

Article

Impacts of Fertilization Type on Soil Microbial Biomass and Nutrient Availability in Two Agroecological Zones of Ghana

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Abstract: The decline in soil productivity amidst efforts to increase crop yield in Sub Saharan Africa (SSA) has made it imperative to assess the current fertilization management approaches. This study was conducted in two agroecological zones (i.e., Guinea Savannah (GS) and Deciduous forest (DF)) of Ghana to evaluate how different fertilization schemes in the long term (>5 years) impacted the soil biochemical properties. Soil samples under four fertilization schemes (inorganic fertilizer only, low-to-medium organic residues only, inorganic fertilizers plus low-to-medium organic residues, and no fertilization) from 20 farmers' field were sampled from March to April 2015. Soil biochemical quality indicators were determined using standard procedures. Overall, the average chemical and microbial biomass contents for most indicators were significantly higher in DF compared to GS. Relative to the reference sites, soil quality improvement were observed under inorganic fertilization in both agroecologies in contrast to significant soil deterioration (26.5%) under sole organic residue application in GS. Furthermore, the results showed that increased inorganic fertilization rate alone or combination with organic residues improved soil quality relative to the reference. The present results suggest the need to raise the current fertilizer application rates, especially in GS in order to enhance optimum soil productivity.

Keywords: deciduous forest; deterioration index; guinea savannah; inorganic; microbial biomass; organic; tropical agroecological zone

1. Introduction

Estimates show that Africa holds 52% of the world's remaining arable land [1] that can be utilized for agricultural production. In recent years, high annual food production rate has been recorded in Sub Saharan Africa (SSA) relative to the other regions [2]. This validates the prospects of agriculture in boosting the economic growth in many African countries. Even though the trend looks promising, agricultural productivity in SSA is faced with many difficulties. For instance, although new crop productivity improvement techniques have been introduced in many parts of the world [3], smallholder farmers predominantly in SSA are largely unable to benefit.

Ghana's economy is hugely dependent on agriculture and an estimated 44% of the population is involved with a significant rural-dwelling proportion [4]. Their farming practices are highly diverse [5] due to their risk averse nature, economic situation and conservative lifestyle. This reflects variability in the crop yield as well as the soil biochemical properties, even though little on farm research exists in SSA focusing on the interaction of soil characteristics and agricultural management practices. Farming in Guinea savannah (GS) and deciduous forest (DF) zones are typically rain-fed and is characterized by the use of traditional management practices [4]. Both zones are major contributors of agricultural products for domestic and international markets. The main crops grown in GS include maize (*Zea mays*), yam (*Dioscorea spp.*), sorghum (*Sorghum bicolor*), millet (*Pennisetum americanum*), rice (*Oryza sativa*) and shea (*Vitellaria paradoxa*). In the DF, maize (*Zea mays*), cassava (*Manihot esculenta*), plantain (*Musa paradisiaca*) and cash crops like cocoa (*Theobroma cacao*) and oil palm (*Elaeis guineensis*) are commonly cultivated. Soil fertility approach is one notable transitioning practice that is differently practiced among farmers in both zones [6,7] and ranges from inorganic to organic amendments or combination of both, fallow periods or even no fertilization. Such differences are reflected in both the type and rates of resource inputs. Consequently, the choice of fertilization practice in the long term impacts the soil physical, chemical and microbial composition differently [8,9].

The efficient use of fertilizer is reflected in nutrient availability to crops [10], as influenced by the synchrony of crop growth with dose and time of application [11], holding other factors constant. For example, soil nitrogen (N) together with total organic carbon (C), extractable phosphorus (P) and bulk density have the ability to reveal changes in microbial community dynamics, mineralization, nutrient availability, humification, and soil porosity [12]. On the other hand, soil microbes have been recognized to influence crop productivity by controlling N dynamics and its synchrony with crop demand [5], hence their presence is important in soil productivity studies especially in SSA soils where high nutrient deficiencies have been reported [13].

This study is one of the few reports that characterize soils under different fertilization and management practices in Ghana. It is important for the development of adaptable standard farming practices aimed at minimizing the negative effects of common crop production practices. The main objective of this study was to examine how long-term fertilization regimes and management practices impact soil chemical and microbial properties in two agroecological zones. The specific objective was to determine the relationship between the current rate of fertilization and soil microbial biomass indicators.

2. Results

2.1. Basic Soil Chemical Properties

Across the 20 farms, soil C and N related indicators, as well as available P, were significantly higher in the DF soils compared to the GS (Table 1). Average soil carbon-to-nitrogen (CN) ratio of DF sites was 6.07 while that of GS was 4.76. At DF, inorganic fertilizer treatment and the reference and respectively had the highest total nitrogen (TN) and total carbon (TC) contents, although no significant differences were observed. Among the GS sites, inorganic fertilizer treatment showed the highest significant soil TC (6.84 g kg^{-1}) and CN ratio contents than the other fertilization types, in contrast to soil TN where no significant differences were observed.

On the other hand, the nitrates and ammonium (NO_3^- -N, NH_4^+ -N) and available P contents of the inorganic fertilizer treatment at DF were significantly higher relative to the other fertilization types. Similarly, at GS, the highest NO_3^- -N ($107.89 \text{ mg kg}^{-1}$) and NH_4^+ -N (65.02 mg kg^{-1}) contents were observed in the inorganic fertilizer treatments.

