

Article

Elevated River Inputs of the Total Alkalinity and Dissolved Inorganic Carbon in the Northern Adriatic Sea

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Abstract: The response of coastal systems to global acidification depends strongly on river inputs, which can alter the total alkalinity (A_T) and dissolved inorganic carbon (DIC) in seawater. The northern Adriatic Sea (NAd) is a shallow continental shelf region that currently receives about 15% of the total freshwater input in the Mediterranean Sea, where the role of riverine discharges on the carbonate system has been poorly studied. In particular, river discharges can alter the carbonate system in the sea, affecting both the equilibrium chemistry and biological processes. For the main rivers flowing into the NAd (the Po, Adige, Brenta, Piave, Livenza, Tagliamento, Isonzo, Timavo and Rižana), data were collected for the pH, concentrations of the total alkalinity (A_T), Ca^{2+} and Mg^{2+} and the isotopic ratio of stable carbon in the dissolved inorganic carbon ($\delta^{13}C_{DIC}$). The DIC fluxes were estimated using the THINCARB (Thermodynamic modeling of INOrganic CARBON) model for the compilation of the A_T and pH data. The results show that the total transport of the A_T in the rivers was 205 Gmol yr^{-1} while the transport of the DIC was 213 Gmol yr^{-1} , of which about 70% was from the Po River. About 97% of the DIC in the river waters was in the form of bicarbonates. The high Mg^{2+}/Ca^{2+} ratios indicate that dolomite weathering is predominant in the Adige, Piave, and Livenza river basins, while lower ratios in the Timavo and Rižana rivers indicate a greater proportion of calcite. The mean $\delta^{13}C_{DIC}$ value was estimated to be $-10.0 \pm 1.7 \text{ ‰}$, a value nowadays considered typical for the DIC flux inputs in oceanic carbon cycle modeling. The DIC flux depends on the mineral weathering and biological activity in each river basin. However, these natural processes can be modified by anthropogenic disturbances that should be better quantified.

Keywords: alkalinity; dissolved inorganic carbon; carbonate system; stable carbon isotope; rivers; runoff; time series; Adriatic Sea; Mediterranean Sea



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1. Introduction

Carbon is continuously displaced along the aquatic land–ocean continuum, which includes freshwaters, estuaries and coastal areas [1,2]. It is transferred in both inorganic (bicarbonates and dissolved CO_2) and organic (dissolved and particulate) forms along these environments. Of the 0.9 Pg C transported annually by rivers to the world ocean, about 0.5 Pg C is organic and 0.4 Pg C is inorganic [3].

In freshwater, dissolved inorganic carbon (DIC) is present in three species: carbonic acid, including dissolved CO_2 (H_2CO_3), bicarbonate (HCO_3^-) and carbonate (CO_3^{2-}) ions. The total alkalinity is defined as $A_T = [HCO_3^-] + 2[CO_3^{2-}]$ + the other weak bases that can be converted to acidic forms at the A_T equivalence point ($pH < 4.5$) through HCl titration [4]. Due to the pH levels in freshwater systems, DIC mainly occurs in the form of bicarbonate ions (HCO_3^-) [5].

CO_2 in the atmosphere and in the soils (i.e., produced by the bacterial decomposition of organic matter) can be converted into organic carbon through biochemical synthesis or removed through chemical rock weathering. This CO_2 is then carried by rivers to the

ocean as particulate organic carbon (POC), dissolved organic carbon (DOC) and DIC. At the same time, continental erosion and rock dissolution cause a natural transfer of carbonates contained in rock from the land to the ocean [6,7].

The chemical processes that occur during the weathering of rocks depend on the rock's composition. The hydrolysis of silicate rocks to form HCO_3^- requires the absorption of CO_2 from an external source (i.e., atmospheric or soil CO_2), whereas the dissolution of carbonate rocks (mainly calcite, CaCO_3 ; and dolomite, $\text{CaMg}(\text{CO}_3)_2$) produces a double amount of HCO_3^- , half from the rock and half from external sources [8]. In the world's rivers, 64% of the total HCO_3^- comes from CO_2 in the soil and 34% from the weathering of carbonate and silicate rocks [6].

On millennial scales, the uptake of atmospheric CO_2 is balanced by carbonate sedimentation in the oceans and by its subsequent return to the atmosphere due to tectonic activity and volcanism [9,10]. However, the natural transport of carbon from the land to the oceans is increasingly affected by human disturbances acting on freshwater ecosystems. These include acid rainfalls enhanced by anthropogenic CO_2 emissions in the atmosphere, air pollution, changes in land use in watersheds and alterations in land surface hydrology [11,12]. For example, it has been shown that the urbanization of river basins can increase the DIC and DOC concentrations in freshwater as a result of water quality degradation, leading to CO_2 partial pressures ($p\text{CO}_2$) in freshwater systems that often exceed the atmospheric equilibrium [13]. These perturbations cause uncertain impacts on the coastal areas and feedbacks on the climate [2,7,14].

In the Mediterranean Sea, the A_T is significantly higher than in the other oceans, and it is linearly correlated with salinity [15]. This correlation is positive in the open sea regions of the Mediterranean where evaporation is higher than precipitation. Negative correlations can be found in marginal areas, where river water strongly contributes to the increase in the A_T [16–20]. The northern Adriatic Sea (NAd) is a marginal sea that receives a relevant contribution of the A_T from rivers [18], which increases its concentration in coastal waters [21,22]. The A_T and DIC of the NAd (A_T mean 2657–2685 $\mu\text{mol kg}^{-1}$; DIC mean: 2350–2378 $\mu\text{mol kg}^{-1}$) [22] are 15–17% higher than the global ocean values (A_T mean: 2302.7 $\mu\text{mol kg}^{-1}$ and DIC mean: 2033.7 $\mu\text{mol kg}^{-1}$) [23]. Therefore, to better assess the fluxes of CO_2 between the atmosphere and sea in northern Adriatic continental shelf, it is essential to quantify the riverine A_T and DIC discharges.

The impact of four rivers (the Isonzo, Timavo, Rižana and Dragonja) on the Gulf of Trieste was investigated by Tamše et al. [24], who estimated a riverine discharge of 1.03×10^{11} g DIC yr^{-1} (8.58 Gmol yr^{-1}) in 2007, with the Isonzo River being the main contributor. Based on the $\delta^{13}\text{C}_{\text{DIC}}$, the authors estimated that the contribution of the rivers to the DIC content in the Gulf of Trieste reached up to 16% during spring. They also pointed out that these rivers had a higher carbonate weathering intensity than the global average due to their calcareous catchments.

Cioce et al. [25] and Fossato [26] found that the A_T was positively correlated with the Ca^{2+} and Mg^{2+} concentrations in the Po River and inversely correlated with the freshwater discharge. Analogous correlations were found for the Adige River [27]. They estimated that the HCO_3^- contributes approx. 90–95% of the A_T .

Despite these publications, river discharges of the A_T , DIC and HCO_3^- in the NAd has been poorly studied, and even recent estimates [18] were still based on historical data that do not reflect the current environmental conditions in this region. In particular, a balance has never been established for the northern shelf area, which is the most affected by riverine inputs. The aim of this study is to (i) create a compilation of river discharge data for the carbonate system parameters in NAd for recent years, (ii) analyze the data in terms of the main processes currently driving these parameters and (iii) estimate the DIC and A_T discharges in the NAd region.

This information is important for the management of valuable river–ocean resources, as well as a better understanding of the underlying processes at local and regional scales for global ocean acidification research.

2. Study Area

The northern Adriatic Sea (NAd), delineated in Figure 1, is a shallow continental shelf with a surface of about 9930 km², a volume of 266 km³ and a mean depth of 27 m [28]. This sea area collects runoff from a catchment area of about 110,600 km², of which 67% of the extent is contributed by the Po River basin [29,30]. The largest freshwater runoff occurs along the west coast, through the Po, Adige and Brenta rivers (84%), while the northern and eastern rivers are less important (10% and 6%, respectively). The NAd drainage basin is a highly industrialized region shared by Italy, Slovenia, Switzerland and Croatia, with large urban settlements, intensive agriculture and a population of about 20 million of inhabitants [29,31,32].

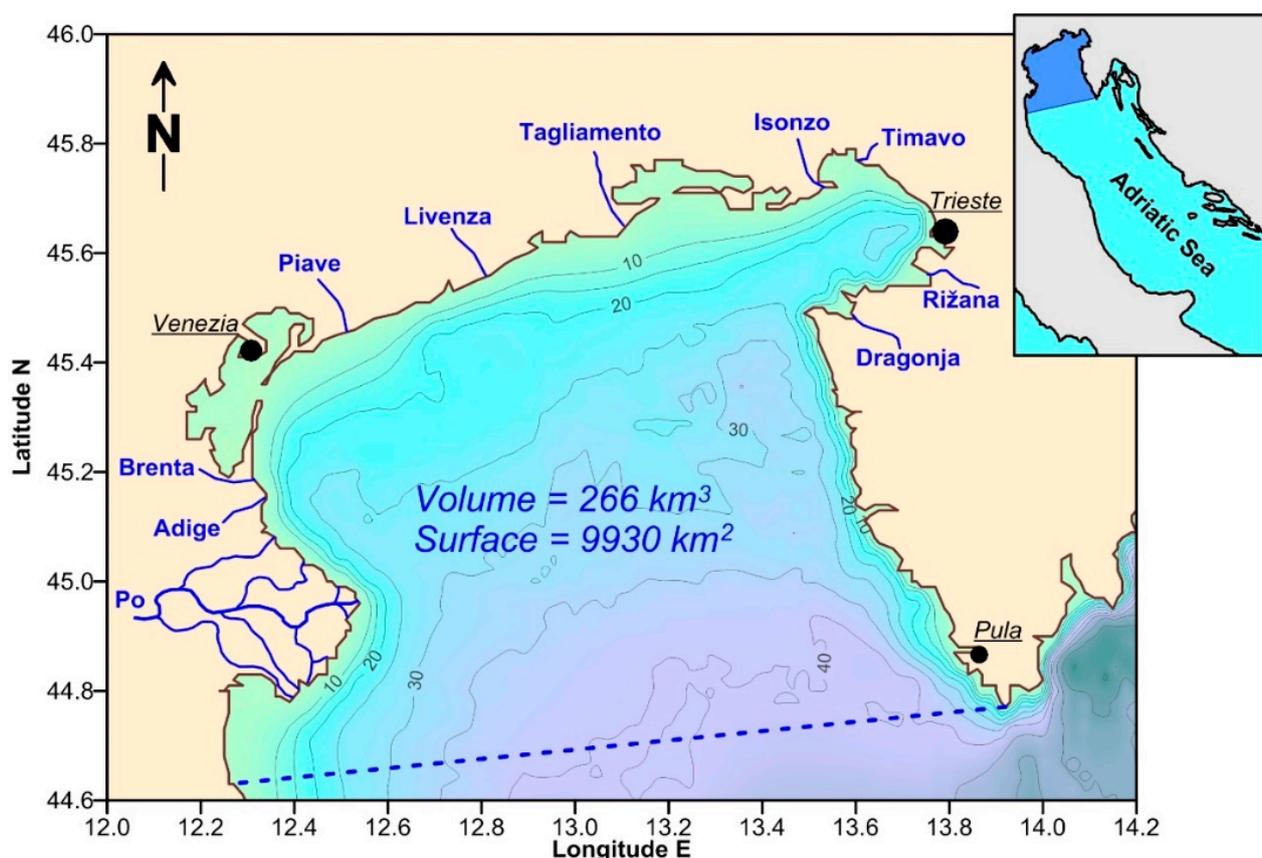


Figure 1. Northern Adriatic Sea (NAd) and the rivers (names in blue fonts) considered in the present study.

In the NAd, the annual cycle of runoff is essentially characterized by two periods of high runoff in the spring and fall, due to seasonal precipitation peaks and snowmelt in the mountainous regions. However, strong interannual and decadal fluctuations of discharge have been observed, with low discharges prevailing in most of the early 2000s and a highly fluctuating discharge during the 2010s [28,33]. These long-term fluctuations in the discharges of the NAd watershed are mainly influenced by the effects of climate change on the precipitation, evapotranspiration and melting of snow and glaciers [34].

