

Nanosponges for Water Treatment: Progress and Challenges

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Abstract: Nanosponges have shown promising capabilities for efficient removal of organic/inorganic pollutants from water based on absorption/adsorption and disinfection processes. The application of nanosponges (especially cyclodextrin-based nanosponges) can be considered a cost-effective strategy with minimal energy and time requirements in comparison to other routinely deployed water treatment modalities. These polymers with unique physicochemical properties, architectures, and highly cross-linked three-dimensional networks need to be further explored for removing pollutants with simultaneous eliminations of microbial contaminants from wastewater. Additionally, the surface functionalization of these nanosponges utilizing magnetic, titanium dioxide, and silver nanomaterials can significantly improve their properties for water remediation purposes, although nanosponges altered with carbon nanotubes and metallic nanomaterials/nanocatalysts for water treatment appliances are barely explored. Notably, crucial factors such as adsorbent type/dosage, contact time, competing ions, adsorption isotherm models, kinetics, thermodynamics, and reaction/experimental conditions (e.g., molar ratios, temperature, and pH) are important aspects affecting the adsorption and removal of pollutants using nanosponges. Furthermore, the nanotoxicity and biosafety of these nanosponge-based systems utilized for water treatment should be comprehensively evaluated. Herein, recent advancements in the design and deployment of nanosponge-based systems for removing organic/inorganic pollutants from water and wastewater are deliberated with an emphasis on challenges and perspectives.

Keywords: nanosponges; water treatment; pollutants; nanosponge-based systems; cyclodextrin-based nanosponges



Citation: Iravani, S.; Varma, R.S. Nanosponges for Water Treatment: Progress and Challenges. *Appl. Sci.* **2022**, *12*, 4182. <https://doi.org/10.3390/app12094182>

Academic Editors: Yanbiao Liu, Ning Li and Xiuwen Cheng

Received: 31 March 2022

Accepted: 19 April 2022

Published: 21 April 2022

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1. Introduction

One of the most critical aspects in human life is access to safe drinking water. Consequently, water treatment is currently vital, particularly because of population growth and the burden of environmental pollution by domestic waste, sewage, chemical materials, industrial waste, and organic/inorganic pesticides, among others. These pollutants cause toxic effects or generate concerns regarding the color, odor, or taste of water. Microbial contaminations are also of serious trepidation, as has been addressed in the field of nanotechnology and nanomaterials sciences [1]. Improvements on water quality using conventional techniques, such as coagulation/settling treatment methods, membrane-based systems, absorption-based systems, and direct filtration, have been widely deployed, but they encompass several disadvantages/limitations of low efficiency, selectivity, and specificity (Table 1) [2,3]. Thus, there is a vital demand for planning efficient strategies or designing smart and low-cost (nano)systems with high sensitivity/specificity for the efficient removal of pollutants from wastewater and for improving the quality of drinking water [4].

Table 1. Some salient advantages and limitations of water treatment strategies.

Strategies	Limitations/Challenges	Advantages	Refs.
Reverse osmosis technique	<ul style="list-style-type: none"> - Not suitable for small organic pollutants - High cost, energy, and pressure 	<ul style="list-style-type: none"> - Seawater desalination - Organic/inorganic pollutant removal - Removal of bacteria from wastewater 	[5,6]
Flocculation/coagulation	<ul style="list-style-type: none"> - Possibility of toxicity and hazardous effects caused by inorganic coagulants - Formation of large amounts of toxic sludge - Not suitable for removing heavy metals or emerging contaminants - Not suitable for the purification of water with dilution of heavy metals at a low level 	<ul style="list-style-type: none"> - Improvements in water clarity to decrease turbidity 	[7,8]
Chemical precipitation	<ul style="list-style-type: none"> - Sludge production - Additional operative cost for the removal of sludge 	<ul style="list-style-type: none"> - Low cost - Simplicity 	[9]
Adsorption	<ul style="list-style-type: none"> - Low selectivity - High cost of adsorbents - Hard separation/isolation 	<ul style="list-style-type: none"> - Ease of processes - High capacities for binding to metals - Reusability - High efficiency/efficacy 	[10]
Electrochemical techniques	<ul style="list-style-type: none"> - High cost - High energy consumption 	<ul style="list-style-type: none"> - Scalability - Flexible operations - High separation - High selectivity/specificity 	[11,12]
Photocatalysis	<ul style="list-style-type: none"> - Long duration time - Restricted appliances - Preparative process of photocatalysts 	<ul style="list-style-type: none"> - Less harmful/toxic by-products - Concurrent elimination of organics and metals 	[13]
Flotation	<ul style="list-style-type: none"> - High cost of operation and maintenance - High cost of initial capital 	<ul style="list-style-type: none"> - Formation of concentrated sludge - Higher discernment - Minimal period of detention - Higher proficiency - Higher rate of overflow 	[14]
Membrane filtration techniques	<ul style="list-style-type: none"> - Complicated processes - Low permeate flux - High cost of membranes - High costs of operation and maintenance - Deterioration of productivity - Frequency of membrane regeneration difficulties 	<ul style="list-style-type: none"> - Requirement of small spaces - High selectivity - Requirement of low pressure 	[15,16]
Ion exchange method	<ul style="list-style-type: none"> - High cost - Renewal of resins - Secondary pollution toxicity/organic contamination from resins - Organic matter adsorption - Possibility of bacterial or chlorine contaminations 	<ul style="list-style-type: none"> - High efficiency of removal - Fast kinetics - High capacity of treatment processes - Suitability for removing nitrates and arsenic from water - Up-scalability 	[17,18]

