

Article

Spatio-Temporal Characteristics of Air Quality Index (AQI) over Northwest China

Shah Zaib ¹, Jianjiang Lu ^{1,*} and Muhammad Bilal ² 

¹ Key Laboratory of Environmental Monitoring and Pollution Control of Xinjiang Bingtuan, School of Chemistry and Chemical Engineering, Shihezi University, Shihezi 832003, China; shah@stu.shzu.edu.cn

² School of Marine Sciences, Nanjing University of Information Science and Technology, Nanjing 210044, China; muhammad.bilal@connect.polyu.hk

* Correspondence: luji@shzu.edu.cn

Abstract: In recent years, air pollution has become a serious threat, causing adverse health effects and millions of premature deaths in China. This study examines the spatial-temporal characteristics of ambient air quality in five provinces (Shaanxi (SN), Xinjiang (XJ), Gansu (GS), Ningxia (NX), and Qinghai (QH)) of northwest China (NWC) from January 2015 to December 2018. For this purpose, surface-level aerosol pollutants, including particulate matter (PM_x, x = 2.5 and 10) and gaseous pollutants (sulfur dioxide (SO₂), nitrogen dioxide (NO₂), carbon monoxide (CO), and ozone (O₃)) were obtained from China National Environmental Monitoring Center (CNEMC). The results showed that fine particulate matter (PM_{2.5}), coarse particulate matter (PM₁₀), SO₂, NO₂, and CO decreased by 28.2%, 32.7%, 41.9%, 6.2%, and 27.3%, respectively, while O₃ increased by 3.96% in NWC during 2018 as compared with 2015. The particulate matter (PM_{2.5} and PM₁₀) levels exceeded the Chinese Ambient Air Quality Standards (CAAQS) Grade II standards as well as the WHO recommended Air Quality Guidelines, while SO₂ and NO₂ complied with the CAAQS Grade II standards in NWC. In addition, the average air quality index (AQI), calculated from ground-based data, improved by 21.3%, the proportion of air quality Class I (0–50) improved by 114.1%, and the number of pollution days decreased by 61.8% in NWC. All the pollutants' (except ozone) AQI and PM_{2.5}/PM₁₀ ratios showed the highest pollution levels in winter and lowest in summer. AQI was strongly positively correlated with PM_{2.5}, PM₁₀, SO₂, NO₂, and CO, while negatively correlated with O₃. PM₁₀ was the primary pollutant, followed by O₃, PM_{2.5}, NO₂, CO, and SO₂, with different spatial and temporal variations. The proportion of days with PM_{2.5}, PM₁₀, SO₂, and CO as the primary pollutants decreased but increased for NO₂ and O₃. This study provides useful information and a valuable reference for future research on air quality in northwest China.

Keywords: northwest China; AQI; primary pollutant; CNEMC; Pearson correlation



Citation: Zaib, S.; Lu, J.; Bilal, M. Spatio-Temporal Characteristics of Air Quality Index (AQI) over Northwest China. *Atmosphere* **2022**, *13*, 375. <https://doi.org/10.3390/atmos13030375>

Academic Editors: Célia Alves, Duanyang Liu, Kai Qin and Honglei Wang

Received: 23 December 2021

Accepted: 17 February 2022

Published: 23 February 2022

Publisher's Note: MDPI stays neutral with regard to jurisdictional claims in published maps and institutional affiliations.



Copyright: © 2022 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (<https://creativecommons.org/licenses/by/4.0/>).

1. Introduction

Unprecedented economic activity, urbanization, industrialization, and motorization have deteriorated the ambient air quality in China [1–6]. China is the manufacturing hub of the world, with the majority of the industries in northwest China (NWC). Several studies have reported higher pollution levels in NWC due to increased industry, coal consumption, distinct topography, and adverse meteorology [2,5,7–11]. Increased pollution levels have attracted the attention of the general public, the scientific community, and relevant authorities because of their detrimental health effects [4,12–16]. To combat increasing pollution levels, China has made significant efforts, e.g., establishing Chinese ambient air quality standards (CAAQS) for six criteria pollutants [17], implementing the Atmospheric Pollution Prevention and Control Action Plan 2013 (APPCAP) [18], technical regulation on ambient air quality index (HJ 633–2012) [19], nationwide air quality monitoring, online data-sharing networks, etc. [20,21]. These measures have helped to reduce pollution to

some extent; e.g., from Ref. [22], a 12.3% reduction in fine particulate matter ($PM_{2.5}$) was observed in China between 2013 and 2015, while other authors [23] observed a one-third reduction in $PM_{2.5}$ from 2013 to 2017 due to APPCAP in China. In Ref. [24], it was observed that the annual average concentration ($PM_{2.5}$), coarse particulate matter (PM_{10}), sulfur dioxide (SO_2), nitrogen dioxide (NO_2), and carbon monoxide (CO) decreased by 27.9%, 23.8%, 51.2%, 10.6%, and 25.3%, respectively, in China from 2015 to 2019. Even after strict environmental regulations, the air pollution in some areas of NWC is beyond certain limits and causes serious health effects [25–28].

Most of the spatio-temporal studies in China focused on central China, north China, considered few parameters, fewer cities, mostly provincial capitals, and lacked detailed assessment, with few exceptions. Multiple studies, e.g., [3,29], observed higher $PM_{2.5}$ pollution in northern China, western China, northwestern China, etc. due to increased industrial emissions, coal combustion, stagnant meteorology, etc. Similarly, other researchers [30,31] carried out a more detailed assessment, covered six criteria pollutants in provincial capitals, and observed higher pollution levels in the north and northwestern region. In Ref. [1], they analyzed the criteria pollutants ($PM_{2.5}$, PM_{10} , SO_2 , NO_2 , CO, and O_3) in 336 cities of China, while other researchers [31] analyzed the criteria pollutants in 367 cities of China and observed higher pollution in industrialized areas of north China. Further, Ref. [2] thoroughly assessed the air pollution in NWC by analyzing the criteria pollutants in six cities of Gansu province in NWC and observed higher pollution in Lanzhou, the provincial capital, with increased energy consumption and industrial activity and facing serious health concerns [25].

$PM_{2.5}$ is more dangerous than PM_{10} and is ranked as the first leading risk factor for disease in China, causing more than 1.1 million premature deaths, with the highest share being stroke, ischemic heart disease (IHD), lung cancer, lower respiratory infections, chronic obstructive pulmonary disease (COPD), etc. [32–34]. Other criteria pollutants, e.g., PM_{10} , CO, O_3 , SO_2 , and NO_2 , cause multiple health disorders, e.g., headaches, dizziness, nausea, respiratory disorders, inflammatory reactions, reduced lung function, hampered neurological function, etc. [35–37]. Apart from jeopardizing human health, air pollution is also responsible for visibility reduction, economic losses, and climate change [13]. $PM_{2.5}$ and O_3 can cause gross domestic production (GDP) losses of 2.09% in health expenditure [38].

In view of such circumstances, we examined the spatial and temporal distribution of $PM_{2.5}$, PM_{10} , SO_2 , NO_2 , CO, and O_3 in 53 cities located in five provinces of NWC for a period of four years (2015 to 2018). Besides the criteria pollutants, we also examined the $PM_{2.5}/PM_{10}$ ratio, Air Quality Index (AQI), AQI class distributions, major pollutant on a representative day, number of pollution days, and correlations among different pollutants to explicate the pollution status, spatial, and temporal distribution of air quality in NWC over time. This study provides useful information and a valuable reference for future research on air quality in NWC and is of considerable significance to environmental protection and human health.

2. Materials and Methods

2.1. Site Selection

In this study, we examined the ambient air quality in the northern and western parts of China, known as northwest China (NWC), from January 2015 to December 2018 to understand the spatio-temporal variation across NWC better. NWC is a mixture of agricultural areas, deserts, mountains, etc. with significant coal reserves, industrial activity, covering an area of 3.1 million sq. km area (32.4% of China), having a population of more than 96.65 million, and experiencing degraded air quality. NWC consists of five provinces, namely Shaanxi (SN), Xinjiang (XJ), Gansu (GS), Ningxia (NX), and Qinghai (QH) (Supplementary Material Figure S1). Table 1 gives detailed information about 5 provinces and 53 cities.

Table 1. Description of cities under observation in five provinces (Shaanxi (SN), Xinjiang (XJ), Gansu (GS), Ningxia (NX), and Qinghai (QH)) of northwest China (NWC).

Province	City	Population (million)	Area (km ²)	Monitoring Stations	Attainment (%)			
					2015	2016	2017	2018
Shaanxi (SX)	Ankang	2.63	23,536	3	20.27	17.49	13.70	8.49
	Baoji	3.717	18,712	8	23.56	36.61	33.42	21.64
	Hanzhong	3.84	27,246	4	20.82	25.68	27.12	14.25
	Shanglou	2.34	19,587	2	15.34	15.85	7.67	3.29
	Tongchuan	0.83	3882	4	24.93	44.54	31.51	21.37
	Weinan	5.52	13,134	4	27.40	54.10	54.52	33.42
	Xian	12	10,097	13	30.14	47.54	50.14	36.16
	Xianyang	5.096	10,213	4	27.12	53.55	55.89	24.93
	Yannan	2.198	37,000	4	22.47	21.04	13.97	3.29
	Yulin	3.38	43,578	4	19.73	19.67	22.74	9.32
Xinjiang (XJ)	Aksu	2.37	127,144.91	2	52.60	72.68	23.56	18.63
	Altay	0.526	117,699.01	2	0.27	0.00	0.00	0.00
	Bortala	0.443	24,934.33	2	9.32	9.56	10.68	6.30
	Crete	0.525	72,468.08	1	49.32	50.27	21.64	16.99
	Changji	1.428	73,139.75	3	18.63	26.78	29.32	24.66
	Hami	0.572	142,094.88	2	16.99	12.57	5.48	1.37
	Hotan	2.014	249,146.59	2	75.07	73.77	19.18	12.05
	Ili	2.482	56,381.53	3	14.79	16.94	23.56	20.55
	Karamy	0.39	8654.08	5	8.49	9.02	11.78	8.22
	Korla	1.278	470,954.25	3	31.51	44.26	34.79	8.77
	Kashgar	3.979	137,578.51	3	71.51	77.32	40.82	30.96
	Shihezi	0.635	456.84	2	21.92	25.41	37.53	25.75
	Tacheng	1.219	94,698.18	1	0.00	0.27	0.27	0.27
	Turpan	0.622	67,562.91	2	36.16	55.19	32.33	21.64
	Urumqi	3.11	13,787.90	7	33.97	31.69	33.15	23.56
	Wujiaqu	0.09	742	1	23.84	33.33	32.60	25.75
Gansu (GS)	Dingxi	3.031	19,609	2	16.71	12.84	8.22	3.29
	Gannan	0.689	40,898	1	18.36	15.03	8.77	2.74
	Jiayuguan	0.231	2935	2	18.08	17.76	4.38	1.92
	Jinchang	0.228	8896	3	20.55	20.77	7.67	2.74
	Jiuquan	1.096	191,342	2	26.58	35.52	31.78	21.10
	Lanzhou	3.61	13,300	5	21.64	15.30	15.89	4.66
	Linxia	0.25	88.6	2	7.12	7.10	6.03	2.47
	Longnan	2.567	27,000	2	18.08	10.66	3.56	0.27
	Pinglian	2.068	11,196	2	21.64	14.75	6.03	2.74
	Qingyang	2.21	27,119	3	15.07	12.30	7.95	1.92
	Silver City	1.708	21,200	2	24.38	17.49	7.95	3.29
	Tianshui	3.262	14,300	3	15.34	16.12	7.40	6.03
	Wuwei	1.815	33,000	2	17.53	15.85	6.85	1.92
	Zhangye	1.2	42,000	2	23.29	14.48	5.48	0.27
Ningxia (NX)	Guyuan	1.45	14,413	3	12.88	13.11	4.38	1.10
	Shizuishan	0.73	5208.13	4	38.08	33.33	24.38	9.86
	Yinchuan	2.293	8874.61	3	25.75	23.50	13.70	4.93
	Wuzhong	1.3	16,758	6	26.03	29.23	30.68	9.59
	Zhongwei	1.041	16,986	3	24.66	21.04	16.16	2.74
Qinghai (QH)	Guoluo/Golog	0.181	76,312	1	8.22	4.10	5.48	1.92
	Haibei	0.273	39,354	1	19.73	15.57	2.47	0.82
	Haidong	1.396	12,810	1	27.67	21.86	12.60	7.40
	Hainan	0.441	45,895	1	16.99	9.29	4.11	1.37
	Haixi	0.515	325,785	1	14.79	6.83	2.19	0.00
	Huanggnan	0.256	17,921	1	33.97	30.33	27.40	21.64
	Xinning	2.208	7372	4	18.90	25.68	13.15	4.66
	Yushu/Gyegu	0.12	13,462	1	5.48	1.91	1.10	0.27

2.2. Data Collection

To meet the objectives of the study, we analyzed the ambient air quality in 53 cities located in five provinces (SN, XJ, GS, NX, and QH) of NWC (Table 1) for a period of four years (2015–2018). The hourly concentration of PM_{2.5}, PM₁₀, SO₂, NO₂, CO, and O₃ was collected from China National Environmental Monitoring Center (CNEMC). The online data-sharing platform covers 367 cities of China, and dispatch/publish information according to the Technical Guideline on Environmental Monitoring Quality Management (HJ 630-2011) [39].