River discharges carry a large variety of pollutants, such as heavy metals, organic and inorganic compounds and nutrients, which are dispersed in coastal marine environments. The large discharges of total nitrogen (median of 181 ktons yr⁻¹) and total phosphorus (10 ktons yr⁻¹) are mostly anthropogenic in origin. Nitrogen emissions are primarily due to diffused sources, such as livestock and agricultural activities, while phosphorus emissions are primarily from point sources, such as municipal and industrial wastewater discharges [29,31]. Nutrient loading from rivers have greatly affected the NAd

marine ecosystem in the past, contributing to the occurrence of eutrophic and dystrophic events [32,35].

3. Materials and Methods

3.1. Freshwater Discharges

The freshwater discharges of the NAd rivers were calculated using the daily averaged flow series ($\text{m}^3 \text{s}^{-1}$) provided, as part of their institutional activity, by the Environmental Protection Agencies (ARPA) of the Regions Emilia Romagna, Veneto and Friuli Venezia Giulia (URL: <http://www.arpa.emr.it/>; <http://www.arpa.veneto.it/>; <http://www.arpa.fvg.it/>, accessed on 15 February 2021) and by a water management company (AcegasApsAmga S.p.A., Trieste, Italy; URL: <http://www.acegasapsamga.it/>, accessed on 22 February 2021) for the Timavo River. The daily averaged flows were calculated through the continuous recording of the hydrometric heights in the gauge stations, which were then transformed into the flow using the discharge equations calculated in each river section of interest. These data were used to identify the river regime in concomitance with the sampling of the chemical parameters (Table A1) and to calculate the annual integrated river water discharges (Tables A2 and A3). The gauge stations were located at the closure of each river drainage basin for the Po (Pontelagoscuro station), Adige (Boara Pisani station), Piave (Ponte di Piave station), Livenza (Meduna di Livenza station), Isonzo (Slovene: Soča; Pieris station) and Timavo (Bocche di Timavo station). The water flow at the Brenta River mouth was calculated as the sum of the main stem flow (Barziza station) and its tributaries: the Bacchiglione River (Montegalda station) and the Gorzone Channel (Stanghella station). The long-term series of flow rates were not available at the river mouth for the Tagliamento River. Instead, the discharge was evaluated in the station of Pioverno, located at the end of its mountainous drainage basin [28]. The water flows of the Rižana (Kubed station) and Dragonja (Podkaštel station) rivers were published by the Environmental Agency of the Republic of Slovenia (ARSO; URL: <http://www.arso.gov.si/en/>, accessed on 8 February 2021).

3.2. Physical and Chemical Parameters

The physical and chemical parameters were collected at the same gauging station locations or at points closer to the river mouths (Table S1) to provide the best estimate of the concentrations of the chemical components discharged into the sea. This dataset was derived from a research activity by the Jožef Stefan Institute that monitored the activities of a water management company (AcegasApsAmga S.p.A., Trieste, Italy) and from the ARPA of the Regions Emilia Romagna, Veneto and Friuli Venezia Giulia. The chemical sampling covered all the seasons, with a 23% distribution of the data in winter, 28% in spring, 24% in summer and 26% in autumn.

The temperature, conductivity and pH were measured in the field using portable thermal conductivity meters and pH meters. The pH data provided by the Jožef Stefan Institute were measured using a Corning 315 high-sensitivity portable pH meter equipped with an Orion Ross combination glass electrode and a thermometer. The pH was calibrated using buffer solutions that were traceable to NIST standards with the pH values of 4.01 and 7.01 (Hanna Instruments, Inc., Woonsocket, RI, USA), obtaining a precision of ± 0.02 pH units. The pH data collected during the monitoring activities (ARPA and AcegasApsAmga S.p.A., Italy) were measured according to a comparable reference method [36] using portable pH meters equipped with combination glass electrodes, temperature correction and two calibration buffers, achieving a typical precision of ± 0.02 pH units. All the pH measurements were given on the NBS scale.

The A_T of the river water samples was determined through potentiometric titration, using a pH glass electrode and a titrating solution of HCl 0.05 N standardized against Na_2CO_3 primary standard solutions with a precision of $\pm 1\%$ [24,36]. The concentrations ($\mu\text{mol L}^{-1}$) of calcium (Ca^{2+}) and magnesium (Mg^{2+}) ions in the river water were deter-

mined using Flame Atomic Absorption Spectrometry or inductively coupled plasma optical emission spectroscopy (ICP-OES) [24,36], with an estimated precision of $\pm 2\%$.

The estimates of the DIC and HCO_3^- concentrations in the river water were calculated applying the THINCARB software [37] to the compilation of the A_T and pH data of the main rivers flowing into the NAd (Table S1) during 2010–2019. A total of 880 data points were collected for the A_T and the carbonate system in the rivers.

The samples for the determination of the stable carbon isotope ratio in the dissolved inorganic carbon ($\delta^{13}\text{C}_{\text{DIC}}$) were collected in the main streams of the rivers in evacuated 12-mL Exetainer[®] glass tubes (Labco Limited, Ceredigion, UK) to prevent gas exchange with the atmosphere. The $\Delta^{13}\text{C}_{\text{DIC}}$ was determined by injecting aliquots of the sample into the evacuated septum tubes containing phosphoric acid. The released $\text{CO}_2(\text{g})$ was then analyzed using a continuous-flow IRMS Europa 20–20 (Crowe, UK) with an ANCA-TG module for trace gas separation. To determine the optimal extraction procedure for the water samples, two Na_2CO_3 standard solutions were prepared with a known $\delta^{13}\text{C}_{\text{DIC}}$ value of $-10.8 \pm 0.2\%$ and $-4.1 \pm 0.2\%$, respectively.

3.3. Data Processing

The annual transport of the DIC, HCO_3^- and A_T (F ; mol yr^{-1}) was estimated only for the major rivers for which a complete data series was available (i.e., the Po, Adige, Brenta, Piave and Livenza rivers). The transport was calculated using the equation based on the discharge weighted means of the daily loads

$$F = \left[\sum_{i=1}^n (C_i \cdot Q_i) / \sum_{i=1}^n Q_i \right] \cdot Q_{\text{yr}} \quad (1)$$

where C_i and Q_i are the concentration of the parameter (mol m^{-3}) and the flow ($\text{m}^3 \text{s}^{-1}$) for each day of sampling “ i ” during the year ($n \geq 4$) and Q_{yr} is the annual water discharge ($\text{m}^3 \text{yr}^{-1}$). This equation allows for the best weighting of the data to compensate for biases due to the flow variability in the concomitance of the sampling and the different time resolutions of the chemical and flow data series [28]. For the other chemical parameters, such as the $\text{Mg}^{2+}/\text{Ca}^{2+}$ ratios and the $\delta^{13}\text{C}_{\text{DIC}}$, all the available data were considered, including those referring to the minor rivers.

The average isotopic signature of all NAd rivers ($\delta^{13}\text{C}_{\text{DIC,R}}$) was estimated through weighing the isotopic signature of each river “ k ” ($\delta^{13}\text{C}_{\text{DIC,k}}$) by its annual average DIC flux ($F_{\text{DIC,k}}$), as shown in Equation (2) [38].

$$\delta^{13}\text{C}_{\text{DIC,R}} = \frac{\sum_{k=1}^n F_{\text{DIC,k}} \cdot \delta^{13}\text{C}_{\text{DIC,k}}}{\sum_{k=1}^n F_{\text{DIC,k}}} \quad (2)$$

The importance of the annual discharges of the freshwater A_T , DIC and HCO_3^- by the NAd rivers was derived by comparing these fluxes with the relevant budgets in the NAd basin and calculating their turnover rates [28]. For freshwater, the turnover rate was calculated as the ratio between the annual freshwater discharge and the volume of the NAd sea area, as shown in Figure 1. For the carbonate system parameters, the inputs from the rivers were compared with the budgets of these components in the sea, which were calculated using the reference mean values in seawater: salinity = 37.32 ± 1.32 , $A_T = 2745.4 \pm 34.5 \mu\text{mol L}^{-1}$, DIC = $2428.4 \pm 43.5 \mu\text{mol L}^{-1}$, $\text{HCO}_3^- = 2181.9 \pm 73.6 \mu\text{mol L}^{-1}$, pH in situ = 8.114 ± 0.077 , $[A_T - \text{DIC}] = 321.6 \pm 131.9 \mu\text{mol L}^{-1}$. These average values were determined from a dataset (Acid.it Project; n. data = 377) collected during 16 cruises carried out at bimonthly intervals, from December 2014 to January 2017, along one transect from the east to the west that was representative of both high-salinity offshore waters and fresher waters within the NAd coastal fronts [22].

4. Results and Discussion

4.1. Chemical Characteristics of the NAd River Waters

The rivers of the NAd showed high A_T concentrations, with mean values of $3115 \pm 536 \mu\text{mol L}^{-1}$ for the Po River, the main freshwater source in this continental shelf, to extremely high values in some smaller karst rivers such as the Timavo, Rižana and Dragonja (Table 1 and Table S1). Based on the pH levels (7.50–8.17), the HCO_3^- was the predominant form of dissolved inorganic carbon in the river water ($2613\text{--}5139 \mu\text{mol L}^{-1}$). The Ca^{2+} concentration was in the range of $860\text{--}2131 \mu\text{mol L}^{-1}$, while the Mg^{2+} concentration was in the range of $306\text{--}638 \mu\text{mol L}^{-1}$, except for the Dragonja River, which had very high concentrations of these two elements at the mouth of the river due to saltwater intrusion. The mean values of the $[A_T - \text{DIC}]$ in the river waters were negative for the Po, Brenta, Piave, Timavo and Dragonja, indicating a frequent excess of the DIC with respect to the A_T , which originated from the CO_2 oversaturation in freshwater systems. In the other rivers, the $[A_T - \text{DIC}]$ was slightly positive.

Table 1. Descriptive statistics of the parameters measured in the river waters (data in Table S1, since the 2000s): number of data (N), mean (AV) and standard deviation (SD).

River		pH	A_T $\mu\text{mol L}^{-1}$	DIC $\mu\text{mol L}^{-1}$	HCO_3^- $\mu\text{mol L}^{-1}$	$[A_T - \text{DIC}]$ $\mu\text{mol L}^{-1}$	Ca^{2+} $\mu\text{mol L}^{-1}$	Mg^{2+} $\mu\text{mol L}^{-1}$	$\delta^{13}\text{C}_{\text{DIC}}$ (‰)
Po	N	122	122	122	122	122	121	82	3
	AV	7.78	3115	3205	3023	−89	1456	480	−9.63
	SD	0.32	536	534	531	164	229	88	0.35
Adige	N	239	237	237	237	237	25	25	15
	AV	8.03	2305	2260	2190	46	1056	489	−6.95
	SD	0.21	901	887	865	58	357	361	0.70
Brenta	N	123	123	123	123	123	15	15	2
	AV	7.98	4473	4578	4428	−105	1346	519	−8.95
	SD	0.18	1823	1855	1804	70	272	160	1.20
Piave	n	122	122	122	122	122	15	15	6
	AV	8.05	4271	4352	4224	−80	1632	850	−7.82
	SD	0.17	1549	1561	1522	78	182	486	1.73
Livenza	n	88	88	88	88	88	7	7	−
	AV	8.02	4804	4766	4614	38	1525	683	−
	SD	0.15	1643	1670	1606	131	270	112	−
Tagliamento	N	47	7	7	7	7	44	44	10
	AV	8.05	3461	3454	3369	8	1872	879	−7.13
	SD	0.16	190	193	185	19	207	109	0.48
Isonzo	N	90	77	76	76	76	18	18	5
	AV	7.99	3695	3658	3537	84	1233	375	−8.12
	SD	0.18	548	561	533	419	236	119	1.20
Timavo	N	196	63	63	63	63	149	149	12
	AV	7.49	4016	4408	3950	−393	1843	281	−11.69
	SD	0.20	276	424	277	249	172	60	0.61
Dragonja	N	5	5	5	5	5	4	4	4
	AV	7.50	5247	5875	5144	−628	2405	2578	−11.60
	SD	0.51	959	766	939	761	447	1724	0.27
Rižana	N	5	5	5	5	5	5	5	4
	AV	8.19	4429	4410	4309	19	2044	244	−11.33
	SD	0.09	325	335	326	13	65	25	0.22

The stable isotopic composition of the dissolved inorganic carbon showed significantly lower values for the Dragonja, Rižana and Timavo than for the other major NAd rivers. The values of the Ca^{2+} , Mg^{2+} and HCO_3^- were high compared to the global weighted average

concentrations (370, 140 and 870 $\mu\text{mol L}^{-1}$, respectively) [6]. The Ca^{2+} and Mg^{2+} in the river waters were primarily derived from rock weathering. The Ca^{2+} sources consisted of carbonate rocks, such as calcite and dolomite, with a minor contribution from Ca-silicate minerals. Dolomite is also a source of dissolved Mg^{2+} in the same ratio as Ca^{2+} . Calcite and dolomite occur almost exclusively in sedimentary rocks and contribute, on average, 65% of the Ca^{2+} concentration in river water [6].

