

## Article

# A Commercial Aircraft Fuel Burn and Emissions Inventory for 2005–2011

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**Abstract:** The commercial aircraft fuel burn and emission estimates of CO<sub>2</sub>, CO, H<sub>2</sub>O, hydrocarbons, NO<sub>x</sub> and SO<sub>x</sub> for 2005–2011 are given as the 4-D Aircraft Fuel Burn and Emissions Inventory. On average, the annual fuel burn and emissions of CO<sub>2</sub>, H<sub>2</sub>O, NO<sub>x</sub>, and SO<sub>x</sub> increased by 2%–3% for 2005–2011, however, annual CO and HC emissions decreased by 1.6% and 8.7%, respectively because of improving combustion efficiency in recent aircraft. Approximately 90% of the global annual aircraft NO<sub>x</sub> emissions were emitted in the NH between 2005 and 2011. Air traffic within the three main industrialised regions of the NH (Asia, Europe, and North America) alone accounted for 80% of the global number of departures, resulting in 50% and 45% of the global aircraft CO<sub>2</sub> and NO<sub>x</sub> emissions, respectively, during 2005–2011. The current Asian fleet appears to impact our climate strongly (in terms of CO<sub>2</sub> and NO<sub>x</sub>) when compared with the European and North American fleet. The changes in the geographical distribution and a gradual shift of the global aircraft NO<sub>x</sub> emissions as well as a subtle but steady change in regional emissions trends are shown in particular comparatively rising growth rates between 0 and 30°N and decreasing levels between 30 and 60°N.

**Keywords:** global and regional aviation; fuel burn; aircraft NO<sub>x</sub> emissions; geographical distribution

## 1. Introduction

The emissions of carbon dioxide (CO<sub>2</sub>), nitrogen oxide (NO<sub>x</sub>), sulphur oxide (SO<sub>x</sub>), carbon monoxide (CO), water vapour (H<sub>2</sub>O), and hydrocarbons (HC) from aircraft can have a significant impact on the atmosphere by changing its oxidising capacity and hence the greenhouse gas removal rates, by increasing the levels of greenhouse gases, and by forming particles, contrails, and cirrus cloud [1–10]. Aircraft emissions have increased over time due to high, sustained, and rising levels of passenger traffic driven by continued economic growth and reduced air fares. Airbus predicts a doubling in the level of demand in the next 15 years [11]. Among the aircraft emissions, the emissions of the long-lived well-mixed greenhouse gas, CO<sub>2</sub>, represent 2.0% to 2.5% of total annual CO<sub>2</sub> emissions [3]. The impact of CO<sub>2</sub> on climate is independent of location due to its long atmospheric life-time of over 100 years. In recent years, the aircraft emissions of CO<sub>2</sub> have become a contentious issue and market based measures (e.g., emission trading, emission related levies, and emission offsetting) have been introduced in certain parts of the world that aim to assign a price to the environmental cost of CO<sub>2</sub> emissions from aircraft. Hence, the EU Emissions Trading System (ETS) was implemented within the European Economic Area (EEA), the 28 EU member states, plus Iceland, Liechtenstein, and Norway [12].

The emissions of NO<sub>x</sub> from aircraft have an indirect effect on our climate because of the production of a very potent greenhouse gas, O<sub>3</sub>, from the reactions of hydrocarbons and NO<sub>x</sub> in the presence

of ultra-violet light. The chemical production of  $O_3$  per  $NO_x$  molecule emitted by an aircraft is a non-linear function of ambient levels of  $NO_x$  [13] and the availability of  $HO_x$  from the photo-oxidation of non-methane hydrocarbons [14,15], both factors depending largely on altitude. Hence, the impact of aircraft  $NO_x$  emissions, unlike that of  $CO_2$ , is dependent on the location of the emissions [16,17]. Near ground level during aircraft take-off, landing, taxi-in, taxi-out, the aircraft are in low power condition leading to higher emissions of CO,  $NO_x$ , HC which may induce smog formation and haze. Aviation  $SO_x$  may play a role in the environment by the development of acid rain and the formation of sulphate aerosol particles [18]. Under some meteorological conditions, the presence of aviation induced  $H_2O$ , sulphate particles can form contrails at high altitudes and promote formation of cirrus clouds, which may have an effect on climate change [2].

The estimation of aircraft  $CO_2$  and  $NO_x$  emissions in the form of inventories is a crucial component in the assessment of the atmospheric and climate impact of aviation. The inventories make up the backbone of these assessments. To date, only a limited number of such inventories [8,19–34] have been produced. Most of the aircraft fuel burn and emissions inventories that were produced in the last two decades cover a handful of years only and few have been able to analyse the trends over a significant number of years.

In this paper, the estimates from an up-to-date aircraft mission fuel burn and emissions inventory are presented. The inventory is referred to as the 4-D Aircraft Fuel Burn and Emissions Inventory and it reflects as closely as possible the composition of the global fleet from 2005 to 2011. The inventory was created using a combination of air traffic data, aircraft performance data, and emissions calculation methods [35]. Further, a global, 3-D distribution of aircraft  $NO_x$  emissions was derived from the inventory for the time period 2005–2011. Because the estimates in the 4-D Aircraft Fuel Burn and Emissions Inventory were derived using one, self-consistent method, they present an illustrative historical trend analysis that narrates a story of markedly deviating growth and decline, and capture the impact of the 2008 global economic crisis. The study also showed the relationship between the intracontinental volume of air traffic and the volume of fuel burn and emissions (measured in Tg) of  $CO_2$  and  $NO_x$  accounted for by the intracontinental traffic within the regions Asia, Europe, and North America.