Nanosponges with cavities and mesh-like/colloidal structures comprising solid nano-materials are suitable for encapsulating various substances/compounds, such as proteins/peptides, drugs, genetic materials, antineoplastic agents, and volatile oils, among others [19]. There are various methods for synthesizing β -cyclodextrin-based nanosponges, including emulsion solvent evaporation, hyper cross-linked cyclodextrin, ultrasound-assisted production, and interfacial phenomenon techniques [20]. Several categories of cyclodextrin-based nanosponges have been introduced, such as plain (urethane-, carbonate-, ester-, and ether-based linkages connecting cyclodextrin to the cross linker), modified (fluor-

rescence and electric charge nanosponges based on specific properties), stimuli-responsive (with the ability of behavior modulation by external environmental changes, such as temperature/pH and reducing/oxidative circumstances), and molecularly-imprinted (with high selectivity/specificity toward molecules) nanosponges [21]. These polymers with unique properties, architectures, and high cross-linked three-dimensional networks are broadly studied in the field of pharmaceuticals and biomedicine [22]. Additionally, nanosponges with suitable modifications or functionalizations using carbon nanotubes, TiO_2 , Ag, etc., have been scrutinized for removing or capturing organic/inorganic pollutants from water and wastewater [23,24]. For instance, cyclodextrin-based nanosponges exhibit promising capacities for removing dissolved organic carbon from water (~84%) [25]. Surfactant-modified zeolite nanosponges are designed for removing nitrate in contaminated water, with the highest nitrate elimination capability ($1338 \text{ mmol Kg}^{-1}/83 \text{ mg g}^{-1}$) [26]. Among the introduced nanosponges, especially derived from cyclodextrin, desirable attributes such as environmental friendliness, sustainability, non-toxicity, cost-effectiveness, and the capability of hosting different molecular agents have shown suitability for water treatment with up-scalable potentials [27]. There have been limited studies regarding the applications of molecular modeling methods for analyzing cyclodextrins, because of the flexibility and size of such molecules. Some limited calculations have been performed based on density functional theory (DFT) methods, with befittingly significant levels of theory and large basis sets. It appears that by uniting the results obtained from experimental and computational studies, the geometry of complexation of these structures can be better analyzed [28]. Notably, some crucial parameters such as adsorbent dosage, pH solution, isotherm, kinetics, thermodynamics, regeneration/desorption studies, recyclability, and adsorption mechanisms need to be further evaluated [29]. Herein, recent developments towards the application of nanosponges for water treatment and pollutant removal are deliberated, focusing on important challenges and future perspectives.

2. Nanosponges for Water Treatment

Cyclodextrin-based nanosponges exhibit excellent potentials as sorbents for removing heavy metal ions from aqueous solutions, providing magnificent opportunities for eliminating organic/inorganic contaminants from water (Table 2) [30]. For instance, cyclodextrin-based nanosponges (0.7–1.2 nm) are synthesized with a strong affinity for absorbing hazardous organic contaminants in wastewater because of the presence of cyclodextrin cavities that provide a hydrophobic environment [31]. Cyclodextrin-based nanosponges can also be applied for fast removal of pollutants from water (~90%), with an utmost adsorption capability of 2 mg g^{-1} . These nanostructures exhibit great potentials for removing some pharmaceutical contaminants from water, namely carbendazim, diclofenac, sulfamethoxazole, and furosemide [32]. Additionally, polymeric nanobiocomposites with multifunctionality are constructed via a cross-linking oligomerization of cyclodextrin deploying phosphorylated multi-walled carbon nanotubes (MWCNTs) trailed by a sol-gel step to incorporate silver (Ag) and titanium dioxide (TiO_2) nanoparticles. These composites can be deployed as potential filter nanosponges for eliminating pollutants (e.g., organic/inorganic materials and pathogenic microorganisms) from water and wastewater. Fourier-transform infrared (FTIR) analysis illustrates that oxygen-containing groups are subsisted on these composites, and carbamate bearing characteristic peaks (at 1645 cm^{-1}) can also be recognized because of the polymerization reactions [33].

Table 2. Some salient advantages and limitations of materials/nanosystems applied for water treatment and pollutant removal from aqueous solutions.