4.2. $\text{Mg}^{2+}/\text{Ca}^{2+}$ Molar Ratios in River Waters

The mean $\text{Mg}^{2+}/\text{Ca}^{2+}$ molar ratios ranged from 0.12 to 1.07, indicating a relevant variation in the contribution of calcite and dolomite dissolution in the different watersheds (Figure 2, Table A4). In particular, they showed a greater contribution of calcite-to-rock weathering for the Po and Isonzo rivers compared to the other major NAd rivers, and a very high contribution for the karst rivers such as the Timavo and Rižana. The catchment area of the Po and its tributaries covers a large part of northern Italy and is surrounded by the Alps in the north and the Apennines in the south. The Apennines are the most affected by rock weathering and sediment transport, where many easily erodible marine sedimentary rocks rich in calcite are found [39,40]. On the other hand, the Isonzo catchment develops in the Eastern Alps, where calcite and dolomite rocks are predominant, and in a karst region, where Cretaceous carbonates predominate, before flowing through the plain [41]. The Timavo and Dragonja catchments are smaller and completely embedded in a karst area (Figure 2).

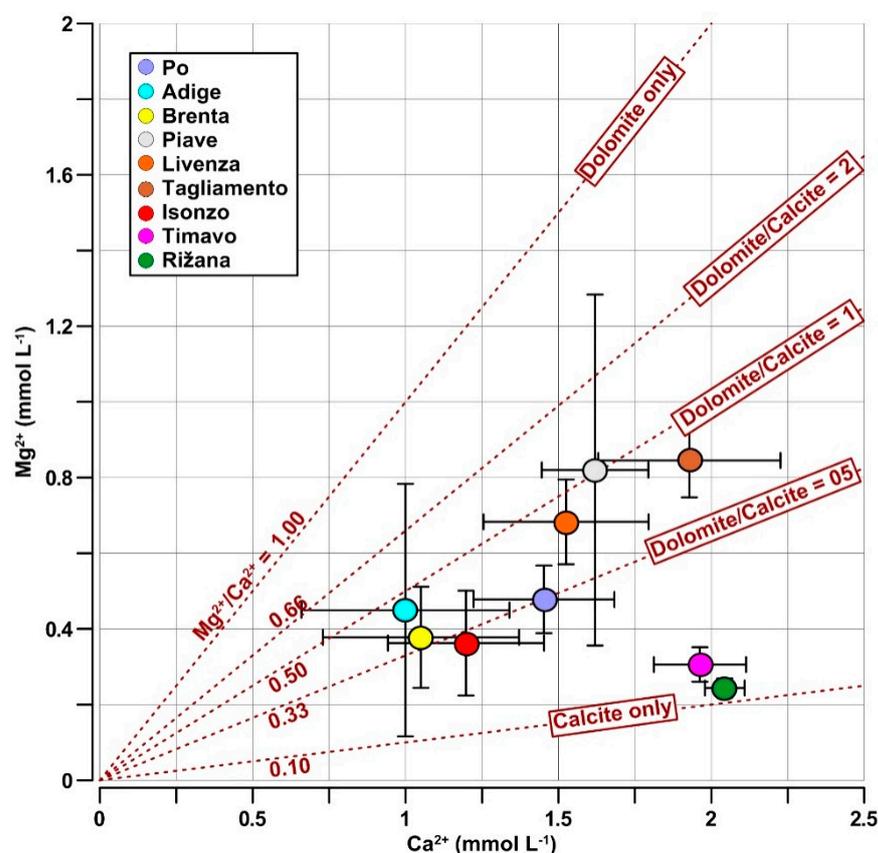


Figure 2. $\text{Mg}^{2+}/\text{Ca}^{2+}$ molar ratios in river water (mean values and standard deviation) compared to theoretical values expected from weathering of carbonate rocks (calcite and dolomite). The data of Dragonja River are not included in the figure because the $\text{Mg}^{2+}/\text{Ca}^{2+}$ ratio (=1.07) was strongly influenced by the intrusion of salt water at the river mouth.

It should be noted, however, that the relationships between Mg^{2+} and Ca^{2+} and HCO_3^- discharges can be further decoupled by other environmental factors. An overflow

of bicarbonates may result from enhanced rock weathering due to acid precipitation, primarily due to the deposition of sulfuric and nitric acids of anthropogenic origin, while outgassing and primary production reduce the HCO_3^- concentrations in river waters relative to the Mg^{2+} and Ca^{2+} concentrations [6]. The relationships between Ca^{2+} and Mg^{2+} relative to HCO_3^- revealed important differences among the NAd rivers. For the major rivers, from the Po to the Isonzo, the Mg^{2+} concentration was linearly related to the HCO_3^- concentration with a ratio of 0.18, indicating a non-negligible contribution of dolomite weathering to the HCO_3^- transport (Figure 3a, Table A1).

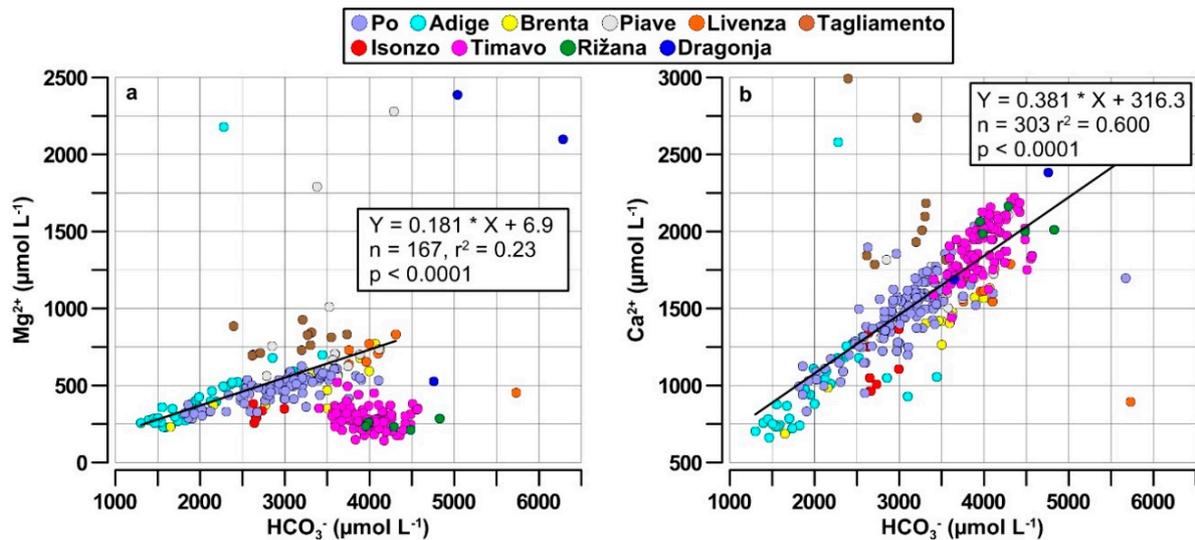


Figure 3. Relationships of the concentrations of (a) Mg^{2+} and (b) Ca^{2+} versus HCO_3^- ($\mu\text{mol L}^{-1}$). Linear regressions were calculated for (a) the rivers from the Po to Isonzo and (b) for all the NAd rivers.

The Timavo, Rižana and Dragonja rivers had low Mg^{2+} concentrations that were not related to HCO_3^- , except for some very high values measured at the mouth of the Dragonja River that were influenced by saltwater intrusion. In contrast, the relationship between Ca^{2+} and HCO_3^- was significant for all the rivers with a ratio of 0.38 (Figure 3b, Table A1). Overall, these ratios indicated simultaneous weathering of dolomite and calcite in the river basins, from the Po to the Isonzo, and predominant calcite weathering in the smaller karst rivers in the easternmost part of the NAd.

4.3. Isotopic Signature and Weathering Intensity of the Riverine DIC

The stable isotopic composition of the DIC in freshwater ($\delta^{13}\text{C}_{\text{DIC}}$) is generally between -25 and 0‰ , depending on the main processes affecting the balance of the bicarbonates in the river basins. The weathering of carbonate rock produces $\delta^{13}\text{C}_{\text{DIC}}$ values around 0‰ , while the respiration of the soil and aquatic organic matter produces negative values (down to -30‰) due to the preferential release of isotopically light carbon (^{12}C) during the remineralization of organic matter. The primary production that preferentially assimilates $^{12}\text{CO}_2$, as well as the equilibration of the freshwater DIC with air CO_2 ($\delta^{13}\text{C}_{\text{DIC}} \approx -8\text{‰}$), generally results in progressively fewer negative $\delta^{13}\text{C}_{\text{DIC}}$ values in the rivers downstream. The $\delta^{13}\text{C}_{\text{DIC}}$ values in freshwater may also change seasonally due to the annual cycles of the biogeochemical processes in the river ecosystems [24,42].

The $\delta^{13}\text{C}_{\text{DIC}}$ data in the NAd rivers fell within a range of values from -12.4 to -5.7‰ (Figure 4, Table A1), and three different groups of data were identified.

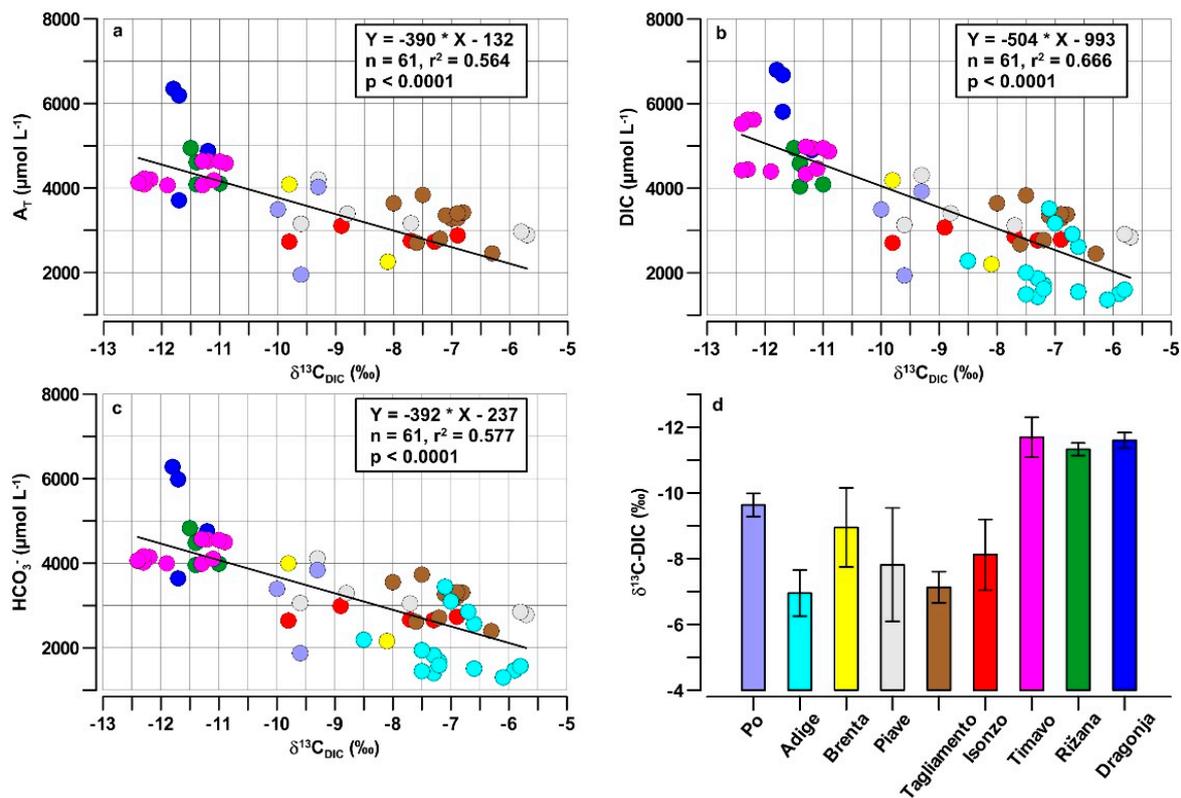


Figure 4. $\delta^{13}\text{C}_{\text{DIC}}$ values (‰) in river water versus the (a) A_T , (b) DIC and (c) HCO_3^- concentrations (μmol L⁻¹). Panel (d) shows the mean and standard deviation of the $\delta^{13}\text{C}_{\text{DIC}}$ for each river.