## 2. Methodology

### 2.1. 4-D Aircraft Fuel Burn and Emission Inventory Derivation

Aircraft Performance Model Implementation (APMI) software containing a database of global aircraft movements, a model of aircraft performance for all phases of a flight, and the equations for the calculation of aircraft emissions was used to estimate aircraft mission fuel burn, mission time, distance, and emissions from 2005 to 2011. The global and regional aircraft movements' statistics from 2005 to 2011 were obtained from the global airline schedules data, CAPSTATS (<http://www.capstats.com/>). Only the commercial, scheduled aircraft information were available in the database, neither military nor non-scheduled traffic, such as business jets and charter flight were included in the database. Each flight was listed by city pair and airline and included the aircraft code and departures frequency. Out of the total 6,622,219 flights collected for the period of 2005 to 2011, each represents the monthly total number of departures on a particular route, made by a particular airline and using a particular type of aircraft. If the same airline operated two different aircraft types on any given route during any given month between 2005 and 2011, the total monthly number of departures was recorded separately for each of the different aircraft types operated by this airline on the same route. Different engine options on a particular aircraft type were also treated separately. The details of the sourcing of the flight records and their incorporation into the database can be found in Wasiuk *et al.* [35]. The CAPSTATS database is similar to the air traffic data contained in OAG (<http://www.oag.com/>), but is much more affordable compared to the traditionally employed OAG. Thus the extracted CAPSTATS air traffic data was compared with an excerpt of the OAG database for a single month, September 2008, and it was

found that 94% of the OAG data (the departure airport and country, destination airport and country, aircraft and airline code) matched with CAPSTATS data. There was a partial match for the remaining 6% of the data rows present in OAG and not in CAPSTATS, with the only field not matching being the aircraft code [35].

The EUROCONTROL Experimental Centre (EEC) developed an aircraft performance model, BADA (Base of Aircraft Data), version 3.9 (<http://www.eurocontrol.int/services/bada>) of which was used to calculate the aircraft performance parameters for all stages of a typical aircraft mission. BADA contains performance and operating procedure data for 338 different aircraft, 117 of which are directly supported. The remaining 221 aircraft are supported by the other models determined by EUROCONTROL. The details of the BADA model can be found in Wasiuk *et al.* [35]. A typical aircraft mission was divided into six stages: taxi out (TOUT), take-off (TO), climb (CL), cruise (CR), descent (DES), and taxi in (TIN). CL, CR, and DES were assigned fuel flow and fuel burn values for a given aircraft, mission type and distance calculated by the BADA Total Energy Model (TEM). The TOUT, TO, TIN phases were modelled using engine specific fuel flow values taken from International Civil Aviation Organization (ICAO) for jet aircraft [36] and from the Swedish Defence Research Agency (FOI) for turboprop aircraft [37]. The ground level emissions were estimated using ICAO specified taxi times and thrust settings. The Operation Performance Model (OPM) of BADA gave the airframe specific performance parameters and the Airline Procedure Model (APM) of BADA gave airframe and airline specific speeds and Mach numbers which were used to calculate the thrust, drag, fuel flow, and fuel burn values at different stages of a typical aircraft mission. The detail of the model including assumptions and limitations is given by Wasiuk *et al.* [35]. The implementation of the mathematical model of aircraft performance was validated against EUROCONTROL performance tables (<http://www.eurocontrol.int/services/bada>).

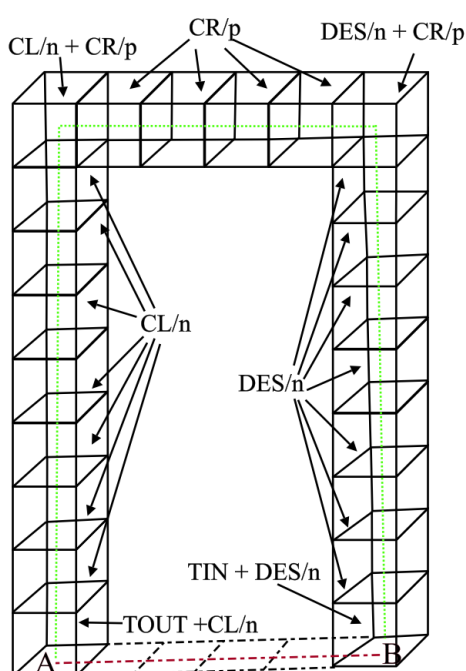
The emissions of CO<sub>2</sub>, CO, NO<sub>x</sub>, SO<sub>x</sub>, H<sub>2</sub>O, and HC were estimated using the Boeing Fuel Flow Method 2 (BFFM2) [38]. Aircraft performance data, reference emissions and fuel flows were used as input to the method. The level of the emissions of CO<sub>2</sub>, CO, NO<sub>x</sub>, SO<sub>x</sub>, H<sub>2</sub>O, and HC is quantified by emission indices (EIs) that specify how many grams of the species are released per one kilogram of fuel burned. EIs are typically represented as a function of an engine power setting and a fuel flow rate. The detailed description of the emission calculation method can be found in Wasiuk *et al.* [35]. The procedure for the calculation of aircraft emissions was validated against an SAE (Society of Automotive Engineers) long-haul flight example [39].

The trajectory simulation of each of the unique missions recorded in the air traffic movements' database with the APMI software assigned a mission distance, time, and fuel burn, a reserve fuel requirement, and the emissions of CO<sub>2</sub>, CO, NO<sub>x</sub>, SO<sub>x</sub>, H<sub>2</sub>O, and HC to each unique flight. These quantities were multiplied by the total number of departures for the month and in this way the monthly totals were calculated. The monthly totals were summed to obtain the yearly estimates.

## 2.2. 3-D NO<sub>x</sub> Emissions Distribution Fields

The total global aircraft NO<sub>x</sub> emissions for each year between 2005 and 2011 were distributed on a grid with a resolution of 5° longitude × 5° latitude. The vertical grid resolution followed the pressure levels and approximate height bands which are based on the vertical model resolution of the 3-D STOCHEM-CRI chemistry and transport model [40,41]. The geographical location of each airport was localised within a grid cell. Each unique route between origin airport (e.g., A) and destination airport (e.g., B) was assigned a shortest path using a direct search algorithm. The algorithm assumed that the mission was flown in a straight line between A and B and that A and B were positioned in the middle of the grid cell in which they had been localised. The path was made up of the grid cells which this straight line crossed. Each unique route was then assigned a number of levels based on the cruise altitude attained during the mission simulated in the APMI. In the mission, TOUT and TIN took place within the grid cell of A and B respectively, CL took place in the column vertically above the grid cell of A until the CR level was reached and DES took place in the column vertically above the grid cell of

