

## Article

# A Consumer Grade UAV-Based Framework to Estimate Structural Attributes of Coppice and High Oak Forest Stands in Semi-Arid Regions

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**Abstract:** Semi-arid tree covers, in both high and coppice growth forms, play an essential role in protecting water and soil resources and provides multiple ecosystem services across fragile ecosystems. Thus, they require continuous inventories. Quantification of forest structure in these tree covers provides important measures for their management and biodiversity conservation. We present a framework, based on consumer-grade UAV photogrammetry, to separately estimate primary variables of tree height (H) and crown area (A) across diverse coppice and high stands dominated by *Quercus brantii* Lindl. along the latitudinal gradient of Zagros mountains of western Iran. Then, multivariate linear regressions were parametrized with H and A to estimate the diameter at breast height (DBH) of high trees because of its importance to accelerate the existing practical DBH inventories across Zagros Forests. The estimated variables were finally applied to a model tree above-ground biomass (AGB) for both vegetative growth forms by local allometric equations and Random Forest models. In each step, the estimated variables were evaluated against the field reference values, indicating practically high accuracies reaching root mean square error (RMSE) of 0.68 m and 4.74 cm for H and DBH, as well as relative RMSE < 10% for AGB estimates. The results generally suggest an effective framework for single tree-based attribute estimation over mountainous, semi-arid coppice, and high stands.

**Keywords:** tree structure; biomass; DBH; *Q. brantii*; Zagros Forests

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## 1. Introduction

Forests significantly contribute to the global carbon cycle by providing the largest reserves of terrestrial carbon [1]. Therefore, quantifying forest structure, particularly across fragile and erosion-prone forest sites, is of major importance for their conservation and management [2,3]. The structural analysis of such vegetation is also a prerequisite to assess their current condition and evaluate their ecosystem services such as carbon stocks, CO<sub>2</sub> uptake [4,5], and aboveground biomass (AGB) [6] as affected by disturbances and land use changes. The primary structural attributes that are common among forest ecosystems include the number of trees, species composition, tree size, and the AGB [7], which are considered inputs to derive secondary attributes such as total leaf area, tree and leaf biomasses, and multiple other ecosystem services [3]. Tree height and diameter at breast height (DBH) are among the basic forest inventory attributes [8,9], which are integrated by many allometric equations for estimating the AGB [10,11] and are therefore essential information for quantifying forest carbon cycle [12], carbon stock [13], and global climate change [14]. Traditional methods of measuring tree height are costly and time-consuming, and they are particularly infeasible for spatial upscaling to larger domains

[15]. In addition, one of the main obstacles of these measurements is the lack of sufficient visibility to the tree top, particularly for broadleaves with circular crowns. One way to bypass this is via the possibility of using high-resolution nadir view to enable two- and three-dimensional measurements by fully observing the surface of the tree canopy, which is commonly enabled by remote sensing data and methods [16–18]. In addition to height estimation, they can be rather intuitively applied to estimate a range of other tree canopy variables, including the crown area [19] and its diameter [20]. The list of studies on tree height estimations, by 3D remote sensing methods, is enormous [21], which includes studies that applied data from both active and passive sensors. LiDAR [13,22,23] and RADAR [24,25] data have been extensively studied, which are yet financially prohibitive despite their generally high estimation accuracies [26]. On the other hand, passive high-resolution sensors, such as photogrammetric Unmanned Aerial Vehicles (UAVs), offer possibilities to derive three-dimensional models at, comparatively, much lower costs [27] that allow for capturing the forest canopy surface via image matching and the subsequent post-processing techniques [28]. On the one hand, multi-rotor drones are easier to take off and land, so they are therefore preferred by researchers because they are usually cheaper and more flexible for scientific purposes [29]. On the other hand, larger-scale studies have been carried out with the help of fixed-wing drones equipped with optical cameras [30]. A number of studies have been performed on fixed-wing drones in measuring the vertical structural attributes of forests for large-scale areas [31–33]. Although passive sensors do not penetrate the crown layer and thus cannot provide information from understories, they are beneficial in measuring the horizontal and vertical structural attributes for single-story forest stands.

Furthermore, DBH is another essential forest structural variable [34,35] that is obtained by measuring the diameter of the trunk at a height of 1.3 m [36]. Although DBH can be conveniently measured with traditional methods such as caliper [35], problems are faced towards its spatial upscaling, particularly across remote and difficult-to-access areas with variations in tree growth forms [37]. In such cases, the use of remote sensing methods is considered as a reliable alternative [38–40]. In general, studies on direct estimation of DBH by remote sensing have mainly focused on using LIDAR data [41,42], while passive optical data fail to deliver sub-canopy observations due to tree foliage cover [37,43]. This entails an indirect estimation of the DBH, using 3D photogrammetric data, for example, by relating DBH and tree height [44] or DBH and canopy area [45]. Additionally, Benjamin and Russell, 2021 [46] estimated DBH using complementary measurements on UAV-derived models. Therefore, they first created a regression model using the area of tree canopies, calculated the canopy diameter by fitting a circle to the canopy of each tree, followed by fitting against the ground reference data. This yielded relative errors ranging between 19.7% and 33.7%. Whereas methods based on under-canopy UAV flights have returned promising results for estimating single tree DBH [47], they commonly entail an additional LiDAR payload that is prohibitive for budget applications designed and applied by most forest administrations. Moreover, AGB reserves are mainly related to forest structure and species composition [48,49]. The AGB is thus an important indicator for measuring the carbon sequestration capacity [50]. Direct methods to estimate AGB include destructive methods [51], while allometric equations are generally used as alternative indirect methods [52].

The UAV technology has enabled producing accurate small-scale estimates based on derivations such as digital surface models (DSMs), which can be further applied to derive side-products such as DTMs, and calibrate spatial models of forest structural attributes. The process of producing DTM in forested areas commonly involves interpolations [53–56] with the main problem being the lack of empty areas, i.e., areas without trees, in dense stands [57]. On the contrary, sparse tree covers enable photogrammetric models that lead to the high accuracy of canopy height models (CHM = DSM-DTM), which are crucial to estimating tree structural variables. For example, Vahid et al. (2021) [58] estimated tree height from CHM using high-pass and local maximum filters across a test site, located

within the Hyrcanian mixed forests of northern Iran, followed by estimation of the tree canopy area using the Invert Watershed Segmentation (IWS) algorithm. This consequently led to calculating crown diameter with a relative RMSE of 10.1% and 7.02% for estimating the height and crown diameter, respectively.

The Middle East and West Asia are currently facing one of the worldwide harshest heatwaves, also termed as a heat dome (see <https://earthobservatory.nasa.gov/images/148430/heatwave-scorches-the-middle-east> access on 15 September 2021). They are considered obvious signs of climate change that have largely affected fauna, flora, and human communities. The Zagros Forests located in western Iran (with marginal expansion into southeastern Turkey and northern Iraq) are among the most fragile ecosystems in west Asia. In Iran, they cover ca. 20% of the total countrywide vegetated area. These forests play a key role in maintaining the harsh climate, supplying water, and economic and social balance of Iran and the region [59]. These forests are dominated by three oak species *Q. infectoria* Olivier, *Q. libani* Olivier, and *Q. brantii* Lindl., with the relative dominance of the latter species increasing along the latitudinal gradient [59–61]. The vertical structures of these species occur in both high and coppice forms, where coppice stands generally have a vertical one-story structure. The tremendous power of sprouting and asexual regeneration of *Q. brantii*, and the incompetence of other species for the establishment, has led to almost pure dominance of *Q. brantii* in central and south Zagros zones, yet with different proportions of high and coppice stands affected by both the climate and lifestyles of local populations (nomadic or sedentary). Due to their unique historical development and multiple anthropogenic influential phenomena, their crucial structural characteristics include species, the origin of trees (high or coppice), number of sprouts per sprout-clumps for coppices, sprout diameter and tree canopy diameters, DBH for high trees, as well as tree height and AGB for trees in both vegetative forms. Despite all associated challenges, continuous monitoring of Zagros Forests is largely constrained by accessibility, lack of logistics, and financial capabilities, leading to hitherto available sporadic and insufficient quantitative structural information, even on small-scale spatial domains.

In this study, we suggest a consumer grade UAV-based framework to retrieve a number of the most important primary and secondary structural attributes of *Q. brantii*-dominated stands in both high and coppice forms along the north-south gradient of Zagros Mountains. We focused on the height and area of tree canopies as directly estimable attributes, whereas the DBH of high trees and AGB were indirectly estimated. We estimated DBH via its relationship with crown area and height. Finally, we estimated the AGB by both allometric models with tree height and DBH as independent variables, as well as nonlinear regression models (random forest regression, RFR) with crown surface area and height as predictors. The aim was to provide both science and practice with objective comparative estimates of single tree-based *Q. brantii* AGB across these extremely fragile ecosystems that have been continuously under climatic and anthropogenic pressures.

The commonly-applied field methods of measuring the structural attributes of Zagros Forests are costly and time consuming due to their mountainous topography and limited accessibility. Therefore, there is an interest in facilitating rapid field measurements of the structural attributes, particularly by means of 3D remote sensing methods. To the best of our knowledge, no international study has been published, so far, on the use of remote sensing in measuring the structural attributes of trees in the Zagros region. Therefore, here, we addressed the basic research question of whether UAV photogrammetric methods can serve as a complementary method in ground measurements of Zagros Forests. In particular, the innovative aspects of this study are that (1) it focuses on a rich dataset of plot measurements that are distributed across the latitudinal gradient of Zagros Forests, (2) it differentiates high and coppice growth forms for all analyses, (3) it compares AGB estimations from machine learning and allometric methods to serve both science and practice. This includes surveys to answer the following questions:

- Are 3D photogrammetric models accurate enough to estimate the height of single trees in Zagros?