Materials or Nanosystems	Limitations/Challenges	Advantages	Refs.
(Nano) zeolites	<ul style="list-style-type: none"> - Low capacity of adsorption - Low permeability - Restrictions in the removal of organic contaminants - Complex adsorption mechanism 	<ul style="list-style-type: none"> - Low cost - High surface area - Appropriate for the elimination of various contaminants 	[34]
(Nano) clays	<ul style="list-style-type: none"> - Low capacity of adsorption - pH sensitivity 	<ul style="list-style-type: none"> - Low cost - Appropriate for the elimination of various organic/inorganic contaminants 	[35]
Carbon-based nanomaterials	<ul style="list-style-type: none"> - High cost - Low selectivity - Poor regeneration 	<ul style="list-style-type: none"> - High efficiency for removing pollutants from water - Good capacity of adsorption - Commercialization potentials 	[36]
(Nano)biosorbent materials	<ul style="list-style-type: none"> - pH sensitivity - Needs modifications to improve properties 	<ul style="list-style-type: none"> - Simplicity - Good sorption - High selectivity - High efficiency - Low cost and competitive tactics - Biodegradability 	[37]
Silica-based (nano)materials	<ul style="list-style-type: none"> - Low resistance towards alkaline solutions - Not suitable for the removal of pollution in solutions with pH < 8 	<ul style="list-style-type: none"> - High abundancy - Availability and low cost 	[38]
Nanophotocatalysts	<ul style="list-style-type: none"> - Preparation of nanophotocatalysts - Commercialization/up-scalable production - Safety evaluations 	<ul style="list-style-type: none"> - Low cost - Stability - Recyclability - Eco-friendliness - High sensitivity - High specific surface area - Tunable porous structures 	[37,39]
Nanosponges	<ul style="list-style-type: none"> - Complexation behavior - Commercialization/up-scalable production - Toxicity/biosafety issues 	<ul style="list-style-type: none"> - Low cost - Good biocompatibility - Easy surface functionalization - Bio-absorbent properties 	

Polymer nanosponges constructed from β -cyclodextrin covalently-cross-linked tannic acid through a condensation reaction can be applied for capturing Pb^{2+} from wastewater with high selectivity/sensitivity, and phenolic hydroxyl groups of tannic acid with the capability of effective binding to Pb^{2+} generate stable structures [40]. A significant adsorption saturation competence of $\sim 136.8 \text{ mg g}^{-1}$ for Pb^{2+} was reported at an original strength of 200 mg L^{-1} , with the highest capability of $\sim 116.5 \text{ mg g}^{-1}$. The efficiency of elimination was $\sim 81\%$ Pb^{2+} within 3 min, and the adsorption was completed within ~ 50 min because of plentiful adsorption locations in tannic acid and proficient mass transfer rates maintained by cyclodextrin [40]. Cross-linking polymerization, sol-gel, and amidation reaction techniques are deployed together for producing phosphorylated multi-walled carbon nanotube-cyclodextrin/Ag-doped TiO_2 nanosponges for removing Pb^{2+} and Co^{2+} metal ions from wastewater [41]. One of the crucial factors with significant effects on the adsorption capacities of nanosponges is the pH solution, which can affect speciation, the ionization degree of pollutants, and the surface charge of adsorbents. By increasing the pH of a solution, the efficacy of Pb^{2+} and Co^{2+} removal is enhanced. On the other hand, by temperature increment, adsorption capability is increased with an enhancement in time of contact. The mechanism of adsorption and removal of these metal ions comprises

repulsion-ion exchange, physical adsorption, and electrostatic attraction. The maximum elimination results can be obtained for removing Co^{2+} ($\sim 7.812 \text{ mg/g}$) and Pb^{2+} (~ 35.86) [41].

Liao et al. [42] designed glycopolymer nanosponges from monosaccharides and β -cyclodextrin by combining a Fischer glycosylation, click reaction, and cross-linking reaction for eliminating boron from water. Secondary bonding, including van der Waals forces and hydrogen bonding, among the incorporated saccharides and the adsorbates, can be accountable for vast improvements in adsorption rates and removal capabilities, providing functional candidates for water treatment and seawater desalination [42]. In addition, cyclodextrin-based nanosponges are synthesized for removing organic pollutants from water, including chlorinated disinfection by-products and 2-methylisoborneol (an odor-causing material in water). High absorption efficiency ($\sim 99\%$) can be attained using these polymers with excellent recyclability [43]. β -cyclodextrin-based nanosponges immobilized with magnetic nanoparticles exhibit excellent potentials for capturing and enriching organic micropollutants (the removal efficiency was $\sim 90\%$ within $\sim 1 \text{ min}$), which can be further evaluated for designing sensitive sensing systems (Figure 1). These ultra-rapid and selective sensing strategies with high efficiency for molecule enrichment can be utilized to design smart nanosystems and devices with portability, flexibility, cost-effectiveness, and simplicity advantages [44].

