girardeau, T.; Riviere, J.P. N-doped photocatalytic titania thin films on active polymer substrates. *J. Nanosci. Nanotechnol.* **2010**, *10*, 1072–1077. [CrossRef] [PubMed]
- 43. Ohsaka, T.; Izumi, F.; Fujiki, Y. Raman spectrum of anatase, TiO₂. *J. Raman Spectrosc.* **1978**, 7, 321–324. [CrossRef]
- 44. Qamar, M.; Muneer, M. A comparative photocatalytic activity of titanium dioxide and zinc oxide by investigating the degradation of vanillin. *Desalination* **2009**, 249, 535–540. [CrossRef]

 Colombo, E.; Li, W.; Bhangu, S.K.; Ashokkumar, M. Chitosan microspheres as a template for TiO₂ and ZnO microparticles: Studies on mechanism, functionalization and applications in photocatalysis and H₂S removal. RSC Adv. 2017, 7, 19373–19383. [CrossRef]

- 46. Barnes, R.J.; Molina, R.; Xu, J.; Dobson, P.J.; Thompson, I.P. Comparison of TiO₂ and ZnO nanoparticles for photocatalytic degradation of methylene blue and the correlated inactivation of gram-positive and gram-negative bacteria. *J. Nanopart. Res.* **2013**, *15*, 1432. [CrossRef]
- 47. Bennett, S.W.; Keller, A.A. Comparative photoactivity of CeO₂, γ-Fe₂O₃, TiO₂ and ZnO in various aqueous systems. *Appl. Catal. B Environ.* **2011**, *102*, 600–607. [CrossRef]
- 48. Martins, P.M.; Miranda, R.; Marques, J.; Tavares, C.J.; Botelho, G.; Lanceros-Mendez, S.; Leong, S.; Razmijou, A.; Wang, K.; Hapgood, K.; et al. TiO₂ based photocatalytic membranes: A review. *J. Membr. Sci.* **2014**, 472, 167–184. [CrossRef]
- 49. Martins, P.M.; Miranda, R.; Marques, J.; Tavares, C.J.; Botelho, G.; Lanceros-Mendez, S. Comparative efficiency of TiO₂ nanoparticles in suspension vs. immobilization into (PVDF–TrFE) porous membranes. *RSC Adv.* **2016**, *6*, 12708–12716. [CrossRef]
- 50. Fischer, K.; Grimm, M.; Meyers, J.; Dietrich, C.; Gläser, R.; Schulze, A. Photoactive microfiltration membranes via directed synthesis of TiO₂ nanoparticles on the polymer surface for removal of drugs from water. *J. Membr. Sci.* **2015**, *478*, 49–57. [CrossRef]
- 51. Starr, B.J.; Tarabara, V.V.; Herrera-Robledo, M.; Zhou, M.; Roualdès, S.; Ayral, A. Coating porous membranes with a photocatalyst: Comparison of LbL self-assembly and plasma-enhanced CVD techniques. *J. Membr. Sci.* **2016**, *514*, 340–349. [CrossRef]
- 52. Song, H.; Shao, J.; He, Y.; Liu, B.; Zhong, X. Natural organic matter removal and flux decline with PEG–TiO₂-doped PVDF membranes by integration of ultra-filtration with photocatalysis. *J. Membr. Sci.* **2012**, 405–406, 48–56. [CrossRef]



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