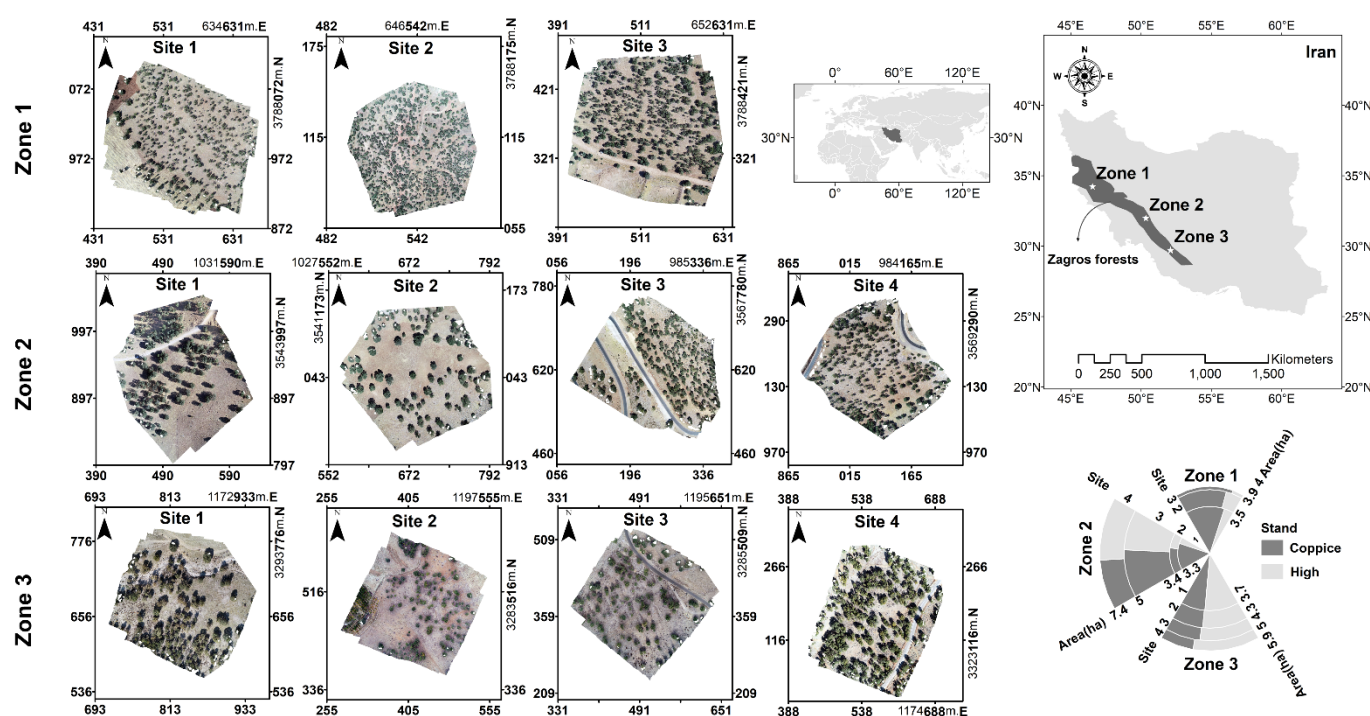
- Can DBH (as an essential variable that is unable to be directly derived from photogrammetric models) be estimated using other UAV-measured primary attributes?
- How do the available single-attribute allometric equations for *Q. brantii* in Zagros Forests perform compared to commonly-applied nonlinear models (RFR) on UAV-extracted variables to predict AGB?

## 2. Materials and Methods

### 2.1. Study Area

Zagros Forests consist of ca. 1300 and ca. 200 km extensions along the latitudinal and longitudinal gradients, respectively. These semi-arid ecosystems comprise an area of ca. 6 million ha and cover about 44% of Iran's total area of forests [59,62]. The average annual temperature ranges between 9 and 25 degrees Celsius, with the maximum 74 degree difference between the coldest and warmest temperatures, which indicates a continental climate [62]. In general, the northern zone of Zagros is associated with comparatively wetter and cooler climate compared with the central and southern zones [59].

We divided Zagros Forests into three regions, based on both general climatic conditions and administrative division, in which we established our sampling locations: northern zone (zone 1), middle zone (zone 2), and southern zone (zone 3). Then, 11 study sites across all three regions with dominance of *Q. brantii* were selected with the aim of deriving their structural attributes in both high and coppice forms by covering a representative latitudinal gradient (Figure 1).



**Figure 1.** The study area includes 11 sites from the three northern (Zone 1), middle (Zone 2), and southern (Zone 3) regions of the Zagros Forests. Each site is represented by UAV orthophotos. The locations are in the UTM zone 38N coordinate system. According to the pie-chart, all sites comprise both high and coppice trees, while in Zone 1, the predominant form is coppice.

### 2.2. Field Measurements

The reference data on a set of tree structural features, including height (H), canopy area (A), diameter at breast height (DBH), and the AGB were either field-sampled or

estimated as described below [63]. The inventories were conducted within the 11 study sites established along the latitudinal gradient of Zagros mountains during 4 months in 2019, i.e., close to UAV imaging dates (Section 2.3) within the framework of provincial sub-projects of “the National Zagros Forest Monitoring Plan” coordinated by the Research Institute of Forests and Rangelands (RIFR) of Iran (Figure 1). The reference AGB was indirectly calculated using the allometric equations (Table 1) [64,65] with the independent variables of H, DBH, and average diameter of the canopy (D) for each of the high and coppice forms separately (Table 2). We relied on the available field records when assigning the stands to each of coppice or high vegetative forms, since single trees in all sites, except those in Zone 1, were tagged with either coppice or high origin, which we further adopted in our analysis. Due to the absence of this designation in all three sites in Zone 1, we took all trees with a single sprout as high trees and those with the number of sprouts > 1 as coppice trees. Thus, we partly deviated from the pure ecological principle in defining the vegetative form, which is primarily based on sexual or asexual growth origin.

**Table 1.** Allometric equations for estimating biomass in Persian oak species.

| Stand   | Independent Variable            | Equation             |
|---------|---------------------------------|----------------------|
| High    | Average tree crown diameter (m) | $Y = 0.881x^{3.228}$ |
|         | DBH (cm)                        | $Y = 0.615x^{1.865}$ |
|         | Tree height (m)                 | $Y = 0.067x^{3.921}$ |
| Coppice | Average tree crown diameter (m) | $Y = 2.534x^{2.383}$ |
|         | Tree height (m)                 | $Y = 1.868x^{2.487}$ |

**Table 2.** Summary of ground measured values.

| Zone | Site | Stand   | Number of Trees | Symbol   | Minimum | Maximum | Mean  | STD  |
|------|------|---------|-----------------|----------|---------|---------|-------|------|
| 1    | 1    | Coppice | 295             | H (m)    | 1.55    | 4.6     | 2.828 | 0.58 |
|      |      |         |                 | D (m)    | 1.1     | 9.55    | 3.9   | 1.43 |
|      |      | High    | 54              | H (m)    | 1.4     | 4.1     | 2.843 | 0.64 |
|      |      |         |                 | D (m)    | 0.85    | 6.75    | 3.549 | 1.67 |
|      |      |         |                 | DBH (cm) | 3       | 24      | 11.02 | 4.68 |
|      | 2    | Coppice | 200             | H (m)    | 1.8     | 6       | 3.775 | 0.85 |
|      |      |         |                 | D (m)    | 1.15    | 6.65    | 3.787 | 1.05 |
|      |      | High    | 50              | H (m)    | 1.4     | 6       | 3.378 | 0.94 |
|      |      |         |                 | D (m)    | 0.85    | 5.55    | 2.902 | 1.2  |
|      |      |         |                 | DBH (cm) | 6       | 31      | 12.14 | 5.05 |
|      | 3    | Coppice | 252             | H (m)    | 2.1     | 6.2     | 3.896 | 0.86 |
|      |      |         |                 | D (m)    | 1.1     | 7.4     | 3.7   | 1.19 |
|      |      | High    | 40              | H (m)    | 2       | 5.2     | 3.538 | 0.85 |
|      |      |         |                 | D (m)    | 0.6     | 3.75    | 2.036 | 0.71 |
|      |      |         |                 | DBH (cm) | 4       | 15      | 8.925 | 2.97 |
| 2    | 1    | Coppice | 59              | H (m)    | 3.2     | 11.1    | 7.436 | 1.78 |
|      |      |         |                 | D (m)    | 2.8     | 13.3    | 6.933 | 1.92 |
|      |      | High    | 10              | H (m)    | 3.2     | 9.6     | 7.08  | 1.96 |
|      |      |         |                 | D (m)    | 3.95    | 10.3    | 6.84  | 1.94 |
|      |      |         |                 | DBH (cm) | 13      | 98      | 44.3  | 22.4 |
|      | 2    | Coppice | 12              | H (m)    | 2.1     | 14.8    | 10.44 | 3.25 |
|      |      |         |                 | D (m)    | 1.9     | 16.6    | 9.504 | 3.37 |
|      |      | High    | 19              | H (m)    | 3.2     | 15      | 10.27 | 2.46 |
|      |      |         |                 | D (m)    | 2.25    | 16.6    | 9.629 | 3.45 |
|      |      |         |                 | DBH (cm) | 12      | 98      | 36.79 | 21.8 |
|      | 3    | Coppice | 49              | H (m)    | 3.2     | 12      | 7.388 | 2.23 |
|      |      |         |                 | D (m)    | 2.4     | 9.6     | 6.349 | 1.95 |
|      |      | High    | 62              | H (m)    | 2.5     | 13.4    | 7.195 | 2.12 |
|      |      |         |                 | D (m)    | 2       | 11.8    | 5.406 | 1.8  |

|   |   |         |     |          |      |       |       |      |
|---|---|---------|-----|----------|------|-------|-------|------|
| 3 | 4 | Coppice | 26  | DBH (cm) | 2    | 59    | 20.27 | 9.41 |
|   |   |         |     | H (m)    | 4.5  | 11.9  | 7.65  | 1.74 |
|   |   | High    | 39  | D (m)    | 3.5  | 12.9  | 8.21  | 2.37 |
|   |   |         |     | H (m)    | 4.5  | 13.4  | 7.697 | 2.01 |
|   |   |         |     | D (m)    | 3.7  | 12.95 | 7.624 | 2.19 |
|   |   |         |     | DBH (cm) | 2    | 59    | 21.49 | 10.9 |
|   | 1 | Coppice | 44  | H (m)    | 3    | 13    | 7.766 | 2.14 |
|   |   |         |     | D (m)    | 1.6  | 14.5  | 6.378 | 2.62 |
|   |   | High    | 23  | H (m)    | 2.1  | 12.5  | 7.835 | 2.86 |
|   |   |         |     | D (m)    | 0.75 | 12    | 6.024 | 3.18 |
|   |   |         |     | DBH (cm) | 3    | 57    | 32.65 | 13.9 |
|   |   |         |     | H (m)    | 4.3  | 8.3   | 6.1   | 0.94 |
|   | 2 | Coppice | 75  | D (m)    | 1.65 | 10.75 | 4.764 | 1.91 |
|   |   |         |     | H (m)    | 3.8  | 8.5   | 6.05  | 1.3  |
|   |   | High    | 36  | D (m)    | 1.25 | 9     | 4.265 | 1.9  |
|   |   |         |     | DBH (cm) | 9    | 42    | 23.17 | 7.56 |
|   | 3 | Coppice | 63  | H (m)    | 3.9  | 8.5   | 6.322 | 1.04 |
|   |   |         |     | D (m)    | 1.5  | 12    | 5.657 | 2.02 |
|   |   | High    | 29  | H (m)    | 3    | 8.2   | 6.183 | 1.12 |
|   |   |         |     | D (m)    | 1    | 11    | 5.545 | 2.08 |
|   |   |         |     | DBH (cm) | 10   | 42    | 22.34 | 7.6  |
|   | 4 | Coppice | 107 | H (m)    | 2.8  | 12    | 7.686 | 2.03 |
|   |   |         |     | D (m)    | 2.75 | 14    | 5.932 | 2.28 |
|   |   | High    | 48  | H (m)    | 3    | 10    | 6.958 | 1.53 |
|   |   |         |     | D (m)    | 1.75 | 9     | 3.738 | 1.57 |
|   |   |         |     | DBH (cm) | 4    | 58    | 19.31 | 8.63 |