1. The first group had the highest A_T , DIC and HCO_3^- and the most negative $\delta^{13}\text{C}_{\text{DIC}}$ values, including the Dragonja, Rižana and Timavo. In the small rivers flowing in the karst catchments, the $\delta^{13}\text{C}_{\text{DIC}}$ values from the weathering of carbonate rocks were significantly shifted toward the negative values by the contribution of the remineralization of organic matter, which also produced the highest DIC and HCO_3^- concentrations in freshwater.
2. The second group had a highly variable A_T , DIC and HCO_3^- and medium to low values of the $\delta^{13}\text{C}_{\text{DIC}}$. This group included rivers such as the Brenta, Piave and Isonzo, which were probably influenced by a combination of processes along their river networks, such as the weathering of carbonate rocks, primary production and CO_2 exchange with the atmosphere, which increased the variability of these parameters. Minimal information was available on the $\delta^{13}\text{C}_{\text{DIC}}$ values in the Po River, but the data showed that these were intermediate values compared to the other NAd rivers, at least at the mouth of this large basin.
3. The third group presented a low A_T , DIC and HCO_3^- and the least negative $\delta^{13}\text{C}_{\text{DIC}}$ values from the Adige and Tagliamento Rivers, which might have been due to the predominance of carbonate weathering in all the available sampling sites.

Despite the large variability in the river ecosystem characteristics of the NAd catchment, significant overall relationships were found between the A_T , DIC, HCO_3^- and $\delta^{13}\text{C}_{\text{DIC}}$ (Figure 4a–c). These relationships were typical of sub-tropical temperate rivers, where the large DIC budgets in freshwater result from a variety of factors and not just from the decomposition of fresh organic carbon, as found in tropical rivers [43].

The NAd rivers basins were also characterized by a high runoff intensity compared to their extent, due to high annual precipitation and snowmelt in the surrounding mountain ranges [44]. Roy et al. [45] also found that interrelated factors such as the lithology, water residence time, mechanical erosion, etc. have a higher influence together than individually. Most studies on weathering in alpine regions concluded that enhanced mechanical erosion

in these environments also increases chemical weathering [46–48]. The stress on the mineral and an increased rock surface area create conditions under which minerals are more easily dissolved.

The global theoretical models of the CO_2 consumption in carbonate watersheds showed a value of the A_T near $3000 \mu\text{mol L}^{-1}$ determined from a best-fit line [49]. Although this value was reasonable as an average for all the NAd watersheds, many drainage basins have waters with much higher HCO_3^- concentrations. The NAd river values ranged from 2.0 to 4.5 mmol HCO_3^- per liter of runoff (Figure 5, Table A3).

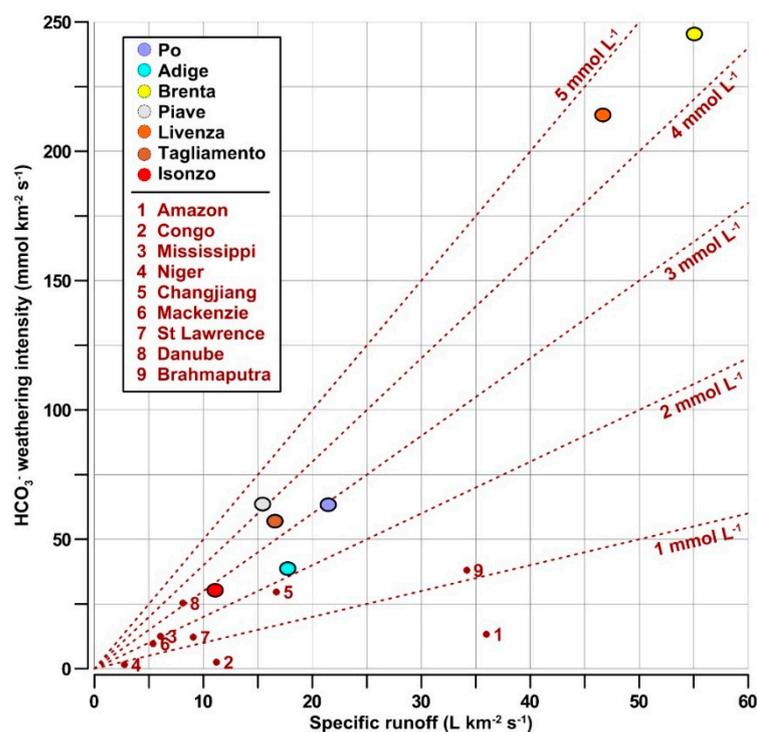


Figure 5. Weathering intensity of the HCO_3^- ($\text{mmol km}^{-2} \text{s}^{-1}$) as a function of the specific runoff ($\text{L km}^{-2} \text{s}^{-1}$) in each drainage basin, in comparison to the most important rivers worldwide (data from [5]).

In particular, the Tagliamento, Piave, Brenta and Livenza rivers showed a high weathering intensity of the HCO_3^- and DIC. This finding was consistent with the fewer negative $\delta^{13}\text{C}_{\text{DIC}}$ values found in some of these rivers, suggesting a stronger contribution of carbonate rock weathering to the total transport of the HCO_3^- . The Po River had a similar ratio for the HCO_3^- weathering intensity to the specific runoff as the Danube ($\sim 3 \text{ mmol L}^{-1}$), although the Danube had a catchment area approx. ten times larger and a median annual flow approx. four times greater [30]. This result confirmed that the largest European rivers have high HCO_3^- and DIC discharges compared to rivers in other continents (Figure 5, Table A3). The DIC weathering intensity is also high in Asian tropical rivers, which have the highest values ($25.85 \text{ mmol C km}^{-2} \text{s}^{-1}$) among tropical rivers worldwide ($8.69 \text{ mmol C km}^{-2} \text{s}^{-1}$) [50]. The DIC concentrations in Asian tropical rivers are lower than in the NAd rivers, possibly due to the abundant vegetation that reduces weathering and greater aquatic photosynthesis [50,51]. However, the contribution of European freshwater discharges should be better evaluated by including the data for medium and small catchments [43,52] and for submarine groundwaters [53,54].

The average isotopic signature of the NAd rivers was estimated to be $-10.0 \pm 1.7\text{‰}$, while the Timavo, Rižana and Dragonja had $-11.6 \pm 0.4\text{‰}$, using the equation in Section 3.3. These data are close to a $\delta^{13}\text{C}_{\text{DIC}}$ signature of -10‰ , which is now considered to be representative of the DIC riverine inputs in ocean carbon cycle modeling [38]. Tropical rivers, such as the Amazon, Congo and Niger, often have strongly negative $\delta^{13}\text{C}_{\text{DIC}}$ values

due to the large contribution of fresh organic carbon decomposition to the freshwater HCO_3^- fluxes. They are also characterized by a rather low HCO_3^- weathering intensity, consistently with the global pattern of continental carbonate rocks and with their high discharges of organic carbon [5,24,43].

The variation in the $\delta^{13}\text{C}_{\text{DIC}}$ signature can be compared to the ratio $[\text{H}_2\text{CO}_3]/[\text{HCO}_3^-]$ in freshwater, as already proposed in the literature [49]. The plot of the $\delta^{13}\text{C}_{\text{DIC}}$ versus the ratio $[\text{H}_2\text{CO}_3]/[\text{HCO}_3^-]$ shows that two groups of rivers can be distinguished (Figure 6).

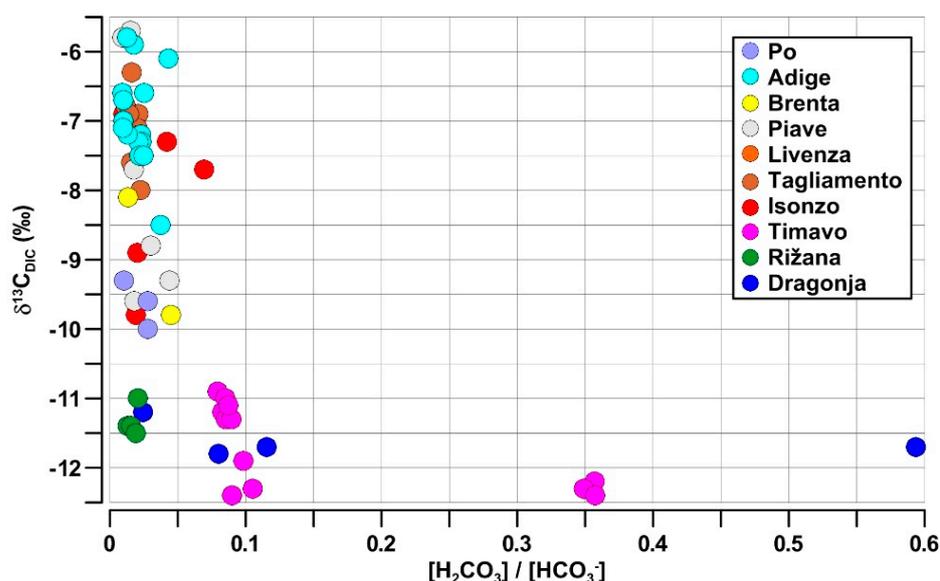


Figure 6. $\delta^{13}\text{C}_{\text{DIC}}$ values (‰) in the NAd rivers as a function of the $[\text{H}_2\text{CO}_3]/[\text{HCO}_3^-]$ ratio in freshwater.

1. Samples from the Timavo Rižana and Dragonja rivers, which show larger variations in the $[\text{H}_2\text{CO}_3]/[\text{HCO}_3^-]$ ratio, ranging between 0 and 0.6 and low values of the $\delta^{13}\text{C}_{\text{DIC}}$ (−11 to −12.5‰), indicating no exchange with atmospheric CO_2 , but the presence of biogenic CO_2 .
2. The second group, with a very low $[\text{H}_2\text{CO}_3]/[\text{HCO}_3^-]$ ratio (<0.1), exhibits greater variations of the $\delta^{13}\text{C}_{\text{DIC}}$ (−10 to −5‰), indicating different biogeochemical processes in the rivers such as an exchange with atmospheric CO_2 , biogenic CO_2 and carbonate weathering.

4.4. Impact of the Riverine A_T and DIC in the NAd

In the 2010s, the annual freshwater discharge of the Po River (27.2–75.1 $\text{km}^3 \text{yr}^{-1}$) was approx. an order of magnitude higher than the other NAd rivers (0.6–10.7 $\text{km}^3 \text{yr}^{-1}$). The year 2014 showed particularly high runoff, while 2017 was the driest year of this decade (Figure 7a, Table A2). For the other rivers, the freshwater discharge was in the order of the Adige > Brenta > Livenza > Piave. The transport of the A_T and carbonate species followed the freshwater discharge, maintaining the same distinction between the Po and the other NAd rivers (Figure 7b–d, Table A2).

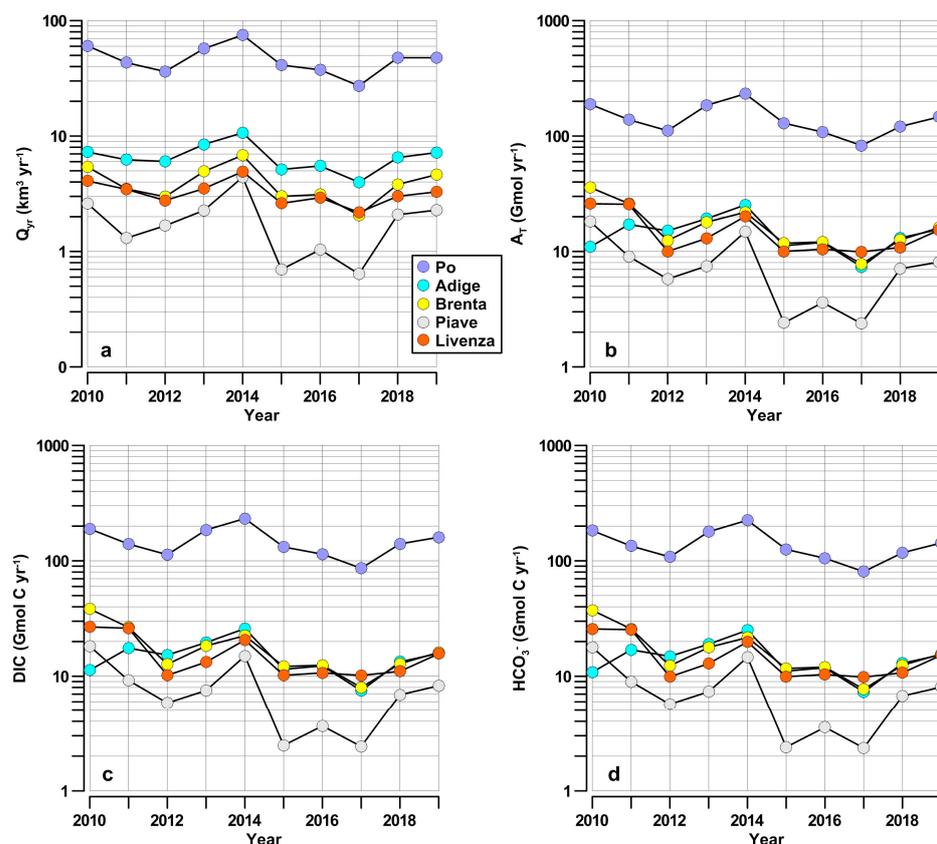


Figure 7. Annual discharges of (a) river water (Q_{yr} ; $\text{km}^3 \text{ yr}^{-1}$), (b) total alkalinity (A_T ; Gmol yr^{-1}), (c) dissolved inorganic carbon (DIC; Gmol C yr^{-1}) and (d) bicarbonate (HCO_3^- ; Gmol C yr^{-1}) in the northern Adriatic by major rivers, in 2010–2019.