Fertilization practice significantly impacted the soil organic matter (SOM) contents (Table 1). At DF, although not significantly different, higher SOM content was observed in the inorganic fertilizer treatments compared to the reference and combined application (organic residue plus chemical fertilizer) sites. Among the GS sites, the highest significant SOM content was observed in the reference.

There were no clear differences in the soil pH among the studied sites in both agroecological zones. The GS sites, although with high variation showed higher soil cation exchange capacity (CEC) values compared to the DF. The highest significant values at GS was observed in the inorganic fertilizer and no fertilizer input treatments while no statistical differences were observed among the DF treatments.

2.2. Microbial Biomass Content-Differences Among Sites

Soils under the DF zone showed higher soil microbial biomass C and N (MBC and MBN) contents but lower MBC-to-MBN ratio compared to GS sites regardless of fertilization type (Figure 1). The average MBC ($904.21 \text{ mg kg}^{-1}$) and MBN (30.32 mg kg^{-1}) contents in DF sites were almost twice higher than that of GS. At GS, the no fertilization input treatment and the reference showed the highest significant MBN contents, compared to the other fertilization types. Additionally, similar observations were made for MBC content as in the case of the sole low-to-medium organic residue amendment.

Similar to the microbial biomass contents, active C contents in DF was significantly higher than that in GS (Figure 2a). At GS, the no fertilizer input ($278.63 \text{ mg kg}^{-1}$) and only inorganic fertilizer treatment ($277.71 \text{ mg kg}^{-1}$) had significantly higher active C contents while the lowest was found in the sole organic residue amendment ($165.73 \text{ mg kg}^{-1}$). Among the DF sites, however, there were no statistical differences in soil active C contents. In contrast to soil MBC, MBN and active C contents and irrespective of fertilization practices, significantly higher average extractable C (Extr C) and potential mineralizable nitrogen (PMN) values were recorded for GS sites compared to the DF (Figures 2b and 3). The highest Extr C content at GS was found in the reference ($120.75 \text{ mg kg}^{-1}$) followed by the co-application of organic residues with inorganic fertilizer treatment ($114.13 \text{ mg kg}^{-1}$) while the lowest was in the sole organic residue amendment (59.72 mg kg^{-1}). In contrast, no significant differences in Extr C were observed among the DF sites. The highest PMN content among the GS sites was observed in no fertilization input treatment followed by the sole organic residue amendment with only inorganic fertilizer treatment being the lowest.

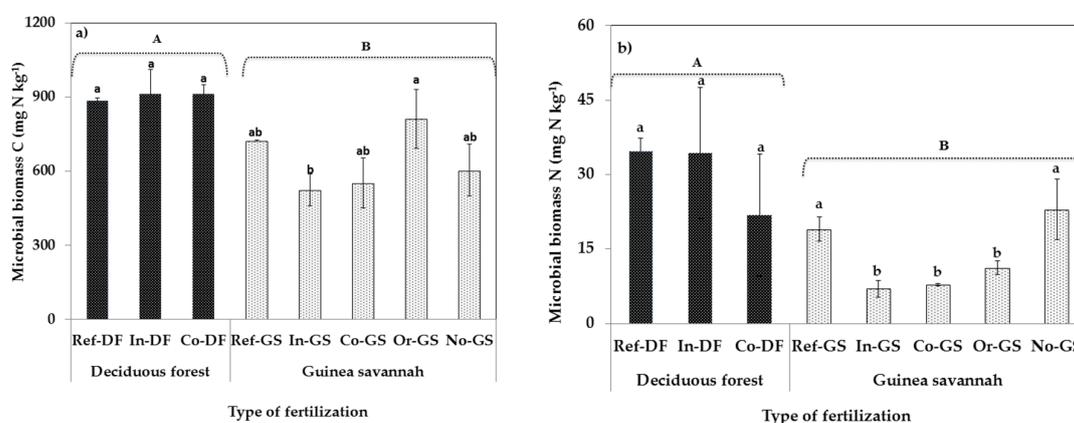


Figure 1. Soil microbial biomass C (a) and microbial biomass N (b) at depth (0–15 cm) of different types of fertilization. Treatment code: Ref-DF, reference DF site; In-DF, inorganic fertilization at DF; Co-DF, inorganic fertilizer, with low-to-medium organic residues at DF; Ref-GS, reference site; In-GS, inorganic fertilization at GS; Co-GS, inorganic fertilizer, with low-to-medium organic residues at GS; Or-GS, sole low-to-medium organic residues at GS and No-GS, no fertilization input at GS. The error bar indicates the standard deviation of three replicates. Small letters (a, b) represent mean differences among fertilization types in each ecological zone and capital letters (A, B) represent mean the difference between the two ecological zones.

Table 1. Biochemical properties of the studied soils.