2.3. Air Quality Index (AQI)

The air quality index (AQI) includes 24-h average measurement of PM_{2.5}, PM₁₀, NO₂, SO₂, CO, and 8-h average concentration of O₃ and reflects the overall air quality [8,28,40]. Individual air quality index (IAQI) for six criteria pollutants is determined by using Equation (1), and the overall AQI is calculated based on the highest IAQI by using Equation (2) according to the instruction given in technical regulation on ambient air quality index (on trial) (HJ-633-2012) [19].

$$IAQI_p = \frac{I_{high} - I_{low}}{C_{high} - C_{low}} \times (C_p - C_{low}) + I_{low} \quad (1)$$

IAQI_p = individual sub air quality index of the pollutant p

C_p = concentration of the pollutant p

C_{high} = concentration breakpoint that is $\geq C_p$

C_{low} = concentration breakpoint that is $\leq C_p$

I_{high} = index breakpoint corresponding to C_{high}

I_{low} = index breakpoint corresponding to C_{low}

$$AQI = \max(I_1, I_2, \dots, I_n) \quad (2)$$

In Equation (2), “n” indicates the number of criteria pollutants. When AQI is higher than 50, the highest IAQI is considered as a major pollutant for that given day [23,28,41–44]. Air Quality Index (AQI) has the following six categories:

Class I: 0–50 (Green), Good

Class II: 51–100 (Yellow), Moderate

Class III: 101–150 (Orange), Unhealthy for Sensitive Groups

Class IV: 151–200 (Red), Unhealthy

Class V: 201–300 (Purple), Very unhealthy

Class VI: 300–500 (Maroon), Hazardous

2.4. Quality Assurance and Quality Control (QA&AR)

Quality assurance and control procedures for ambient air quality data were strictly in accordance with Chinese Ambient Air Quality Standards (CAAQS) (GB-3095-2012) [17]. The daily average value was calculated when we have valid data for more than 16 h of that day (except for ozone, minimum 6-h value for 8-h ozone value); the monthly average was calculated only when we have 27 daily mean values; an annual value was calculated only when we have 324 daily mean values. Besides this, manual inspection was carried out to remove abnormal values, e.g., PM_{2.5} values higher than PM₁₀ values.

2.5. Inverse Distance Weighted (IDW) Spatial Interpolation

Many spatial interpolation methods, such as kriging (universal or ordinary) and inverse distance weighted (IDW) spatial interpolation, have been used in different studies [45,46]. IDW geospatial interpolation is a type of deterministic method for multivariate interpolation with a known scattered set of points. IDW assigns values to unknown points according to the weighted average of the values of the known points and is more suitable

for regional interpolation [47]. In this study, we used IDW spatial interpolation technique to interpolate spatial distribution of PM_{2.5}, PM₁₀, SO₂, NO₂, CO, O₃, AQI, and PM_{2.5}/PM₁₀ ratio in NWC. Equation (3) describes the interpolation analysis.

$$Z_p = \frac{\sum_i^n = 1 \frac{z_i}{dip}}{\sum_i^n = 1 \frac{1}{dip}} \quad (3)$$

where Z_p refers to the value of unknown point, Z_i is the value observed at the point of I ; I represents the nearest neighborhood of interpolated point produced; p is the weighting absolute value, and p is equal to inverse distance weight, respectively.

2.6. Statistical Analysis

In this study, we used Statistical Package for Social Sciences (SPSS) for Windows (IBM SPSS Statistics, Version 25) to find the Pearson's correlation coefficient for criteria pollutants on an annual and seasonal basis [48], and used RStudio for graphical representation [49]. The high (low) value of the Pearson's correlation represents the same (different) variations in one criteria pollutant with respect to another pollutant. The effect of a certain variable was considered statistically significant for P (0.01, and 0.05) (two-tailed). Annual mean values, mean absolute deviation (MAD), mean square error (MSE), root mean square error (RMSE), mean absolute percentage error (MAPE), and mean percentage error (MPE) of six criteria pollutants between 2015 and 2018 were calculated by Microsoft Excel 2016.

3. Results

3.1. Spatial and Temporal Variation of Six Criteria Pollutants

During the study period (2015–2018), the average concentration of PM_{2.5}, PM₁₀, SO₂, NO₂, and CO decreased by 28.2% (14.2%, 25.8%, 31.1%, 34.4%, 47.1%) (Figure 1a), 32.7% (18.5%, 32.5%, 31.8%, 32.7%, 46.3%) (Figure 1b), 41.9% (30.6%, 43.8%, 46.8%, 48.7%, 32.3%) (Figure 1c), 6.2% (+3.59%, 4.44%, 18.8%, +3.72%, 8.45%) (Figure 1d), and 27.3% (32.7%, 21%, 38.4%, 17.2%, 16%) (Figure 1e), respectively, in NWC (SN, XJ, GS, NX, QH). In contrast to the other pollutants, the ozone levels increased by 3.69% (5.4%, 6.04%, 1.32%, 19.3%, 5.66%) in NWC (SN, XJ, GS, NX, QH) between 2015 and 2018 (Figure 1f). The annual average concentration of PM_{2.5} and PM₁₀ failed to comply with CAAQS Grade II standards (35 µg/m³ and 70 µg/m³, annual mean) and exceeded them by 25% and 31.9%, respectively, and exceeded them by 3.37 and 3.61 times, respectively, for the WHO air quality guidelines (10 µg/m³ and 20 µg/m³, annual mean) in NWC. PM_{2.5} and PM₁₀ failed to comply with CAAQS Grade II standards in SN, XJ, GS, NX, and NWC (Figure 1a,b), while SO₂ and NO₂ complied with CAAQS Grade II standards (60 µg/m³ and 40 µg/m³, annual mean) in SN, XJ, GS, NX, QH, and NWC (Figure 1c,d). CO and O₃ do not have annual standards under CAAQS; CO decreased in SN, XJ, GS, NX, QH, and NWC, while O₃ decreased in GS and QH during 2018 as compared with 2015 (Figure 1e,f). During the study period, the highest concentration of PM_{2.5}, PM₁₀, SO₂, NO₂, CO, and O₃ occurred in SN, XJ, NX, SN, SN, and QH, respectively. Figure 2 explains the spatial distribution of the criteria pollutants in 53 cities of NWC during 2015 to 2018, obtained by the inverse distance weighted (IDW) interpolation technique. The obtained results from spatial interpolation were quite similar to the actual values. In the case of spatial distribution, 92.5%, 96.2%, 92.5%, 64.5%, 88.7%, and 11.3% of the cities of NWC experienced a reduction in PM_{2.5} (Figure 2a–d), PM₁₀ (Figure 2e–h), SO₂ (Figure 2i–l), NO₂ (Figure 2m–p), CO (Figure 2q–t), and O₃ (Figure 2u–x), respectively, during 2018 as compared 2015. Similarly, 66%, 72.5%, and 13.2% of the cities failed to meet the CAAQS Grade II for PM_{2.5}, PM₁₀, and NO₂ (35 µg/m³, 70 µg/m³, 40 µg/m³, annual mean), respectively (Table S1). Most of the cities that were not complying with the CAAQS are cities with a larger population and increased industrial activities.

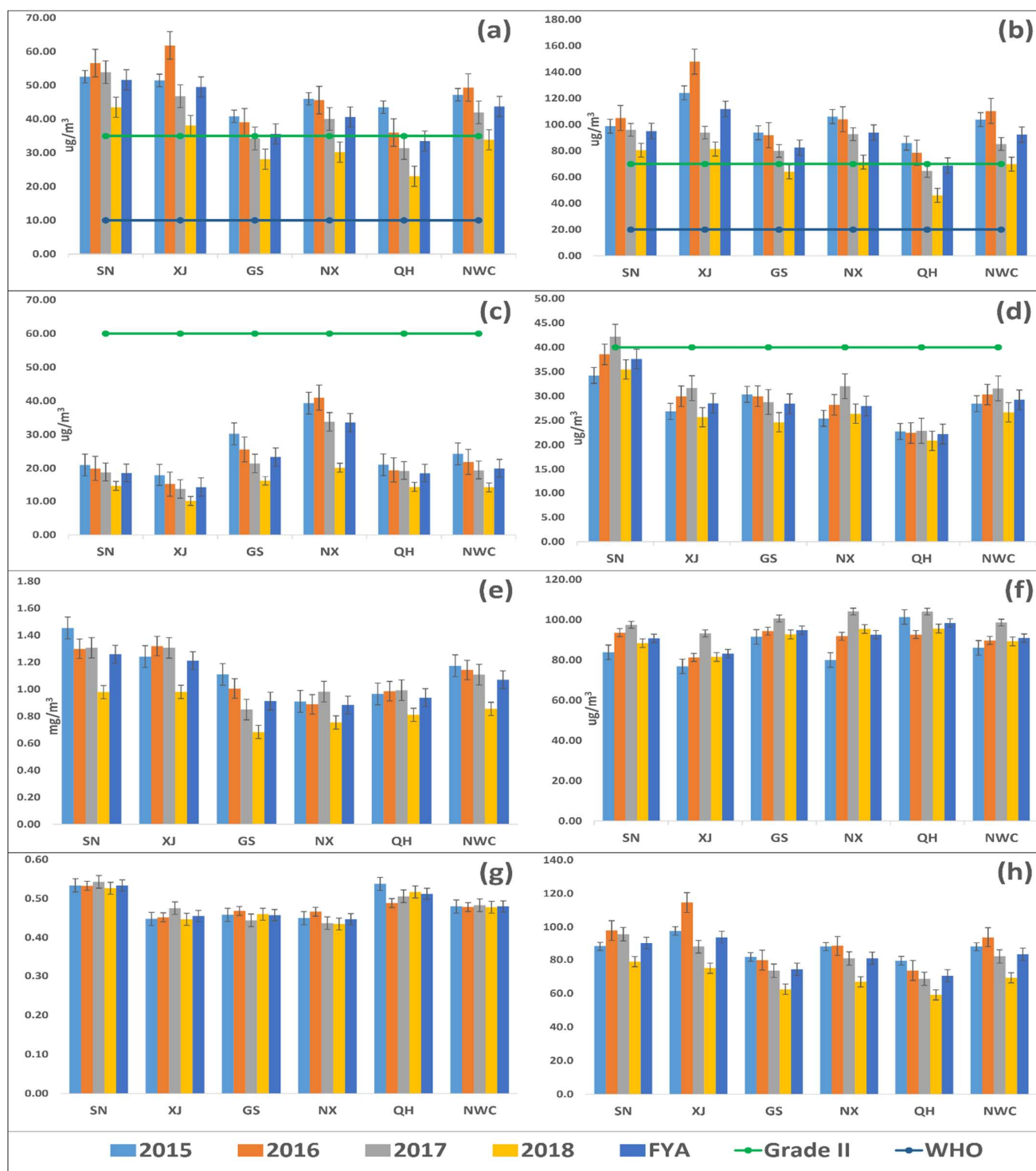


Figure 1. Annual variation of $PM_{2.5}$ (a), PM_{10} (b), SO_2 (c), NO_2 (d), CO (e), O_3 (f), $PM_{2.5}/PM_{10}$ (g), and AQI (h) in five provinces (Shaanxi (SN), Xinjiang (XJ), Gansu (GS), Ningxia (NX), and Qinghai (QH)) of northwest China (NWC) and NWC as a whole between 2015 and 2018. Descriptions are as follows: light blue bar (2015), orange bar (2016), grey bar (2017), yellow bar (2018), blue bar (four-year average (FYA)), parrot line with dots (CAAQS Grade II standards, annual mean), and dark blue with dots (WHO standards). The abbreviations are as follows: $PM_{2.5}$ (fine particulate matter), PM_{10} (coarse particulate matter), SO_2 (sulfur dioxide), NO_2 (nitrogen dioxide), CO (carbon monoxide), O_3 (ozone), $PM_{2.5}/PM_{10}$ (ratio of $PM_{2.5}$ with PM_{10}), and AQI (air quality index).