B (see Figure 1). A grid cell covers an area of approximately  $400 \text{ km} \times 600 \text{ km}$  (latitude  $\times$  longitude) at midlatitudes and as an aircraft would not exceed a distance of 400 km in a direct, unobstructed climb out to its cruise altitude, this was deemed to be a sufficient vertical distribution of the emissions in the climb phase. When the CR level grid cell vertically above grid cell A was reached, the cruise emissions were distributed in a straight line on the CR level image of the ground level path until the CR level grid cell of the column directly above B was reached, *i.e.*, cruise was assumed to be at a constant altitude for each flight. The minimum fuel burn for an aircraft in cruise at a constant Mach number is usually obtained by allowing height to change as fuel is burnt. In practical terms, this means a stepped-height cruise. In the APMI, a constant cruise altitude is used, with the height chosen to minimize fuel burn on each mission. If, instead of altitude, a stepped cruise were adopted then the difference in fuel burn would probably be of order 5% for a long haul route and less for shorter range flights. In this way, the  $\text{NO}_x$  emissions of all the flights between 2005 and 2011 were distributed on a 3-D grid forming 3-D  $\text{NO}_x$  emissions distribution fields which give a trajectory of the global aircraft  $\text{NO}_x$  emissions from 2005 to 2011.



**Figure 1.**  $\text{NO}_x$  emissions trajectory on the STOCHEM grid. TOUT stands for taxi out, CL for climb, CR for cruise, DES for descent, TIN for taxi in, p for the number of grids that the path from A to B spans and n for the number of levels.

### 3. Results and Discussion

#### 3.1. Global Totals

The global totals of the number of departures, distance flown, mission time, mission fuel, reserve fuel, and the emissions of  $\text{CO}_2$ ,  $\text{CO}$ ,  $\text{H}_2\text{O}$ ,  $\text{HC}$ ,  $\text{NO}_x$ , and  $\text{SO}_x$  between 2005 and 2011 as estimated in the 4-D Aircraft Fuel Burn and Emissions Inventory are summarised in Tables 1 and 2. The underlying number of departures on which the estimates of the emissions are based is lower by 8% compared with Wilkerson *et al.* [30] who presented estimates based on OAG air traffic movement statistics complemented with radar data capturing cargo, military, charter, and unscheduled flights [28]. The difference in the number of departures between estimates given here and those given in Wilkerson *et al.* [30] can effectively be thought of as approximately the number of military, cargo, and unscheduled departures not captured by CAPSTATS [35]. The fuel burn estimates generated within this work are compared with those of previous emission inventory studies (see Table 2). Such a

comparison is limited both by the quality of the emission inventory, and the quality and availability of fuel production data, as there is no perfect database with which to validate or evaluate the jet fuel consumption data [23,24,42]. Nevertheless, the International Energy Agency (IEA) publishes yearly figures of the total global crude oil production, as well as the total global jet kerosene production. It estimates that in 2009 the total global crude oil production was 3518.5 Tg, while the total global jet kerosene production was 233.6 Tg, constituting 6.6% of the total global crude oil production [43]. This figure does not take into account jet fuel production from coal (synthetic jet fuel) which could be significant in some regions (e.g., South Africa).

The fuel match is defined as the ratio of the fuel consumption estimate from an inventory and the global supply of aviation fuel [42]. Using our estimation of the global total fuel use for the year 2009, we calculated the fuel match as  $258.4/233.6 = 1.1$ , a 10% overestimate. This overestimation may depend on the fact that the IEA relies on reported information. Total fuel use, including the estimated mission fuel burn and reserve fuel requirement was adopted for the calculation described. Using the estimated mission fuel burn only, the fuel match is 0.68, an underestimate of 32%, taking into account that mission fuel burn was estimated with flight trajectory simulations optimising for least fuel burn. It is most likely that the actual fuel use figure falls somewhere between these two extremes, *i.e.*, a fuel use figure with a reserve fuel allowance and a fuel burn figure with no reserve fuel allowance. Military aviation and cargo could be the significant source of global aircraft fuel [27,30] which was not considered in the fuel consumption estimate resulting in underestimating the mission fuel burn in this study.

The total global number of departures in this study increased by an annual average of 2% for the period of 2005 to 2011. The global economic downturn of 2008 is reflected in the near zero (0.2%) rise from 2007 to 2008 and the knock on effect on the demand in the following year, 2009, is seen in the 2.8% fall in the number of departures from 2008 to 2009. The annual changes in the mission distance, mission time, mission fuel, and the aircraft emissions of CO<sub>2</sub>, CO, H<sub>2</sub>O, NO<sub>x</sub>, and SO<sub>x</sub> follow the same trend as the changes in the number of departures: an increase between 2005 and 2007, near zero growth in 2008, a decline in 2009, followed by a recovery in 2010 and 2011. On average, between 2005 and 2011, the number of departures, mission time, mission fuel, reserve fuel, CO<sub>2</sub>, H<sub>2</sub>O, NO<sub>x</sub>, and SO<sub>x</sub> emissions increased by between 2 and 3% (Tables 1 and 2). On the other hand, the emissions of CO and HC decreased by 1.6% and 8.7% on average between 2005 and 2011. NO<sub>x</sub> emissions increased significantly by over 20% in total between 2005 and 2011, hence, the complex CO<sub>2</sub>/NO<sub>x</sub> trade-off related to the engine pressure ratio is evidenced by the continuing increase in annual NO<sub>x</sub> emissions. The effect of increasing the pressure ratios has largely offset significant improvements in NO<sub>x</sub> control technology, resulting in only a slight decrease in the NO<sub>x</sub> Emission Index (EI) over the decade between 1990 and 2002 [26]. Conversely, the benefits of improving combustion efficiency in recent aircraft are now seen in a substantial reduction in annual CO and HC emissions and even greater reductions in fleet EI CO and EI HC for civil aviation [26].

**Table 1.** Global totals and percent changes of the number of departures, mission distance and time, total fuel, mission fuel, and reserve fuel during 2005–2011.

Year	2005	2006	2007	2008	2009	2010	2011
Departures (10 <sup>6</sup> )	28.4	28.9 (1.9)	30.5 (5.6)	30.6 (0.2)	29.7 (−2.8)	30.7 (3.2)	31.8 (3.8)
Distance (10 <sup>9</sup> km)	31.5	32.7 (3.8)	35.0 (7.2)	35.6 (1.7)	34.8 (−2.4)	36.4 (4.5)	38.5 (5.9)
Mission time (10 <sup>6</sup> h)	56.3	58.0 (3.1)	61.9 (6.6)	62.6 (1.1)	60.9 (−2.6)	63.4 (4.0)	66.7 (5.2)
Mission fuel burn (Tg)	147.6	152.2 (3.1)	160.9 (5.7)	163.0 (1.3)	158.1 (−3.0)	163.9 (3.6)	173.2 (5.7)
Reserve fuel (Tg)	94.0	96.0 (2.1)	101.6 (5.8)	102.1 (0.6)	99.0 (−3.1)	102.0 (3.0)	106.4 (4.3)

Note: The percentage annual changes are shown in parenthesis.