The annual transports of the A_T , DIC and HCO_3^- by the Adige, Brenta and Livenza rivers were very similar, despite the differences in the freshwater discharges among these rivers. The Piave River almost always had the lowest freshwater and carbonate loads. The interannual variability of the freshwater and carbonate discharges was pronounced, reaching approx. 65% of these amounts when the extremely wet (2014) and dry (2017) years were compared.

On average, the transport of the A_T , DIC and HCO_3^- through the Po River reached the values of 144 ± 43 , 149 ± 41 and 140 ± 41 Gmol yr^{-1} , respectively in the 2010s (Figure 8, Table A2). The other NAd rivers were in the order of the Brenta, Livenza, Adige and Piave. Despite the limited availability of the data, the Tagliamento and Isonzo rivers were considered to be of minor importance at a basin scale, although they had a significant impact on the carbonate marine system at the sub-regional scale [24,55]. Historical data for the A_T river concentrations in the 1970s were used by Copin Montegut [15] to estimate the A_T discharge in the Mediterranean basin and reused by Cossarini et al. [18] for the Adriatic Sea. These estimates were made only for the alkalinity in a few rivers using old data. Our current estimate is 5–15% higher than that calculated on the basis of the Copin Montegut [15] data, primarily due to a higher A_T concentration in the riverine waters. The NAd rivers contributed to 64% of the total A_T discharged in the Adriatic basin, according to the estimate of Cossarini et al. [18] and to 12% of the total A_T discharged by all Mediterranean rivers, according to the estimate of Copin Montegut [15], or to 23%, according to the estimate of Cossarini et al. [18].

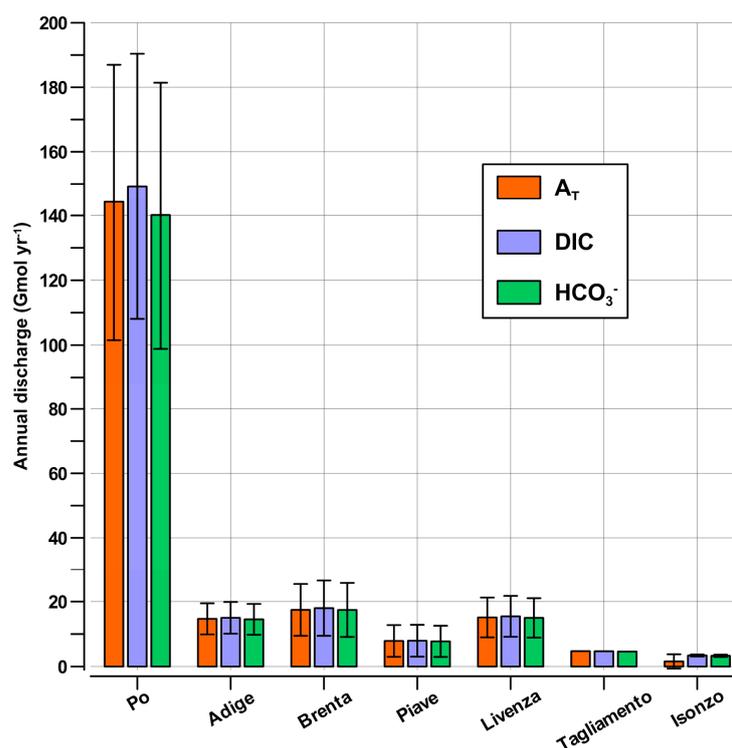


Figure 8. Mean annual transport of A_T, DIC and HCO₃⁻ (Gmol yr⁻¹) from the NAd rivers. Data for the Tagliamento and Isonzo rivers were estimated for 2007 and 2011 only.

The mean DIC flux of the Po was slightly lower than the flux of the Rhône (162 G mol C yr⁻¹) [56], but almost four times lower than the flux of the Danube (629 G mol C yr⁻¹) [5]. The total DIC flux into the NAd represented 13–15% of the total DIC input by rivers in the Mediterranean Sea, estimated by Sempéré et al. [56].

The considered rivers did not represent the total discharge of carbonates in the NAd, which also consisted of small watercourses and freshwater discharges through coastal lagoons and submarine springs along the east coast. Nevertheless, a comparison can be made between the marine budgets of the A_T, DIC and HCO₃⁻ in the NAd by calculating the turnover rates of these components. In the 2010s, the mean freshwater discharge of the NAd rivers was 65.7 km³ yr⁻¹, which represented 24.7% yr⁻¹ of the seawater volume of the region shown in Figure 1 (i.e., 266 km³). These rivers also discharged an average of 205.2 Gmol yr⁻¹ of A_T, 212.8 Gmol yr⁻¹ of DIC and 202.6 Gmol yr⁻¹ of HCO₃⁻. Assuming the typical concentrations in seawater of A_T = 2745.4 ± 34.5 μmol L⁻¹, DIC = 2428.4 ± 44.7 μmol L⁻¹ C and HCO₃⁻ = 2181.9 ± 73.6 μmol L⁻¹ (see Section 3.3), the total budgets of these parameters in the NAd can be estimated to be 729.2, 645.0 and 579.5 Gmol, respectively. These data indicate that river discharges have an important influence on the carbonate system in the NAd, accounting for 28, 33 and 35% yr⁻¹ of the budget of these parameters in the sea, respectively (Table 2).

Table 2. Comparison between the volume of the North Adriatic (NAd) and the annual river discharge, as well as the budgets (Gmol) and mean annual river discharges (Gmol yr⁻¹) of the A_T, DIC and HCO₃⁻. Turnover rates of all the parameters are expressed as percentage per year (% yr⁻¹).

	Water	A _T	DIC	HCO ₃ ⁻
Bulk NAd	266 km ³	729.2 ± 9.2 Gmol	645.0 ± 11.9 Gmol	579.5 ± 19.5 Gmol
River discharge	65.7 ± 24.5 km ³ yr ⁻¹ 24.7 ± 9.2% yr ⁻¹	205.2 ± 75.8 Gmol yr ⁻¹ 28.1 ± 10.7% yr ⁻¹	212.8 ± 72.6 Gmol yr ⁻¹ 33.0 ± 11.9% yr ⁻¹	202.6 ± 72.2 Gmol yr ⁻¹ 35.0 ± 13.6% yr ⁻¹

The Northern Adriatic region under consideration receives the largest runoff in the entire Mediterranean compared to its seawater volume, since it is a shallow continental shelf surrounded by an extremely large catchment area. As a result, the mean annual outflows of the total nitrogen and total phosphorus can be as much as seven times higher than the budgets of these biogenic elements in the NAd [28,29]. The present study showed, for the first time, the importance of continental inputs of the A_T , DIC and HCO_3^- in the context of river discharge, as they account for approx. one-third of the marine budgets of these parameters at the annual scale. However, human activities can alter the terrestrial weathering and land surface hydrology, and thus, the HCO_3^- fluxes in the rivers [14,57]. Pollution, acid rainfall and the use of high N fertilizers can increase the acidification of the soils and promote carbonate dissolution [5].

The potential effect of river discharges on ocean acidification in the NAd marine ecosystem can also be inferred from the composite parameter [$A_T - DIC$], which is a property conservative with the mixing and almost linearly correlated to the changes in the pH and the saturation state of the calcium carbonate [58]. The river waters in the NAd often have higher concentrations of the A_T and DIC than the seawater (Table 1), but their values of the [$A_T - DIC$] (-628 to $46 \mu\text{mol L}^{-1}$) are often lower than the values usually found in seawater ($321.6 \pm 131.9 \mu\text{mol L}^{-1}$). Negative [$A_T - DIC$] values characterized the Po, Brenta, Piave, Timavo and Dragonja, while the Adige, Livenza, Tagliamento Isonzo and Rižana had slightly positive values. Negative values are common to most of the rivers worldwide as they originate through the CO_2 oversaturation in freshwater systems [58]. This feature means that, in these cases, the river discharges can potentially decrease the pH in the NAd coastal waters [7,14,59]. However, it should also be kept in mind that large phytoplankton blooms occur in the river plumes due to the discharge of land-borne nutrients, often reaching eutrophic and hypertrophic conditions in the NAd [28,35]. Primary production can reduce the DIC concentrations, thereby increasing the pH in seawater.

River discharges are also characterized by a multiscale spatio-temporal variability due to the effects of climatic fluctuations and a continuous evolution of anthropogenic pressure in their catchments. For these reasons, more comprehensive monitoring of the carbonate system in freshwater ecosystems should be conducted to better model the future trends and variations of ocean acidification in the coastal areas [1,2].

5. Conclusions

This study shows that river discharge has an important influence on the carbonate balance and dynamics in the NAd, but it also suggests that more comprehensive monitoring of the carbonate system in the river basins should be conducted to better quantify the contributions of smaller watercourses, coastal lagoons, coastal groundwaters and submarine springs. Although the Po River was by far the largest contributor to the A_T and DIC discharges, the impacts of the other sources of continental water cannot be ignored, at least at the sub-regional scale.

The NAd and its watershed is also region that is changing under the influence of climate trends and human activities. For these reasons, the river discharges of the A_T and DIC might be easily altered in the future, having possible unexpected impacts on the seasonal and interannual variations of the carbonate system in the coastal waters.

The comparison between the inorganic carbon parameters, the stable isotopic composition of the DIC and the Ca^{2+} and Mg^{2+} concentrations showed that the carbonate system in the river drainage basins depends on a variety of factors, including rock and soil weathering, biological processes in the river networks, air and freshwater pollution and changes in the river regime. These factors, which are common in the temperate watersheds in industrialized regions, establish the NAd region as an important site for understanding the evolution of river–sea systems.

The following key points were highlighted in this study.

- Strong bicarbonate weathering occurred in these mountainous watersheds.

- Calcite weathering was prevalent in the Po, Isonzo, Timavo and Rižana catchments, while dolomite weathering was more important in the Adige, Piave and Livenza rivers.
- The rivers with an A_T higher than the DIC concentration have the potential to buffer the NAd waters, while those, such as the Po River, have the potential to enhance the acidification process in the marine environment.
- The carbonate weathering intensity in the NAd watershed was among the highest and was approx. four to 35 times higher than the global average ($7 \text{ meq km}^{-2} \text{ s}^{-1}$);
- Global climate models predicted that the precipitation pattern in southern Europe is likely to change. If the rivers in the NAd maintain a relatively unchanged concentration of carbonate mineral weathering products, the carbonate inputs to the coastal areas are expected to depend primarily on the overall changes in freshwater discharge.

Supplementary Materials: The following supporting information can be downloaded at: <https://www.mdpi.com/article/10.3390/w15050894/s1>, Key parameters measured in the NAd rivers are available in the file: Suppl-River-data-Tab S1.xlsx. Table S1: Physical and chemical parameters measured in the northern Adriatic rivers.

Author Contributions: M.G. supervised the data mining and processing, M.G., N.O., S.T. processed the data set, M.G. and S.C. analyzed the data and prepared the figures, M.G., N.O., S.T. and S.C. wrote the paper, M.G. and S.C. performed the final editing of the paper. All authors have read and agreed to the published version of the manuscript.

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Conflicts of Interest: The authors declare that the research was conducted in the absence of any commercial or financial relationship that could be construed as a potential conflict of interest.