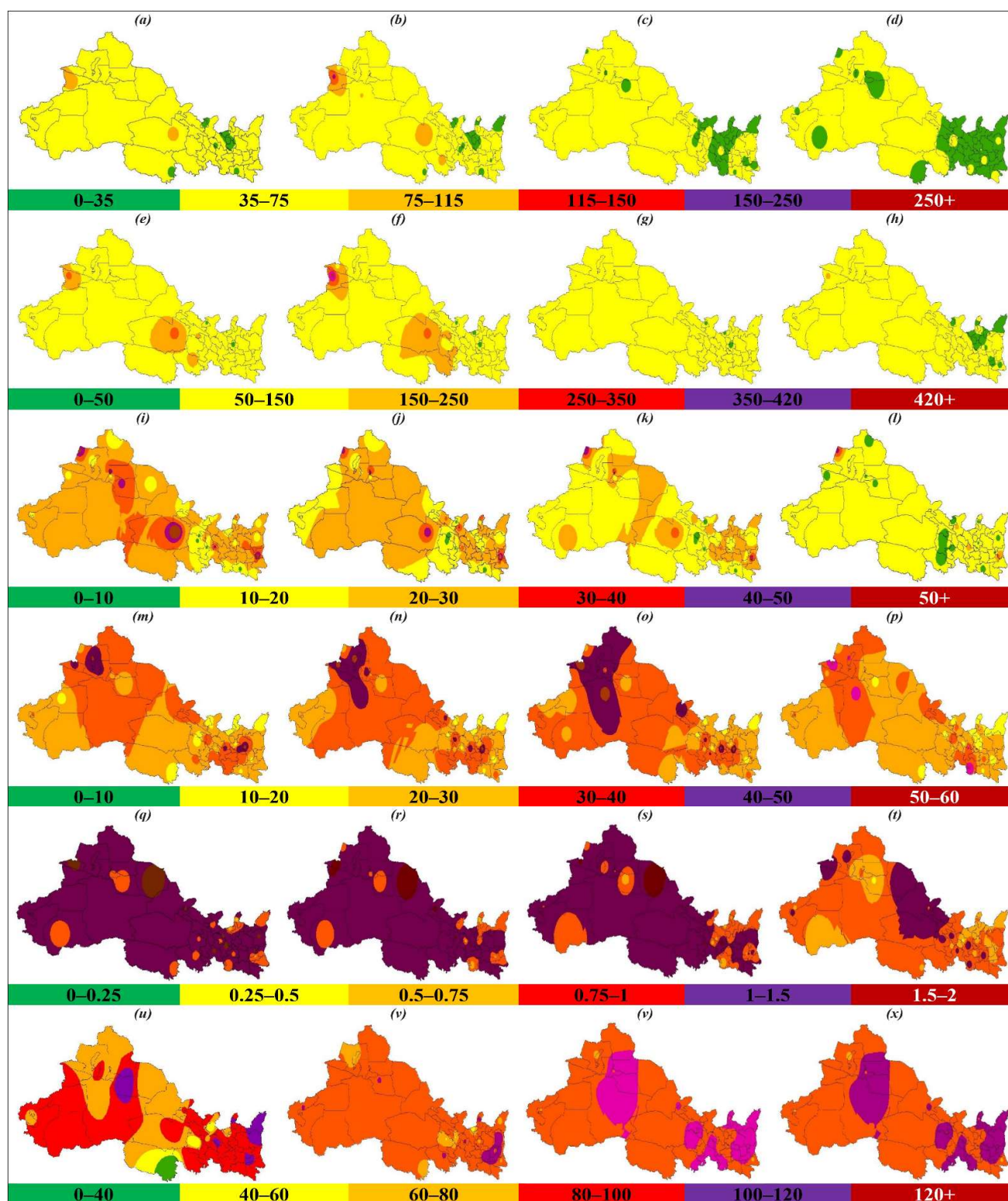


Figure 2. The spatial distribution of PM_{2.5} (a–d), PM₁₀ (e–h), SO₂ (i–l), NO₂ (m–p), CO (q–t), and O₃ (u–x) between 2015 and 2018 in northwest China (NWC). Color represents the different pollution levels, e.g., green (good), yellow (moderate), orange (unhealthy for the sensitive group), red (unhealthy for all), purple (very unhealthy), and maroon (hazardous). The abbreviations are as follows: PM_{2.5} (fine particulate matter), PM₁₀ (coarse particulate matter), SO₂ (sulfur dioxide), NO₂ (nitrogen dioxide), CO (carbon monoxide), and O₃ (ozone).

3.2. Seasonal Variation of Six Criteria Pollutants

In terms of seasonality, all the pollutants ($PM_{2.5}$, PM_{10} , SO_2 , NO_2 , and CO) observed the highest concentration in winter and the lowest occurred in summer except O_3 (vice versa) between 2015 and 2018 (Figure 3). $PM_{2.5}$ exceeded CAAQS Grade II ($75 \mu\text{g}/\text{m}^3$, daily mean) in SN, and XJ during winter (Figure 3a), while PM_{10} , SO_2 , NO_2 , CO, and O_3 complied with CAAQS Grade II ($150 \mu\text{g}/\text{m}^3$, $150 \mu\text{g}/\text{m}^3$, $80 \mu\text{g}/\text{m}^3$, $4 \text{ mg}/\text{m}^3$, and $160 \mu\text{g}/\text{m}^3$, daily mean) in SN, XJ, GS, NX, QH, and NWC during spring, summer, autumn, and winter (Figure 3b–f). The average concentration of $PM_{2.5}$, PM_{10} , SO_2 , NO_2 , and CO decreased in all seasons, e.g., spring, summer, autumn, and winter, while the average concentration of O_3 increased in all seasons between 2015 and 2018. Figure S2 explains the spatial distribution of the criteria pollutants in 53 cities of NWC during different seasons. $PM_{2.5}$ exceeded CAAQS Grade II standards (daily mean) in 1.89%, 5.56%, and 32.1% of the cities during spring, summer, and winter, respectively (Figure S2a). Similarly, PM_{10} exceeded the daily standard in 5.66%, 7.55%, 1.89%, and 26.4% of the cities in spring, summer, autumn, and winter, respectively (Figure S2b), while, SO_2 , NO_2 , CO, and O_3 complied with CAAQS Grade II standards (daily mean) in all the cities of NWC during all seasons (Figure S2c–f).

Figure 4 illustrates the monthly variation in the criteria pollutants, $PM_{2.5}/PM_{10}$ ratio, and AQI in NWC between 2015 and 2018. $PM_{2.5}$, PM_{10} , NO_2 , SO_2 , and CO explicated “U” shaped curves, with the highest concentration in winter (October to January) due to increased coal combustion for civil heating and stagnant meteorological conditions, e.g., lower wind speed, low temperature, etc., while the lowest concentration occurred in summer (June to August) due to seasonal rains and favorable atmospheric conditions that help in pollution dispersion. In the case of PM_{10} , higher pollution levels also occurred in spring (March to May) due to haze events. In contrast to other pollutants, the O_3 levels were higher in June to August (summer) and lower in winter.

3.3. $PM_{2.5}/PM_{10}$ Ratio

The $PM_{2.5}/PM_{10}$ ratio normally reflects the composition and quality of air, e.g., a higher $PM_{2.5}/PM_{10}$ ratio indicates the increased proportion of $PM_{2.5}$ and a lower $PM_{2.5}/PM_{10}$ ratio indicates a higher concentration of PM_{10} in the atmosphere. The annual average $PM_{2.5}/PM_{10}$ ratio in NWC during 2015 to 2018 was 0.480 ± 0.08 , 0.478 ± 0.07 , 0.483 ± 0.08 , and 0.478 ± 0.07 , respectively, and experienced a reduction of 0.43% over time (Figure 1g). The highest $PM_{2.5}/PM_{10}$ ratio occurred in SN followed by QH, GS, XJ, and NX. In terms of seasonality, the highest $PM_{2.5}/PM_{10}$ ratio occurred in winter, followed by autumn, spring, and summer, and experienced an average change of -2.77% , -4.98% , 1.82% , and 5.31% , respectively (Figure 3g). In the case of monthly variation, a “U” shaped curve was observed with the highest value in winter and the lowest value in the summer Figure 4g. Figure 5 illustrates the annual (a–d) and seasonal (spring (e–h), summer (i–l), autumn (m–p), and winter (q–t)) spatial distribution of $PM_{2.5}/PM_{10}$ ratio in NWC during 2015 to 2018. In 2018, 49.1% of the cities of NWC experienced an increase in the $PM_{2.5}/PM_{10}$ ratio as compared with 2015. Similarly, 49.1%, 62.3%, 35.9%, and 39.6% of the cities of NWC experienced increased $PM_{2.5}/PM_{10}$ ratio in spring, summer, autumn, and winter, respectively, in 2018 against 2015. From 2015 to 2018, approximately 1.89%, 1.89%, 3.74%, and 26.4% of the cities experienced $PM_{2.5}/PM_{10}$ ratio higher than 0.60 in spring, summer, autumn, and winter, respectively, in NWC.

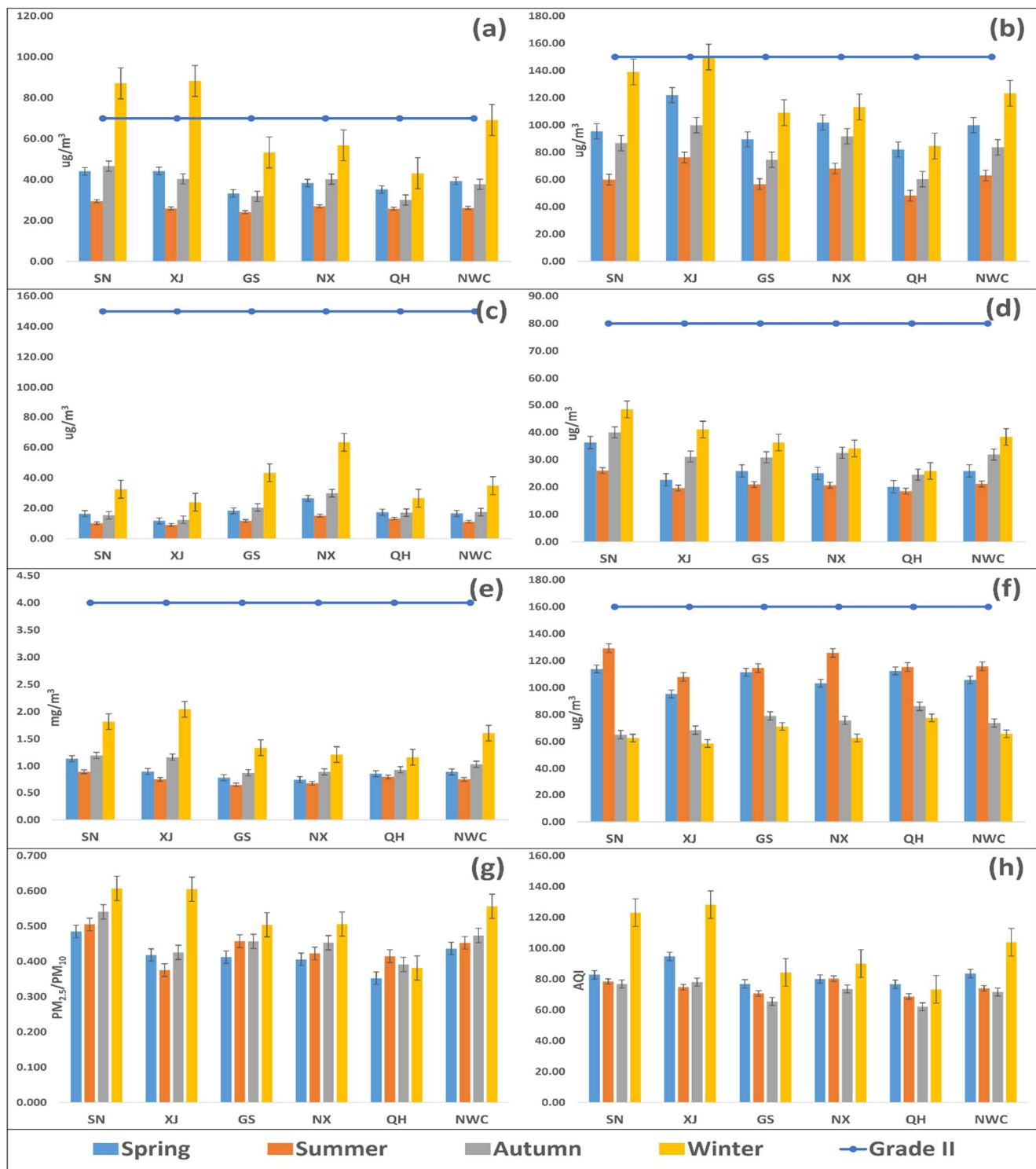


Figure 3. Seasonal variation of PM_{2.5} (a), PM₁₀ (b), SO₂ (c), NO₂ (d), CO (e), O₃ (f), PM_{2.5}/PM₁₀ ratio (g), and AQI (h) in five provinces (Shaanxi (SN), Xinjiang (XJ), Gansu (GS), Ningxia (NX), and Qinghai (QH)) of northwest China (NWC) and NWC as a whole between 2015 and 2018. Descriptions are as follows: light blue bar (spring), orange bar (summer), grey bar (autumn), yellow bar (winter), and blue line with dots (CAAQS Grade II standards, daily mean). The abbreviations are as follows: PM_{2.5} (fine particulate matter), PM₁₀ (coarse particulate matter), SO₂ (sulfur dioxide), NO₂ (nitrogen dioxide), CO (carbon monoxide), O₃ (ozone), PM_{2.5}/PM₁₀ (ratio of PM_{2.5} with PM₁₀), and AQI (air quality index).