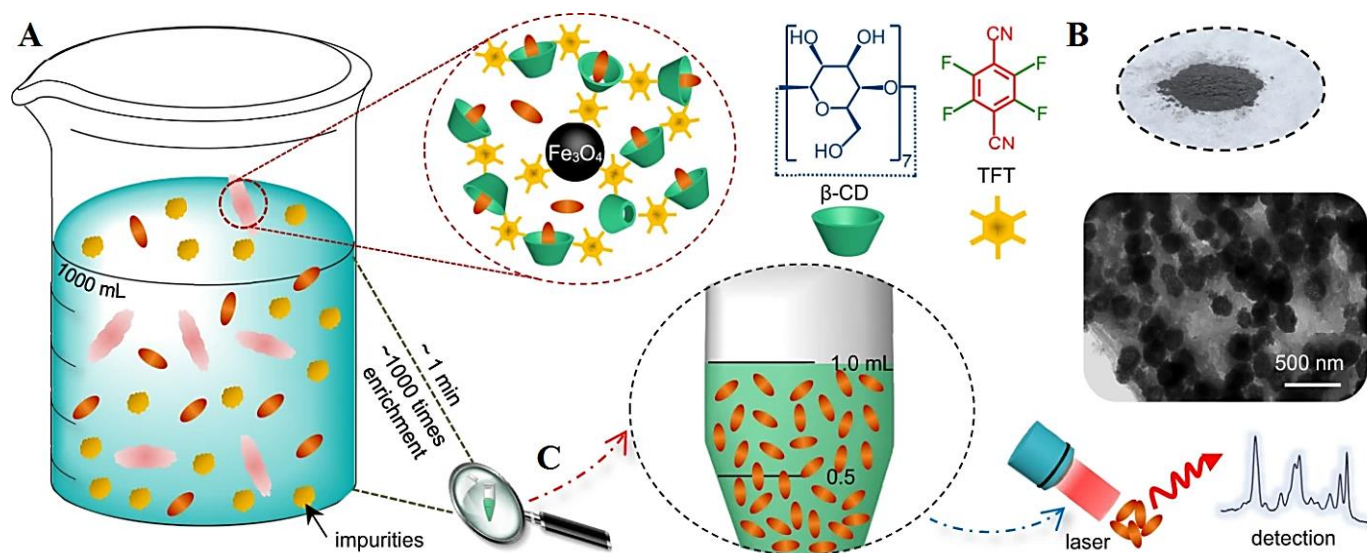


Figure 1. (A–C) The processes of adsorption/desorption by applying magnetic nanoparticle immobilized β -cyclodextrin-based nanosponges with highly efficient enrichment potentials. Adapted with permission from Ref. [44]. Copyright 2021, Springer Nature, CC BY 4.0.

Cyclodextrin-based nanosponges are designed for the elimination of *p*-nitrophenol from aqueous streams through the adsorption process [45], and they are prepared via the reaction of β -cyclodextrin with hexamethylenediisocyanate as the cross-linking agent. Notably, the efficiency of adsorption does not noticeably alter at various temperatures; however, adsorption proficiency can be affected by the concentration and type of cross-linking agents. Consequently, the maximum adsorption energy and capacity for these nanosponges, endowed with high porosity and rigidity, are reported to be 1.837 L mg^{-1} and $\sim 1.0 \text{ mg g}^{-1}$, respectively [45]. In addition, surface-functionalized cis diol comprising mesoporous nanosponges were evaluated for rapid elimination of organic micropollutants and boric acid from water (Figure 2) [46]. These nanosponges exhibit an efficient capacity for boron adsorption, like the commercial resins, in addition to faster adsorption (up to 60 times). Furthermore, bisphenol A can be up-taken by these nanosponges with the equilibrium adsorption achieved in less than two min. The deployment of immobilized polyols can induce synergistic effects because of secondary bonding, simultaneously reducing the time of adsorption and enhancing the capacity of adsorption. Polyol-functionalized

mesoporous nanosponges with high sensitivity/selectivity can be considered promising candidates for water treatment appliances [46].