### 2.3. UAV Imaging

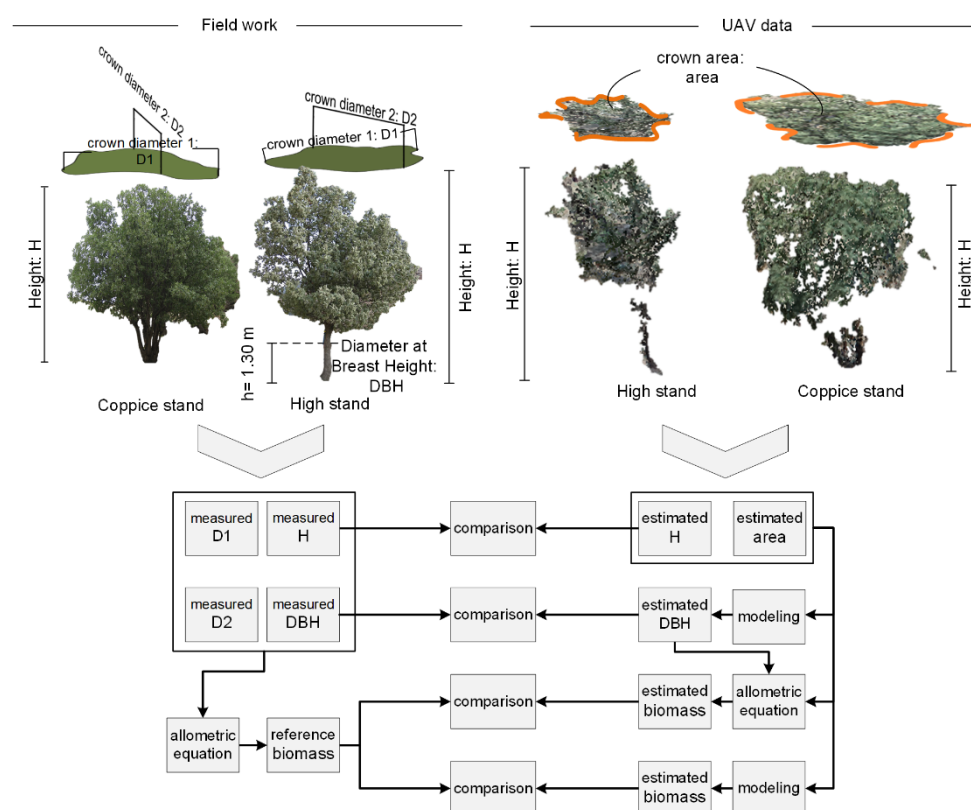
We applied a consumer-grade DJI Phantom 4 pro multi-rotor UAV (DJI, 2016) for aerial imaging. The device included a 3-axis stabilization gimbal, 1" CMOS sensor camera and FOV 84° 8.8 mm/24 mm lens. The full list of technical specifications can be found at <https://www.dji.com/phantom-4-pro/info> (access on 15 September 2021). The flights for different sites were designed according to the site-specific topographic conditions and tree cover but generally as double-grid networks with 80% horizontal/vertical overlaps conducted with an iOS version of Pix4DCapture (<https://support.pix4d.com/hc/en-us/articles/204010419-iOS-Pix4DCapture-Manual-and-Settings> access on 15 September 2021) installed on an iPad 2018 tablet. The flights were carried out during 2nd–5th of September 2019 for northern and central zones and 6th–7th of November 2019 for southern zone with a resolution of 5472 × 3648 pixels at 70 m flight altitude. In general, images were taken with a resolution of 5 cm, although there were slight differences in image resolution at various sites. This resulted in ca. 125 image tiles per forest site.

Five control points that were recorded per site (four corners and a centroid) that included GPS-measured latitudes, longitudes, and altitudes. The coordinates were measured using the real-time kinematic (RTK) method for northern and central plots and the post-processing kinematic (PPK) method for southern plots by applying two GPS receivers of Trimble 5700 and 5800 and a Zephyr Geodetic antenna. For the PPK method, the recorded points were post-processed using geodetic fixed reference points within Shamim National Geodetic Platform (<https://shamim.ssaa.ir/sbc> access on 15 September 2021).

### 2.4. Methodology

Figure 2 shows the general flowchart of this research. Based on the field-measured variables of tree height, DBH, the first diameter, and the second diameter of the tree crown, allometric equations (Table 1) were parametrized to estimate the AGB values (see the left side of Figure 2). Furthermore, UAV-derived CHMs and orthophotos were applied

to extract the height and crown surface as primary variables, using which multivariate linear regressions were developed to estimate the DBH with H and A as independent variables. Consequently, the AGB values were predicted by (1) allometric equations as a function of H, DBH, and D (Table 1) and (2) nonlinear regressions with H and A as independent variables (see Figure 2 right wing). We divided the entire reference data into separate training and test datasets for all three stages for which estimations were carried out. These three stages included (1) derivation of UAV-based models, (2) segmentation of individual tree crowns, and (3) derivation of single tree structural variables.



**Figure 2.** General workflow of the study.

#### 2.4.1. UAV-Derived Terrain and Surface Models

The raw UAV images were processed in Agisoft Metashape 1.7 (<https://www.agisoft.com>, accessed on 15 September 2021). The products from UAV data comprised DTM, DSM, and orthophoto for each site. Then, we derived the CHM using  $CHM = DSM - DTM$ .

#### 2.4.2. Individual Tree Crown Segmentation

Individual trees were segmented in order to extract the height and crown area of individual trees. Despite the high dominance of *Q. brantii* within all sites, the individual trees were highly diverse in terms of their height and crown shape, resulting in the use of CHM for tree segmentation [19,66]. We applied the Marker-Controlled Watershed Segmentation (MCWS,) as a previously-tested and efficient method, across sparse tree stands [67,68]. We abstained from using the classical watershed as an unsupervised classification method due to the over-estimations raised by the presence of noise in images [69], which necessitates variable levels of pre- and post-processing. The MCWS method has been previously used to detect single tree crowns by applying a set of pre- and post-processors [67,70], with more technical details in [69] to which the reader is referred.

#### 2.4.3. Extracting the Height and Canopy Area of Individual Trees

After delineating the crown boundary of each tree, the variables of height and crown area for each tree were calculated as the maximum value of CHM in the tree polygon and the polygon area, respectively. The fact that the field reference data included the measured tree heights enabled a comparison of the CHM-derived and field-measured heights for high and coppice trees separately.

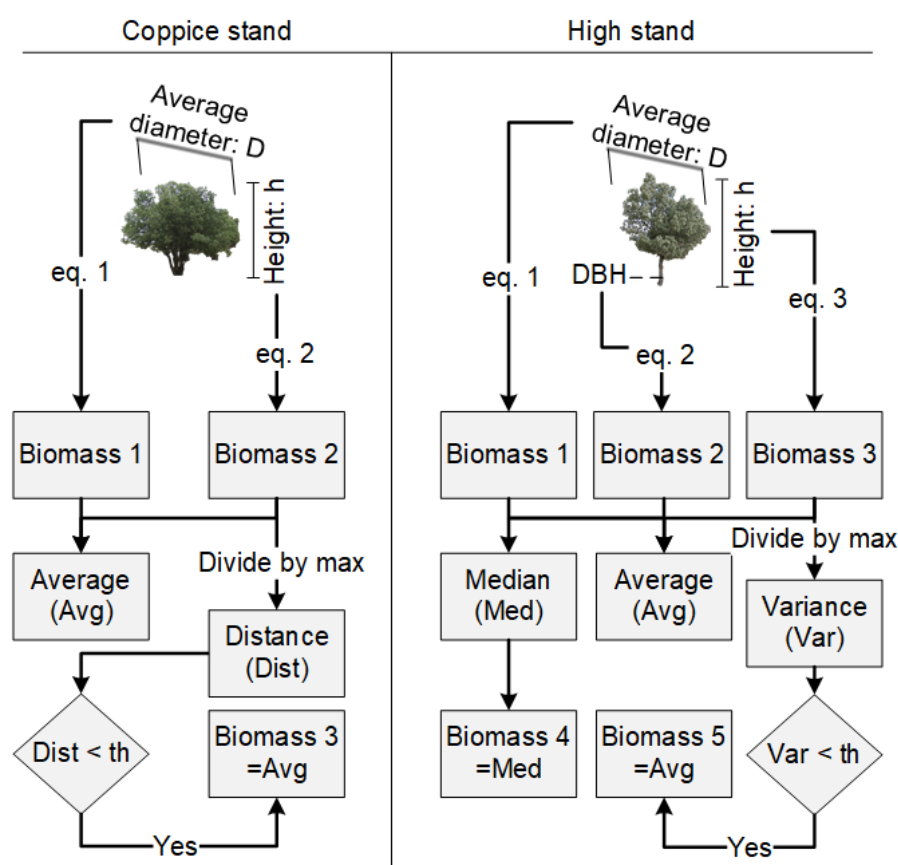
#### 2.4.4. DBH Estimation

Since the DBH was infeasible to be directly derived from UAV photogrammetric data, we estimated it as a function of H and A [71–73] to serve its practical implementation. To this aim, multivariate linear regression models, using these two parameters (H and A) that were extracted in the previous steps, were applied [74]. The field-measured DBH values (for high trees) were used to train and test the model with random 50% partitioning within each forest site.

#### 2.4.5. AGB Estimation

Allometric equations for estimating the biomass of *Q. brantii* trees in both high and coppice forms, using independent variables of H, DBH, and D followed [64] (Table 1). The H was directly extracted from UAV-based CHM, while the DBH was indirectly derived as described above. We directly applied the last two equations in Table 1 to estimate the AGB for trees in high stands, while the remainder of the equations were applied to the coppice stands. Moreover, we additionally included the A variable in AGB estimation, whereas it was not originally included in the allometric equations listed in Table 1. Therefore, a nonlinear machine learning model, with H and A as independent variables, was applied for each of the tree vegetative forms. We employed RFR, a supervised algorithm based on ensemble learning by using a combination of a set of regression trees [75,76]. It converges by (1) randomly drawing numerous initial bootstrap samples with replacements from the original training dataset, (2) fitting each regression tree to each of the bootstrap samples, and (3) predicting all trees by constructing decision trees during training [77]. The RFR was numerous reported as a powerful and accurate approach, followed by its compatibility with nonlinear relationships [78]. The number of regression trees, identified as the number of estimators, was defined as hyperparameter by testing the RMSE on training data for values ranging between 100 and 1000 at intervals of 50 [79]. This resulted in selecting 700 trees as the optimal number of estimators. According to the number of existing allometric equations that were used to calculate the reference AGB values, the total three and two AGB values were obtained for each high and coppice tree, respectively. Moreover, the average of those models was also calculated as an additional AGB value.