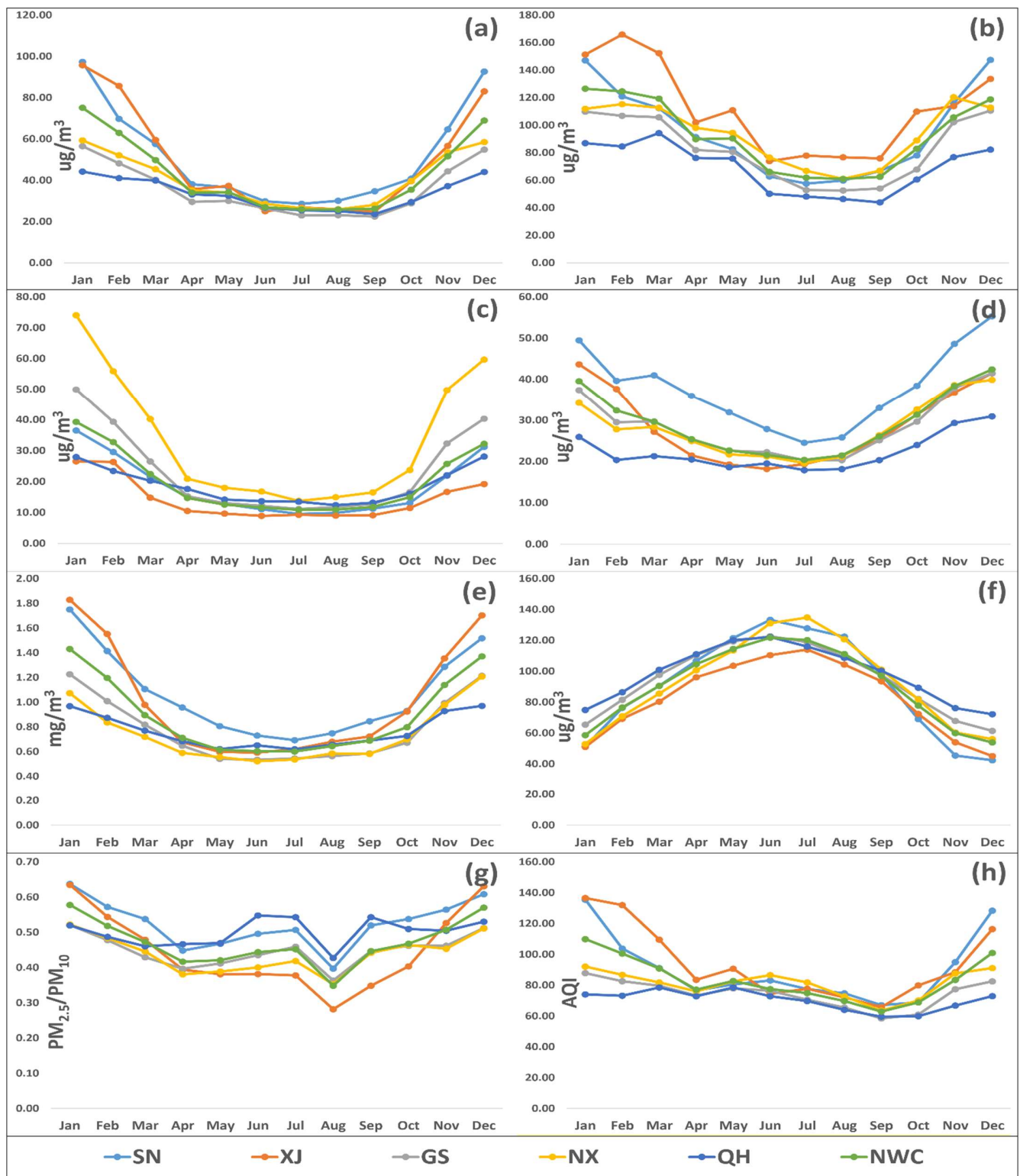


Figure 4. The monthly average concentration of PM_{2.5} (a), PM₁₀ (b), SO₂ (c), NO₂ (d), CO (e), O₃ (f), PM_{2.5}/PM₁₀ (g), and AQI (h) in five provinces (Shaanxi (SN), Xinjiang (XJ), Gansu (GS), Ningxia (NX), and Qinghai (QH)) of northwest China (NWC) between 2015 and 2018. Descriptions are as follows: light blue line with dots (SN), orange line with dots (XJ), grey line with dots (GS), yellow line with dots (NX), blue line with dots (QH), and parrot line with dots (NWC). The abbreviations are as follows: PM_{2.5} (fine particulate matter), PM₁₀ (coarse particulate matter), SO₂ (sulfur dioxide), NO₂ (nitrogen dioxide), CO (carbon monoxide), and O₃ (ozone).

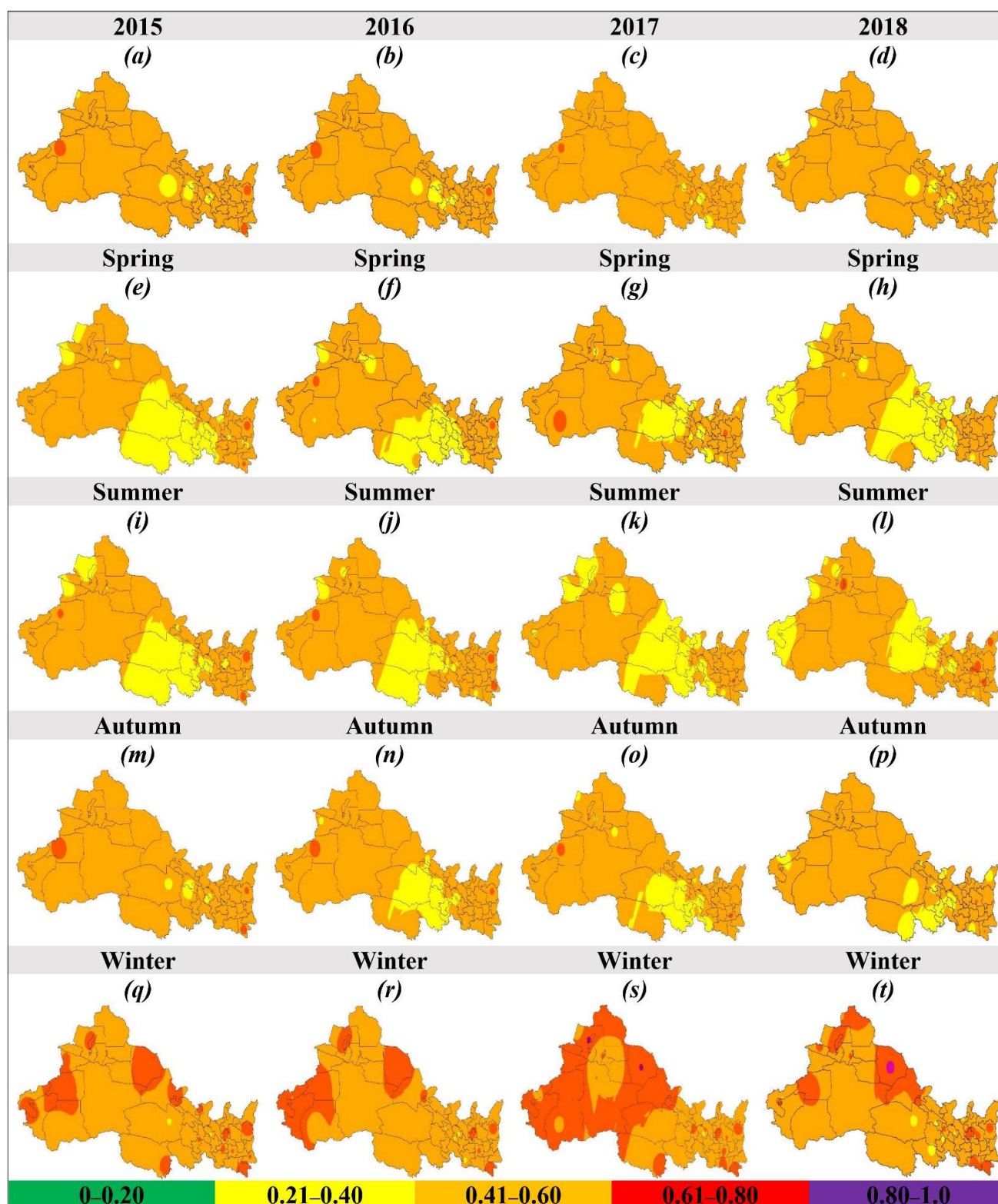


Figure 5. Annual (a–d) and seasonal (spring (e–h), summer (i–l), autumn (m–p), and winter (q–t)) spatial distribution of $PM_{2.5}/PM_{10}$ in 53 cities of northwest China (NWC) between 2015 and 2018. Color represents the different pollution levels, e.g., green (good), yellow (moderate), orange (unhealthy for a sensitive group), red (unhealthy for all), purple (very unhealthy), and maroon (hazardous).

3.4. Air Quality Index (AQI)

Air quality index (AQI) is a color-coded scale that simplifies different pollutants concentrations into a single numerical value that reflects overall air quality, health effects, sensitive groups, and required precautionary measures. During the study period (2015–2018), the annual average AQI in NWC was 88.1 ± 24.1 , 93.5 ± 36.3 , and 82 ± 18.7 , 69.2 ± 14.8 , respectively, and improved by 21.3% (Figure 1h). The highest AQI occurred in XJ, followed by SN, NX, GS, and QH. The AQI improved in all the cities except a few cities in SN (Weinan, Xian), and XJ (Changji, Ili, Shihezi, Tacheng, Wujiaqu). In 2018, the average AQI was under the threshold value of 100 in all the cities except Shihezi and Wujiaqu in Xinjiang (Figure 6d).

In the case of seasonal variation, the highest AQI occurred in winter, followed by spring, summer, and autumn, respectively, and improved by 17.5%, 30.8%, 18.7%, 17.4%, respectively, in NWC during 2018 as compared with 2015. The seasonal variation was consistent throughout NWC, e.g., highest AQI in winter and lowest in autumn, except XJ (Figure 3h). Figure 6 illustrates the seasonal (spring (e–h), summer (i–l), autumn (m–p), and winter (q–t)) spatial distribution of AQI in NWC between 2015 and 2018. In different seasons, e.g., spring, summer, autumn, and winter, the number of cities exceeding the AQI threshold value of 100 decreased from 24.5% to 0% (Figure 6e–h), 7.55% to 0% (Figure 6i–l), 7.55% to 0% (Figure 6m–p), and 50.9% to 22.6% (Figure 6q–t), respectively, in NWC. In the case of the monthly variation, a “U” shaped curve was observed with the highest value in winter and the lowest value in summer Figure 4h.

3.5. Proportion of Six Air Quality Index (AQI) Classes

Figure 7 explains the annual (a–d) and seasonal (spring (e–h), summer (i–l), autumn (m–p), and winter (q–t)) proportion of AQI classes in NWC during 2015 to 2018. During the study period, the average proportion of Class I, Class II, Class III, Class IV, Class V, and Class VI accounted for 17.2%, 63.1%, 13.1%, 3.01%, 2.14%, and 1.03% of the days, respectively. In NWC, the proportion of Class I, Class II, Class III, Class IV, Class V, and Class VI experienced an average change of 114.1%, -1.08% , -55.2% , -29.5% , -69.2% , and -58.3% , respectively, in 2018, with respect to 2015. The combined proportion of Class I and II increased by 18.5%, with the highest increase in spring (28.73%), followed by winter (27.2%), autumn (11.2%), and summer (10.9%), indicating significant improvement in air quality over the time span.

3.6. The Major Pollutants/Primary Pollutants

During the study period, PM_{10} was a major pollutant, accounting for more than 32.9% of the days, followed by O_3 (25.9%), $PM_{2.5}$ (16.4%), NO_2 (3.52%), CO (1.43%), and SO_2 (1.01%) in NWC (Figure 8). In 2018, the number of days with PM_{10} , $PM_{2.5}$, SO_2 , and CO as major pollutants decreased by 35%, 38%, 52%, and 90%, respectively, and increased by 46% and 11% for O_3 and NO_2 , respectively. PM_{10} was a major pollutant in autumn (41.3%), spring (39.1%), and a second major pollutant in winter (34.3%), while $PM_{2.5}$ was a major pollutant in winter (42.1%), and O_3 was a major pollutant (59.7%) in summer. The number of days with O_3 as a major pollutant was higher in the hotter months March–September), and, for $PM_{2.5}$, it was higher in the colder months (November–February). During the study period, the number of days with PM_{10} as a major pollutant decreased by 47.2%, 76.3%, and 19.7% in spring, summer, and autumn, respectively, and increased by 4.58% in winter 2018 as compared with 2015. Similarly, the number of days with $PM_{2.5}$ as a major pollutant decreased by 62.3%, 90.8%, 42.9%, and 24.6% in spring, summer, autumn, and winter, respectively. The number of days with SO_2 as a major pollutant decreased in all the seasons except summer, NO_2 decreased in summer and winter, CO decreased in spring, summer, autumn, and winter, while O_3 increased in all the seasons, e.g., in spring, summer, autumn, and winter 2018 as compared with 2015.

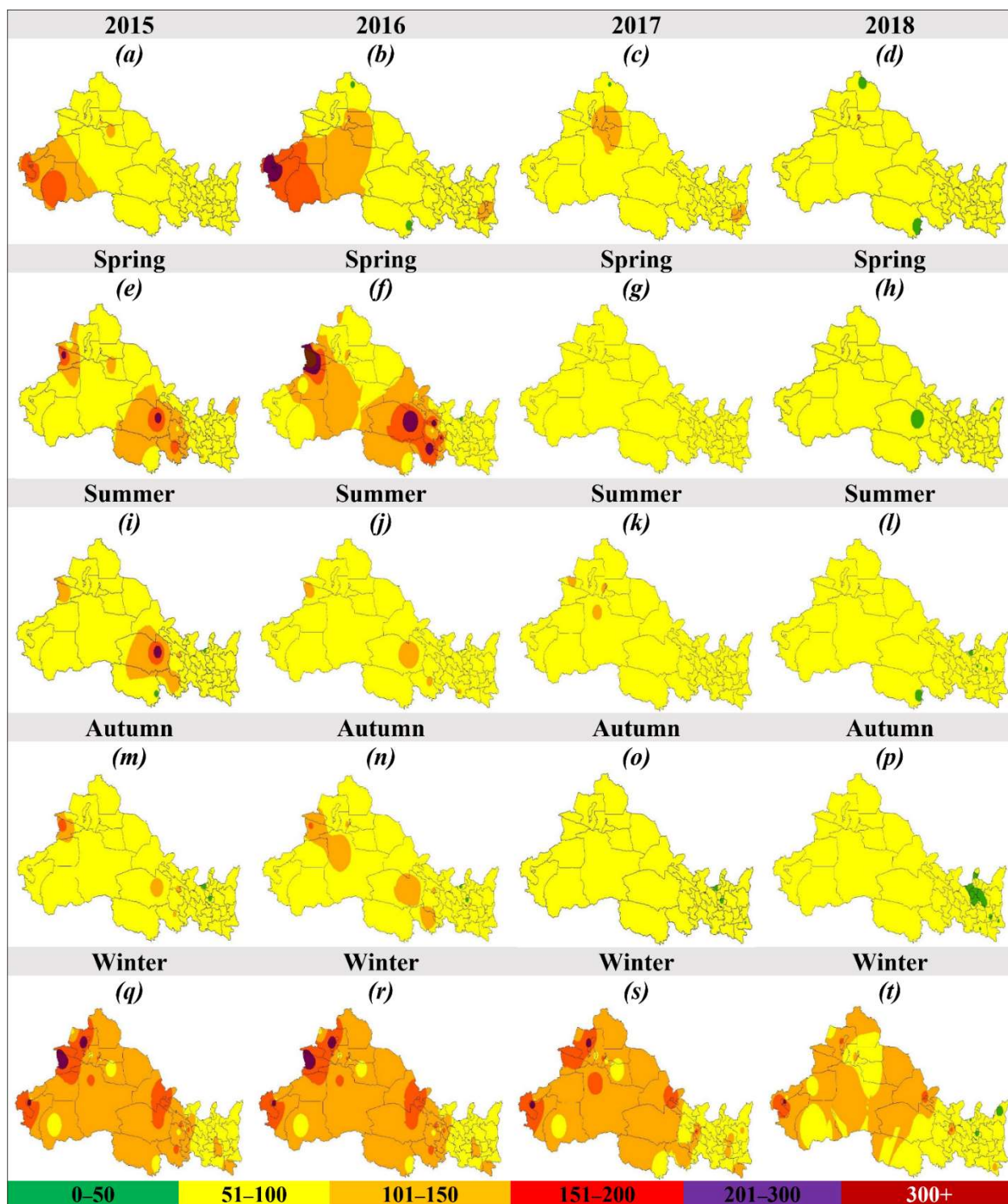


Figure 6. Annual (a–d) and seasonal (spring (e–h), summer (i–l), autumn (m–p), and winter (q–t)) spatial distribution of AQI between 2015 and 2018 in 53 cities of northwest China (NWC). Color represents the different classes of air quality index, e.g., green (0–50, good), yellow (51–100, moderate), orange (101–150, unhealthy for the sensitive group), red (151–200, unhealthy for all), purple (201–300, very unhealthy), and maroon (300+, hazardous).