**Table 2.** Global annual fuel burn and aircraft CO<sub>2</sub>, CO, H<sub>2</sub>O, HC, NO<sub>x</sub> and SO<sub>x</sub> emissions in this study and their comparison with previous inventories.

Year	Fuel Burn (Tg)	CO <sub>2</sub> (Tg)	CO (Tg)	H <sub>2</sub> O (Tg)	HC (Tg)	NO <sub>x</sub> (Tg)	SO <sub>x</sub> (Tg)	Reference
1990	92.8	293.0	0.53	115.0	0.14	1.2	0.074	[22]
1992	110.0	347.0	n/a	135.0	n/a	n/a	0.13	[8]
1992	94.9	423.0	0.50	165.0	0.195	1.2	n/a	[24]
1999	128.0	n/a	0.69	n/a	0.19	1.7	n/a	[44]
1999	136.0	430.0	0.67	167.0	0.23	1.4	0.16	[8]
2000	181.0	572.0	0.54	224.0	0.08	2.5	0.15	[28]
2000	214.0	677.0	n/a	n/a	n/a	2.9	n/a	[31]
2000	152.0	480.0	n/a	187.0	n/a	2.0	0.18	[8]
2001	170.0	536.0	0.46	210.0	0.06	2.4	0.14	[28]
2002	171.0	539.0	0.48	211.0	0.06	2.4	0.14	[28]
2002	156.0	492.0	0.51	193.0	0.06	2.1	n/a	[26]
2002	154.0	486.0	n/a	190.0	n/a	n/a	0.18	[8]
2003	176.0	557.0	0.49	218.0	0.06	2.5	0.14	[28]
2004	188.0	595.0	0.51	233.0	0.06	2.7	0.15	[28]
2004	174.1	549.7	0.63	215.3	0.09	2.5	0.20	[30]
2005	203.0	641.0	0.55	251.0	0.07	2.9	0.16	[29]
2005	180.6	570.5	0.75	n/a	0.2	2.7	0.22	[33]
<b>2005</b>	<b>147.6</b>	<b>464.7</b>	<b>0.78</b>	<b>181.5</b>	<b>0.28</b>	<b>3.4</b>	<b>0.12</b>	<b>This study</b>
2006	188.2	594.3	0.68	232.8	0.10	2.7	0.22	[8,30]
<b>2006</b>	<b>152.2</b>	<b>479.3</b>	<b>0.74</b>	<b>187.2</b>	<b>0.24</b>	<b>3.5</b>	<b>0.13</b>	<b>This study</b>
<b>2007</b>	<b>160.9</b>	<b>506.8</b>	<b>0.75</b>	<b>198.0</b>	<b>0.23</b>	<b>3.7</b>	<b>0.14</b>	<b>This study</b>
2008	229.0	725.0	0.69	282.0	0.09	3.2	0.18	[32]
<b>2008</b>	<b>163.0</b>	<b>513.4</b>	<b>0.74</b>	<b>200.5</b>	<b>0.21</b>	<b>3.8</b>	<b>0.14</b>	<b>This study</b>
<b>2009</b>	<b>158.1</b>	<b>498.0</b>	<b>0.69</b>	<b>194.5</b>	<b>0.18</b>	<b>3.7</b>	<b>0.13</b>	<b>This study</b>
2010	240.0	n/a	1.92	n/a	0.3	3.02	n/a	[34]
<b>2010</b>	<b>163.9</b>	<b>516.0</b>	<b>0.69</b>	<b>201.5</b>	<b>0.17</b>	<b>3.9</b>	<b>0.14</b>	<b>This study</b>
<b>2011</b>	<b>173.2</b>	<b>545.3</b>	<b>0.70</b>	<b>213.0</b>	<b>0.16</b>	<b>4.1</b>	<b>0.15</b>	<b>This study</b>
2015	255.0	803.0	1.14	315.0	0.10	2.3	0.10	[22]
2015	282.0	n/a	1.44	n/a	0.23	3.9	n/a	[25]
2020	336.0	n/a	1.39	n/a	0.23	4.9	n/a	[25]
2025	327.0	1029.0	1.15	404.0	0.15	3.3	n/a	[26]
2030	440.0	n/a	2.33	n/a	0.29	4.95	n/a	[34]
2050	770.0	n/a	2.64	n/a	0.25	7.5	n/a	[34]

Notes: n/a denotes no available data; NO<sub>x</sub> (Tg) and SO<sub>x</sub> (Tg) are reported as NO<sub>2</sub> (Tg) and SO<sub>2</sub> (Tg), respectively.

A number of aircraft fuel burn and emission inventories has been compiled to date and then shown together with the 4-D Aircraft Fuel Burn and Emissions Inventory (this study) for 2005–2011 in Table 2. Table 2 highlights the difficulties in correctly estimating the proportion of the yearly emissions changes due to differences in the methodology used in the derivation of the emissions inventory. The use of varying data sources in the compilation of the inventories also leads to different results for the same year and the scarcity of consistent time lines of estimates.