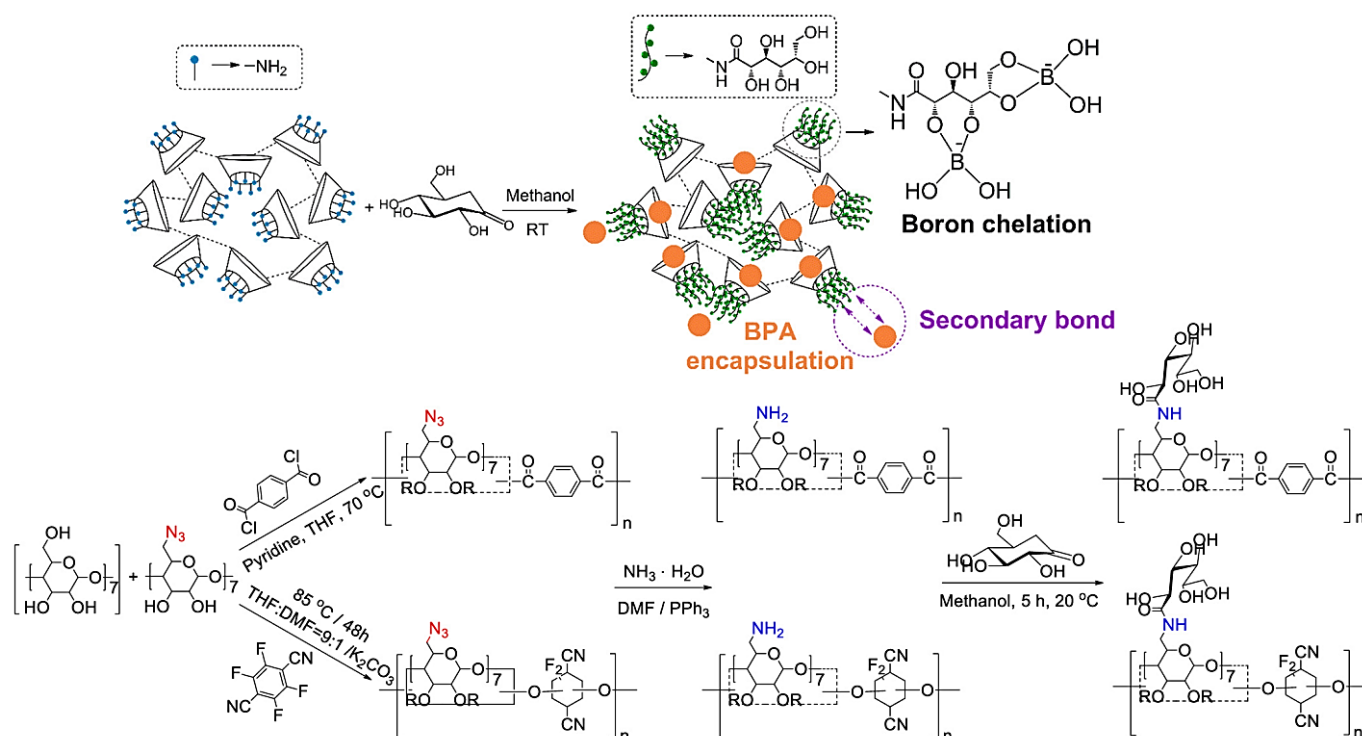


Figure 2. The preparative process for mesoporous cyclodextrin polymers using a copper-free synthesis technique and functionalization of polyols immobilized on their surfaces. Adapted with permission from Ref. [46], Copyright 2019, American Chemical Society.

Nanosponges constructed from β -cyclodextrin polyurethane with insolubility, recyclability, and great surface area ($352.5 \text{ m}^2 \text{ g}^{-1}$) features are additionally altered with phosphorylated MWCNTs and adorned with Ag nanoparticles and TiO_2 via the amidation reaction, cross polymerization deploying diisocyanate (as a linker), and the sol-gel process. The designed nanosponges can adsorb Congo red dye and trichloroethylene from wastewater with maximum capacities of 146.96 mg g^{-1} and $27,507 \text{ mg g}^{-1}$, respectively [29]. Arkas et al. [47] synthesized hyper-branched polymers functionalized with long aliphatic chains for encapsulating/capturing lipophilic polycyclic aromatic pollutants such as pyrene and fluoranthene from water with inclusion formation constants of 2.0×10^8 – $6.3 \times 10^6 \text{ M}^{-1}$ and 3.8×10^6 – $4 \times 10^5 \text{ M}^{-1}$, respectively. Notably, the chemical structure of the parent hyper-branched polymers and the type of polycyclic aromatic compounds can affect the loading capacity of nanosponges for fluoranthene (6 – 31 mg g^{-1}), phenanthrene (15 – 54 mg g^{-1}), and pyrene (6 – 35 mg g^{-1}) [47].

Cyclodextrin nanosponges with reusability are designed via cross linking of 1,2,3,4-butanetetracarboxylic acid with β -cyclodextrin in the company of poly(vinyl alcohol) for the adsorptive elimination of cationic contaminants from water [48]. The maximum adsorption was reported for the removal of paraquat (~ 120.5 , 97.0%), safranin (~ 92.6 , 96.7%), and malachite green ($\sim 64.9 \text{ mg/g}$, 98.3%) [48]. In addition, biodegradable cyclodextrin-based nanosponges with high biosafety features are constructed using a one-step solvothermal technique using β -cyclodextrin and diphenyl carbonate for removing dyestuffs from the waste stream. The highest adsorption capacity for Basic red 46 and Rhodamine B was $\sim 101.43 \text{ mg g}^{-1}$ and 52.33 mg g^{-1} , respectively. The amount of adsorbent, the molar ratio of β -cyclodextrin and diphenyl carbonate, pH solution, initial concentration, and time of contact can affect the efficiency of nanosponges for pollutant removal [49]. Fenyvesi et al. [50] utilized cyclodextrin bead polymers for the removal of micropollutants with $>80\%$ effi-

ciency from effluent. Notably, a correlation between the sorption efficacy and the binding constant of micropollutants and cyclodextrin polymers was reported, revealing that complexes of pollutants with cyclodextrin polymers can play a crucial role in sorption mechanisms [50].