Treatment Code	TC (g kg ⁻¹)	TN (g kg ⁻¹)	CN Ratio	NO ₃ (mg kg ⁻¹)	NH ₄ (mg kg ⁻¹)	Phosphorus (mg kg ⁻¹)	SOM (%)	CEC	pH
Ref-DF	16.85 ± 0.14a	2.38 ± 0.08a	7.09 ± 0.30a	176.17 ± 1.17b	76.0 ± 0.48ab	5.82 ± 0.00b	2.63 ± 0.00a	9.60 ± 0.01a	6.12
In-DF	15.97 ± 4.95a	2.64 ± 0.62a	6.00 ± 0.61a	362.32 ± 42.25a	78.2 ± 0.58a	19.77 ± 4.26a	3.22 ± 0.58a	10.66 ± 1.49a	5.94
Co-DF	10.04 ± 4.42a	1.91 ± 0.22a	5.14 ± 1.65a	307.88 ± 21.39a	68.3 ± 1.18abc	16.53 ± 8.24ab	2.13 ± 0.81a	11.07 ± 3.09a	5.85
Mean-DF	14.29 ± 4.62A	2.31 ± 0.46A	6.07 ± 1.23A	282.12 ± 86.2A	77.62 ± 8.19A	14.04 ± 7.84A	2.66 ± 0.69A	10.44 ± 0.76B	5.97
Ref-GS	6.05 ± 0.43ab	1.30 ± 0.04a	4.68 ± 0.47b	38.60 ± 0.06b	49.4 ± 0.74bc	16.89 ± 0.01a	2.48 ± 0.01a	16.82 ± 4.46ab	ND
In-GS	6.84 ± 0.62a	1.19 ± 0.08a	5.73 ± 0.16a	107.89 ± 5.36a	65.0 ± 2.14a	14.55 ± 3.12b	1.15 ± 0.13c	19.72 ± 5.88a	5.80
Co-GS	5.71 ± 1.12ab	1.26 ± 0.170a	4.51 ± 0.33b	24.44 ± 0.45b	44.0 ± 0.48bc	11.05 ± 1.36bc	0.89 ± 0.12cd	ND	5.90
Or-GS	4.32 ± 0.68b	1.05 ± 0.09a	4.11 ± 0.51b	39.19 ± 1.27b	31.4 ± 1.16c	13.68 ± 1.19bc	0.46 ± 0.08d	10.53 ± 1.86b	6.40
No-GS	6.24 ± 1.88ab	1.28 ± 0.22a	4.79 ± 0.63b	33.44 ± 0.73b	43.9 ± 0.81bc	17.25 ± 8.44a	1.82 ± 0.76b	19.20 ± 3.26a	5.45
Mean-GS	5.83 ± 1.26B	1.22 ± 0.15B	4.76 ± 0.67B	62.45 ± 44.38B	46.72 ± 15.21B	13.35 ± 2.21A	1.36 ± 0.79B	16.57 ± 4.21A	5.89

Values represent the mean ± standard deviation of three replicates per fertilization type. Bold values indicate the average of the respective fertilization type in each ecological zone. Treatment code: Ref-DF, reference DF site; In-DF, inorganic fertilization at DF; Co-DF, inorganic fertilizer, with low-to-medium organic residues at DF; Ref-GS, reference site; In-GS, inorganic fertilization at GS; Co-GS, inorganic fertilizer, with low-to-medium organic residues at GS; Or-GS, sole low-to-medium organic residues at GS and No-GS, no fertilization input at GS. SOM, Soil organic matter; TN, total nitrogen; TC, total nitrogen; P, phosphorus; CN, carbon to nitrogen ratio; NH₄, Ammonium; NO₃, Nitrate. In each column, small letters (a, b) represent mean differences among fertilization types in each respective ecological zone and capital letters (A, B) represent mean the difference between the two ecological zones at $p < 0.05$ using Duncan's Multiple Range Test.

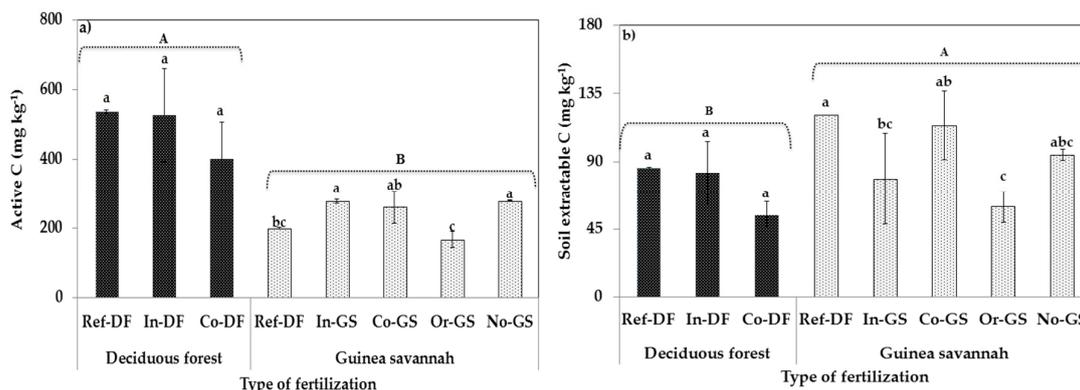


Figure 2. Active Carbon (a), soil extractable C (b) at depth (0–15 cm) of different types of fertilization. Treatment code: Ref-DF, reference DF site; In-DF, inorganic fertilization at DF; Co-DF, inorganic fertilizer, with low-to-medium organic residues at DF; Ref-GS, reference site; In-GS, inorganic fertilization at GS; Co-GS, inorganic fertilizer, with low-to-medium organic residues at GS; Or-GS, sole low-to-medium organic residues at GS and No-GS, no fertilization input at GS. The error bar indicates the standard deviation of three replicates. Small letters (a, b) represent mean differences among fertilization types in each ecological zone and capital letters (A, B) represent mean the difference between the two ecological zones.