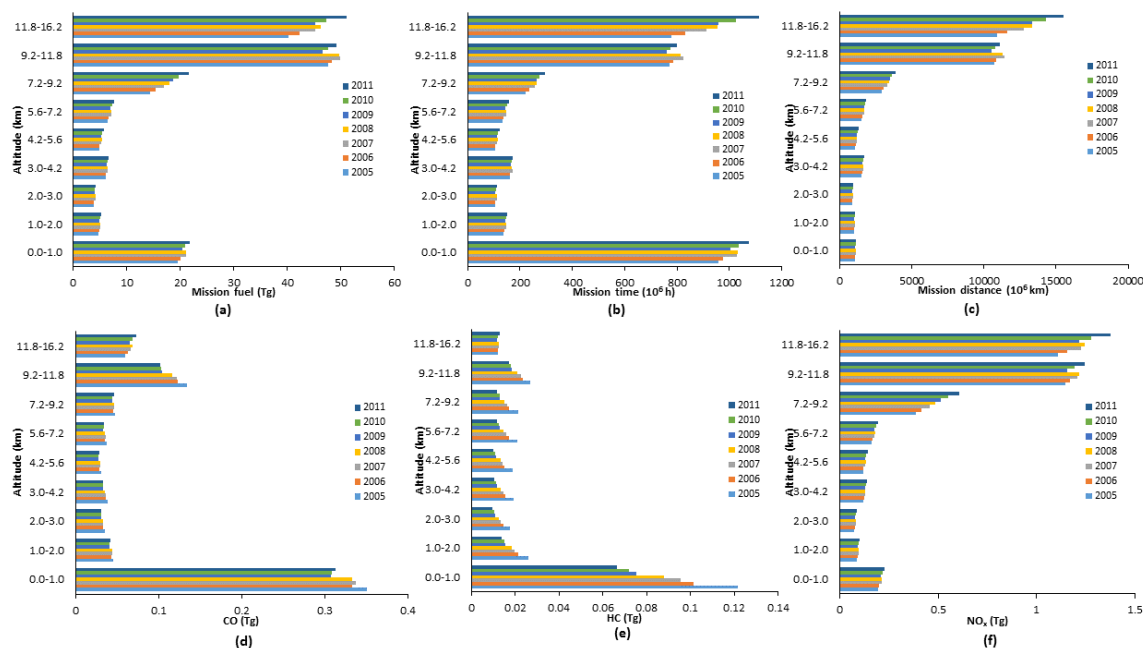
The unscheduled departures were extracted and the flight trajectories were optimised for least fuel burn in the APMI, so the mission fuel burn estimate in this study is found to be 18% (in 2005), 27% (in 2005), 19% (in 2006), 29% (in 2008), and 32% (in 2010) lower than that given in Simone *et al.* [33], Kim *et al.* [29], Wilkerson *et al.* [30], Chèze *et al.* [32] and Yan *et al.* [34], respectively. It is uncertain what fraction of underestimation of the aircraft emissions is due to excluding unscheduled departures and to the fuel burn optimisation. The emissions of CO<sub>2</sub>, H<sub>2</sub>O, and SO<sub>x</sub> are calculated as a function of fuel burn and so the estimates of the emissions are dependent on the carbon, hydrogen and sulphur content in the fuel. The estimates of the emissions of CO<sub>2</sub> and H<sub>2</sub>O are found to be lower by the same ratios of fuel burn for 2005 (~18% and ~27%), 2006 (~19%), and 2008 (~29%), respectively. The emission estimate of SO<sub>x</sub> found in our study is ~24% (in 2005), ~40% (in 2006), and ~22% (in 2008) lower than that found in Kim *et al.* [29], Wilkerson *et al.* [30], and Chèze *et al.* [32], respectively because of the lower fuel burn, and slightly (5%) higher EI SO<sub>x</sub> (0.84 g/kg in this study vs 0.8 g/kg in Kim *et al.* [29]) used in the calculation. The estimates of the emissions of CO, HC, and NO<sub>x</sub> are found to be higher in this work compared with those in Simone *et al.* [33], by 4%, 40%, and 26%, respectively, in Kim *et al.* [29], by 42%, 300%, and 17%, respectively, in Wilkerson *et al.* [30], by 9%, 140%, and 30%, respectively,

and in Chèze *et al.* [32], by 7%, 133%, and 19%, respectively. The emissions of CO, HC, and NO<sub>x</sub> are calculated as a function of fuel burned, altitude, and engine type. As these emissions do not depend on the mission fuel burn alone, hence not on the number of departures modelled, there are other reasons (e.g., airframe/engine match, the use of ICAO prescribed taxi in and taxi out times, reduced thrust take-off, the use of different data sources for turboprop engine fuel flow and EI data, modelling a large number of jet engines which exhibit a non-standard behaviour with respect to EICO and EIHC, and errors associated with modelling EICO and EIHC below the 7% power setting) for these differences. Care would need to be exercised in extending the current methodology to newer aircraft with staged combustor engines, since the Boeing Fuel Flow method used to assign emissions estimates from the APMI results is not appropriate for engines with such combustors [35]. However, the inventory generated within this work is helpful in extending this timeline with estimates based on a self-consistent method.

### 3.2. Global Vertical Profiles

Figure 2 shows the global vertical distribution between 0 and 16.2 km of the fuel burned, mission time, the distance flown, and the emissions of CO<sub>2</sub>, CO, H<sub>2</sub>O, HC, NO<sub>x</sub>, and SO<sub>x</sub> for the time period of 2005–2011. In this study, 20 Tg (13%) fuel is burned below 1.0 km (Figure 2a) which is comparable to the studies of Wilkerson *et al.* [30] and Simone *et al.* [33] who reported about 17–18 Tg of fuel consumption within 1.0 km above ground level. About 30% of the total global mission time was spent between 0–1 km on average between 2005 and 2011 (Figure 2b) which caused higher fuel burned at 0–1 km. Over the mid-altitude region (1–7.2 km), the fuel burn is minimum (~3% per altitude level) which is comparable with the study of Olsen *et al.* [8]. About 11, 30 and 28% fuel was burned at 7.2–9.2, 9.2–11.8, and 11.8–16.2 km, respectively on average between 2005 and 2011 (Figure 2a). 9–12 km is the typical cruise altitude, so the highest fuel burn was found at 9.2–11.8 km. The gradual shift with the introduction of new generation technology aircraft attaining higher cruise altitudes (>12 km) resulted in a rise in mission time (6% on average per year between 2005 and 2011) and mission distance (7% on average per year between 2005 and 2011) (Figure 2b,c), and an increased fuel burn at 11.8–16.2 km. The largest increase (7%) in fuel burn every year on average between 2005 and 2011 took place between 7.2 and 9.2 km, whereas the small increase in fuel burn (1%) on average between 2005 and 2011 took place between 9.2 and 11.8 km (Figure 2a).