Here, obtaining different AGB values for a given tree was inherent, since each allometric equation in Table 1 features a different independent variable (H, D, and DBH). Therefore, we performed a numerical modification to derive an average instead of an average that possibly does not incorporate different possible values for AGB. We did this modification by first defining a threshold value and then checking the variance of different AGB values prior to calculating the final reference value. This ensured that the AGB value was ignored, i.e., it was treated as an outlier, if the variance (for high trees) or distance (for coppice trees) was higher than the defined threshold value. Otherwise, we calculated the mean value of the applied multiple AGB equations and considered it as the modified average. This resulted in the final four AGB values for high stands and three values for coppice stands. In addition, the median of the three AGB values of high stands was considered to be an additional AGB value (Figure 3).



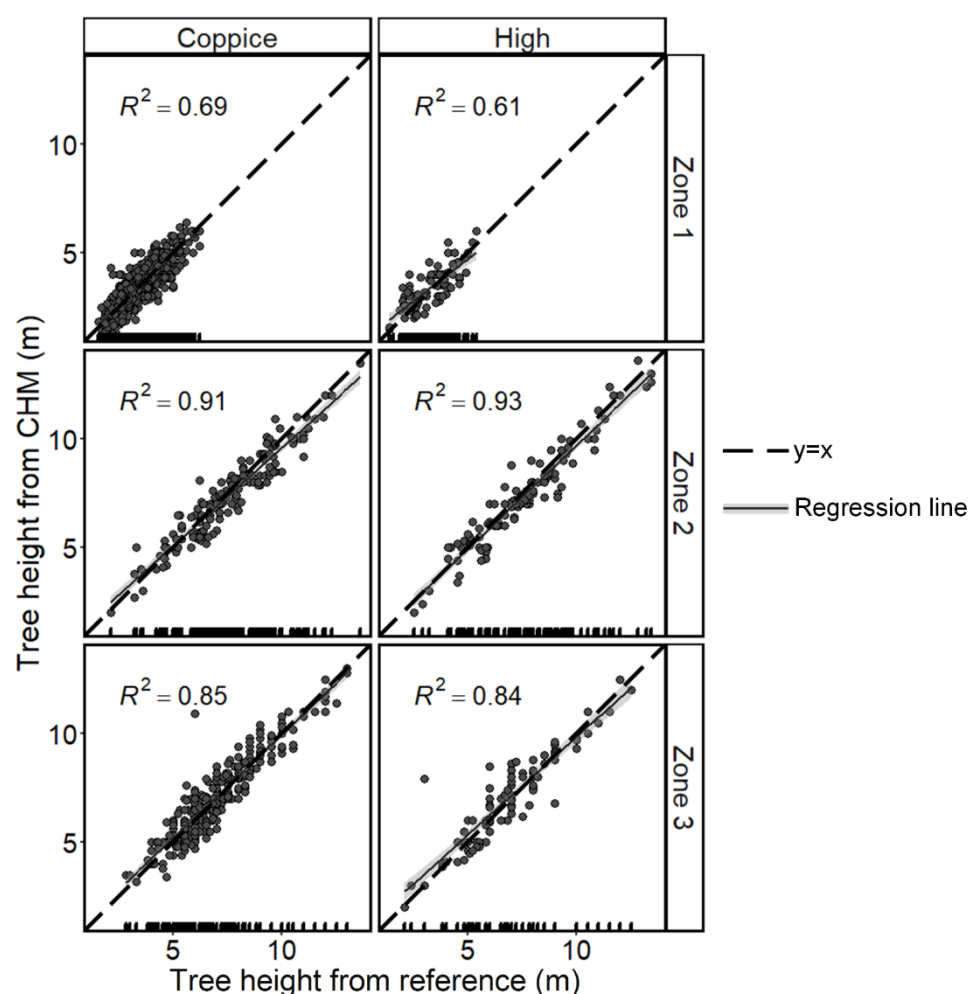
**Figure 3.** Flowchart of generating different AGB values.

Following the generation of reference AGB values, the RFR was trained for each group. These models were evaluated using the Leave-One-Out-Cross-Validation (LOOCV) [80].

### 3. Results

#### 3.1. Tree Height Estimation

The results of tree height estimation for both vegetative forms were obtained in each zone (Figure 4). The reader is referred to the Appendix A for the site-specific results. The results suggested the  $R^2$  values of 0.69, 0.91, and 0.85 for high stands and 0.61, 0.93, and 0.84 for coppice stands for CHM-derived tree heights compared with the reference values in Zones 1–3, respectively. Apart from the practically accurate fits, it also suggested a high-quality segmentation that was previously carried out by MCWS prior to height estimation.

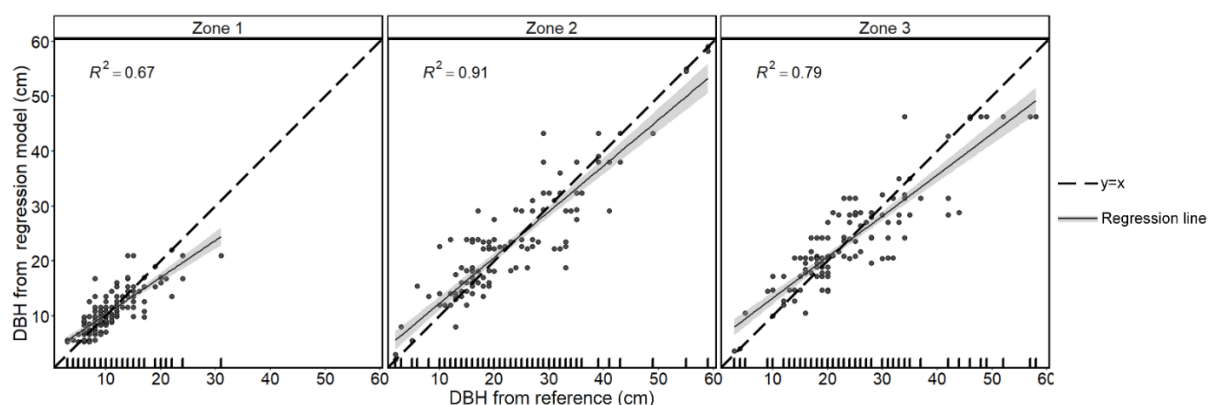


**Figure 4.** Comparison of H from CHM with actual measured values (reference values).

As shown in Figure 4, different values of tree height, from 1 m to 15 m in all three zones, were extracted with high accuracy.

### 3.2. DBH Estimation

The results of multivariate linear regression, as a comparison between the modelled DBH and the terrestrial reference values, showed a high linear correlation of estimated values with the reference values (Figure 5). The site-specific results are included in the Appendix B.



**Figure 5.** Comparison of DBH, estimated from the multivariate linear regression model with independent variables of height (H) and canopy area (A) in trees with high stand form.

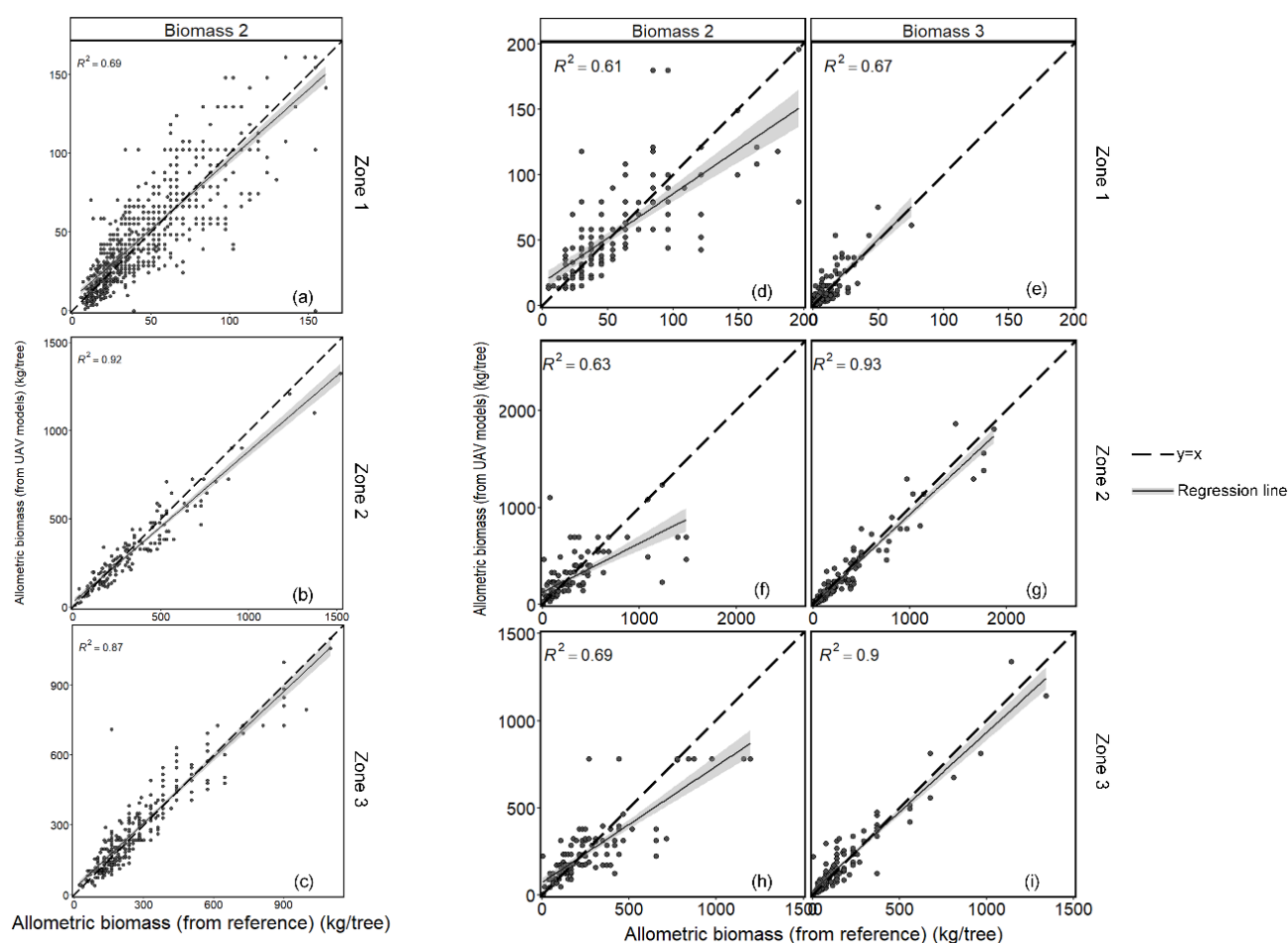
### 3.3. AGB Estimation

#### 3.3.1. AGB estimation by allometric equations

The results of allometric equations, with H, D, and DBH as independent variables, were partially affected by the prediction accuracy of H and DBH from UAV data (Table 3). In addition, the results of AGB estimations, when comparing field- and UAV-derived variables, are summarized in Figure 6. We directly extracted the crown area (A) and tree height (H) from the UAV models (CHM and orthophoto), while we indirectly estimated the DBH using the regression models on H and A. Therefore, only allometric equations on DBH and H were used. For the site-specific results, the reader is referred to the Appendix C.