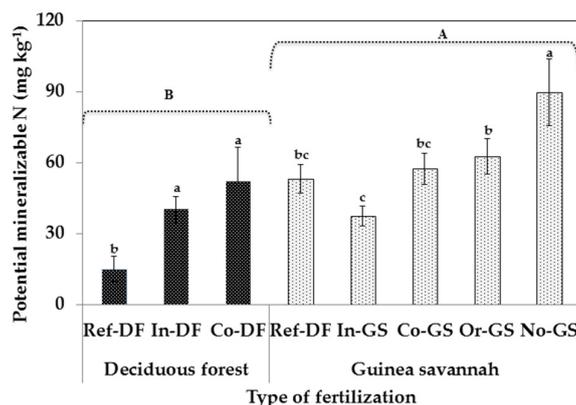


Figure 3. Potential mineralizable nitrogen at depth (0–15 cm) of different types of fertilization. Treatment code: Ref-DF, reference DF site; In-DF, inorganic fertilization at DF; Co-DF, inorganic fertilizer, with low-to-medium organic residues at DF; Ref-GS, reference site; In-GS, inorganic fertilization at GS; Co-GS, inorganic fertilizer, with low-to-medium organic residues at GS; Or-GS, sole low-to-medium organic residues at GS and No-GS, no fertilization input at GS. The error bar indicates the standard deviation of three replicates. Small letters (a, b) represent mean differences among fertilization types in each ecological zone and capital letters (A, B) represent mean the difference between the two ecological zones.

The average deterioration index (DI) ranged from −31.0% to 26.5% with DF sites showing lower deterioration relative to the GS (Figure 4a). The lowest DI was observed in the inorganic fertilization treatment while the highest was in the organic residue amendment. Both fertilization schemes in the DF sites showed negative DI values. However, at GS, only the inorganic fertilization and no fertilization treatments showed negative deterioration. The carbon management index (CMI) at DF was higher in the inorganic fertilizer treatment, though not significantly different from the combined application amendment (Figure 4b). Similarly, the highest CMI at the GS was observed in the inorganic fertilizer treatment while the lowest was in the sole organic residue amendment.

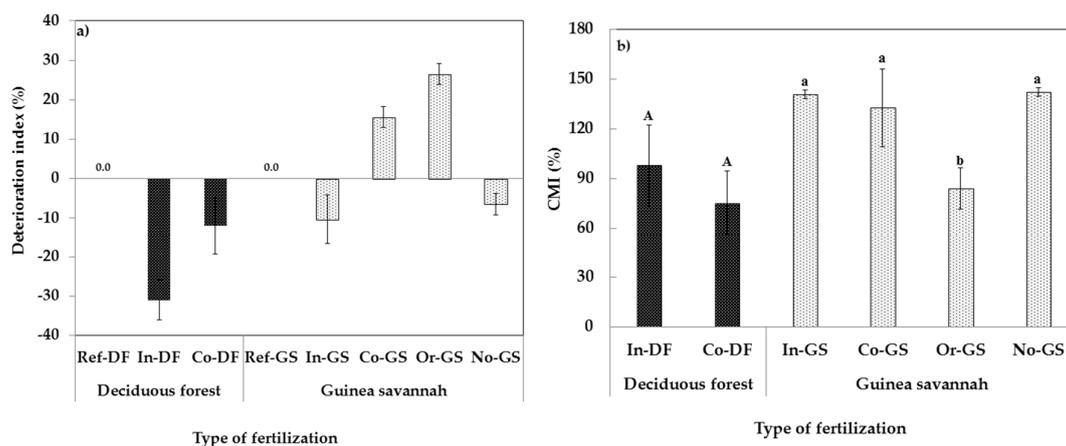


Figure 4. Deterioration index (DI) (a) and Carbon management index (CMI) (b) for different types of fertilization in the DF and GS agro-ecological zones. Treatment code: Ref-DF, reference DF site; In-DF, inorganic fertilization at DF; Co-DF, inorganic fertilizer, with low-to-medium organic residues at DF; Ref-GS, reference site; In-GS, inorganic fertilization at GS; Co-GS, inorganic fertilizer, with low-to-medium organic residues at GS; Or-GS, sole low-to-medium organic residues at GS and No-GS, no fertilization input at GS. Each deterioration index is an average of percentage changes of TN, TC, NO_3^- -N, NH_4^+ -N, Extr C, active C, PMN, MBN, MBC, and SOM relative to their respective reference sites. The error bar indicates the standard deviation of the soil parameters.

3. Discussion

3.1. Fertilization Type and Climate Effects on Soil Chemical Properties

The basic soil fertility status for most of the studied sites (Table 1) was relatively higher than values reported by Issaka et al. [14]. However, these concentrations are generally below the threshold for attaining the optimum crop yield [15]. Inorganic N and extractable P measure the two most important and limiting soil nutrients in typical productive land management systems [5,12]. Similarly, CN ratio is indicative of the capacity of the soil to store and recycle nutrients [16]. As such, the reported low soil nutrient contents threaten crop and soil productivity. Although this study has no information on yield output, a report by Benin et al. [17] revealed enhanced crop yield in the country during seasons of increased inorganic fertilizer application due to subsidy placement.