The vertical distributions of the global total CO<sub>2</sub>, H<sub>2</sub>O, and SO<sub>x</sub> emissions qualitatively mirror that of fuel burn which have been shown in Figures S1, S2, and S3, respectively. Over 40% of the total global aircraft CO and HC emissions were released at ground level on average between 2005 and 2011 (Figure 2d,e). The CO and HC EIs decrease as a function of the engine power setting, thus CO and HC emissions predominate at idle and other low engine power settings [45]. Idle and low power settings are used during ground operations, *i.e.*, during taxiing, so the majority of the CO and HC emissions are released on the ground. Between 2005 and 2011, the global emissions of aircraft CO decreased by ~10% in total below 7.2 km (Figure 2d). Between 7.2 and 9.2 km, there was a near zero change, between 9.2 and 11.8 km, they decreased by over 20% in total between 2005 and 2011, and above 11.8 km, they increased by nearly 20%. The total global aircraft HC emissions were reduced by nearly a half in total between 2005 and 2011 in the entire modelling domain except between 9.2 and 11.8 km where they decreased by nearly 40% and above 11.8 km where they increased by nearly 10%, respectively (Figure 2e). The combustor inlet pressure at idle power is higher in modern large engines, with a resulting tendency to cause decreased CO and HC emissions at ground level.



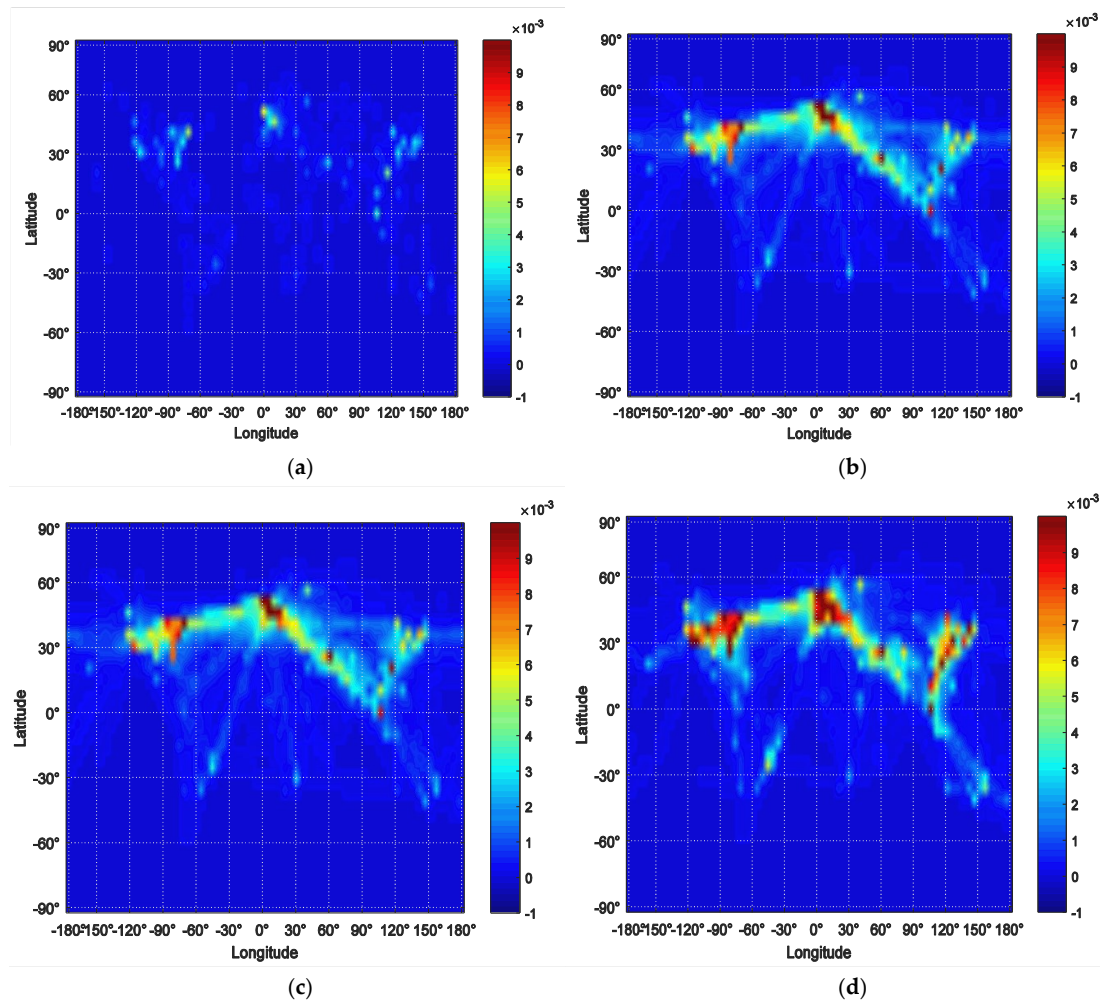
**Figure 2.** Vertical distribution of the global total (a) mission fuel burn, (b) mission time, (c) mission distance, (d) CO emissions, (e) HC emissions, and (f) NO<sub>x</sub> emissions during 2005–2011.

Aircraft NO<sub>x</sub> emissions occur primarily at high engine power settings, hence during take-off and during the cruise portion of a flight [44], thus ~80% of the total global annual aircraft NO<sub>x</sub> emissions was distributed between 7.2 and 16.2 km during the period of 2005–2011 (Figure 2f). In 2005 and 2006, the highest fraction (over one-third) of the total global annual aircraft NO<sub>x</sub> emissions was released at an altitude between 9.2 and 11.8 km and from 2007 onwards between 11.8 and 16.2 km which is the most pronounced change in the vertical distribution (see Figure 2f). This could be due to a shift to aircraft types that are able to attain higher cruise altitudes from 2007 onwards. The changes in the amount of NO<sub>x</sub> emitted between 9.2 and 11.8 km on average between 2005 and 2011 was significantly below the global average, despite the fact that the second highest proportion of the total global annual aircraft NO<sub>x</sub> emissions were deposited between these altitudes. The highest rise (7.7%) took place above 11.8 km on average between 2005 and 2011, nearly double the global average.