**Table 3.** RMSE value of H and DBH parameters, estimated from photogrammetric models.

| Parameter | Stand   | UAV Variables   | Best RMSE | Unit |
|-----------|---------|-----------------|-----------|------|
| Height    | Coppice | CHM             | 0.573     | m    |
|           | High    |                 | 0.572     |      |
| DBH       | High    | CHM, Orthophoto | 2.620     | cm   |

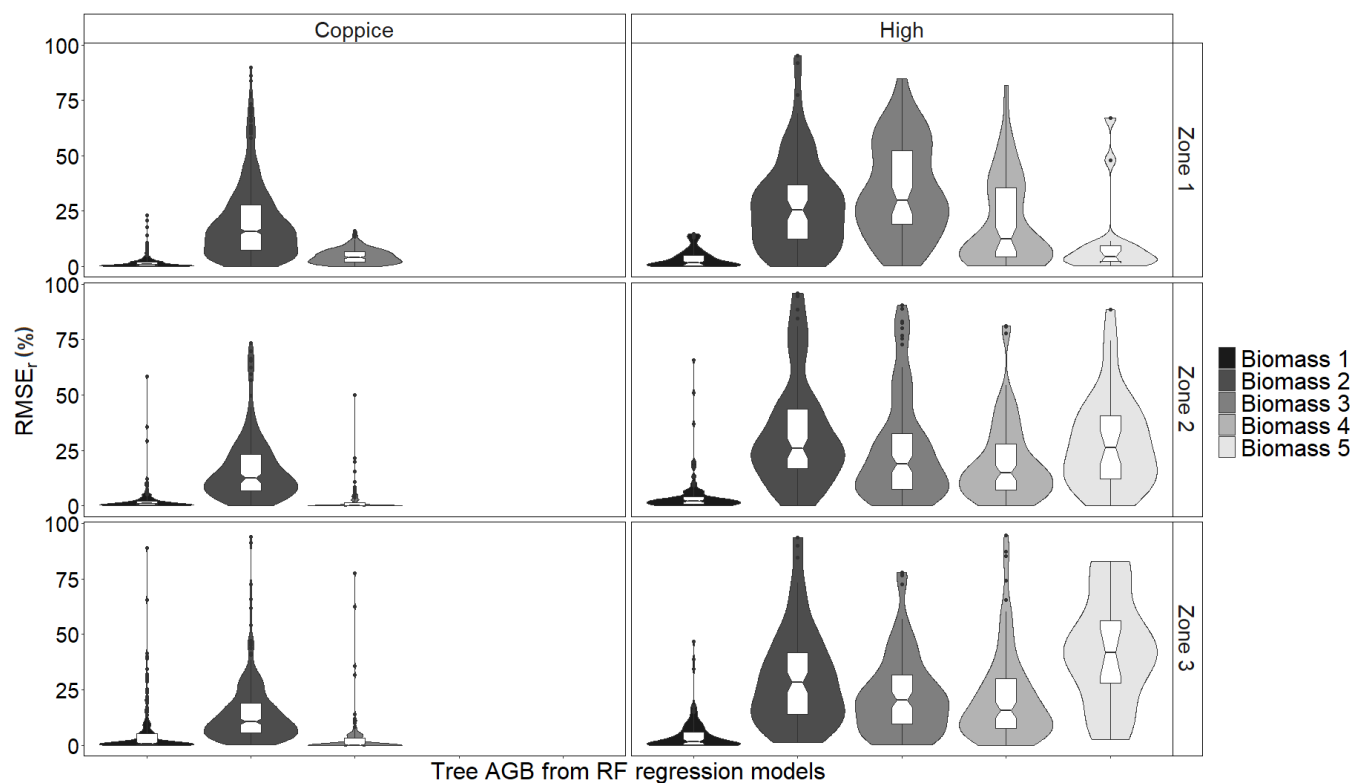


**Figure 6.** Biomass values obtained from allometric equations on variables measured by UAV (vertical axis) and reference values (horizontal axis) in coppice stands (a–c) and high stands (d–i).

The results suggested that the AGB model 3, with H as an independent variable, showed a better fit with the reference data compared with the model 2 that used DBH as an independent variable.

### 3.3.2. AGB Estimation by RFR

We only used H and A to estimate AGB using RFR, since the UAV-modelled DBH was a function of H and A. The relative RMSE values of LOOCV, for different reference AGB models, suggested comparatively higher performance of model 1 in mean rRMSE of AGB estimation for both high and coppice stands (Figure 7).



**Figure 7.** Relative RSME values of RFR models for AGB estimation.

As expected, model 3 returned intermediate results for coppice trees, since it was averaged from models 1 and 2. This was also partially the case for high stands. Among models 2 and 3, i.e., models based on DBH and H, the model with allometry on height in high trees outperformed the model with DBH for Zones 2 and 3. This was indicated by the lower mean rRMSE and the higher density of error distribution in the lower values for model 3. The allometric model 4, i.e., the median of models 1, 2, and 3, showed higher accuracy than model 5 for Zones 2 and 3, as resulted by its comparatively lower average value and the higher density of error distribution in the mid-range. In all three zones, the DBH model for the high stands was shown to perform weaker than almost all other models (except for Zone 1). Furthermore, model 1, i.e., the allometry on D, performed better than all other models. For coppice stands in all three zones, D and H were the most and least influential variables in allometric models, respectively (the site-specific results are presented in Appendix D).

## 4. Discussion

Zagros Forests are among the most ecologically diverse, yet most degraded ecosystems in Iran and the Middle East. Upon first reports from early 2000, the region has been severely affected by a multi-agent oak dieback. This phenomenon mostly occurred across the central and southern parts of the Zagros region [81,82]. This necessitates rapid and cost-effective quantification and monitoring of their structural attributes to maintain their sustainability. However, harsh topography and infeasibility of field works, over remote and partly inaccessible areas, majorly constrain providing detailed structural information, which can be alleviated by locally adjustable UAV-based approaches. These benefits are

mainly valid across small-scale areas. Due to the problems raised by limited flight duration, limited areal coverage and fixed gimbals, the use of consumer grade UAVs are commonly limited to small-scale areas and plot-based inventories [83,84]. On the other hand, due to the sensitivity to weather conditions, it is necessary to ensure the weather stability before starting the flight operation [29]. Another major limitation of UAVs concerns the legal restrictions on flying, which avoids their use in many areas, including those close to country borders [85]. Despite those critical limitations, the advantages of using UAVs in small areas outweigh their disadvantages [86] and thus argument their use as alternatives or complements to other remote sensing data.

Here, we mainly focused on indirect methods for estimating essential structural variables of trees in natural *Q. brantii* dominated stands, in high and coppice vegetative forms that were spread over three distant biogeographical zones, along the latitudinal gradient of Zagros Forests. The canopy surface parameters and tree height for *Q. brantii* dominated stands were directly measured on UAV-based photogrammetric models. However, the AGB is generally calculated in practice by allometric equations, based on either height (H), average crown diameter (D), or diameter at breast height (DBH) of the trees in single-factor allometric equations [64]. Whereas we relied on a range of previously available local AGB models for trees grown within coppice and high stands, we primarily tested the possibility to parametrize the models with UAV-estimated tree variables to serve the practical implementation of our results. To this aim, DBH was estimated, by linear regression of the H and A, and was then used within the allometric equations to derive the AGB. Apart from that, we also followed the commonly-tested RFR approach based on nonparametric spatial models with UAV-estimated input variables. We deliberately avoided using a large feature space, with multiple textural and spectral indices, to maintain the parsimony of the models as well as their comparability against the allometric models (see, for example, [87]). This might be contrary to some relevant studies in which larger feature spaces were applied at the cost of applying exhaustive variable selection techniques (see [88]), which, however, are associated with limited comparability between allometric and RFR modeling approaches.

With the rationale to minimize the potential error of tree height that occurred in field measurements (due to insufficient visibility of top crown in broadleaves, human errors, and device errors) [89–92] numerous former studies applied high resolution remote sensing data (mostly LiDAR). Other relevant studies using UAV photogrammetry reported RMSEs of 0.28 m [28] 0.34 m [93] 1.30 m [19] and 3.04 m, respectively. Our results of 0.82 and 0.64 m RMSE, within high- and coppice-dominated stands, were in-line with those studies, though we generally do not recommend direct comparisons of absolute error rates among studies conducted in areas with different ecosystem properties and varying standards for field data collection.

One of the main challenges in estimating the tree height is the production of CHM in dense forests [94]. To produce an accurate CHM, Vahid et al. (2021) [58] acquired UAV data in two different seasons corresponding to leaf-off and leaf-on conditions, followed by estimating single-tree heights. The estimations resulted in  $R^2 = 0.808$ . Although leaf abundance was stated to be an influential factor on single tree height estimation [95], we used only leaf-on season images in our study within multiple sites of the Zagros Forests, due to different density of trees ranging from semi-dense to sparse, while there was no fully dense forest stand in our study sites. We reached  $R^2$  values ranging between 0.61 and 0.93, depending on both factors of site and growth form. Despite the fact that the tree species were largely similar across all surveyed forest sites, the difference in accuracy across the three zones of Zagros Forests can be attributed to two factors of (1) the imaging conditions and (2) the applied RFR machine learning method. Although we attempted to capture imagery, possibly under constant weather conditions and flight planning in all three Zones, the difference in accuracy may be due to differences in the number of trees, i.e., density in each area (see Table 2), slightly different times of the flight and partially different topography of the sites. The latter factor can introduce variable levels of artifacts

in digital terrain modeling [96], single tree detection [97], and the subsequent AGB estimation [98] when generally working on all 3D sources of airborne data. However, this was out of the scope of our study.

In terms of DBH, it is among the most conveniently-measured variables in the field. Its direct derivation from passive UAV imagery is infeasible, but it is possible to be estimated by models incorporating image-based UAV measurements, such as tree height and canopy area [99]. Here, we estimated H and A from UAV data across areas with high trees, which was in line with previous studies that estimated DBH using passive photogrammetry data [45]. Our linear model reached an  $R^2$  in range of [0.67, 0.91] for Zones 1, 2, and 3, which was comparatively in line with [100] that reported  $R^2$  of 0.738 in a nonlinear model of DBH as a function of canopy diameter and tree height.