Appendix A

Table A1. Physical and chemical parameters, including $\delta^{13}\text{C}_{\text{DIC}}$, measured in river waters (data source: Jožef Stefan Institute).

River	Sample Site	Sampling Date dddd/mm/yy	Water Flow $\text{m}^3 \text{s}^{-1}$	Specific Runoff $\text{L km}^2 \text{s}^{-1}$	T $^{\circ}\text{C}$	pH NBS Scale	EC $\mu\text{S cm}^{-1}$	A_{T} $\mu\text{mol L}^{-1}$	Ca^{2+} $\mu\text{mol L}^{-1}$	Mg^{2+} $\mu\text{mol L}^{-1}$	Cl ⁻ $\mu\text{mol L}^{-1}$	$\delta^{13}\text{C-DIC}$ ‰	DIC $\mu\text{mol L}^{-1}$	HCO_3^- $\mu\text{mol L}^{-1}$	CO_3^{2-} $\mu\text{mol L}^{-1}$	H_2CO_3 $\mu\text{mol L}^{-1}$
Isonzo	Gradisca	30/5/07	50.7	24.9	13.5	7.80	262	2736	1048	306	40	-7.3	2769	2650	7	112
Isonzo	Gradisca	31/10/07	26.4	16.4	13.5	8.05	318	2715	1250	381	50	-	2699	2625	12	62
Isonzo	Gradisca	6/6/08	153.8	63.0	12.8	8.15	257	2732	1340	257	69	-9.8	2704	2639	15	50
Isonzo	Gradisca	26/5/11	2.3	14.8	19.1	8.38	226	2877	1007	334	49	-6.9	2788	2729	31	27
Isonzo	Gradisca	13/10/11	0.4	12.5	14.2	8.11	264	3103	1104	348	50	-8.9	3070	2992	16	61
Isonzo	Pieris	30/5/07	50.7	-	15.0	7.57	274	2748	961	288	49	-7.7	2852	2663	4	185
Isonzo	Pieris	31/10/07	26.4	-	11.3	8.03	336	3082	1365	411	63	-	3081	2990	12	77
Dragonja	Mouth	29/5/07	0.3	3.2	14.3	6.64	7490	3713	1688	5300	54,505	-11.7	5809	3644	1	2164
Dragonja	Mouth	29/10/07	0.1	0.7	12.5	8.06	570	5119	2704	2387	22,085	-	5178	5035	24	117
Dragonja	Mouth	9/6/08	1.2	12.4	19.9	7.98	641	4871	2381	525	1395	-11.2	4897	4757	23	116
Dragonja	Mouth	27/5/11	0.1	0.9	18.6	7.47	2225	6346	2846	2098	17,107	-11.8	6794	6279	9	503
Dragonja	Mouth	13/10/11	0.0	0.1	17.6	7.33	12,560	6185	-	-	-	-11.7	6674	5979	6	690
Rižana	Bridge	23/5/07	0.2	1.0	21.1	8.05	449	4108	1986	259	192	-11.0	4087	3982	23	82
Rižana	Bridge	29/10/07	3.6	17.4	11.4	8.27	429	4397	2162	230	89	-	4377	4281	32	63
Rižana	Bridge	9/6/08	4.0	19.5	14.6	8.30	345	4081	2063	233	104	-11.4	4040	3954	34	51
Rižana	Bridge	27/5/11	0.5	2.5	15.3	8.21	349	4612	1997	212	109	-11.4	4585	4482	32	70
Rižana	Bridge	13/10/11	0.2	0.8	14.9	8.13	432	4948	2011	285	163	-11.5	4946	4825	29	91
Timavo	Spring 1	30/5/07	33.0	87.3	11.9	6.88	454	4200	2104	268	206	-12.2	5625	4143	1	1479
Timavo	Spring 1	31/10/07	10.0	26.5	13.6	7.36	464	4135	2155	282	195	-	4542	4072	4	465
Timavo	Spring 1	10/6/08	23.0	60.8	12.1	7.41	373	4082	2065	267	190	-12.3	4446	4018	4	422
Timavo	Spring 1	26/5/11	15.0	39.7	12.5	7.51	332	4640	1843	352	208	-11.2	4946	4561	6	378
Timavo	Spring 1	13/10/11	12.0	31.7	13.6	7.52	397	4587	1749	380	236	-10.9	4866	4502	6	356
Timavo	Spring 2	30/5/07	33.0	-	11.9	6.89	455	4221	2080	262	203	-12.3	5619	4163	1	1453
Timavo	Spring 2	31/10/07	10.0	-	12.9	7.47	457	4121	2110	279	195	-12.4	4427	4056	5	365
Timavo	Spring 2	10/6/08	23.0	-	12.2	7.4	377	4034	2025	264	190	-	4400	3969	4	426
Timavo	Spring 2	26/5/11	15.0	-	12.4	7.5	318	4631	1823	342	208	-11.0	4945	4552	6	386
Timavo	Spring 2	13/10/11	12.0	-	13.3	7.49	395	4065	1716	368	240	-11.3	4333	3986	5	341
Timavo	Spring 3	30/5/07	33.0	-	11.9	6.88	456	4124	2036	263	203	-12.4	5520	4065	1	1453
Timavo	Spring 3	31/10/07	10.0	-	13.0	7.39	467	4041	2124	280	198	-	4413	3979	4	429
Timavo	Spring 3	10/6/08	23.0	-	15.5	7.41	392	4067	2019	263	190	-11.9	4393	3995	5	392
Timavo	Spring 3	26/5/11	15.0	-	12.0	7.48	321	4642	1839	346	212	-11.3	4980	4564	6	409
Timavo	Spring 3	13/10/11	12.0	-	13.4	7.48	393	4183	1752	372	241	-11.1	4468	4104	5	358

Table A1. Cont.

River	Sample Site	Sampling Date dddd/mm/yy	Water Flow m ³ s ⁻¹	Specific Runoff L km ² s ⁻¹	T °C	pH NBS Scale	EC μS cm ⁻¹	A _T μmol L ⁻¹	Ca ²⁺ μmol L ⁻¹	Mg ²⁺ μmol L ⁻¹	Cl ⁻ μmol L ⁻¹	δ ¹³ C-DIC ‰	DIC μmol L ⁻¹	HCO ₃ ⁻ μmol L ⁻¹	CO ₃ ²⁻ μmol L ⁻¹	H ₂ CO ₃ μmol L ⁻¹
Tagliamento	Invilino	3/6/08	35.0	-	15.8	8.19	598	2452	2993	886	52	-6.3	2449	2393	17	39
Tagliamento	Tolmezzo	18/10/11	38.2	-	12.6	8.15	653	3275	2737	926	88	-7.0	3286	3205	19	60
Tagliamento	Pioverno	2/6/11	23.0	16.0	14.8	8.09	429	3289	1930	729	98	-6.9	3283	3198	17	67
Tagliamento	Pioverno	18/10/11	38.2	11.7	10.5	8.14	492	3352	2007	829	95	-7.1	3351	3267	17	66
Tagliamento	Dignano	3/6/08	38.4	-	14.8	8.22	413	2704	1842	695	203	-7.6	2675	2615	19	40
Tagliamento	Dignano	2/6/11	23.0	-	17.5	8.32	477	3426	2097	762	92	-6.8	3375	3304	32	39
Tagliamento	Dignano	18/10/11	38.2	-	11.3	8.29	527	3408	2183	845	94	-6.9	3384	3311	26	47
Tagliamento	Madrisio	6/6/08	36.0	-	16.0	8.20	428	2803	1785	710	94	-7.2	2771	2709	19	43
Tagliamento	Madrisio	30/5/11	26.1	-	16.2	8.05	433	3836	1917	831	90	-7.5	3836	3733	19	83
Tagliamento	Madrisio	20/10/11	91.0	-	13.6	8.07	494	3643	1817	812	74	-8.0	3644	3546	18	79
Piave	Provagna	17/6/08	100.0	-	14.0	7.94	337	3392	1529	654	80	-8.8	3411	3299	12	99
Piave	Provagna	1/6/11	26.7	-	15.0	8.23	290	2893	1308	562	44	-5.7	2846	2783	20	42
Piave	Provagna	18/10/11	7.3	-	10.6	8.49	429	2969	1817	753	57	-5.8	2910	2849	34	26
Piave	San Dona	6/6/08	168.0	42.8	15.4	8.15	336	3158	1531	542	88	-9.6	3127	3053	19	55
Piave	San Dona	30/5/11	68.0	13.0	16.5	8.15	334	3162	1461	582	77	-7.7	3125	3052	19	54
Piave	San Dona	20/10/11	13.8	9.5	13.2	7.78	457	4201	1724	735	107	-9.3	4303	4111	10	180
Brenta	Valstagna	1/6/11	100.5	-	12.8	8.40	144	1742	686	230	48	-	1685	1651	16	18
Brenta	Valstagna	20/10/11	67.8	-	11.5	8.32	256	2255	985	373	66	-8.1	2203	2157	17	29
Brenta	Limena	30/5/11	104.9	39.9	16.2	8.18	215	2766	963	316	86	-	2713	2651	17	44
Brenta	Limena	20/10/11	67.8	22.0	13.1	7.77	424	4086	1567	592	214	-9.8	4183	3993	10	180
Po	Pollesella	8/6/08	3160.0	45.1	18.8	7.94	286	1949	1054	307	310	-9.6	1932	1872	8	52
Po	Pollesella	30/5/11	1030.0	14.7	23.7	8.33	477	4024	1652	562	621	-9.3	3921	3837	44	40
Po	Pollesella	20/10/11	873.0	12.5	14.5	7.97	469	3490	1518	540	557	-10.0	3500	3392	14	94
Adige	San Michele	17/6/08	404.2	-	10.8	8.05	164	1566	751	287	86	-6.6	1547	1502	6	38
Adige	San Michele	31/5/11	259.3	-	13.7	8.18	158	1533	662	257	82	-5.9	1496	1462	9	26
Adige	San Michele	19/10/11	175.2	-	8.1	8.12	234	1739	867	347	116	-7.2	1716	1670	8	38
Adige	Mattarello	15/6/08	484.7	-	10.8	7.82	151	1351	703	257	79	-6.1	1358	1299	3	56
Adige	Chizzola	15/6/08	484.7	-	11.2	8.08	164	1455	756	275	98	-7.3	1433	1394	6	33
Adige	Chizzola	1/6/11	260.8	-	13.4	8.11	166	1903	741	273	107	-7.3	1870	1822	9	38
Adige	Chizzola	19/10/11	175.2	-	9.6	7.89	269	2266	1010	396	150	-8.5	2280	2192	6	82
Adige	Peri	15/6/08	484.7	-	11.6	8.11	169	1515	780	285	93	-7.5	1490	1451	7	31
Adige	Peri	30/5/11	312.6	-	13.9	8.14	174	1817	720	269	103	-	1780	1737	10	33
Adige	Peri	19/10/11	175.2	-	8.7	8.08	262	2018	941	399	138	-7.5	2000	1943	8	48
Adige	Avisio	17/6/08	404.2	-	12.7	8.32	163	1666	743	228	147	-7.2	1619	1585	13	21
Adige	Avisio	31/5/11	259.3	-	16.3	8.3	169	1647	732	229	117	-5.8	1596	1563	13	20
Adige	Avisio	19/10/11	175.2	-	9.7	8.5	329	2687	1314	449	151	-6.6	2619	2565	30	23
Adige	Ala	15/6/08	484.7	-	12.3	8.44	240	2993	1049	679	29	-6.7	2914	2854	31	28
Adige	Ala	1/6/11	260.8	-	14.2	8.43	216	3257	929	593	27	-7.0	3165	3100	35	30
Adige	Ala	19/10/11	175.2	-	9.5	8.48	293	3601	1055	696	28	-7.1	3511	3440	38	33

Table A2. Annual discharges of freshwater, total alkalinity (A_T), dissolved inorganic carbon (DIC) and bicarbonates (HCO_3^-) by the northern Adriatic rivers. Averages (AV) and standard deviations (SD) are reported.