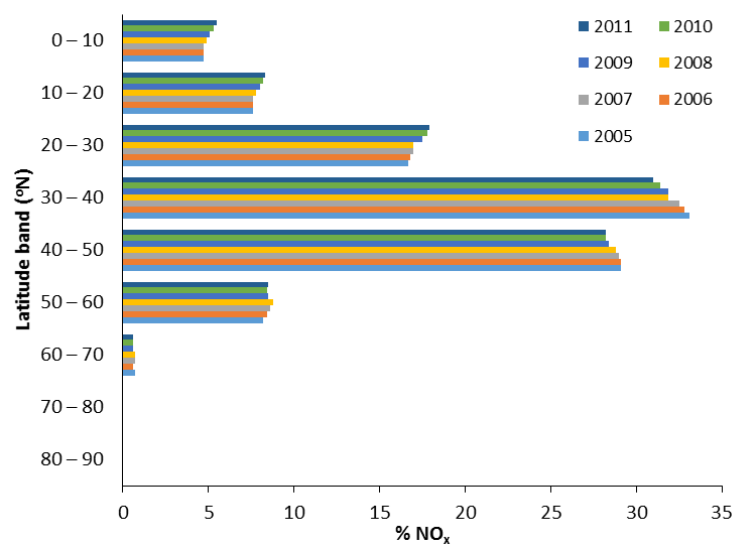
### 3.3. Spatial Distribution of Global Aircraft NO<sub>x</sub> Emissions

The average (2005–2011) spatial distribution of the total global NO<sub>x</sub> emissions from aircraft between 5.6 and 16.2 km (Figure 3) highlights the areas of activity and major global flight paths. Approximately 90% of the total global annual aircraft NO<sub>x</sub> emissions was allocated to the northern hemisphere (NH) on average between 2005 and 2011 for all altitudes from 0 to 16.2 km. During this period, the highest proportion, approximately one-third, of the NO<sub>x</sub> emissions allocated to the NH was found between 30°N and 40°N. Nearly 90% of all the NO<sub>x</sub> allocated to the NH was emitted between 20°N and 60°N (Figure 4). Two features of the distribution of the annual aircraft NO<sub>x</sub> emissions in the NH between 2005 and 2011 were highlighted by the inventory. The first feature is a small but consistent decline in the fraction emitted between 30°N and 60°N (Figure 4). 30°N coincides with the southern border of the US, while 49°N coincides with the northern border of the USA. Europe except Scandinavia and Iceland is contained within these latitudes. Secondly, and conversely, there is a small but consistent rise in this share between 0°N and 30°N (Figure 4). The 0–30°N latitude belt covers South Asia. This points to a subtle shift in the distribution of the annual aircraft NO<sub>x</sub> emissions in the NH.





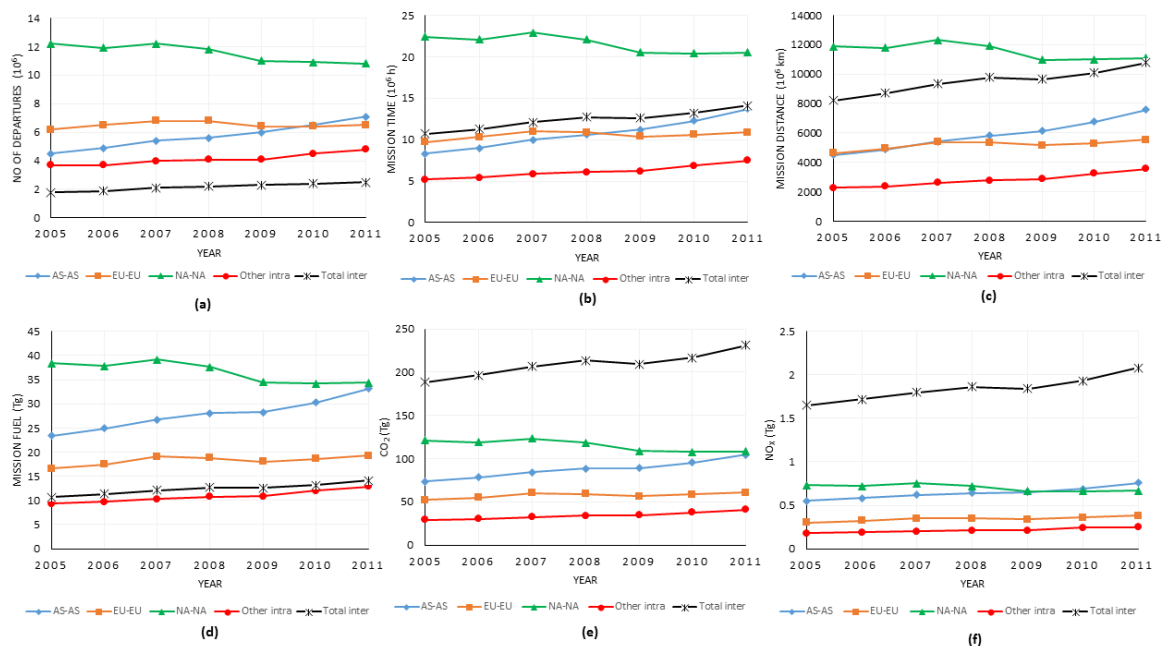
**Figure 3.** The average (2005–2011) distribution of the global aircraft  $\text{NO}_x$  emissions (in Tg), (a) between 5.6 and 7.2 km, (b) between 7.2 and 9.2 km, (c) between 9.2 and 11.8 km, (d) between 11.8 and 16.2 km.



**Figure 4.** The fraction (%) of the total global annual emissions of aircraft  $\text{NO}_x$  allocated to the NH in each  $10^\circ$  latitude band from  $0^\circ\text{N}$  to  $90^\circ\text{N}$  for the period of 2005–2011.

### 3.4. Regional Totals

Global air traffic, in terms of the number of departures, was dominated by intracontinental traffic within the regions of Asia (AS), Europe (EU), and North America (NA) which accounted for ~80% of the total each year between 2005 and 2011, resulting in ~50% and ~45% of the total global aircraft CO<sub>2</sub> and NO<sub>x</sub> emissions, respectively (Figure 5). Historically NA is the world's largest civil aviation market but by 2025, both EU (27%) and AS (32%) are predicted to have larger aviation markets than NA (25%) [45]. The regional results in the study show that in 2005, the AS and NA shares of the market (in terms of the number of departures) were 16% and 43%, respectively. By 2011, the respective shares were 22% and 34%, hence much closer to each other, with the AS share already bigger than the EU (21%) and rising, while the NA share is decreasing.