Superparamagnetic Fe₃O₄ nanoparticles are decorated on the surface of β -cyclodextrin-derived carbonate nanosponges with good reusability for the elimination of dinotefuran from water (Figure 3), and the highest adsorption is $\sim 4.53 \times 10^{-3} \text{ mmol g}^{-1}$ [51]. It appears that these magnetic nanosponges can be considered promising candidates for the removal of neonicotinoids from aqueous solutions, with high efficiency, cost-effectiveness, non-toxicity, and reusability. Poly(vinyl alcohol)-cyclodextrin nanosponges are synthesized via citric acid with β -cyclodextrin in the presence of poly(vinyl alcohol). These nanosponges are employed for removing paraquat from water by the adsorption process. Consequently, the maximum adsorption capacity is $\sim 112.2 \text{ mg g}^{-1}$, and the reuse of these nanosponges is $\sim 90.3\%$ for paraquat remediation after five successive cycles [52]. Cataldo et al. [53] prepared cyclodextrin-calixarene nanosponges as sorbents with high efficiency to eliminate Pb²⁺ ions from aqueous solutions. These nanosponges can be considered promising candidates for the elimination of inorganic/organic contaminants, especially after suitable surface functionalization/modification to improve sorbent capabilities [53].

Halloysite clay (halloysite nanotubes) and organic cyclodextrin derivatives are utilized to synthesize inorganic–organic nanosponges through microwave irradiation techniques under solvent-free conditions (Figure 4) [54]. These nanosponges are deployed as nanoadsorbents for removing Rhodamine B and some cationic/anionic dyes. Notably, the pH solution and electrostatic interactions can influence the adsorption procedure. High efficiency of adsorption can be attained for cationic dyes relative to anionic ones, offering these nanosponge hybrids for the adsorption of dyes with high selectivity [54]. Additionally, phosphorylated carbon nanotube- β -cyclodextrin nanosponges are fabricated utilizing hexamethylene diisocyanate as a cross linker. These polymers are decorated with Ag and TiO₂ nanoparticles via a sol-gel technique for eliminating metal ions pollutants from wastewater [55].

Nanosponges with environmentally-benign properties and higher compositions of carboxyl groups are obtained through the cross linking of β -cyclodextrin with linecaps utilizing an aqueous citric acid solution [56]. Pyromellitic nanosponges are fabricated via the reaction of β -cyclodextrin and linecaps with pyromellitic dianhydride in dimethyl sulfoxide. Consequently, pyromellitic nanosponges demonstrate better retention potentials than citrate nanosponges at a metal concentration of 500 ppm. However, these citrate and pyromellitic nanosponges exhibit higher retention potentials ($\sim 94\%$) at low metal concentrations ($\leq 50 \text{ ppm}$). The citrate nanosponge can selectively adsorb significant amounts of heavy metals in the attendance of meddlesome salts from sea water compared to the other constructed nanosponge [56]. Raja et al. [57] synthesized Co₃O₄/NiO nanosponges through a precipitation method for the photocatalytic degradation of rhodamine B and Congo red dye, wherein they exhibited high photocatalytic efficacy, offering them as attractive candidates with superb photocatalytic performance against organic contaminants [57]. It appears that future works should be envisioned for the optimization of conditions (e.g., temperature, pH, amount of adsorbents used, contact time, competing ions, etc.) to improve the wastewater treatment processes. Moreover, detailed characterizations, simple/cost-effective production techniques, and deep molecular analyses as well as biosafety, commercialization, and nanotoxicological evaluations are still awaiting a thorough investigation, especially for translating the ensuing results from laboratory scales to industrial stages [58–60].