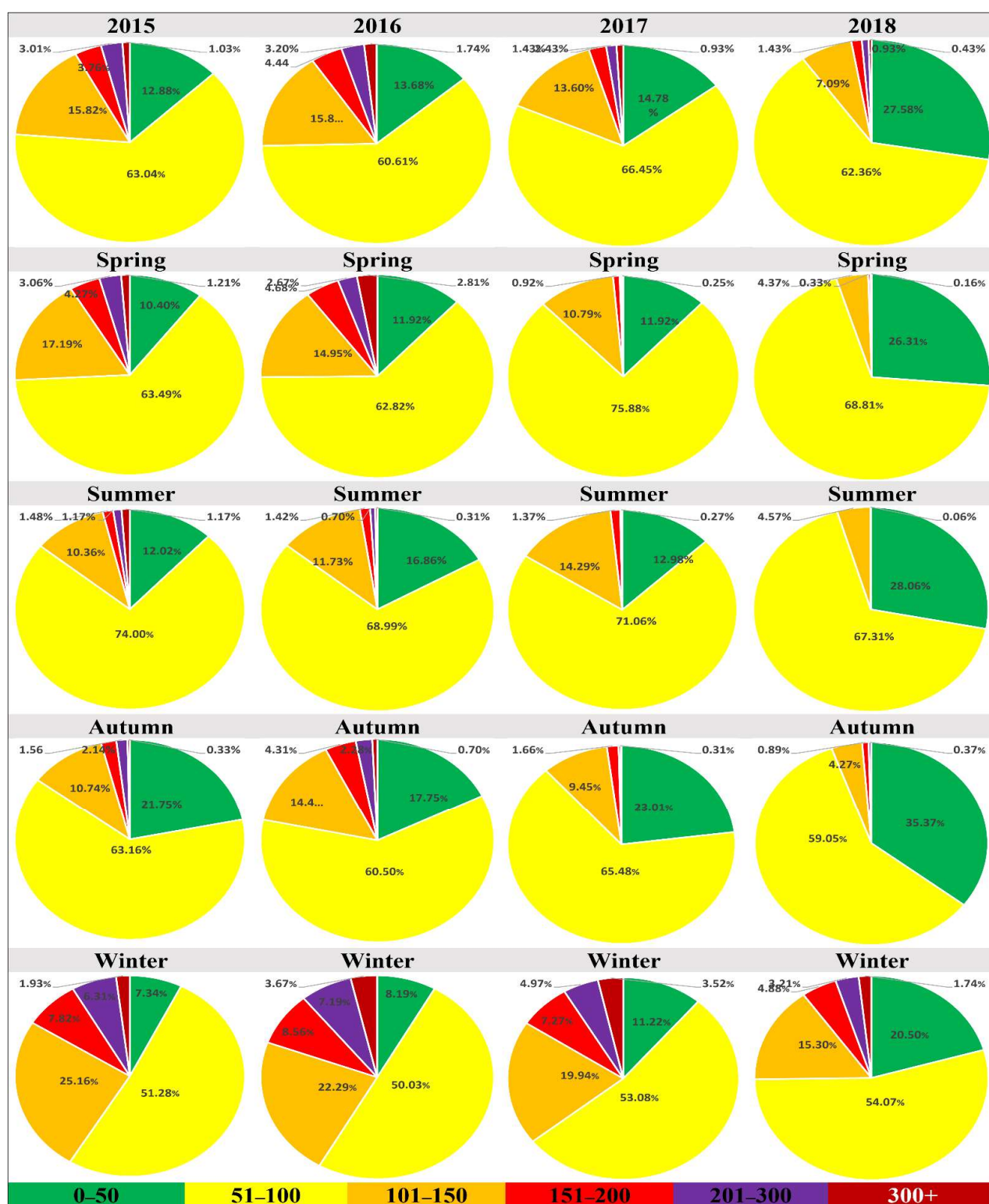


Figure 7. The annual (1st row) and seasonal (spring (2nd row), summer (3rd row), autumn (4th row), and winter (5th row)) distribution of AQ class, e.g., Class I (0–50, good, green), Class II (51–100, moderate, yellow), Class III (101–150, unhealthy for a sensitive group, orange), Class IV (151–200, unhealthy for all, red), Class V (201–300, very unhealthy, purple), and Class VI (300+, hazardous, maroon) in northwest China (NWC) between 2015 and 2018.

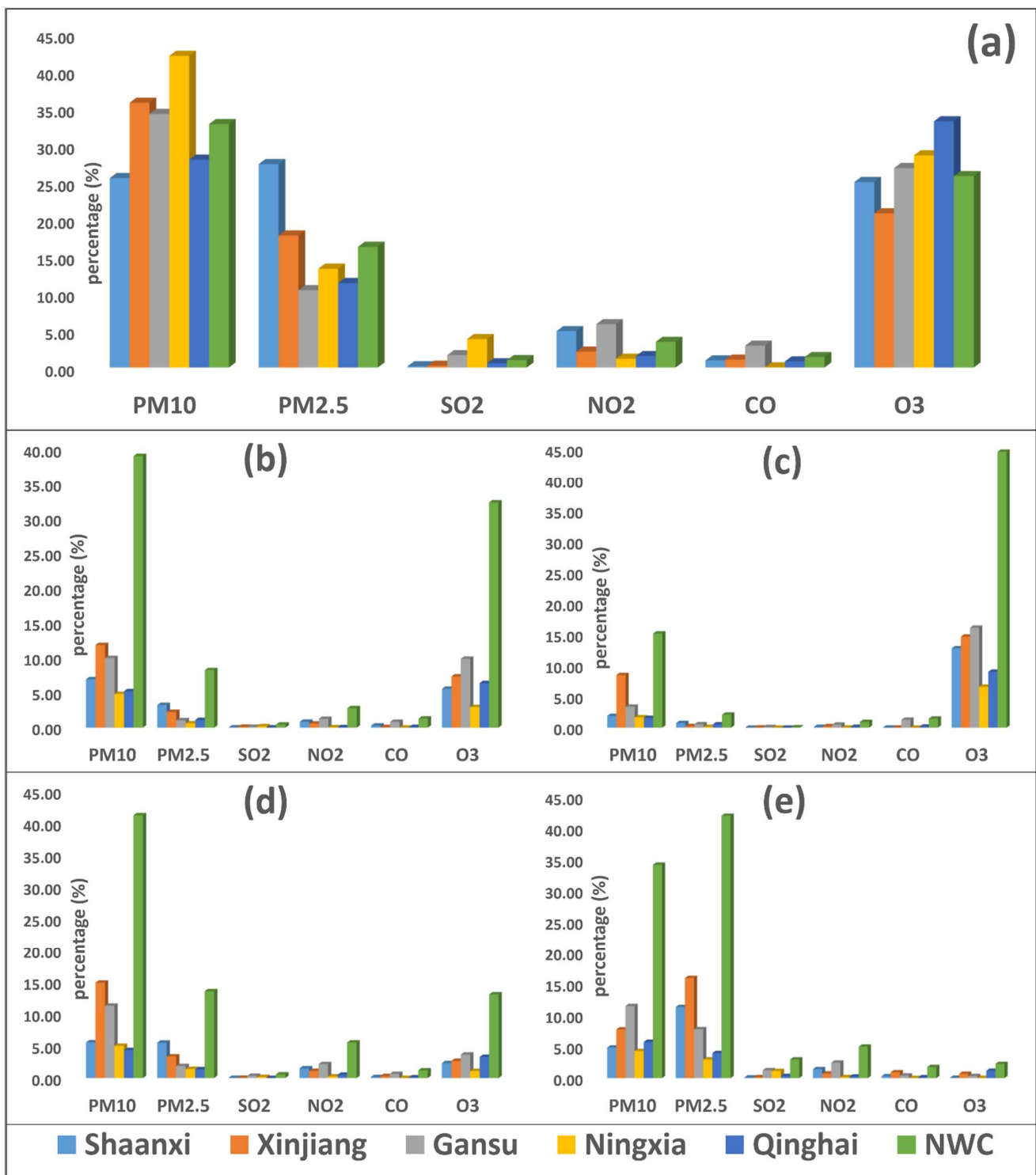


Figure 8. The annual (a) and seasonal (spring (b), summer (c), autumn (d), winter (e)) percentage of days with different primary pollutants (PM_{2.5}, PM₁₀, SO₂, NO₂, CO, and O₃) in five provinces (Shaanxi (SN), Xinjiang (XJ), Gansu (GS), Ningxia (NX), and Qinghai (QH)) of northwest China (NWC) between 2015 and 2018. Descriptions are as follows: light blue bar (Shaanxi), orange bar (Xinjiang), grey bar (Gansu), yellow bar (Ningxia), blue bar (Qinghai), and parrot bar (NWC). The abbreviations are as follows: PM_{2.5} (fine particulate matter), PM₁₀ (coarse particulate matter), SO₂ (sulfur dioxide), NO₂ (nitrogen dioxide), CO (carbon monoxide), and O₃ (ozone).

3.7. Pollution Days/Non-Attainment Days

Any day with one or more pollutants exceeding CAAQS (Grade II) standards is considered as a non-attainment/pollution day. In NWC, the proportion of non-attainment days were 23.3%, 23.9%, 16.2%, and 8.9% during 2015 to 2018, respectively. The proportion of non-attainment days decreased by 61.77% (24%, 47.1%, 79.1%, 77.9%, and 73.9%) in NWC (SN, XJ, GS, NX, and QH) during 2015 to 2018 (Table 1). The highest reduction in the proportion of non-attainment days occurred in spring (79.6%), summer (63.1%), autumn (60.4%), and winter (37.6%), respectively.

3.8. Statistical Analysis

The result of Pearson's correlation (Table S2) indicated that AQI was strongly positively correlated ($R > 0.5$) with $PM_{2.5}$, PM_{10} , SO_2 , NO_2 , and CO on an annual basis and strongly anti-correlated ($R > -0.5$) with O_3 in the NWC (Figure S3a). The seasonal variation in the correlation between AQI and different pollutants was evident (Figure S3b–e). In spring, AQI was strongly correlated ($R > -0.5$) with $PM_{2.5}$, PM_{10} , SO_2 , and CO, moderately correlated ($R > -0.3$) with NO_2 , while strongly anti-correlated ($R > -0.5$) with O_3 (Figure S3b). In summer, AQI was strongly correlated ($R > 0.5$) with $PM_{2.5}$, PM_{10} , and O_3 , weakly correlated ($R > 0.2$) with SO_2 , weakly anti-correlated ($R > -0.1$) with CO, and strongly anti-correlated ($R > -0.5$) with NO_2 (Figure S3c). In autumn, AQI was strongly correlated ($R > 0.5$) with $PM_{2.5}$, PM_{10} , SO_2 , NO_2 , and CO, while strongly anti-correlated ($R > -0.5$) with O_3 (Figure S3d). In winter, AQI was strongly correlated ($R > 0.5$) with $PM_{2.5}$, PM_{10} , SO_2 , and CO, moderately correlated ($R > 0.3$) with NO_2 , and moderately anti-correlated ($R > -0.3$) with O_3 (Figure S3e). Throughout the study period, all the pollutants were positively correlated with each other except O_3 (Table S2).