Partly relevant studies such as [99] reported inferior predictions of AGB as a function of UAV-based tree height and canopy area compared with those based on DBH. However, the latter study was conducted in a mangrove ecosystem, in which both height and canopy area are less influential on the AGB compared to most terrestrial tree ecosystems, mainly due to their comparatively more complex structure and coppice-only vegetative forms. Here, our AGB models for both vegetative forms showed superiority of tree height to DBH when estimating AGB from UAV data by allometric equations. To the best of our knowledge, this is the first study that (1) differentiated between estimates of coppice and high stands and (2) estimated AGB based on photogrammetric data in the Zagros region, whereas even field-based studies on AGB in the Zagros region are rare [101]. Allometric equations, with DBH as an independent variable, were reported to be mostly correlated with AGB [102]. This was mostly the case for average canopy diameter (D) in field-based studies carried out within the Zagros ecosystem [103], which can be mainly attributed to the coppice and multi-stem origin of the majority of its forest stands. Similarly, the highest performances in both high and coppice forms in our UAV-based study were observed for models that were parametrized with average crown diameter. Apart from the dominant coppice origin, a further reason could be the generally extensive crown of *Q. brantii* trees within commonly sparse Zagros stands, which leads to the formation of a major part of the variance described for the AGB. Comparable UAV-based studies that combined spectral information and H, extracted from UAV photogrammetric data, include, for example, [104] for poplar plantations, with  $R^2$  values reaching 0.54 and RMSE = 0.23 kg/m<sup>2</sup> for the AGB (in a site with average AGB of 1.68 kg/m<sup>2</sup>). Furthermore, Lin et al. (2018) [105] suggested a nonlinear regression model for estimating AGB with height as a single independent variable extracted from UAV oblique photography, which resulted in RMSE = 54 kg and rRMSE = 34% when compared with allometric equations on H. Here, our UAV-based models delivered RMSE values < 10%, using only H and A as the independent variables. Although we are aware of the general limitation of reference data calculations using allometric equations [106], we attempted to compensate for this by (1) using diverse allometries, including three equations for high-dominated and two for coppice-dominated stands, and (2) additionally averaging the resulted reference values via using the modified mean and median values.

Finally, our UAV photogrammetric models to estimate a set of essential structural attributes can be tested across not only the Zagros region, but also over other similarly structured semi-arid forest stands. In forests similar to the Zagros Forests, the single trees extraction is more accurate due to the sparse tree distribution. Moreover, the height of each tree can be estimated with higher accuracy compared with denser tree ecosystems. In dense forests, deriving an accurate UAV-based DTM is mostly challenged by (1) difficulty in separating the trees and (2) ground invisibility caused by tree density. Both constraints have a direct negative effect on the estimation of crown area and height, as well as indirectly on the estimation of DBH and AGB. However, in some cases, high resolution images can be helpful in extracting dense trees, but this adds another limitation: namely, the infeasibility of imaging large areas with consumer-grade UAVs. We tested direct and indirect methods to estimate the height, crown area, DBH, and AGB, separately, for stands

dominated by trees with coppice and high vegetative forms. When comparing the estimates against detailed site-based results (see the Appendices A–D), the results suggested comparatively high accuracies, achieved by the photogrammetric method, applied on consumer grade UAV datasets. This can further support both forest scientific and practical administrative applications within the Zagros region, which already largely lacks such estimates and thus cannot be quantitatively monitored following the occurrence of multiple diminishing factors including land-use change, tree decline, increased cattle grazing, and frequent wildfires.

## 5. Conclusions

Sustainable management of fragile semi-arid mountainous forests entails quantitative monitoring of tree attributes on various spatial scales. Currently, sparsely available field estimates in Zagros is exacerbated by weak technical infrastructure, e.g., infeasibility of using Terrestrial LiDAR, GeoSLAM, or airborne LiDAR, which entail relying on methods based on a combination of limited field data and passive photogrammetry or remote sensing data. We suggested a UAV-assisted workflow to measure and estimate a range of primary and secondary structural attributes on single tree-level across multiple forest sites, located along the latitudinal gradient of Zagros Forests. Differentiation between high and coppice growth forms is highly relevant in Zagros and other oak-dominated ecosystems due to the strong anthropogenic pressure on tree regeneration, as well as high ability of oak trees for asexual regeneration. Despite the reported uncertainties, the results largely support the use of UAVs for plot-based measurements of single trees over such difficult-to-reach areas. Our results for direct estimation of H and A from three-dimensional models were practically accurate over almost the entire forest sites, which justified their further application in estimating the single tree AGB using both allometric equations and machine learning. Moreover, estimating the DBH by regression on A and H as input variables was done to suggest a practical workflow for both science and forest administration in Zagros, where regular field-based methods are infeasible due to inaccessibility and lack of logistic facilities for frequent monitoring. However, this is associated with an unknown amount of aggregated error. We also showed that high resolution UAV data enable an indirect retrieval of single tree DBH using UAV-extracted A and H as independent variables with negligible correlation. Finally, we suggest further research on (1) the use of consumer-grade UAV photogrammetry for tree measurements, in particular across coppice trees to reduce the existing uncertainties via testing the effects of topography, data acquisition time, and shadow on model performance and (2) to follow spatial upscaling methods, based on a combination of UAV and satellite multispectral data for larger spatial scales.

**Author Contributions:** Conceptualization, H.L. and S.A.F.; methodology, S.A.F. and H.L.; software, S.A.F.; validation, S.A.F.; formal analysis, S.A.F. and H.L.; investigation, S.A.F. and H.L.; data curation, H.L.; writing—original draft preparation, S.A.F. and H.L.; writing—review and editing, S.A.F. and H.L.; visualization, S.A.F.; supervision, H.L. All authors have read and agreed to the published version of the manuscript.

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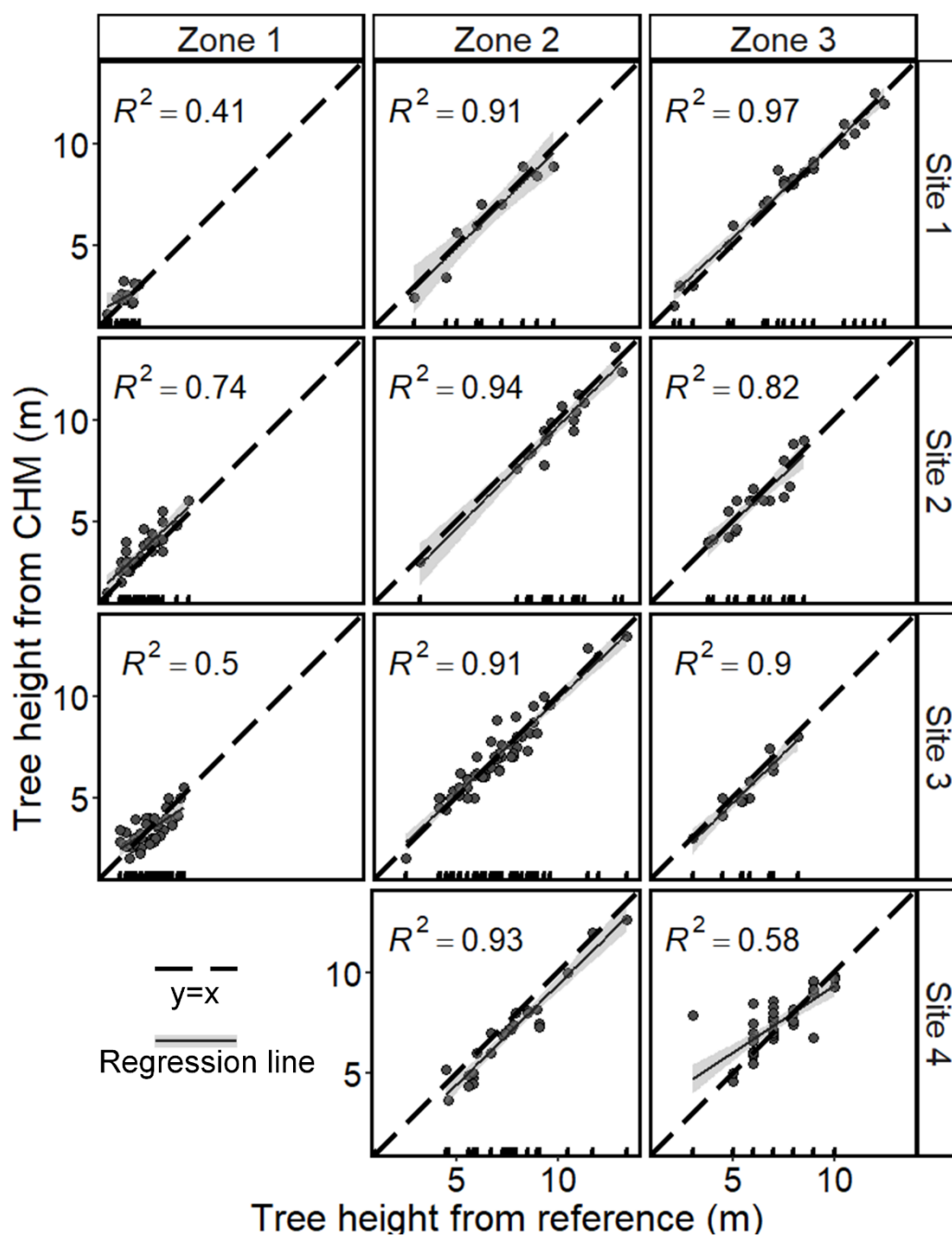
**Data Availability Statement:** Not applicable.

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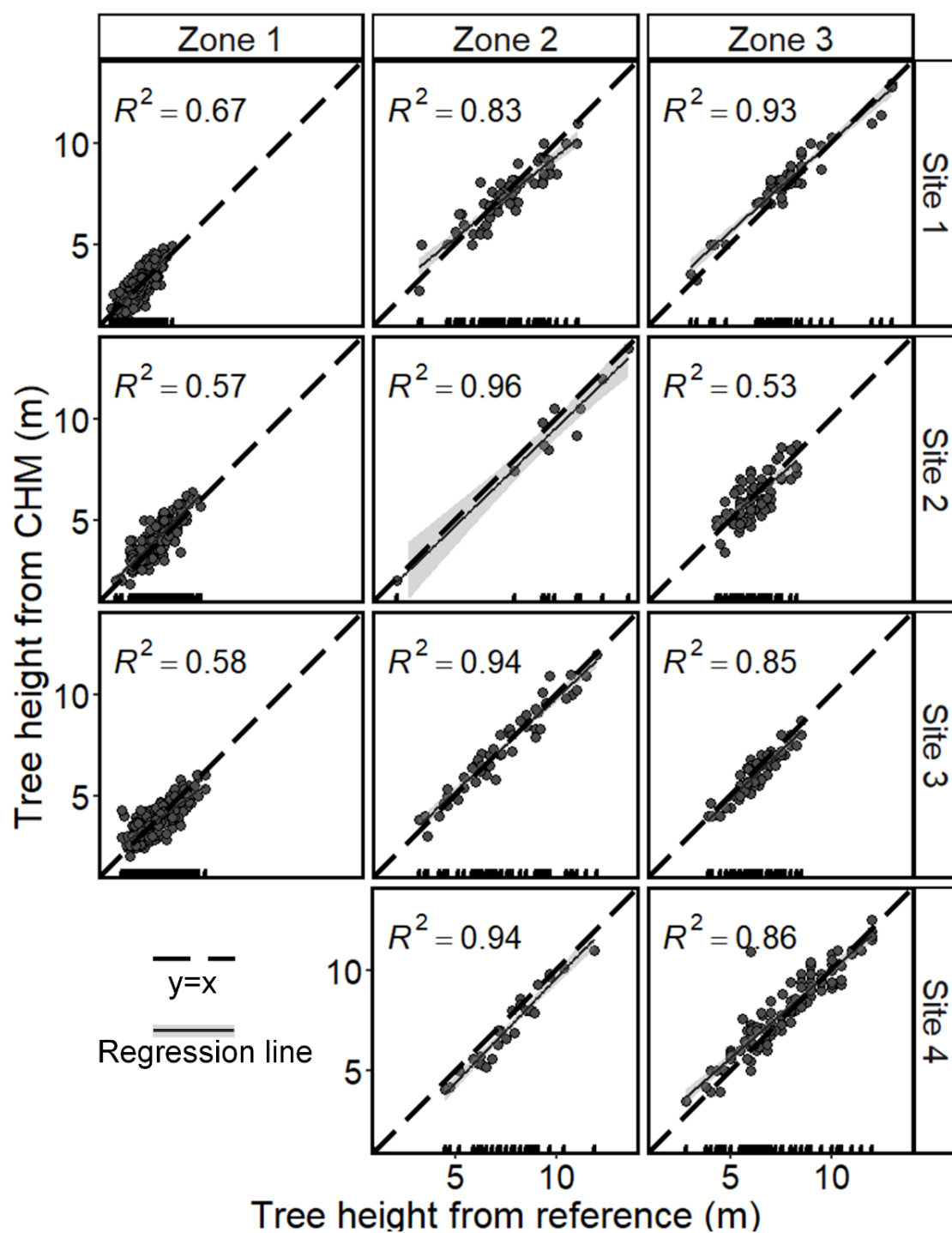
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**Conflicts of Interest:** The authors declare no conflict of interest.