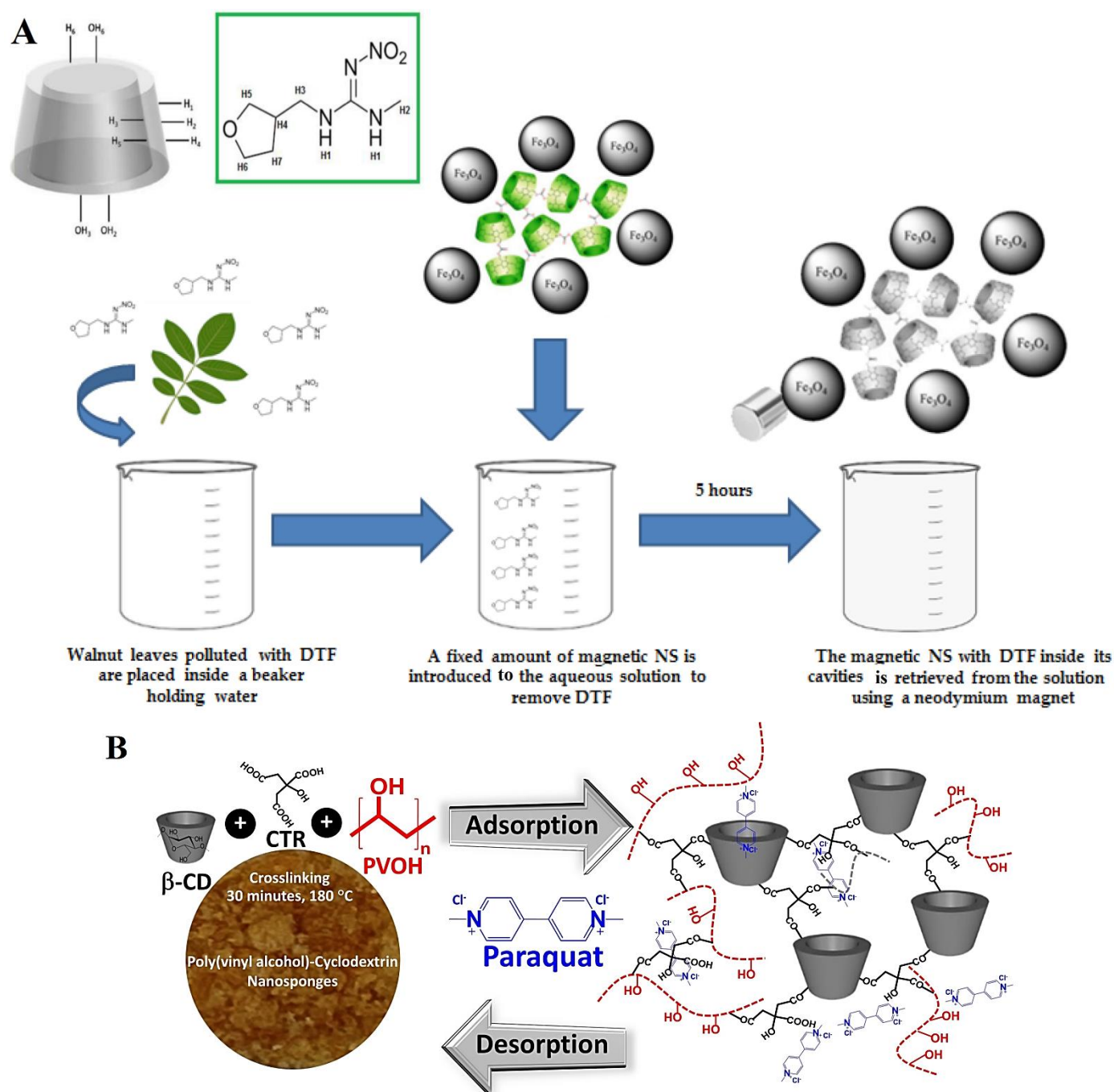


Figure 3. (A) β -cyclodextrin-based nanosponges (NS) are functionalized by magnetic nanoparticles for the removal of dinotefuran (DTF) pollutant. Adapted with permission from Ref. [51]. Copyright 2020, Multidisciplinary Digital Publishing Institute (CC BY 4.0). (B) Poly(vinyl alcohol)-cyclodextrin nanosponges for the adsorption of paraquat (a hazardous agrochemical) from water. CTR: citric acid; β -CD: β -cyclodextrin. Adapted with permission from Ref. [52]. Copyright 2021, Multidisciplinary Digital Publishing Institute (CC BY 4.0).