4. Discussion

During the study period, the average concentrations of $PM_{2.5}$, PM_{10} , SO_2 , NO_2 , CO, and O_3 , decreased in 92.5%, 96.2%, 92.5%, 64.5%, 88.7%, and 11.3% of the 53 cities in NWC. Based on the results above, we concluded that strict environmental regulations have significantly improved the air quality in NWC between 2015 and 2018 [3,20–23]. $PM_{2.5}$ mainly originates from industrial activities, coal consumption, power generation, biomass burning, automobile exhausts, construction activities, road dust, etc. [28–30,50–52]. From 2015 to 2018, the average concentration of $PM_{2.5}$ decreased in all the cities except a few cities (Changji, Ili, Shihezi, Wujiaqu) in the northern part of XJ. Higher pollution in XJ, SN, and GS, $PM_{2.5}$ hotspots [53,54], is because of increased coal-based industry, vehicular emission, civil heating, construction activities, natural sources (dust storms), and adverse meteorology [55–60]. PM_{10} mainly originates from natural sources, e.g., sand storms, haze events, etc., as well as from anthropogenic sources, e.g., developmental activities, industrial emissions, traffic emissions, road dust, etc. [15,26–28]. The highest pollution levels occurred in XJ, followed by SN, NX, GS, and QH. Elevated pollution levels in southern Xinjiang (Kashgar) indicate the influences of emissions from natural sources, e.g., Taklimakan deserts, dust storms, haze events, etc. [61–64]. Similarly, higher particulate pollution in Shaanxi (FWP region) is associated with increased anthropogenic emissions, e.g., industrial activities, construction activities, etc. [55]. All the cities of NWC experienced a reduction in PM_{10} except Shihezi and Wujiaqu in northern XJ, indicating the influence of both manmade and natural emission sources. This decrease is associated with strict environmental regulations, e.g., Chinese Ambient Air Quality Standards (CAAQS) (GB 3095–2012) [17], Atmospheric Pollution Prevention and Control Action Plan (APPCAP, 2013) [18], technical regulation on ambient air quality index (HJ 633–2012) [19], the establishment of nationwide air quality monitoring stations, etc. [3,20–23].

In the case of gaseous pollutants, the average concentration of SO_2 , NO_2 , and CO decreased by 41.9%, 6.19%, and 27.3%, respectively, in NWC. Industrial emissions, coal burning, fossil fuel burning, power generation, traffic exhausts, etc. are major sources of SO_2 , NO_2 , and CO [65–69]. The concentration of SO_2 decreased in all the cities except

Ili, Hami (XJ), Jiuquan (GS), and Yushu (QH), indicating the influence of increased coal combustion, vehicular exhaust, and industrial emission [27,28,57,70]. In the case of NO_2 , the average concentration decreased in 71.7% of the cities of NWC between 2015 and 2018. The highest pollution level occurred in the provincial capitals, e.g., Xi'an (SN), Urumqi (XJ), Lanzhou (GS), etc., and major cities (Ili, Hami, Jiuquan, Yushu, etc.), indicating increased fossil fuel combustion, e.g., automobile exhaust, industrial emission, etc. [34,63,65–69]. Similarly, the highest concentration of CO occurred in SN, followed by XJ, QH, GS, and NX. Higher CO indicates incomplete combustion of fossil fuels, biomass burning, industrial emission, and causes multiple health disorders, e.g., hypoxia, major heart and neural disorders [28,67,68]. In contrast to other pollutants, O_3 increased by 3.69% in NWC during the study period. All the cities experienced an increase in O_3 concentration except Ankang, Shanglou, Longnan, Haibei, Hainan, and Haixi. This increase in O_3 is associated with a decrease in $\text{PM}_{2.5}$ and other pollutants, which slows down the sink of hydroperoxy radicals and helps in the accumulation of ozone [55].

In terms of seasonality, $\text{PM}_{2.5}$, PM_{10} , SO_2 , NO_2 , and CO experienced the same seasonal variation, e.g., highest in winter and lowest in summer. Higher pollution in winter is associated with increased coal combustion, civil heating, power generation, fossil fuel burning, industrial activity, vehicular exhausts, and adverse/stagnant meteorology [13–15,26–30,50–52,71–75]. In the case of PM_{10} , higher pollution levels also occurred from March to May (spring) due to haze events [76,77]. In contrast to other pollutants, the concentration of O_3 was highest in summer and lowest in winter [14,28,30,78]. Ozone is a secondary pollutant, formed due to a photochemical reaction between VOCs and NO_x [43,44,79]. The concentration of ozone in the summer was approximately twice that in winter due to lower NO_x levels in winter as NO_x levels decrease the O_3 depletion and enhance the accumulation of O_3 [80]. Similarly, higher temperatures, e.g., in summer, favor the accumulation of ozone [81–83].

The $\text{PM}_{2.5}/\text{PM}_{10}$ ratio reflects air quality, pollution sources, and origin, e.g., a higher $\text{PM}_{2.5}/\text{PM}_{10}$ ratio indicates the increased proportion of $\text{PM}_{2.5}$, mainly emitted from anthropogenic activities, and a lower ratio indicates an increased proportion of PM_{10} , mainly from natural activities [28,61–64]. During the study period (2015–2018), the $\text{PM}_{2.5}/\text{PM}_{10}$ ratio slightly decreased by 0.43% and in 50.9% of the cities of NWC. This decrease is associated with a reduction in $\text{PM}_{2.5}$ over time. In general, the $\text{PM}_{2.5}/\text{PM}_{10}$ ratio was higher in winter (low temperature) as compared with summer (high temperature) due to increased anthropogenic activities that release a significant amount of $\text{PM}_{2.5}$ and stable atmospheric conditions that help the accumulation of pollution [26–29,53,54,84].

During the study period (2015–2018), the AQI improved by 21.3%, and 86.8% of the cities of NWC experienced AQI improvement. This improvement is associated with a reduction in the criteria pollutants over time. AQI crossed the threshold value of 100 in 10 cities, out of which seven cities are in Xinjiang (Kashgar, Hotan, Aksu, Wujiaqu, Crete, Urumqi, Turpan) and three cities are in Shaanxi (Xiangyang, Xi'an, Weinan) (Table S1). The higher AQI values in Xinjiang and Shaanxi are associated with the increased coal-based industry, civil heating, and vehicular exhaust [3,14,30,51,52,56,69,85]. In the case of seasonal variations, the highest AQI occurred in winter due to increased anthropogenic emissions and stable atmospheric conditions [15,58,84]. In NWC, the proportion of AQI “Class I” improved by 114.1%, while the proportion of Class II, Class III, Class IV, Class V, and Class VI decreased by 1.08%, 55.2%, 29.5%, 69.2%, and 58.3%, respectively, during 2015 to 2018. The proportion of AQI “Class I” improved from 12.9% in 2015 to 27.6% in 2018. Similarly, the proportion of AQI “Class I” improved in all the provinces, e.g., SN, XJ, GS, NX, and QH, in all the seasons, and improved by 2.53 times, 2.33 times, 1.63 times, and 2.79 times in spring, summer, autumn, and winter in 2018 as compared with 2015, which indicates improvement in air quality [28,57].

During the study period (2015–2018), the proportion of days with PM_{10} , $\text{PM}_{2.5}$, SO_2 , and CO as a major pollutant decreased by 35%, 38%, 52%, and 90%, respectively, due to strict environmental legislation [20–23]. $\text{PM}_{2.5}$ was a major pollutant in winter (42%), indicating

anthropogenic emissions, e.g., coal burning, civil heat, industrial emissions, and vehicular emissions [28,56–58]. Similarly, O₃ was a major pollutant (44.7%) in summer due to lower NO_x levels as lower NO_x levels prevent ozone depletion and the higher temperature in summer favors ozone generation and accumulation [78,80,82,83,86–88], while PM₁₀ was a major pollutant in autumn (41.3%), spring (39.1%), and the second major pollutant in winter (34.3%). The number of days with PM₁₀ as a major pollutant was higher in southern Xinjiang due to emissions from natural sources, e.g., the Taklimakan desert, sand storms [61–64,89]. Any day with one or more pollutants exceeding CAAQS (Grade II) is considered as a non-attainment/pollution day [2,30]. During the study period, the proportion of non-attainment days decreased by 61.8% in NWC. Similarly, the proportion of non-attainment days decreased in all the provinces, e.g., SN, XJ, GS, NX, and QH, and experienced a reduction of 79.6%, 63.1%, 60.4%, and 37.6% in spring, summer, autumn, and winter, respectively, which clearly indicates that the ambient air quality improved significantly.

Rapid economic development, industrialization, haze events, dust storms, and adverse meteorological conditions play a crucial role in air quality deterioration [2,4–10]. The Chinese government is working proactively to combat the pollution levels by revising and implementing strict environmental regulations [3,20,21]. According to this study, the concentration of PM_{2.5}, PM₁₀, SO₂, NO₂, and CO, AQI, the proportion of AQI “Class I”, and pollution days decreased significantly in NWC between 2015 and 2018.

5. Conclusions

In this study, we examined the spatial and temporal variation of ambient air quality in northwest China (NWC) for a period of four years (2015–2018). During the study period, the average concentration of PM_{2.5}, PM₁₀, SO₂, NO₂, CO, and O₃ decreased in 92.5%, 96.2%, 92.5%, 64.5%, 88.7%, and 11.3% of the cities in NWC. The annual average concentration of particulate matter (PM_{2.5} and PM₁₀) exceeded the CAAQS Grade II standards and WHO recommended air quality guidelines in NWC, while the annual average concentration of SO₂ and NO₂ complied with the CAAQS Grade II standards in NWC. In the case of seasonality, the highest pollution level occurred in winter except for ozone, with varying degrees of spatial distribution. The AQI, the proportion of AQI Class I, and the number of pollution days improved by 21.3%, 114.1%, and 61.8%, respectively, in NWC. The AQI improved in all the seasons, with the maximum improvement in spring followed by summer, winter, and autumn. In NWC, PM₁₀ was a major pollutant for most of the days, followed by O₃, PM_{2.5}, NO₂, CO, and SO₂ with different spatial and temporal variations. A strong correlation occurred between AQI and all the pollutants except O₃. Stricter regulations, e.g., a three-year action plan to win the blue sky defense war, sector-specific guidelines, and strict enforcement of environmental legislation, are the keys to pollution-free and breathable air. This paper comprehensively discussed the spatio-temporal characteristics of the ambient air quality in NWC and calls for future detailed assessment focusing on source apportionment, health risk assessment, the impact of meteorology, dispersion modeling, and impact of the chemical processes that influence the air quality.

Supplementary Materials: The following are available online at <https://www.mdpi.com/article/10.3390/atmos13030375/s1>, Figure S1: The locations of 53 cities in five provinces (Shaanxi (SN), Xinjiang (XJ), Gansu (GS), Ningxia (NX), and Qinghai (QH)) northwest China (NWC). Color represents the different classes of air quality index, e.g., green (0–50, good), yellow (51–100, moderate), orange (101–150, unhealthy for a sensitive group), red (151–200, unhealthy for all), purple (201–300, very unhealthy), and maroon (300+, hazardous); Figure S2: The seasonal (spring (light blue line), summer (orange line), autumn (grey line), and winter (yellow line)) spatial distribution of PM_{2.5} (a), PM₁₀ (b), SO₂ (c), NO₂ (d), CO (e), and O₃ (f) in 53 cities of northwest China between 2015 and 2018. Descriptions are as follows: light blue line with dots (spring), orange line with dots (summer), grey line with dots (autumn), yellow line with dots (winter), and the blue line (CAAQS, daily mean). The abbreviations are as follows: PM_{2.5} (fine particulate matter), PM₁₀ (coarse particulate matter), SO₂ (sulfur dioxide), NO₂ (nitrogen dioxide), CO (carbon monoxide), and O₃ (ozone); Figure S3: Annual

(a) and seasonal (spring (b), summer (c), autumn (d), winter (e)) relationship between air quality index (AQI) and criteria pollutants (PM_{2.5}, PM₁₀, SO₂, NO₂, CO, and O₃). The abbreviations are as follows: AQI (air quality index), PM_{2.5} (fine particulate matter), PM₁₀ (coarse particulate matter), SO₂ (sulfur dioxide), NO₂ (nitrogen dioxide), CO (carbon monoxide), and O₃ (ozone); Table S1: Lists of cities, their rankings in five provinces (Shaanxi (SN), Xinjiang (XJ), Gansu (GS), Ningxia (NX), and Qinghai (QH)) of northwest China (NWC) between 2015 and 2018; Table S2: Pearson correlation between AQI and six criteria pollutants (PM_{2.5}, PM₁₀, SO₂, NO₂, CO, and O₃) in northwest China (NWC) between 2015 and 2018. The abbreviations are as follows: AQI (air quality index), PM_{2.5} (fine particulate matter), PM₁₀ (coarse particulate matter), SO₂ (sulfur dioxide), NO₂ (nitrogen dioxide), CO (carbon monoxide), and O₃ (ozone).

Author Contributions: Conceptualization, S.Z.; data curation/accumulation, S.Z.; investigation and data validation, J.L. and M.B.; methodology, S.Z.; data modelling, S.Z.; GIS mapping, S.Z.; writing—original draft preparation, S.Z.; review and editing, J.L. and M.B.; supervision, J.L.; funding acquisition, J.L. All authors have read and agreed to the published version of the manuscript.

Funding: This research work is supported by the National Natural Science Foundation of China (No. 21667026), the Social Science Foundation of Xinjiang Production and Construction Corps (No. 18YB13), and the Startup Foundation for Introduction Talent of NUIST (2017r107).

Institutional Review Board Statement: Not applicable.

Informed Consent Statement: Not applicable.

Data Availability Statement: The data presented in this study are available upon request from the corresponding author.

Acknowledgments: I would like to acknowledge the China National Environmental Monitoring Center (CNEMC) for the provision of air quality data. I am also very thankful to Jianjiang Lu and research scholar Anam Arshad from the School of Chemistry and Chemical Engineering, Shihezi University for helping and guiding me throughout the preparation of this paper.

Conflicts of Interest: The authors declare no conflict of interest.