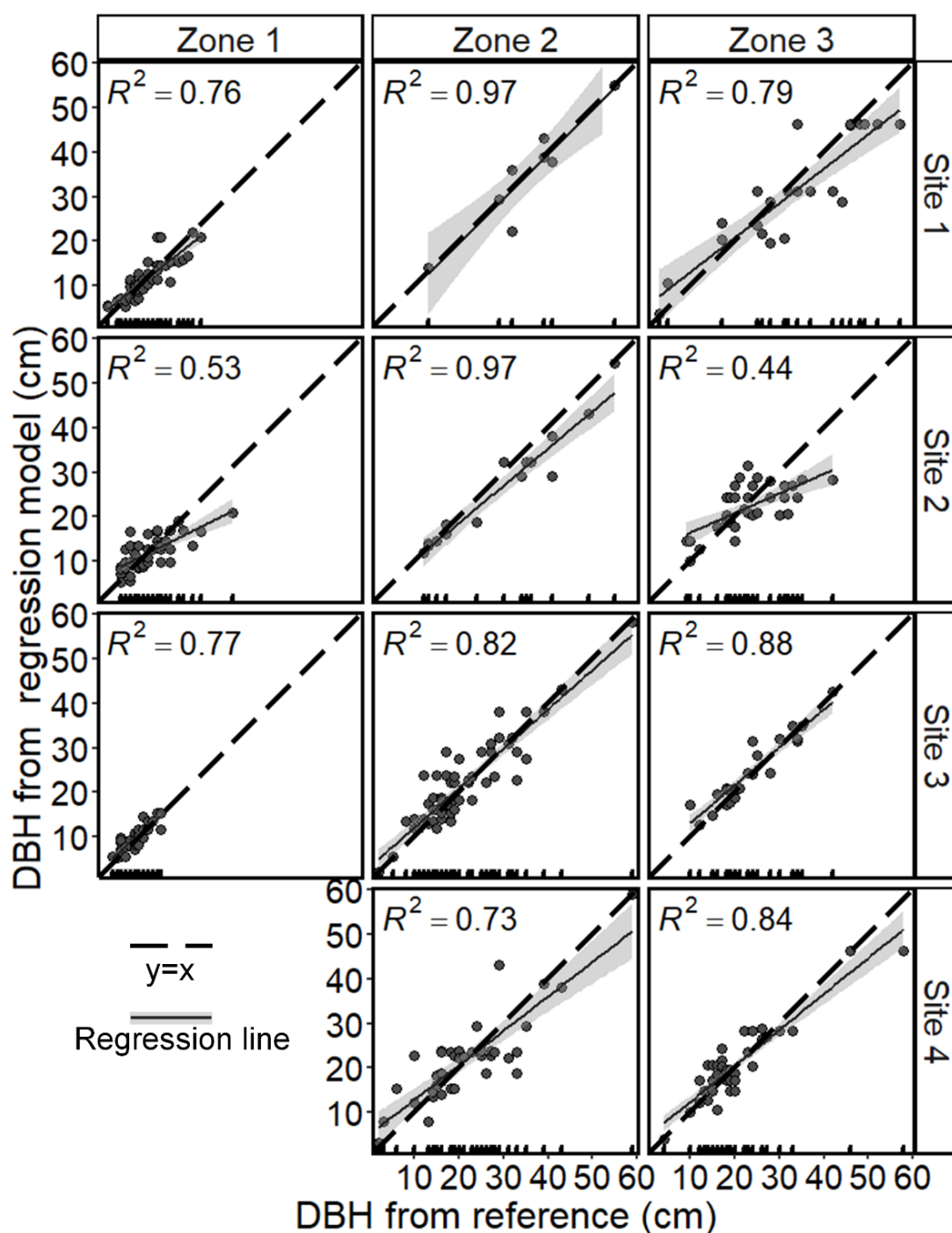
## Appendix A



**Figure A1.** Site-specific comparison of H from CHM with actual measured values (reference values) in trees with high stand form.

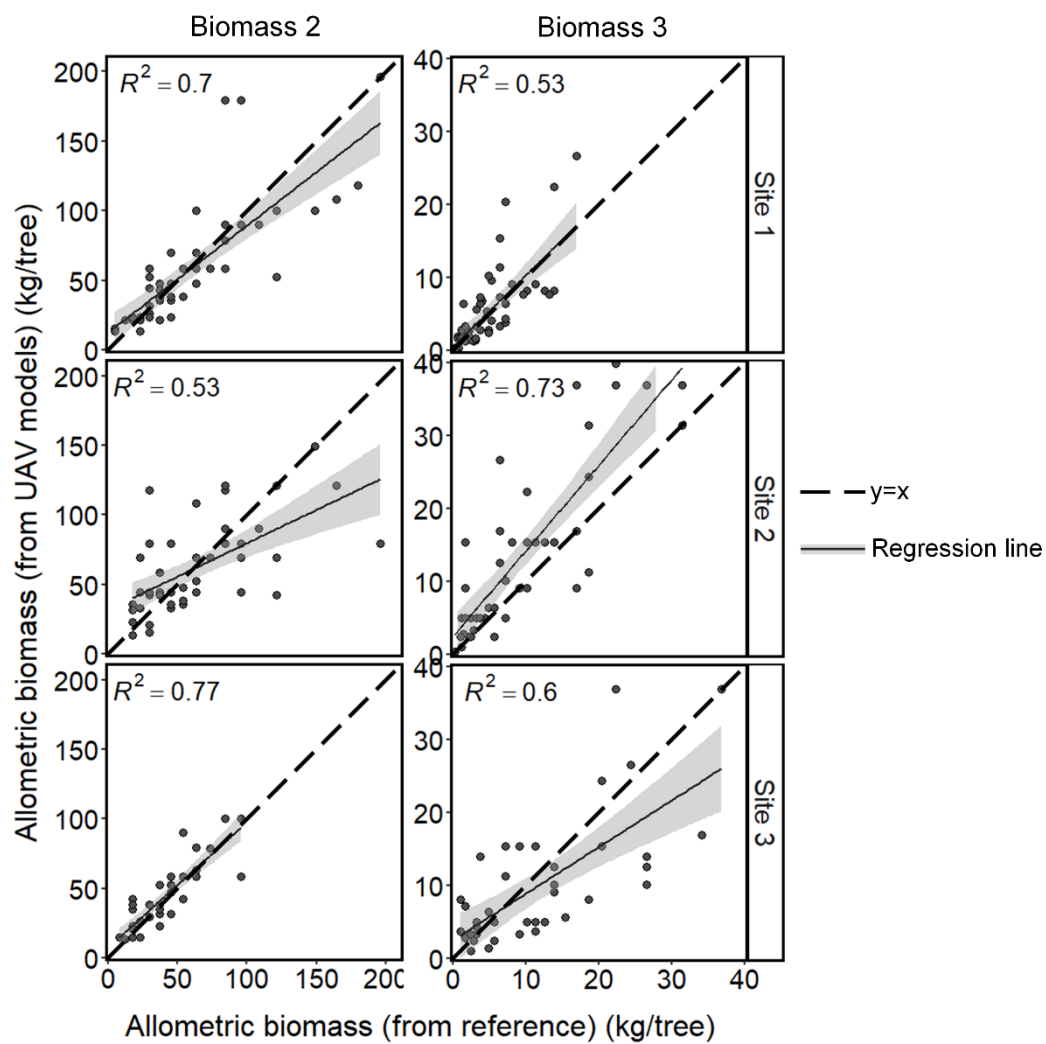


**Figure A2.** Site-specific comparison of H from CHM with actual measured values (reference values) in trees with coppice stand form.