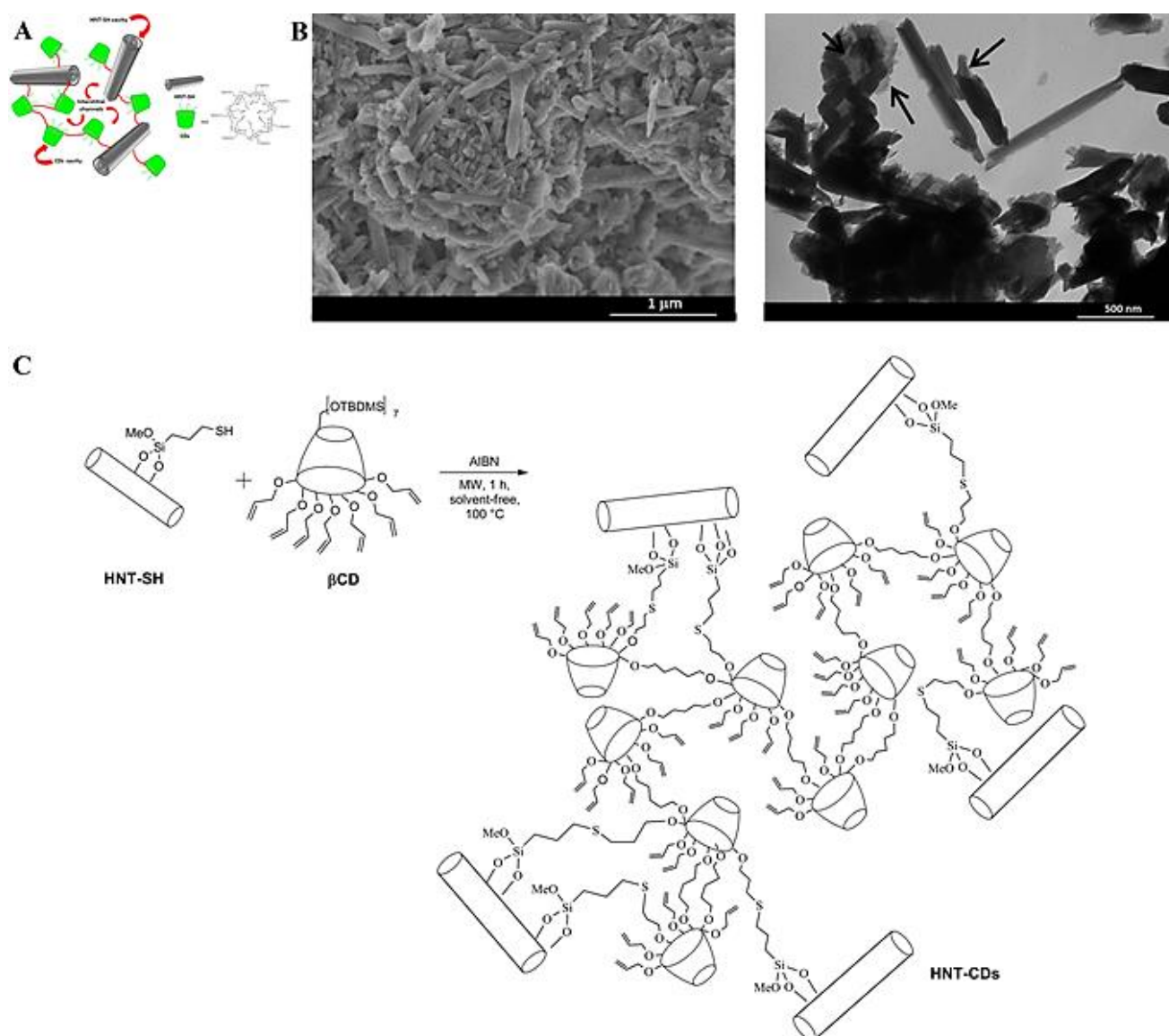


Figure 4. (A) Halloysite nanotubes (HNT)-cyclodextrin derivatives (CDs) nanosponge hybrids. (B) Scanning electron microscopy (SEM) and transmission electron microscopy (TEM) images of the designed nanosponge hybrids. (C) The preparative process of HNT-CDs nanosponge hybrids. Adapted with permission from Ref. [54]. Copyright 2017, American Chemical Society.

3. Conclusions and Future Outlooks

Compared to other nanosystems/materials, nanosponges exhibit high porosity, easy functionalization, simplicity, and cost-effectiveness, which make them attractive candidates for capturing and removing organic/inorganic pollutants/contaminants (e.g., dyes, pharmaceuticals, and heavy metals) from water and wastewater. Notably, cyclodextrin-based nanosponges with unique physicochemical properties and architectures, including good biocompatibility, low/non-toxicity, easy surface functionalization, and bio-absorbent features, can be employed for eliminating a wide variety of pollutants from water via adsorption/inclusion. However, more elaborative studies are warranted to critically evaluate and overcome some crucial aspects/challenges, such as adsorption mechanisms, functionalization processes, removal activities/efficiencies, toxicity/biosafety issues, hydrophobic interactions, electrostatics, and complexation agents for improved targeting of pollutants. Additionally, the type of applied materials (e.g., native/modified polymers), adsorbent dosage, contact time, competing ions, adsorption isotherm models, kinetics, thermodynam-

ics, and reaction/experimental conditions (e.g., molar ratios, temperature, and pH solution) are important factors affecting the adsorption and removal of pollutants using nanosponges. The physicochemical properties, optimization/fabrication processes, complexation behaviors, low-cost production techniques, commercialization, structural variations, and safety evaluations of these nanosystems are essential for their future appliances in water treatment. For antimicrobial and antiviral potentials of these nanosponge-based systems, more elaborative studies are still required for clarifying the mechanism of activities and their efficacy, as well as their long-term biocompatibility and biosafety issues.

Author Contributions: S.I. and R.S.V.: conceptualization, writing-review and editing. All authors have read and agreed to the published version of the manuscript.

Funding: This study was self-funded.

Institutional Review Board Statement: Not applicable.

Informed Consent Statement: Not applicable.

Conflicts of Interest: There are no conflict of interests.

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