References

- Li, R.; Wang, Z.; Cui, L.; Fu, H.; Zhang, L.; Kong, L.; Chen, J. Air pollution characteristics in China during 2015–2016: Spatiotemporal variations and key meteorological factors. *Sci. Total Environ.* **2019**, *648*, 902–915. [[CrossRef](#)] [[PubMed](#)]
- Luo, H.; Guan, Q.; Lin, J.; Wang, Q.; Yang, L.; Tan, Z.; Wang, N. Air pollution characteristics and human health risks in key cities of northwest China. *J. Environ. Manage.* **2020**, *269*, 110791. [[CrossRef](#)] [[PubMed](#)]
- Maji, K.J. Substantial changes in PM_{2.5} pollution and corresponding premature deaths across China during 2015–2019: A model prospective. *Sci. Total Environ.* **2020**, *729*, 138838. [[CrossRef](#)] [[PubMed](#)]
- Qiu, Y.; Yang, T.; Seyler, B.C.; Wang, X.; Wang, Y.; Jiang, M.; Li, F. Ambient air pollution and male fecundity: A retrospective analysis of longitudinal data from a Chinese human sperm bank (2013–2018). *Environ. Res.* **2020**, *186*, 109528. [[CrossRef](#)] [[PubMed](#)]
- Yang, J.; Ji, Z.; Kang, S.; Zhang, Q.; Chen, X.; Lee, S.Y. Spatiotemporal variations of air pollutants in western China and their relationship to meteorological factors and emission sources. *Environ. Pollut.* **2019**, *254*, 112952. [[CrossRef](#)]
- Zhao, S.; Yin, D.; Yu, Y.; Kang, S.; Qin, D.; Dong, L. PM_{2.5} and O₃ pollution during 2015–2019 over 367 Chinese cities: Spatiotemporal variations, meteorological and topographical impacts. *Environ. Pollut.* **2020**, *264*, 114694. [[CrossRef](#)]
- Pu, H.; Luo, K.; Wang, P.; Wang, S.; Kang, S. Spatial variation of air quality index and urban driving factors linkages: Evidence from Chinese cities. *Environ. Sci. Pollut. Res.* **2017**, *24*, 4457–4468. [[CrossRef](#)]
- Wang, W.; Cui, K.; Zhao, R.; Hsieh, L.T.; Lee, W.J. Characterization of the air quality index for Wuhu and Bengbu cities, China. *Aerosol Air Qual. Res.* **2018**, *18*, 1198–1220. [[CrossRef](#)]
- Shao, Z.; Jia, X.; Zhong, X.; Wang, D.; Wei, J.; Wang, Y.; Chen, L. Detection, extinguishing, and monitoring of a coal fire in Xinjiang, China. *Environ. Sci. Pollut. Res.* **2018**, *25*, 26603–26616. [[CrossRef](#)]
- Soleimani, M.; Amini, N.; Sadeghian, B.; Wang, D.; Fang, L. Heavy metals and their source identification in particulate matter (PM_{2.5}) in Isfahan City, Iran. *J. Environ. Sci.* **2018**, *72*, 166–175. [[CrossRef](#)]
- Zhao, S.; Yu, Y.; Yin, D.; He, J.; Liu, N.; Qu, J.; Xiao, J. Annual and diurnal variations of gaseous and particulate pollutants in 31 provincial capital cities based on in situ air quality monitoring data from China National Environmental Monitoring Center. *Environ. Int.* **2016**, *86*, 92–106. [[CrossRef](#)] [[PubMed](#)]
- Fan, H.; Zhao, C.; Yang, Y. A comprehensive analysis of the spatio-temporal variation of urban air pollution in China during 2014–2018. *Atmos. Environ.* **2020**, *220*, 117066. [[CrossRef](#)]

13. Feng, J.; Yu, H.; Mi, K.; Su, X.; Li, Y.; Li, Q.; Sun, J. One-year study of PM_{2.5} in Xinxiang city, North China: Seasonal characteristics, climate impact and source. *Ecotoxicol. Environ. Saf.* **2018**, *154*, 75–83. [\[CrossRef\]](#) [\[PubMed\]](#)
14. Liu, B.; Wu, J.; Zhang, J.; Wang, L.; Yang, J.; Liang, D.; Zhang, Q. Characterization and source apportionment of PM_{2.5} based on error estimation from EPA PMF 5.0 model at a medium city in China. *Environ. Pollut.* **2017**, *222*, 10–22. [\[CrossRef\]](#)
15. Turap, Y.; Rekefu, S.; Wang, G.; Talifu, D.; Gao, B.; Aierken, T.; Nuerla, A. Chemical Characteristics and Source Apportionment of PM_{2.5} during winter in the Southern Part of Urumqi, China. *Aerosol Air Qual. Res.* **2019**, *19*, 1325–1337. [\[CrossRef\]](#)
16. Wang, H.; Chen, H. Understanding the recent trend of haze pollution in eastern China: Roles of climate change. *Atmos. Chem. Phys.* **2016**, *16*, 4205–4211. [\[CrossRef\]](#)
17. GB 3095-2012. *Chinese Ambient Air Quality Standards (CAAQS)*; Ministry of Environmental Protection of the Peoples Republic of China (MEP): Beijing, China, 2012.
18. Zhu, J.; Chen, L.; Liao, H. Multi-pollutant air pollution and associated health risks in China from 2014 to 2020. *Atmos. Environ.* **2022**, *268*, 118829. [\[CrossRef\]](#)
19. HJ 633-2012. *Technical Regulation on Ambient Air Quality Index (on Trial)*; Ministry of Environmental Protection of the Peoples Republic of China (MEP): Beijing, China, 2012.
20. Ye, W.F.; Ma, Z.Y.; Ha, X.Z. Spatial-temporal patterns of PM_{2.5} concentrations for 338 Chinese cities. *Sci. Total Environ.* **2018**, *631*, 524–533. [\[CrossRef\]](#)
21. Song, C.; Wu, L.; Xie, Y.; He, J.; Chen, X.; Wang, T.; Dai, Q. Air pollution in China: Status and spatiotemporal variations. *Environ. Pollut.* **2017**, *227*, 334–347. [\[CrossRef\]](#)
22. Guo, H.; Cheng, T.; Gu, X.; Wang, Y.; Chen, H.; Bao, F.; Meng, C. Assessment of PM_{2.5} concentrations and exposure throughout China using ground observations. *Sci. Total Environ.* **2017**, *601*, 1024–1030. [\[CrossRef\]](#)
23. Shen, Y.; Zhang, L.; Fang, X.; Ji, H.; Li, X.; Zhao, Z. Spatiotemporal patterns of recent PM_{2.5} concentrations over typical urban agglomerations in China. *Sci. Total Environ.* **2019**, *655*, 13–26. [\[CrossRef\]](#) [\[PubMed\]](#)
24. Guo, P.; Umarova, A.B.; Luan, Y. The spatiotemporal characteristics of the air pollutants in China from 2015 to 2019. *PLoS ONE* **2020**, *15*, e0227469. [\[CrossRef\]](#) [\[PubMed\]](#)
25. Li, T.; Guo, Y.; Liu, Y.; Wang, J.; Wang, Q.; Sun, Z.; He, M.Z.; Shi, X. Estimating mortality burden attributable to short-term PM_{2.5} exposure: A national observational study in China. *Environ. Intl.* **2019**, *125*, 245–251. [\[CrossRef\]](#)
26. Turap, Y.; Talifu, D.; Wang, X.; Abulizi, A.; Maihemuti, M.; Tursun, Y.; Rekefu, S. Temporal distribution and source apportionment of PM_{2.5} chemical composition in Xinjiang, NW-China. *Atmos. Res.* **2019**, *218*, 257–268. [\[CrossRef\]](#)
27. Chen, J.; Lu, J.; Ning, J.; Yan, Y.; Li, S.; Zhou, L. Characteristics and sources analysis of PM_{2.5} in a major industrial city of northern Xinjiang, China. *SN Appl. Sci.* **2020**, *2*, 1–9. [\[CrossRef\]](#)
28. Yin, Z.; Cui, K.; Chen, S.; Zhao, Y.; Chao, H.R.; Chang-Chien, G.P. Characterization of the air quality index for Urumqi and Turfan cities, China. *Aerosol Air Qual. Res.* **2019**, *19*, 282–306. [\[CrossRef\]](#)
29. Zhao, X.; Zhou, W.; Han, L.; Locke, D. Spatiotemporal variation in PM_{2.5} concentrations and their relationship with socioeconomic factors in China's major cities. *Environ. Intl.* **2019**, *133*, 105145. [\[CrossRef\]](#)
30. Wang, Y.; Ying, Q.; Hu, J.; Zhang, H. Spatial and temporal variations of six criteria air pollutants in 31 provincial capital cities in China during 2013–2014. *Environ. Intl.* **2014**, *73*, 413–422. [\[CrossRef\]](#)
31. Kuerban, M.; Waili, Y.; Fan, F.; Liu, Y.; Qin, W.; Dore, A.J.; Zhang, F. Spatio-temporal patterns of air pollution in China from 2015 to 2018 and implications for health risks. *Environ. Pollut.* **2020**, *258*, 113659. [\[CrossRef\]](#)
32. Cohen, A.J.; Brauer, M.; Burnett, R.; Anderson, H.R.; Frostad, J.; Estep, K.; Feigin, V. Estimates and 25-year trends of the global burden of disease attributable to ambient air pollution: An analysis of data from the Global Burden of Diseases Study 2015. *Lancet.* **2017**, *389*, 1907–1918. [\[CrossRef\]](#)
33. Zhu, G.; Hu, W.; Liu, Y.; Cao, J.; Ma, Z.; Deng, Y.; Wang, H. Health burdens of ambient PM_{2.5} pollution across Chinese cities during 2006–2015. *J. Environ. Manage.* **2019**, *24*, 250–256. [\[CrossRef\]](#) [\[PubMed\]](#)
34. Wang, Q.; Wang, J.; Zhou, J.; Ban, J.; Li, T. Estimation of PM_{2.5} associated disease burden in China in 2020 and 2030 using population and air quality scenarios: A modelling study. *Lancet Planet. Health.* **2019**, *3*, 71–80. [\[CrossRef\]](#)
35. Li, R.; Kou, X.; Geng, H.; Xie, J.; Tian, J.; Cai, Z.; Dong, C. Mitochondrial damage: An important mechanism of ambient PM_{2.5} exposure-induced acute heart injury in rats. *J. Hazard. Mater.* **2015**, *287*, 392–401. [\[CrossRef\]](#)
36. Mannucci, P.M.; Franchini, M. Health effects of ambient air pollution in developing countries. *Int. J. Environ. Res. Public Health* **2017**, *14*, 1048. [\[CrossRef\]](#) [\[PubMed\]](#)
37. Zhang, J.; Liu, Y.; Cui, L.L.; Liu, S.Q.; Yin, X.X.; Li, H.C. Ambient air pollution, smog episodes and mortality in Jinan, China. *Sci. Rep.* **2017**, *7*, 1–8. [\[CrossRef\]](#)
38. Xie, Y.; Dai, H.; Zhang, Y.; Wu, Y.; Hanaoka, T.; Masui, T. Comparison of health and economic impacts of PM_{2.5} and ozone pollution in China. *Environ. Int.* **2019**, *130*, 104881. [\[CrossRef\]](#)
39. HJ 630-2011. *Technical Guideline on Environmental Monitoring Quality Management*; Ministry of Environmental Protection of the People Republic of China (MEP): Beijing, China, 2011.
40. Thach, T.Q.; Tsang, H.; Cao, P.; Ho, L.M. A novel method to construct an air quality index based on air pollution profiles. *Int. J. Hyg. Environ. Health.* **2018**, *221*, 17–26. [\[CrossRef\]](#)
41. Cheng, W.L.; Chen, Y.S.; Zhang, J.; Lyons, T.J.; Pai, J.L.; Chang, S.H. Comparison of the revised air quality index with the PSI and AQI indices. *Sci. Total Environ.* **2007**, *382*, 191–198. [\[CrossRef\]](#)