Year	Freshwater $km^3 yr^{-1}$	A_T $Gmol yr^{-1}$	DIC $Gmol C yr^{-1}$	HCO_3^- $Gmol C yr^{-1}$	Year	Freshwater $km^3 yr^{-1}$	A_T $Gmol yr^{-1}$	DIC $Gmol C yr^{-1}$	HCO_3^- $Gmol C yr^{-1}$
Po River					Piave River				
2010	60.47	188.52	188.99	182.90	2010	2.60	18.16	18.20	17.80
2011	43.26	138.46	139.74	134.43	2011	1.30	9.04	9.21	8.96
2012	36.21	111.32	113.07	108.11	2012	1.67	5.77	5.85	5.71
2013	57.16	184.26	185.38	178.91	2013	2.27	7.46	7.52	7.35
2014	75.09	232.17	232.63	224.85	2014	4.40	14.77	14.95	14.61
2015	41.20	129.24	131.83	125.60	2015	0.70	2.41	2.47	2.39
2016	37.44	108.25	114.35	105.27	2016	1.04	3.61	3.65	3.55
2017	27.23	82.79	86.19	80.65	2017	0.64	2.37	2.42	2.35
2018	47.68	120.92	140.20	117.63	2018	2.08	3.09	7.18	7.01
2019	47.70	146.61	160.06	142.96	2019	2.28	8.10	8.27	8.05
AV	47.34	144.25	149.25	140.13	AV	1.90	7.48	7.97	7.78
SD	13.11	42.73	41.15	41.29	SD	1.06	5.09	4.90	4.80
Adige River					Livenza River				
2010	7.27	10.93	11.30	10.85	2010	4.08	25.92	26.75	25.70
2011	6.23	17.14	17.53	16.96	2011	3.46	25.61	26.07	25.37
2012	6.03	15.08	15.27	14.88	2012	2.76	10.00	10.18	9.90
2013	8.47	19.20	19.61	19.00	2013	3.50	12.98	13.27	12.88
2014	10.67	25.29	25.90	25.06	2014	4.92	20.11	20.60	19.86
2015	5.14	11.23	11.50	11.09	2015	2.61	10.02	10.19	9.92
2016	5.51	11.96	12.31	11.85	2016	2.92	10.46	10.70	10.37
2017	3.98	7.35	7.50	7.29	2017	2.18	9.92	10.17	9.84
2018	6.51	13.10	13.34	12.97	2018	3.01	10.84	11.03	10.71
2019	7.19	15.55	15.79	15.36	2019	3.29	15.34	15.67	15.22
AV	6.70	14.68	15.00	14.53	AV	3.27	15.12	15.46	14.98
SD	1.77	4.79	4.89	4.74	SD	0.74	6.13	6.30	6.07

Table A2. Cont.

Year	Freshwater km ³ yr ⁻¹	A _T Gmol yr ⁻¹	DIC Gmol C yr ⁻¹	HCO ₃ ⁻ Gmol C yr ⁻¹	Year	Freshwater km ³ yr ⁻¹	A _T Gmol yr ⁻¹	DIC Gmol C yr ⁻¹	HCO ₃ ⁻ Gmol C yr ⁻¹
Brenta River									
2010	5.39	35.91	38.28	37.27					
2011	3.49	25.98	26.52	25.74					
2012	2.98	12.45	12.71	12.33					
2013	4.93	17.92	18.35	17.79					
2014	6.82	21.87	22.46	21.68					
2015	3.00	11.77	12.18	11.68					
2016	3.11	12.09	12.45	12.01					
2017	2.06	7.79	7.98	7.71					
2018	3.81	12.60	12.86	12.48					
AV	3.95	17.60	18.20	17.63					
SD	1.40	8.40	8.96	8.74					

Table A3. Comparison of different river fluxes and weathering intensities of dissolved inorganic carbon and bicarbonates in this study (2010–2019) and in the literature.

River Name	Basin Area km ²	Water Flow km ³ yr ⁻¹	Specific Runoff L km ⁻² s ⁻¹	Carbonate in Soil %	Basin Latitude Degree	HCO ₃ ⁻ μmol L ⁻¹	DIC Flux Gmol yr ⁻¹	HCO ₃ ⁻ Flux Gmol yr ⁻¹	HCO ₃ ⁻ Weathering Intensity mmol km ⁻² s ⁻¹	DIC Weathering Intensity mmol km ⁻² s ⁻¹	Reference
Po	70,091	47.30	21.4	-	45 N	2929	148.0	139.82	63.25	66.98	This study
Adige	11,954	6.70	17.8	-	45 N	2131	15.0	14.53	38.55	39.80	This study
Brenta	2280	3.95	55.0	-	45 N	3325	18.2	17.63	245.23	253.11	This study
Piave	3899	1.90	15.4	-	46 N	3629	8.0	7.78	63.27	64.84	This study
Livenza	2222	3.27	46.7	-	46 N	4309	15.5	14.98	213.73	220.66	This study
Tagliamento	2582	1.35	16.5	-	46 N	3231	4.7	4.62	56.70	57.89	This study
Isonzo	3452	1.20	11.0	-	46 N	2723	3.4	3.30	30.29	31.30	This study
Amazon	5,854,000	6642	36.0	3.9	2 S	369	2450	2450.90	13.28	13.27	[5]
Congo	3,699,000	1308	11.2	10.1	4 S	224	293	292.99	2.51	2.51	[5]
Mississippi	3,203,000	610	6.0	18.1	36 N	2074	1265	1265.14	12.52	12.52	[5]
Niger	2,240,000	193	2.7	6.3	10 N	550	106	106.15	1.50	1.50	[5]
Changjiang	1,794,000	944	16.7	44.0	30 N	1780	1680	1680.32	29.70	29.69	[5]
Mackenzie	1,713,000	290	5.4	20.6	64 N	1800	522	522.00	9.66	9.66	[5]
St Lawrence	1,267,000	363	9.1	24.9	47 N	1339	486	486.06	12.16	12.16	[5]
Danube	788,000	202	8.1	14.5	48 N	3115	629	629.23	25.32	25.31	[5]
Brahmaputra	583,000	628	34.2	33.8	25 N	1114	700	699.59	38.05	38.07	[5]
Asian tropical rivers	11,342,854	8694	24.3	-	30 N-30 S	1064 *	9241	-	-	25.85	[57]

* Concentration of DIC.

Table A4. Calcium and magnesium concentrations (mmol L^{-1}) in the riverine waters. Average (AV), standard deviation (SD) and number of measurements (N) are reported.

River	Period	N	Ca ²⁺ AV	Ca ²⁺ SD	Mg ²⁺ AV	Mg ²⁺ SD
			mmol L^{-1}	mmol L^{-1}	mmol L^{-1}	mmol L^{-1}
Po	2010–2019	82	1.45	0.23	0.48	0.09
Adige	2010–2019	28	1.00	0.34	0.45	0.33
Brenta	2010–2018	15	1.05	0.32	0.38	0.13
Piave	2010–2019	15	1.62	0.17	0.82	0.46
Livenza	2010–2019	7	1.52	0.27	0.68	0.11
Tagliamento	2008–2019	47	1.93	0.30	0.85	0.10
Isonzo	2007; 2008; 2011; 2016	18	1.20	0.26	0.36	0.14
Timavo	1998–2003; 2006–2012; 2016	173	1.96	0.15	0.31	0.05
Rižana	2007, 2008, 2011	5	2.04	0.06	0.24	0.03
Dragonja	2007, 2008, 2011	3	2.40	0.45	2.58	1.72

References

1. Probst, J.L.; Mortatti, J.; Tardy, Y. Carbon river fluxes and weathering CO₂ consumption in the Congo and Amazon river basins. *Appl. Geochem.* **1994**, *9*, 1–13. [\[CrossRef\]](#)
2. Salisbury, J.; Vandemark, D.; Hunt, C.; Campbell, J.; Jonsson, B.; Mahadevan, A.; McGillis, W.; Xue, H. Episodic riverine influence on surface DIC in the coastal Gulf of Maine. *Estuar. Coast. Shelf Sci.* **2009**, *82*, 108–118. [\[CrossRef\]](#)
3. Drake, T.W.; Raymond, P.A.; Spencer, R.G.M. Terrestrial carbon inputs to inland waters: A current synthesis of estimates and uncertainty. *Limnol Oceanogr. Lett.* **2018**, *3*, 132–142. [\[CrossRef\]](#)
4. Dickson, A.G. The development of the alkalinity concept in marine chemistry. *Mar. Chem.* **1992**, *40*, 49–63. [\[CrossRef\]](#)
5. Cai, W.-J.; Guo, X.; Chen, C.-T.A.; Dai, M.; Zhang, L.; Zhai, W.; Lohrenz, S.E.; Yin, K.; Harrison, P.J.; Wang, Y. A comparative overview of weathering intensity and HCO₃⁻ flux in the world's major rivers with emphasis on the Changjiang, Huanghe, Zhujiang (Pearl) and Mississippi Rivers. *Cont. Shelf Res.* **2008**, *28*, 1538–1549. [\[CrossRef\]](#)
6. Berner, E.K.; Berner, R.A. *Global Environment. Water, Air, and Geochemical Cycles*; Princeton University Press: Princeton, NJ, USA, 2012; p. 444. ISBN 9780691136783.
7. Kitidis, V.; Shutler, J.D.; Ashton, I.; Warren, M.; Brown, I.; Findlay, H.; Hartman, S.E.; Sanders, R.; Humphreys, M.; Kivimäe, C.; et al. Winter weather controls net influx of atmospheric CO₂ on the north-west European shelf. *Sci. Rep.* **2019**, *9*, 20153. [\[CrossRef\]](#)
8. Amiotte Suchet, P.; Probst, J.L.; Ludwig, W. Worldwide distribution of continental rock lithology: Implications for the atmospheric/soil CO₂ uptake by continental weathering and alkalinity river transport to the oceans. *Glob. Biogeochem. Cy.* **2003**, *17*, 1038. [\[CrossRef\]](#)
9. Ludwig, W.; Amiotte Suchet, P.; Munhoven, G.; Probst, J.L. Atmospheric CO₂ consumption by continental erosion: Present-day controls and implications for the last glacial maximum. *Glob. Planet. Chang.* **1998**, *16–17*, 95–108. [\[CrossRef\]](#)
10. Ludwig, W.; Amiotte Suchet, P.; Probst, J.-L. Enhanced chemical weathering of rocks during the last glacial maximum: A sink for atmospheric CO₂? *Chem. Geol.* **1999**, *159*, 147–161. [\[CrossRef\]](#)
11. Barnes, R.T.; Raymond, P.A. The contribution of agricultural and urban activities to inorganic carbon fluxes within temperate watersheds. *Chem. Geol.* **2009**, *266*, 318–327. [\[CrossRef\]](#)
12. Xuan, Y.; Cao, Y.; Tang, C.; Li, M. Changes in dissolved inorganic carbon in river water due to urbanization revealed by hydrochemistry and carbon isotope in the Pearl River Delta, China. *Environ. Sci. Pollut. Res.* **2020**, *27*, 24542–24557. [\[CrossRef\]](#) [\[PubMed\]](#)
13. das Neves Lopes, M.; Decarli, C.J.; Pinheiro-Silva, L.; Lima, T.C.; Leite, N.K.; Petrucio, M.M. Urbanization increases carbon concentration and pCO₂ in subtropical streams. *Environ. Sci. Pollut. Res.* **2020**, *27*, 18371–18381. [\[CrossRef\]](#) [\[PubMed\]](#)
14. Raymond, P.A.; Hamilton, S.K. Anthropogenic influences on riverine fluxes of dissolved inorganic carbon to the oceans. *Limnol. Oceanogr. Letters.* **2018**, *3*, 143–155. [\[CrossRef\]](#)
15. Copin-Montegut, C. Alkalinity and carbon budgets in the Mediterranean Sea. *Glob. Biogeochem. Cy.* **1993**, *7*, 915–925. [\[CrossRef\]](#)
16. Schneider, A.; Wallace, D.W.R.; Körtzinger, A. Alkalinity of the Mediterranean Sea. *Geophys. Res. Lett.* **2007**, *34*, L15608. [\[CrossRef\]](#)
17. Álvarez, M.; Sanleón-Bartolomé, H.; Tanhua, T.; Mintrop, L.; Luchetta, A.; Cantoni, C.; Schroeder, K.; Civitarese, G. The CO₂ system in the Mediterranean Sea: A basin wide perspective. *Ocean Sci.* **2014**, *10*, 69–92. [\[CrossRef\]](#)
18. Cossarini, G.; Lazzari, P.; Solidoro, C. Spatiotemporal variability of alkalinity in the Mediterranean Sea. *Biogeosciences* **2015**, *12*, 1647–1658. [\[CrossRef\]](#)
19. Rivaro, P.; Messa, R.; Massolo, S.; Frache, R. Distributions of carbonate properties along the water column in the Mediterranean Sea: Spatial and temporal variations. *Mar. Chem.* **2010**, *121*, 236–245. [\[CrossRef\]](#)
20. Ingrosso, G.; Giani, M.; Cibic, T.; Karuza, A.; Kralj, M.; Del Negro, P. Carbonate chemistry dynamics and biological processes along a river-sea gradient (Gulf of Trieste, northern Adriatic Sea). *J. Mar. Syst.* **2016**, *155*, 35–49. [\[CrossRef\]](#)