Soils in both zones are similar in characteristics but exist in different climates [18]. In the present study, however, soils in the DF showed high nutrient content compared to GS. Similar to this finding, Issaka et al. [14] reported high soil TC, TN, available P and SOM in DF relative to GS zone. The relative high soil nutrient content in DF sites is associated with the high fertilizer application rates as shown by a positive correlation between the rate of fertilizer application and soil chemical content (Table 2). In contrast, the low fertilizer application rate in a continuous cropping manner under climate-driven high SOM decomposition in GS [19], may have contributed to the observed low soil C, N, and P composition.

Only inorganic fertilizer application followed by co-application of organic residues with inorganic fertilizers proved superior among the DF sites. Similarly, at the GS, sole chemical fertilization resulted in relatively higher soil nutrients status, in contrast to the lower values observed for the sole organic residues (Table 1). The enhanced productivity of inorganic fertilizers in both zones relates to its readily available forms, thereby ensuring quick absorption by plants. Additionally, the rate of input application correlated with a number of soil properties (Table 2). Hence, increased inorganic fertilization in both agroecological zones is promising although concerns of its high cost and price instability [20] coupled with the low economic status of Ghanaian farmers may limit its optimal use. Moreover, problems regarding the potential topsoil acidification due to continuous application of inorganic fertilizers in soils with poor buffering capacity have been reported [21], and the subsequent long-term soil fertility enhancement is still not clear.

Table 2. Linear correlation among the analyzed biological and chemical soil parameters ($n = 45$).

	Rate	TN	TC	CN	NO ₃	Extr C	MBC	MBN	ACC	P	PMN	SOM	AM
Rate													
TN	0.57 **												
TC	0.60 **	0.96 **											
CN	0.54 **	0.52 **	0.73 **										
NO ₃	0.73 **	0.82 **	0.76 **	0.43 **									
Extr C	−0.38 *	−0.14 NS	−0.09 NS	−0.03 NS	−0.54 **								
MBC	0.43 **	0.83 **	0.72 **	0.24 NS	0.84 **	−0.42 **							
MBN	0.17 NS	0.65 **	0.66 **	0.45 **	0.60 **	−0.16 NS	0.62 **						
ACC	0.57 **	0.95 **	0.96 **	0.62 **	0.80 **	−0.16 NS	0.73 **	0.66 **					
P	0.21 NS	0.44 **	0.40 *	0.15 NS	0.48 **	−0.19 NS	0.47 **	0.38 *	0.50 **				
PMN	−0.83 **	−0.30 *	−0.35 *	−0.42 **	−0.50 **	0.27 NS	0.14 NS	0.03 NS	−0.34 *	−0.11 NS			
SOM	0.15 NS	0.52 **	0.60 **	0.50 **	0.24 NS	−0.12 NS	0.38 *	0.24 NS	0.55 **	0.26 NS	−0.02 NS		
AM	0.71 **	0.60 **	0.56 **	0.38 *	0.79 **	−0.41 **	0.61 **	0.35 *	0.62 **	0.45 **	0.63 **	0.13 NS	
CEC	−0.55 **	−0.56 **	−0.50 **	−0.19 NS	−0.62 **	0.25 NS	−0.76	−0.32 *	−0.46 *	0.01	0.32 *	−0.16 NS	−0.28 NS

SOM, Soil organic matter; TN, total nitrogen; TC, total nitrogen; SOM, Organic matter; TN, total nitrogen; CN, carbon to nitrogen ratio; AM, Ammonium; NO₃, nitrate; PMN, P, phosphorus; Potential mineralizable nitrogen; MBC, Microbial biomass carbon; MBN, Microbial biomass nitrogen; Extr C, Soil extractable carbon; CEC, cation exchange capacity. Significance levels: NS: $p > 0.05$; * $p < 0.05$; ** $p < 0.01$.

In the present study, non-fertilizer input treatment at GS as regulated by short fallow periods enhanced the soil chemical composition. However, accelerated human population growth, which consequently reduces the length of the fallow period makes non-fertilization based agronomic practice a less sustainable option.

It must be mentioned that the organic inputs used by farmers either in sole organic residue systems or in combined application with inorganic fertilizers (Table S1) are materials of low to medium quality [21]. The nutrient composition of such organic materials is usually low, thus their slow N inputs to the soil are inadequate. Additionally, the nutrient release pattern of such organic residues is not commonly matched to the nutrient demands of many cultivated crops [22]. Besides, all year availability of these organic residues in SSA has been questioned due to its competitive use as roofing material, fuel, and fodder for livestock [23].

The results of the present study show that combined input application of organic and inorganic resources in both agroecological zones gave better responses in basic soil fertility status compared to sole organic residue application. This reiterates the need to encourage co-application of inorganic and organic fertilization [5]. Hence, inherent quality characteristics and decomposition trends of locally available organic materials must be investigated to allow for their efficient incorporation into crop production systems [24].