42. She, Q.N.; Peng, X.; Xu, Q.; Long, L.B.; Wei, N.; Liu, M.; Jia, W.X. Air quality and its response to satellite-derived urban form in the Yangtze River Delta, China. *Intl. J. Environ. Res. Public Health*. **2017**, *75*, 297–306. [\[CrossRef\]](#)
43. Xu, K.; Cui, K.; Young, L.H.; Hsieh, Y.K.; Wang, Y.F.; Zhang, J.; Wan, S. Impact of the COVID-19 Event on Air Quality in Central China. *Aerosol Air Qual. Res.* **2020**, *20*, 915–929. [\[CrossRef\]](#)
44. Xu, K.; Cui, K.; Young, L.H.; Wang, Y.F.; Hsieh, Y.K.; Wan, S.; Zhang, J. Air Quality Index, Indicator Air Pollutants and Impact of COVID-19 Event on the Air Quality near Central China. *Aerosol Air Qual. Res.* **2020**, *20*, 1204–1221. [\[CrossRef\]](#)
45. Apak, S.; Atay, E. Industrial Policy and Climate Change Management of Turkey as an EU candidate country. *Procedia-Soc. Behav. Sci.* **2013**, *75*, 246–254. [\[CrossRef\]](#)
46. Das, M.; Das, A.; Ghosh, S.; Sarkar, R.; Saha, S. Spatio-temporal concentration of atmospheric particulate matter (PM_{2.5}) during pandemic: A study on most polluted cities of indo-gangetic plain. *Urban. Climate* **2021**, *35*, 100758. [\[CrossRef\]](#)
47. Zhang, H.; Tripathi, N.K. Geospatial hot spot analysis of lung cancer patients correlated to fine particulate matter (PM_{2.5}) and industrial wind in Eastern Thailand. *J. Clean. Prod.* **2018**, *170*, 407–424. [\[CrossRef\]](#)
48. Zaib, S.; Lu, J.; Shahid, M.Z.; Ahmar, S.; Shahid, I. Impact of SARS-CoV-2 on Ambient Air Quality in Northwest China (NWC). *Atmosphere* **2021**, *12*, 518. [\[CrossRef\]](#)
49. RStudio Team. RStudio: Integrated Development for R. R Studio, PBC, Boston, MA 2020. Available online: <http://www.rstudio.com> (accessed on 28 January 2022).
50. Chen, J.; Lu, J.; Ning, J.; Yan, Y.; Li, S.; Zhou, L. Pollution characteristics, sources, and risk assessment of heavy metals and perfluorinated compounds in PM_{2.5} in the major industrial city of northern Xinjiang, China. *Air Qual. Atmos. Health*. **2019**, *12*, 909–918. [\[CrossRef\]](#)
51. Liu, T.; Zhuang, G.; Huang, K.; Lin, J.; Wang, Q.; Deng, C.; Fu, J.S. A Typical Formation Mechanism of Heavy Haze-Fog Induced by Coal Combustion in an Inland City in North-Western China. *Aerosol Air Qual. Res.* **2017**, *17*, 98–107. [\[CrossRef\]](#)
52. Du, W.; Zhang, Y.; Chen, Y.; Xu, L.; Chen, J.; Deng, J.; Xiao, H. Chemical characterization and source apportionment of PM_{2.5} during spring and winter in the Yangtze River Delta, China. *Aerosol Air Qual. Res.* **2017**, *17*, 2165–2180. [\[CrossRef\]](#)
53. Zhao, Z.; Lv, S.; Zhang, Y.; Zhao, Q.; Shen, L.; Xu, S.; Jin, C. Characteristics and source apportionment of PM_{2.5} in Jiaying, China. *Environ. Sci. Pollut. Res.* **2019**, *26*, 7497–7511. [\[CrossRef\]](#)
54. Zhao, D.; Chen, H.; Yu, E.; Luo, T. PM_{2.5}/PM₁₀ Ratios in Eight Economic Regions and Their Relationship with Meteorology in China. *Adv. Meteorol.* **2019**. [\[CrossRef\]](#)
55. Shen, F.; Zhang, L.; Jiang, L.; Tang, M.; Gai, X.; Chen, M.; Ge, X. Temporal variations of six ambient criteria air pollutants from 2015 to 2018, their spatial distributions, health risks and relationships with socioeconomic factors during 2018 in China. *Environ. Int.* **2020**, *137*, 105556. [\[CrossRef\]](#) [\[PubMed\]](#)
56. Da, Y.L.M.A.B.; Limu, D.L.N.T.L.; Miti, A.B.L.Y.; Wang, X.; Ding, X. Autumn and wintertime polycyclic aromatic hydrocarbons in PM_{2.5} and PM_{2.5–10} from Urumqi, China. *Aerosol Air Qual. Res.* **2013**, *13*, 407–414. [\[CrossRef\]](#)
57. Ding, Y.; Lu, J.; Liu, Z.; Li, W.; Chen, J. Volatile organic compounds in Shihezi, China, during the heating season: Pollution characteristics, source apportionment, and health risk assessment. *Environ. Sci. Pollut. Res.* **2020**, *27*, 16439–16450. [\[CrossRef\]](#) [\[PubMed\]](#)
58. Xu, G.; Jiao, L.; Zhang, B.; Zhao, S.; Yuan, M.; Gu, Y.; Tang, X. Spatial and temporal variability of the PM_{2.5}/PM₁₀ ratio in Wuhan, Central China. *Aerosol Air Qual. Res.* **2016**, *17*, 741–751. [\[CrossRef\]](#)
59. Zhang, Q.; Streets, D.G.; Carmichael, G.R.; He, K.B.; Huo, H.; Kannari, A.; Chen, D. Asian emissions in 2006 for the NASA INTEX-B mission. *Atmos. Chem. Phys.* **2009**, *9*, 5131–5153. [\[CrossRef\]](#)
60. Zhao, Y.; Nielsen, C.P.; Lei, Y.; McElroy, M.B.; Hao, J. Quantifying the uncertainties of a bottom-up emission inventory of anthropogenic atmospheric pollutants in China. *Atmos. Chem. Phys.* **2011**, *10*, 29075–29111. [\[CrossRef\]](#)
61. Pey, J.; Querol, X.; Alastuey, A.; Forastiere, F.; Stafoggia, M. African dust outbreaks over the Mediterranean Basin during 2001–2011: PM₁₀ concentrations, phenomenology and trends, and its relation with synoptic and mesoscale meteorology. *Atmos. Chem. Phys.* **2013**, *13*, 1395–1410. [\[CrossRef\]](#)
62. Wang, S.; Zhang, M.; Minguillón, M.C.; Zhang, X.; Feng, F.; Qiu, X. PM₁₀ concentration in urban atmosphere around the eastern Tien Shan, Central Asia during 2007–2013. *Environ. Sci. Pollut. Res.* **2015**, *22*, 6864–6876. [\[CrossRef\]](#)
63. Wang, Y.; Zhang, J.; Bai, Z.; Yang, W.; Zhang, H.; Mao, J.; Chen, L. Background concentrations of PMs in Xinjiang, West China: An estimation based on meteorological filter method and Eckhardt algorithm. *Atmos. Res.* **2019**, *215*, 141–148. [\[CrossRef\]](#)
64. Wang, W.; Samat, A.; Abuduwaili, J.; Ge, Y. Spatio-Temporal Variations of Satellite-Based PM_{2.5} Concentrations and Its Determinants in Xinjiang, Northwest of China. *Int. J. Environ. Res. Public Health*. **2020**, *17*, 2157. [\[CrossRef\]](#)
65. Ashraf, A.; Butt, A.; Khalid, I.; Alam, R.U.; Ahmad, S.R. Smog analysis and its effect on reported ocular surface diseases: A case study of 2016 smog event of Lahore. *Atmos. Environ.* **2019**, *198*, 257–264. [\[CrossRef\]](#)
66. Bourdrel, T.; Bind, M.A.; Béjot, Y.; Morel, O.; Argacha, J.F. Cardiovascular effects of air pollution. *Arch. Cardiovasc. Dis.* **2017**, *110*, 634–642. [\[CrossRef\]](#) [\[PubMed\]](#)
67. Guo, H.; Wang, T.; Simpson, I.J.; Blake, D.R.; Yu, X.M.; Kwok, Y.H.; Li, Y.S. Source contributions to ambient VOCs and CO at a rural site in eastern China. *Atmos. Environ.* **2004**, *38*, 4551–4560. [\[CrossRef\]](#)
68. Kato, S.; Shiobara, Y.; Uchiyama, K.; Miura, K.; Okochi, H.; Kobayashi, H.; Hatakeyama, S. Atmospheric CO, O₃, and SO₂ measurements at the summit of Mt. Fuji during the summer of 2013. *Aerosol Air Qual. Res.* **2016**, *16*, 2368–2377. [\[CrossRef\]](#)

69. Wang, Y.; Jia, C.; Tao, J.; Zhang, L.; Liang, X.; Ma, J.; Zhang, K. Chemical characterization and source apportionment of PM_{2.5} in a semi-arid and petrochemical-industrialized city, Northwest China. *Sci. Total Environ.* **2016**, *573*, 1031–1040. [\[CrossRef\]](#)
70. Wang, M.; Zheng, S.; Nie, Y.; Weng, J.; Cheng, N.; Hu, X.; Bai, Y. Association between short-term exposure to air pollution and dyslipidemias among type 2 diabetic patients in northwest China: A population-based study. *Intl. J. Environ. Res. Public Health.* **2018**, *15*, 631. [\[CrossRef\]](#)
71. Ma, X.; Xiao, Z.; He, L.; Shi, Z.; Cao, Y.; Tian, Z.; Liu, J. Chemical Composition and Source Apportionment of PM_{2.5} in Urban Areas of Xiangtan, Central South China. *Int. J. Environ. Res. Public Health* **2019**, *16*, 539. [\[CrossRef\]](#)
72. Tao, J.; Gao, J.; Zhang, L.; Zhang, R.; Che, H.; Zhang, Z.; Hsu, S.C. PM_{2.5} pollution in a megacity of southwest China: Source apportionment and implication. *Atmos. Chem. Phys.* **2014**, *14*, 8679–8699. [\[CrossRef\]](#)
73. Alam, K.; Mukhtar, A.; Shahid, I.; Blaschke, T.; Majid, H.; Rahman, S.; Rahman, N. Source Apportionment and Characterization of Particulate Matter (PM₁₀) in Urban Environment of Lahore. *Aerosol Air Qual. Res.* **2014**, *14*, 1851–1861. [\[CrossRef\]](#)
74. Chatterjee, A.; Sarkar, C.; Adak, A.; Mukherjee, U.; Ghosh, S.K.; Raha, S. Ambient air quality during Diwali festival over Kolkata—a mega-city in India. *Aerosol Air Qual. Res.* **2013**, *13*, 1133–1144. [\[CrossRef\]](#)
75. Bilal, M.; Mhawish, A.; Nichol, J.E.; Qiu, Z.; Nazeer, M.; Ali, M.A.; Ke, S. Air pollution scenario over Pakistan: Characterization and ranking of extremely polluted cities using long-term concentrations of aerosols and trace gases. *Remote Sens. Environ.* **2021**, *264*, 112617. [\[CrossRef\]](#)
76. Huang, X.; Zhang, J.; Luo, B.; Wang, L.; Tang, G.; Liu, Z.; Wang, Y. Water-soluble ions in PM_{2.5} during spring haze and dust periods in Chengdu, China: Variations, nitrate formation and potential source areas. *Environ. Pollut.* **2018**, *243*, 1740–1749. [\[CrossRef\]](#) [\[PubMed\]](#)
77. Li, H.; Duan, F.; Ma, Y.; He, K.; Zhu, L.; Ma, T.; Kimoto, T. Case study of spring haze in Beijing: Characteristics, formation processes, secondary transition, and regional transportation. *Environ. Pollut.* **2018**, *242*, 544–554. [\[CrossRef\]](#) [\[PubMed\]](#)
78. Chai, F.; Gao, J.; Chen, Z.; Wang, S.; Zhang, Y.; Zhang, J.; Ren, C. Spatial and temporal variation of particulate matter and gaseous pollutants in 26 cities in China. *J. Environ. Sci.* **2014**, *26*, 75–82. [\[CrossRef\]](#)
79. Zhang, H.; Ying, Q. Contributions of local and regional sources of NO_x to ozone concentrations in Southeast Texas. *Atmos. Environ.* **2011**, *45*, 2877–2887. [\[CrossRef\]](#)
80. Chameides, W.L.; Fehsenfeld, F.; Rodgers, M.O.; Cardelino, C.; Martinez, J.; Parrish, D.; Wang, T. Ozone precursor relationships in the ambient atmosphere. *J. Geophys. Res. Atmos.* **1992**, *97*, 6037–6055. [\[CrossRef\]](#)
81. Zhao, R.; Cui, K.P.; Wang, W.W.; Wang, L.C.; Yan, P. Atmospheric PM_{2.5} and total PCDD/FsWHO2005-TEQ level: A case of Handan and Kaifeng cities, China. *Aerosol Air Qual. Res.* **2018**, *18*, 994–1007. [\[CrossRef\]](#)
82. Biswas, M.S.; Ghude, S.D.; Gurnale, D.; Prabhakaran, T.; Mahajan, A.S. Simultaneous Observations of Nitrogen Dioxide, Formaldehyde and Ozone in the Indo- Gangetic Plain. *Aerosol Air Qual. Res.* **2019**, *19*, 1749–1764. [\[CrossRef\]](#)
83. Rasheed, A.; Aneja, V.P.; Aiyer, A.; Rafique, U. Measurements and analysis of air quality in Islamabad, Pakistan. *Earth's Future* **2014**, *2*, 303–314. [\[CrossRef\]](#)
84. Huang, W.; Long, E.; Wang, J.; Huang, R.; Ma, L. Characterizing spatial distribution and temporal variation of PM₁₀ and PM_{2.5} mass concentrations in an urban area of southwest China. *Atmos. Pollut. Res.* **2015**, *6*, 842–848. [\[CrossRef\]](#)
85. Shen, F.Z.; Ge, X.L.; Hu, J.L.; Nie, D.Y.; Tian, L.; Chen, M.D. Air pollution characteristics and health risks in Henan Province, China. *Environ. Res.* **2017**, *156*, 625–634. [\[CrossRef\]](#) [\[PubMed\]](#)
86. Filonchyk, M.; Peterson, M. Air quality changes in Shanghai, China, and the surrounding urban agglomeration during the COVID-19 lockdown. *J. Geovis. Spat. Anal.* **2020**, *4*, 1–7. [\[CrossRef\]](#)
87. Ghauri, B.; Lodhi, A.; Mansha, M. Development of baseline (air quality) data in Pakistan. *Environ. Monit. Assess.* **2007**, *127*, 237–252. [\[CrossRef\]](#)
88. Sharma, A.; Sharma, S.K.; Mandal, T.K. Influence of ozone precursors and particulate matter on the variation of surface ozone at an urban site of Delhi, India. *Sustain. Environ. Res.* **2016**, *26*, 76–83. [\[CrossRef\]](#)
89. Sugimoto, N.; Shimizu, A.; Matsui, I.; Nishikawa, M. A method for estimating the fraction of mineral dust in particulate matter using PM_{2.5}-to- PM₁₀ ratios. *Particuology* **2016**, *28*, 114–120. [\[CrossRef\]](#)