21. Luchetta, A.; Cantoni, C.; Catalano, G. New observations of CO₂ induced acidification in the Northern Adriatic Sea, over the last quarter century. *Chem. Ecol.* **2010**, *26*, 1–17. [[CrossRef](#)]
22. Urbini, L.; Ingrassio, G.; Djakovac, T.; Piacentino, S.; Giani, M. Temporal and spatial variability of the CO₂ system in a riverine influenced area of the Mediterranean Sea, the northern Adriatic. *Front. Mar. Sci.* **2020**, *7*, 679. [[CrossRef](#)]
23. Feely, R.A.; Doney, S.C.; Cooley, S.R. Ocean Acidification: Present conditions and future changes in a high-CO₂ World. *Oceanography* **2009**, *22*, 36–47. [[CrossRef](#)]
24. Tamše, S.; Ogrinc, N.; Walter, L.M.; Turk, D.; Faganeli, J. River Sources of Dissolved Inorganic Carbon in the Gulf of Trieste (N Adriatic): Stable Carbon Isotope Evidence. *Estuaries Coast.* **2014**, *38*, 151–164. [[CrossRef](#)]
25. Cioce, F.; Stocco, G.; Toniolo, R. Hydrological and physical–chemical investigations on the Po river at Posella. February 1973–February 1975. *Atti Dell’istituto Veneto Di Sci. Lett. Ed Arti* **1977**, *135*, 119–132.
26. Fossato, V.U. Hydrological, chemical and physical investigations on the Po river at Posella. June 1968–June 1970. *Archo Oceanogr. Limnol.* **1971**, *17*, 125–139.
27. Fossato, V.U. Hydrological, chemical and physical investigations on the Adige river at Boara Pisani. June 1968–June 1970. *Archo Oceanogr. Limnol.* **1971**, *17*, 105–123.
28. Cozzi, S.; Giani, M. River water and nutrient discharges in the Northern Adriatic Sea: Current importance and long term changes. *Cont. Shelf Res.* **2011**, *31*, 1881–1893. [[CrossRef](#)]
29. Volf, G.; Atanasova, N.; Kompare, B.; Ožanić, N. Modelling nutrient loads to the northern Adriatic. *J. Hydrol.* **2013**, *504*, 182–193. [[CrossRef](#)]
30. Cozzi, S.; Ibáñez, C.; Lazar, L.; Raimbault, P.; Giani, M. Flow regime and nutrient-loading trends from the largest South European watersheds: Implications for the productivity of Mediterranean and Black Sea’s coastal areas. *Water* **2019**, *11*, 1. [[CrossRef](#)]
31. Viaroli, P.; Soana, E.; Pecora, S.; Laini, A.; Naldi, M.; Fano, E.A.; Nizzoli, D. Space and time variations of watershed N and P budgets and their relationships with reactive N and P loadings in a heavily impacted river basin (Po River, Northern Italy). *Sci. Total Environ.* **2018**, *639*, 1574–1587. [[CrossRef](#)]
32. Zhang, Q.; Cozzi, S.; Palinkas, C.; Giani, M. Recent status and long-term trends in freshwater discharge and nutrient inputs. In *Coastal Ecosystems in Transition: A Comparative Analysis of the Northern Adriatic and Chesapeake Bay*, 1st ed.; Geophysical Monograph; Malone, T., Malej, A., Faganeli, J., Eds.; Wiley & Sons Ltd.: Hoboken, NJ, USA, 2021; p. 256. ISBN 978-1-119-54358-9.
33. Grilli, F.; Accoroni, S.; Acri, F.; Bernardi Aubry, F.; Bergami, C.; Cabrini, M.; Campanelli, A.; Giani, M.; Guicciardi, S.; Marini, M.; et al. Seasonal and interannual trends of oceanographic parameters over 40 years in the northern Adriatic Sea in relation to nutrient loadings from EMODnet Chemistry data portal. *Water* **2020**, *12*, 2280. [[CrossRef](#)]
34. Ravazzani, G.; Barbero, S.; Salandin, A.; Senatore, A.; Mancini, M. An integrated hydrological model for assessing climate change impacts on water resources of the upper Po River basin. *Water Resour. Manage.* **2015**, *29*, 1193–1215. [[CrossRef](#)]
35. Giani, M.; Djakovac, T.; Degobbis, D.; Cozzi, S.; Solidoro, C.; Fonda Umani, S. Recent changes in the marine ecosystems of the northern Adriatic Sea. *Estuar. Coast. Shelf Sci.* **2012**, *115*, 1–13. [[CrossRef](#)]
36. APAT, IRSA-CNR. *Metodi analitici per le acque*; Manuali e guide APAT, 29/2003; IRSA-CNR: Brugheri, Italy, 2003; Volume 2, pp. 493–839. ISBN 88-448-0083-7.
37. Jarvie, P.; King, S.M.; Neal, C. Inorganic carbon dominates total dissolved carbon concentrations and fluxes in British rivers: Application of the THINCARB model—Thermodynamic modelling of inorganic carbon in freshwaters. *Sci. Tot. Environ.* **2017**, *575*, 496–512. [[CrossRef](#)] [[PubMed](#)]
38. Brunet, F.; Gaiero, D.; Probst, J.L.; Depetris, P.J.; Gauthier Lafaye, F.; Stille, P. $\delta^{13}\text{C}$ tracing of dissolved inorganic carbon sources in Patagonian rivers (Argentina). *Hydrol. Process.* **2005**, *19*, 3321–3344. [[CrossRef](#)]
39. Ravaioli, M.; Alvisi, F.; Menegazzo Vitturi, L. Dolomite as a tracer for sediment transport and deposition on the northwestern Adriatic continental shelf (Adriatic Sea, Italy). *Cont. Shelf Res.* **2003**, *23*, 1359–1377. [[CrossRef](#)]
40. Tesi, T.; Miserocchi, S.; Acri, F.; Langone, L.; Boldrin, A.; Hatten, J.A.; Albertazzi, S. Flood-driven transport of sediment, particulate organic matter, and nutrients from the Po River watershed to the Mediterranean Sea. *J. Hydrol.* **2013**, *498*, 144–152. [[CrossRef](#)]
41. Szramek, K.; Walter, L.M.; Kanduč, T.; Ogrinc, N. Dolomite versus Calcite weathering in hydrogeochemically diverse watersheds established on bedded carbonates (Sava and Soča Rivers, Slovenia). *Aquat. Geochem.* **2011**, *17*, 357–396. [[CrossRef](#)]
42. Barešić, J.; Horvatinčić, N.; Roller-Lutz, Z. Spatial and seasonal variations in the stable C isotope composition of dissolved inorganic carbon and in physico-chemical water parameters in the Plitvice Lakes system. *Isot. Environ. Health Stud* **2011**, *47*, 316–329. [[CrossRef](#)]
43. Chaplot, V.; Mutema, M. Sources and main controls of dissolved organic and inorganic carbon in river basins: A worldwide meta-analysis. *J. Hydrol.* **2021**, *603*, 126941. [[CrossRef](#)]
44. Ludwig, W.; Dumont, E.; Meybeck, M.; Heussner, S. River discharges of water and nutrients to the Mediterranean and Black Sea: Major drivers for ecosystem changes during past and future decades? *Prog. Oceanogr.* **2009**, *80*, 199–217. [[CrossRef](#)]
45. Roy, S.; Gaillardet, J.; Allègre, C.J. Geochemistry of dissolved and suspended loads of the Seine River, France: Anthropogenic impact, carbonate and silicate weathering. *Geochim. Cosmochim. Acta.* **1999**, *63*, 1277–1292. [[CrossRef](#)]
46. Fairchild, I.J.; Killawee, J.A.; Sharp, M.J.; Spiro, B.; Hubbard, B.; Lorrain, R.D.; Tison, J.-L. Solute generation and transfer from a chemically reactive alpine glacial-proglacial system. *Earth Surf. Process. Landf.* **1999**, *24*, 1189–1211. [[CrossRef](#)]
47. Anderson, S.P.; Drever, J.I.; Frost, C.D.; Holden, P. Chemical weathering in the foreland of a retreating glacier. *Geochim. Cosmochim. Acta* **2000**, *64*, 1173–1189. [[CrossRef](#)]

48. Jacobson, A.D.; Blum, J.D. Relationship between mechanical erosion and atmospheric CO₂ consumption in the New Zealand Southern Alps. *Geology* **2003**, *31*, 865–868. [[CrossRef](#)]
49. Amiotte Suchet, P.; Aubert, D.; Probst, J.L.; Gauthier-Lafaye, F.; Probst, A.; Andreux, F.; Viville, D. δ¹³C pattern of dissolved inorganic carbon in a small granitic catchment: The Strengbach case study (Vosges Mountain, France). *Chem. Geol.* **1999**, *159*, 129–145. [[CrossRef](#)]
50. Huang, T.H.; Fu, Y.H.; Pan, P.Y.; Chen, C.T.A. Fluvial carbon fluxes in tropical rivers. *Curr. Opin. Environ. Sustain.* **2012**, *4*, 162–169. [[CrossRef](#)]
51. Huang, T.H.; Chen, C.T.A.; Tseng, H.C.; Lou, J.Y.; Wang, S.L.; Yang, L.; Kandasamy, S.; Gao, X.; Wang, J.T.; Aldrian, E.; et al. Riverine carbon fluxes to the South China Sea. *J. Geophys. Res. Biogeosci.* **2017**, *122*, 1239–1259. [[CrossRef](#)]
52. Maier, M.-S.; Teodoru, R.C.; Wehrli, B. Spatio-temporal variations in lateral and atmospheric carbon fluxes from the Danube Delta. *Biogeosciences* **2021**, *18*, 1417–1437. [[CrossRef](#)]
53. Rodellas, V.; Garcia-Orellana, J.; Masqué, P.; Feldman, M.; Weinstein, Y. Submarine groundwater discharge as a major source of nutrients to the Mediterranean Sea. *Proceeding Natl. Acad. Sci.* **2015**, *112*, 3926–3930. [[CrossRef](#)]
54. Trezzi, G.; Garcia-Orellana, J.; Rodellas, V.; Masqué, P.; Garcia-Solson, E.; Andersson, P.S. Assessing the role of submarine groundwater discharge as a source of Sr to the Mediterranean Sea. *Geochim. Cosmochim. Acta* **2017**, *200*, 42–54. [[CrossRef](#)]
55. Cantoni, C.; Luchetta, A.; Celio, M.; Cozzi, S.; Raicich, F.; Catalano, G. Carbonate system variability in the Gulf of Trieste (North Adriatic Sea). *Estuar. Coast. Shelf Sci.* **2012**, *115*, 51–62. [[CrossRef](#)]
56. Sempéré, R.; Charrière, B.; Van Wambeke, F.; Cauwet, G. Carbon inputs of the Rhône River to the Mediterranean Sea: Biogeochemical implication. *Glob. Biogeochem. Cy.* **2000**, *14*, 669–681. [[CrossRef](#)]
57. Stets, E.G.; Kelly, V.J.; Crawford, C.G. Long-term trends in alkalinity in large rivers of the conterminous US in relation to acidification, agriculture, and hydrologic modification. *Sci. Tot. Environ.* **2014**, *488–489*, 280–289. [[CrossRef](#)] [[PubMed](#)]
58. Xue, L.; Cai, W.J. Total alkalinity minus dissolved inorganic carbon as a proxy for deciphering ocean acidification mechanisms. *Mar. Chem.* **2020**, *222*, 103791. [[CrossRef](#)]
59. Brush, M.J.; Giani, M.; Totti, C.; Testa, J.M.; Faganeli, J.; Ogrinc, N.; Kemp, W.M.; Fonda Umani, S. Eutrophication, Harmful Algae, Oxygen Depletion, and Acidification. In *Coastal Ecosystems in Transition: A Comparative Analysis of the Northern Adriatic and Chesapeake Bay*, 1st ed.; Geophysical, Monograph; Malone, T., Malej, A., Faganeli, J., Eds.; Wiley & Sons Ltd.: Hoboken, NJ, USA, 2021; Volume 256, pp. 75–104. ISBN 978-1-119-54358-9.

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