3.2. Fertilization Type on Soil Microbial Biomass Indicators

Across the wide array of different fertilization systems, microbial composition differed (Figures 1–3). The high rate of fertilizer application in DF relative to GS sites partly as a result of farmers' low economic status [25], enhanced the growth of vegetation in DF and subsequently resulted in an increased SOM contents. The soluble C compounds from SOM consequently induced microbial biomass community [26]. This finding was supported by significant relationships between the rate of fertilizer application and soil microbial biomass indicators (Table 3). Moreover, soil microbial biomass was largely influenced by soil nutrients as shown by the significant positive association with the resultant soil TN. In support of this result, Lovell et al. [27] reported that long-term addition of N at 200 kg N ha⁻¹ to previously unfertilized soils increased MBC by 60% and MBN by 63%. The observed low PMN at DF sites is related to the high SOM composition which triggered soil microbial population during incubation, resulting in inorganic N immobilization. This assumption was justified by the observed low PMN content in most DF sites coupled with the negative correlation with other microbial parameters (Table 2).

Table 3. Summary of forward step-wise regression analysis between rates of fertilizer application, soil TN and soil microbial properties ($n = 45$).

Variable	Microbial Properties							
	MBN	MBC	BIO CN	ACC	PMN	EXTR C	MBC/TC	MBN/TN
Rate								
R ²	0.03	0.18	0.02	0.33	0.69	0.14	0.23	0.10
Sig.	0.3 ^{NS}	0.004 [*]	0.4 ^{NS}	0.001 ^{**}	0.001 ^{**}	0.01 [*]	0.001 ^{**}	0.04 [*]
Soil TN								
R ²	0.42	0.69	0.10	0.91	0.09	0.02	0.25	0.04
Sig.	0.001 ^{**}	0.001 ^{**}	0.03 [*]	0.001 ^{**}	0.04 [*]	0.4 ^{NS}	0.001 ^{**}	0.2 ^{NS}

SOM, Soil organic matter; TN, total nitrogen; ACC, active carbon; PMN, Potential mineralizable nitrogen; MBC, Microbial biomass carbon; MBN, Microbial biomass nitrogen; Extr C, Soil extractable carbon; Sig; statistical significance of regression. Significance levels: NS: $p > 0.05$; * $p < 0.05$; ** $p < 0.01$.

The low microbial composition at the GS sites (Figures 1–3) can be also attributed to the severe bush burning practices during harmattan, together with extreme drought conditions [25]. A number of reports have highlighted the effects of soil moisture and temperature [28], burning [9] on soil microbial biomass and fertility. According to this study, GS sites showed high Extr C contents (Figure 2b), predominantly from the burning regimes but did not stimulate the soil microbial biomass.

The rampant bush burning activity rather released various charred biomass forms into the soil mass, which subsequently contributes to soil CEC. According to Liang et al. [29], black carbon content in soil is effective in retaining soil CEC. In line with the present finding, Ajwa et al. [30] reported that long-term burning and N fertilization reduced MBC and MBN contents although total C was not significantly altered.

The type of fertilizer inputs in each zone impacted the soil microbial biomass (Figure 1). This was especially evident in the only inorganic fertilizer or the combined application of organic residues with inorganic fertilizer treatments. Applications of inorganic fertilizers have often produced inconsistent effects on soil microbial biomass or its fractions; enhancement [31], suppression [32], little or no effect [16,33]. The present results of DF are in accord with Grego et al. [33], where inorganic N application showed minimal effects on the soil microbial biomass compared to other fertilization types. Such discrepancies may be attributed to the soil type differences and their interactive effects with fertilization inputs as well as the time of measurement after inorganic fertilizer addition. In contrast, none of the fertilization practices in GS showed consistent effects on the soil microbial biomass indicators. Although the non-fertilization input approach showed encouraging results on soil microbial biomass content, human population pressure which subsequently makes agricultural lands less available has rendered fallow agriculture unsustainable. Sole low-medium quality organic residue amendment for extended period coupled with burning activity is less sustainable as depicted in the low soil chemical and microbial biomass contents.

The computed soil DI reflects the impacts of land management on the studied soil properties compared to the respective reference sites [34,35]. The higher the DI value, the higher the soil deterioration and vice versa. Irrespective of the agroecological zone differences, increased rate of inorganic fertilizer application in either inorganic or combined fertilization enhanced soil quality (Figure 4). This relates to the increased readily available mineral N which alone or synergistically resulted in enhanced growth of vegetation and subsequently increased C inputs to the soil. The contradiction in CMI of the present study compared to Blair et al. [36] and Yang et al. [37] are ascribed to the differences in the rate of application and organic material quality. Unlike the sole low-to-medium organic residues incorporation in the GS, the relatively high application rate with rich dairy manure as observed in Blair et al. [36] and Yang et al. [37] induced shifts in C dynamics as a percentage of total C and that subsequently increased the CMI.