**Figure 5.** Regional totals of (a) the number of departures, (b) mission time, (c) mission distance, (d) mission fuel, (e) CO<sub>2</sub> emissions, (f) NO<sub>x</sub> emissions for 2005–2011. Note: Other intra consists of AF↔AF, CA↔CA, CB↔CB, ME↔ME, SA↔SA, SW↔SW. AS, EU, NA, AF, CA, CB, ME, SA, SW represent Asia, Europe, North America, Africa, Central America, Caribbean, Middle East, South America, Australasia, respectively.

In 2005, there were 63% fewer departures, mission time and mission distance in AS than in NA, but the mission fuel, emissions of CO<sub>2</sub>, and NO<sub>x</sub> were only 39%, 40%, and 25% lower, respectively. By 2011, the number of departures, mission time, and mission distance in AS were 34%, 32%, and 33% lower than the number of departures, mission time, and mission distance in NA. However, mission fuel and CO<sub>2</sub> emissions were only 3% and 4% lower in AS respectively, while the emissions of NO<sub>x</sub> were 13% higher. Considering the time period from 2005 to 2011, on average, there were 50% fewer departures (with 49% lower mission distance and 50% lower mission time) taking place within AS than in NA each year, but they were burning 34% less fuel and generating only 24% less CO<sub>2</sub> and 8% less NO<sub>x</sub> each year (Figure 5).

Comparing AS with EU, there were 27% fewer departures and only 14 and 3% lower mission time and distance, respectively within AS than within EU in 2005, but the smaller number of intracontinental AS flights burned 41% more fuel and generated 40% more CO<sub>2</sub> and 80% more NO<sub>x</sub> emissions than those within EU. By 2011, the number of the departures, mission time and distance in AS had overtaken the number of departures, mission time and distance in EU by 9%, 25%, and 37%, respectively and

AS air traffic was generating 72% more CO<sub>2</sub> emissions and 100% more NO<sub>x</sub> emissions by burning 72% more fuel than EU air traffic. On average for the time period of 2005–2011, there were 12% fewer departures, but 2% increased mission time and 13% increased mission distance within AS than within the EU each year, however they were generating 52% and 87% more CO<sub>2</sub> and NO<sub>x</sub> per year (Figure 5). Comparing EU to NA, the results were more uniform. Between 2005 and 2011, on average, EU annual intracontinental departures, mission distance, mission time were 44%, 55%, and 51%, respectively lower than NA, but the mission fuel, CO<sub>2</sub>, and NO<sub>x</sub> annual emissions were also 50% lower within the EU than within NA (Figure 5).

The variabilities in the regional emission intensities of CO<sub>2</sub> between AS, EU, NA during the time period of 2005–2011 suggest that AS air traffic is contributing more to aviation's fraction of the total global CO<sub>2</sub> burden. The regional and global atmospheric impact of the differences in the regional emissions intensities of aircraft NO<sub>x</sub>, in terms of increased tropospheric O<sub>3</sub> production, is heterogeneous. Gilmore *et al.* [46] showed that the magnitude of sensitivity for southeast AS is a factor of 2 higher relative to NA for global O<sub>3</sub> due to changes in regional NO<sub>x</sub> emissions. In this study, the annual average increase in the number of departures within AS was 8%, 1% in EU, and −1.9% in NA. Aviation is growing significantly faster in AS compared with EU and NA, so the emissions of NO<sub>x</sub> from aircraft grow at a higher rate and with a higher intensity in AS, which also has the strongest response in terms of O<sub>3</sub> production to the addition of the NO<sub>x</sub> emissions.

#### 4. Conclusions

On the 20th anniversary of the first comprehensive aircraft emission inventory, we presented a timeline of seven consecutive years (2005–2011) of estimates of the mission fuel burn, a reserve fuel requirement, and the emissions of CO<sub>2</sub>, CO, H<sub>2</sub>O, HC, NO<sub>x</sub>, and SO<sub>x</sub> of present day global and regional air traffic. The estimation referred to as the 4-D Aircraft Fuel Burn and Emissions Inventory was based on a detailed representation of the global fleet derived from actual records of air traffic movements and a detailed distribution of the fuel consumption and the emissions throughout the entire flight cycle. The 4-D Aircraft Fuel Burn and Emissions Inventory enabled a consistent global and regional trend analysis (2005–2011) and comparison of fuel burn and emissions for the first time since Kim *et al.* [27] who presented estimates for the time period 2000–2005, thus picking up the timeline in 2005, extending it to 2011, and bringing it up-to-date. An estimate of a global reserve fuel requirement was given for the first time within this work, which together with the mission fuel burn estimate provides an improved quantification of the global aviation fuel consumption with a fuel match of 1.1 (a 10% overestimate) relative to an IEA estimate for the total global jet kerosene production for 2009. The inventory results revealed that global estimates mask substantial regional fluctuations and differences in the quantity and distribution of the aircraft NO<sub>x</sub> emissions and in the number of departures, in particular, a persistent decline in North America and persistent growth in Asia. Moreover, a trend of substantially higher CO<sub>2</sub> and NO<sub>x</sub> emission intensities within Asia is found if we compare their emissions within Europe and North America. A gradual shift in the global distribution of the NO<sub>x</sub> emissions from aircraft, and a subtle but steady change in the regional emission trends, are found with comparatively higher and rising growth rates in latitudes 0–30°N. These factors may have potentially large atmospheric implications in light of recent findings which indicate that the sensitivities to aircraft NO<sub>x</sub> emissions in Southeast Asia are more than double those in North America and Europe [46].

**Supplementary Materials:** The following are available online at <http://www.mdpi.com/2073-4433/7/6/78/s1>, Figure S1: Vertical distribution of the global aircraft CO<sub>2</sub> emissions during 2005–2011; Figure S2: Vertical distribution of the global aircraft H<sub>2</sub>O emissions during 2005–2011; Figure S3: Vertical distribution of the global aircraft SO<sub>x</sub> emissions during 2005–2011.

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**Author Contributions:** D.K. Wasiuk, D.E. Shallcross, and M.H. Lowenberg conceived the idea, designed the experiments; D.K. Wasiuk and M.A.H. Khan analysed the data and wrote the paper; M.H. Lowenberg and D.E. Shallcross provided important suggestions and approved the final manuscript.

**Conflicts of Interest:** The authors declare no conflict of interest.

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