4. Materials and Methods

4.1. Site Description and Soil Samplings

This study was carried out in 8 farming communities in the GS and DF zones of Ghana. The communities in GS comprised; Bannayilli, Janton Dabongshie, Dimabi, Nyankpala and Zugu, all located in the northern region of Ghana (9°15'1" N, 0°52'49" W; 170 m above sea level (a.s.l.)). In the DF, the communities constituted Akroso and Kade (6°8'48" N, 0°53'58" W; 170 m (a.s.l.) in the eastern region, and Mankessim (5°17'26" N, 0°58'57" W; 10 m (a.s.l.)) in the central region. Mean monthly rainfall (1983–2012) and mean temperature (1983–2012) obtained from Ghana Meteorological Agency showed a bimodal rainfall pattern and relatively lower temperature in DF relative to GS (Figure 5). The soils in GS are defined as weathered Lixisols/Luvisols/Plinthosols [38]. The soils in DF are Forest Ochrosols and are characteristically similar to those in GS [18].

In each agroecological zone, common fertilization schemes being adopted by farmers were investigated. In the DF, the two notable fertilization practices were sole inorganic fertilizer application (NPK and sulphate of ammonia (SOA)) and inorganic fertilizer (NPK and urea), with low-to-medium organic residues.

In the GS, however, the four fertilization practices comprised; sole inorganic fertilizer application (NPK and SOA), inorganic fertilizer, (NPK and SOA) with low-to-medium organic residues, sole low-to-medium organic residues, and no fertilization. It must be emphasized that the total nutrient

input in the studied fertilization practices varied and were all applied below the conventional recommended rates. This study, however, sought to evaluate the effects of each ‘low-input’ fertilization approach on soil productivity under specific agronomic schemes in their respective agroecological zones. Common farmland improvement approaches in each zone were selected as reference sites. At GS, the soil was taken from an average of five years old agroforestry field (Ref-GS) while neighboring 10-year fallow fields (Ref-DF) were chosen for the DF. For each fertilization type, 3 different farms with similar agronomic practices and input application rates under at least 5 years of continuous cultivation were chosen as replicates. Table S1 shows the detailed fertilization inputs and agronomic characteristics of the studied fields.

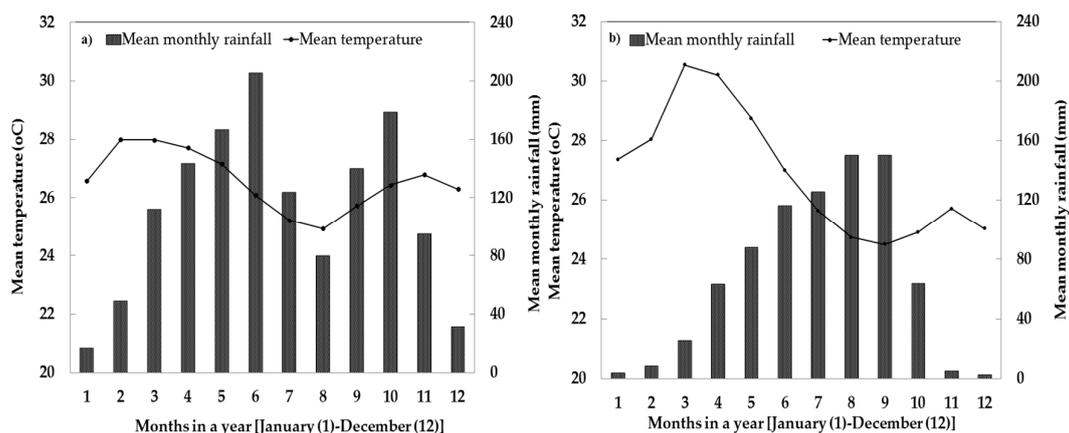


Figure 5. Mean monthly rainfall (1983–2012) and mean temperature (1983–2012) in (a) Deciduous forest and (b) Guinea savannah agroecological zone of Ghana.

Soil sampling was performed on each field in the dry season from March–April, 2015 using hand auger. On each field, ten core samples were randomly collected from 0 to 15 cm soil depth and pooled to form a composite sample. Soil samples were passed through a 2-mm sieve, stored in plastic bags at temperatures below 4°C before further biochemical analysis.

4.2. Soil Physicochemical Analyses

Soil Extr C was determined as described by Hu et al. [39]. Briefly, a 20 g dry weight soil equivalent was extracted with 50 mL 0.5 M K_2SO_4 after shaking mechanically at 120 rpm for 1 h. The dissolved organic C in filtrates was analyzed using TOC-L analyzer (TOC-L CPH, Shimadzu Corporation, Kyoto, Japan). Inorganic N was estimated by extracting a 10 g equivalent dry weight soil with 100 mL of 2 M KCl. The NH_4^+-N and NO_3^--N contents in the filtrates were respectively determined using the indo-phenol-blue colorimetric method [40] and continuous flow injection analysis method [41], on a UV-VIS spectrophotometer (UV mini 1240, Shimadzu, Kyoto, Japan). Total N and total C contents in soil (0.1 g dry weight) were determined using an automatic sensitive CN analyzer (Sumigraph NC-80 Auto analyzer; Sumika Chemical Analysis Service Ltd., Osaka, Japan). Available *p* content in soil samples was first extracted with 0.05 N sulphuric acid solution (H_2SO_4) buffered with 0.02 M ammonium sulphate ($NH_4(SO_4)_2$, pH 3) followed by colorimetric determination using the truog-soluble P method [42].

Soil pH was measured in the supernatant suspension of a 1:2.5 solid-liquid mixture (pH- H_2O) with a Beckman PKG-260 pH meter (Beckman Coulter Instruments Inc., Fullerton, CA, USA). The CEC was determined by extracting soil samples with 1 M ammonium acetate (NH_4OAc , pH 7) following the method by Schollenberger and Simon [43].

4.3. Soil Microbial Analyses

MBC and MBN were estimated using the modified fumigation extraction method proposed by Hobbie [44]. Briefly, 20 g dry weight equivalent sub soil samples designated as fumigated and non-fumigated were weighed out from each soil mass. Samples designated as fumigated were kept in a desiccator clouded with alcohol-free chloroform for 72 h. Non-fumigated samples and fumigated samples were extracted with 0.5 M of 50 mL K_2SO_4 . The dissolved organic C and N in 0.5 M K_2SO_4 extracts were respectively analyzed with TOC-L and TNM-L analyzer (TOC-L CPH, Shimadzu Corporation, Kyoto, Japan). Soil MBC was estimated from the relationship $MBC = C/K_{EC}$. The calibration value (K_{EC}) estimated as 0.45 [45], was used to convert the extracted organic C to microbial biomass C. Similarly, soil MBN was determined in the same extract from the relationship $MBN = N/K_{EN}$, where K_{EN} given as 0.54 according to Brookes et al. [46] was used to convert the extracted organic N to microbial biomass N. PMN was determined using the method described by Gugino et al. [47]. Briefly, NH_4^+ -N contents in soil were measured at time = 0 and $t = 7$ days for different sub samples of the same soil mass. At time = 0, NH_4^+ -N contents in 10 g dry weight soil was determined in 2 M KCl extract. For time = 7, 10 g dry weight soil was incubated with 10 mL distilled water at 30 °C for 7 days. The incubated soil mixture was extracted with 30 mL of 2.67 M KCl solution. NH_4^+ -N content in both extracts (at time = 0 and time = 7) were determined on a UV-VIS spectrophotometer. PMN is the difference between KCl-extractable NH_4^+ -N contents before (time = 0) and after (time = 7) incubation.

Active C was determined by following the procedure described by Weil et al. [48]. Soil sample (5 g) was mixed with 2 mL of 0.02 M $KMnO_4$ and 18 mL distilled H_2O in a centrifuge tube. The mixture was shaken for approximately 2 min and allowed to sit for 10 min. After settling, 0.5 mL of supernatant was mixed with 49.5 mL deionized water in a centrifuged tube and the active C content was determined on a UV-VIS spectrophotometer at an absorbance of 550 nm. SOM per unit mass of soil was determined by dry combustion [49] using the Electric muffle furnace (FUL 230 FA, Advantech Toyo Co., Ltd., Tokyo, Japan).

4.4. Calculation of Deterioration Index (DI) and Carbon Management Index (CMI)

DI was calculated following the procedure of Adejuwon and Ekanade [34]

$$DI = \sum \left(\frac{Ref - x}{Ref} \right) \times 100.$$

DI is the percentage changes of the individual fertilization schemes for the different soil properties relative to their respective reference sites. This was based on the assumption that the individual fertilization sites once had a similar quality composition as their reference sites [35]. The Ref is the reference mean value of a specific soil property and x is the corresponding soil property for the comparable fertilization scheme. The different soil properties comprised TN, TC, NO_3^- -N, and NH_4^+ -N, Extr C, active C, PMN, MBN, MBC, and SOM.

CMI, which is a measure of the soil C dynamics relative to the reference site [50] was calculated for each treatment in both agroecological zones as follows:

- (1) C pool index (CPI) was expressed as $CPI = \frac{\text{Sample total C}}{\text{Reference sample total C}}$
- (2) Lability index (LI) was calculated as $LI = \frac{\text{Lability of C in sample soil}}{\text{Lability of C in reference soil}}$ where,
- (3) Lability of C = $\frac{\text{Active C content}}{\text{TOC-Active C content}}$
- (4) And CMI is given as $CMI = CPI \times LI \times 100$

4.5. Data Analysis

The data were subjected to analysis of variance (ANOVA) using SAS software (SAS Institute Inc., Cary, NC, USA. Version 9.0) to detect significant differences among the different fertilization practices. Mean differences were compared using Duncan's multiple range test at $p < 0.05$. Correlations between the measured soil parameters were done using the Pearson's correlation coefficient. Multiple regression analysis was done among the soil parameters using the SigmaPlot program, version 11 (Systat Software Inc., San Jose, CA, USA).

5. Conclusions

The present study provides evidence showing how different types of fertilization influence the resultant soil fertility status of farmlands. Overall, the advantage of chemical fertilization, although applied below recommended rates was evident on soil biochemical composition in both agroecologies relative to the reference sites where conventional soil improvement approaches were being practiced. Additionally, high rates of fertilizer application in DF relative to GS regulated the resultant soil chemical composition and further contributed to enhanced soil microbial biomass dynamics. This was especially evident in either the sole inorganic fertilizer application treatment or its combined application with organic resources. On the other hand, organic residues incorporation as sole soil resource input in GS induced low soil productivity. Hence, it is suggested that the current low fertilization rate, especially in GS, must be increased to reasonable thresholds as in the deciduous forest in order to ensure its sustained effects on soil productivity.

Supplementary Materials: The following are available online at <http://www.mdpi.com/2073-4395/7/3/55/s1>. Table S1. Main management and agronomic characteristics of the studied sites.

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