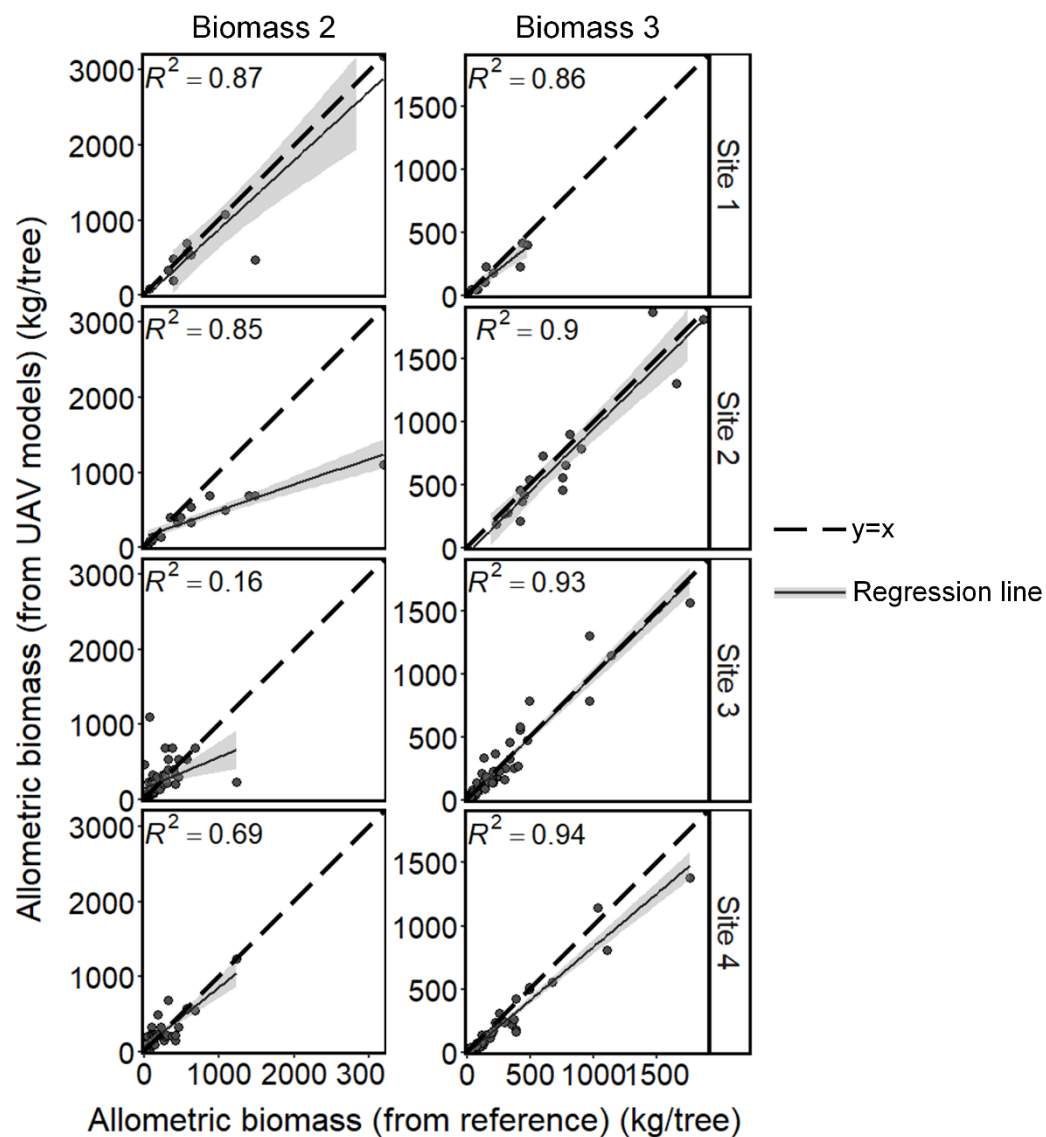
**Figure A3.** Site-specific comparison of DBH estimated from multivariate linear regression model with independent variables of height (H) and canopy area (A) in trees with high stand form.

## Appendix B

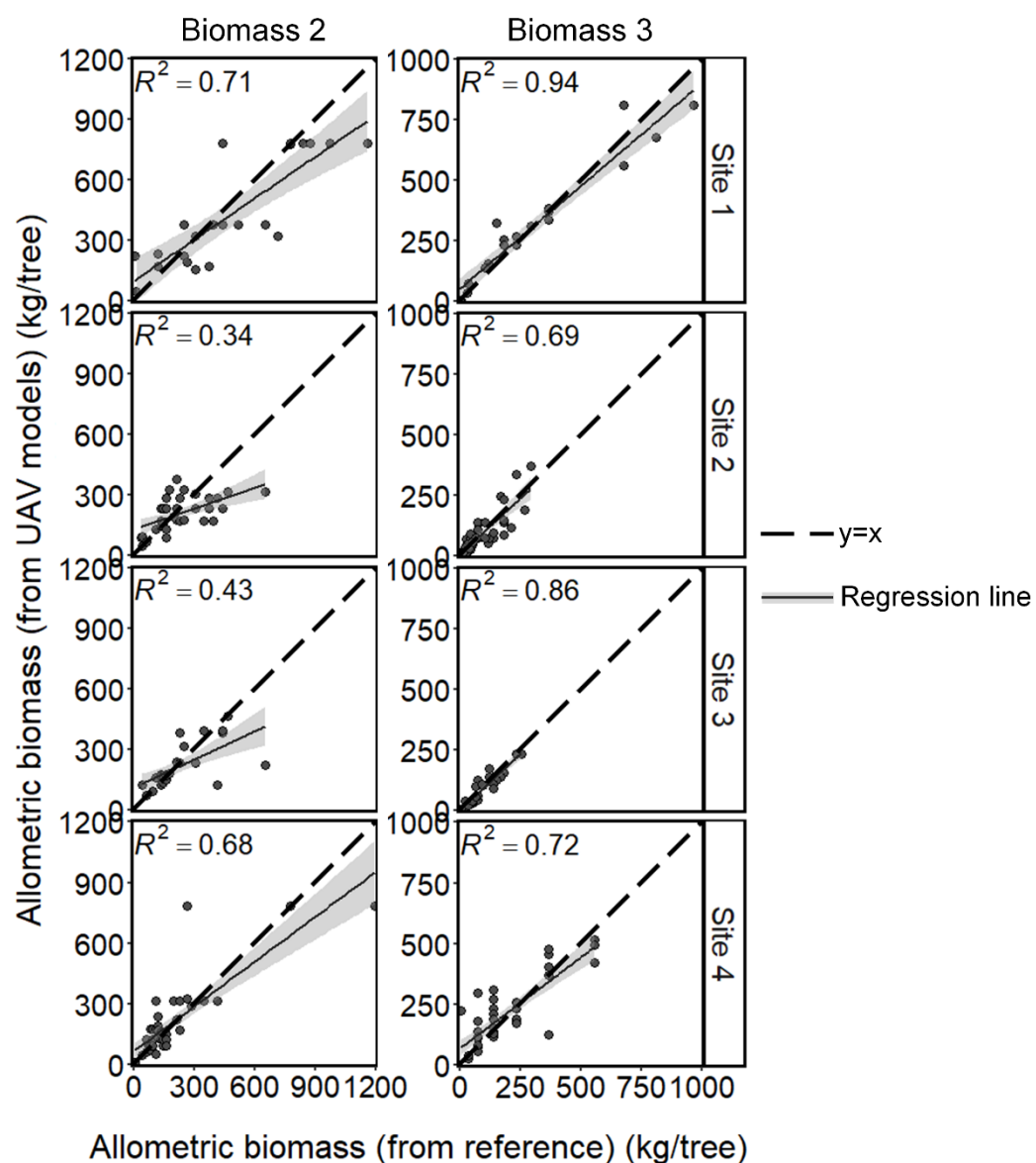


**Figure A4.** Site-specific AGB values obtained from allometric equations on variables measured by UAV (vertical axis) and reference values (horizontal axis) in high stands in Zone 1.

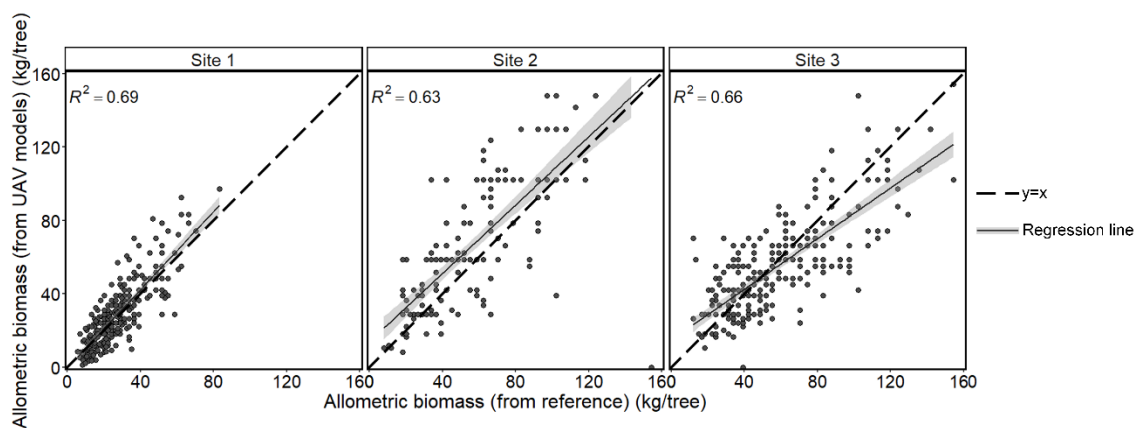
## Appendix C



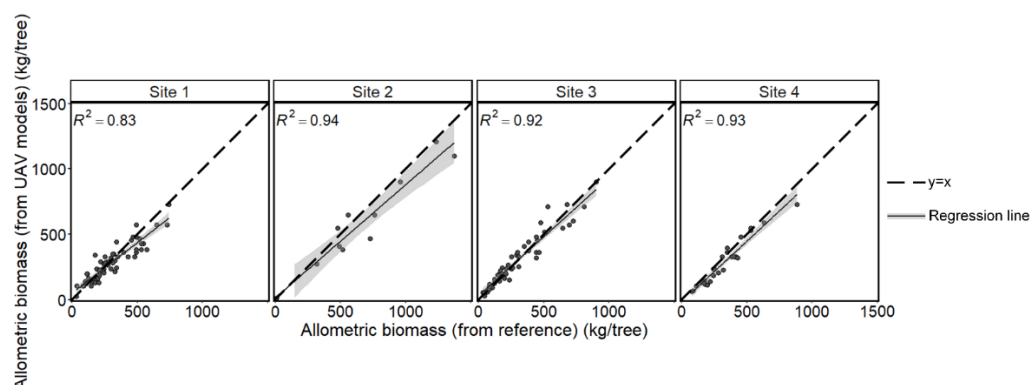
**Figure A5.** Site-specific AGB values obtained from allometric equations on variables measured by UAV (vertical axis) and reference values (horizontal axis) in high stands in Zone 2.



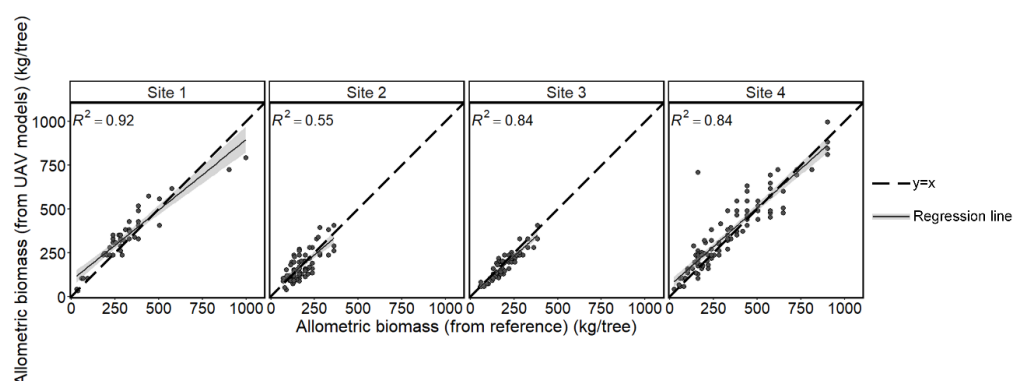
**Figure A6.** Site-specific AGB values obtained from allometric equations on variables measured by UAV (vertical axis) and reference values (horizontal axis) in high stands in Zone 3.



**Figure A7.** Site-specific AGB values obtained from allometric equations on variables measured by UAV (vertical axis) and reference values (horizontal axis) in coppice stands in Zone 1.

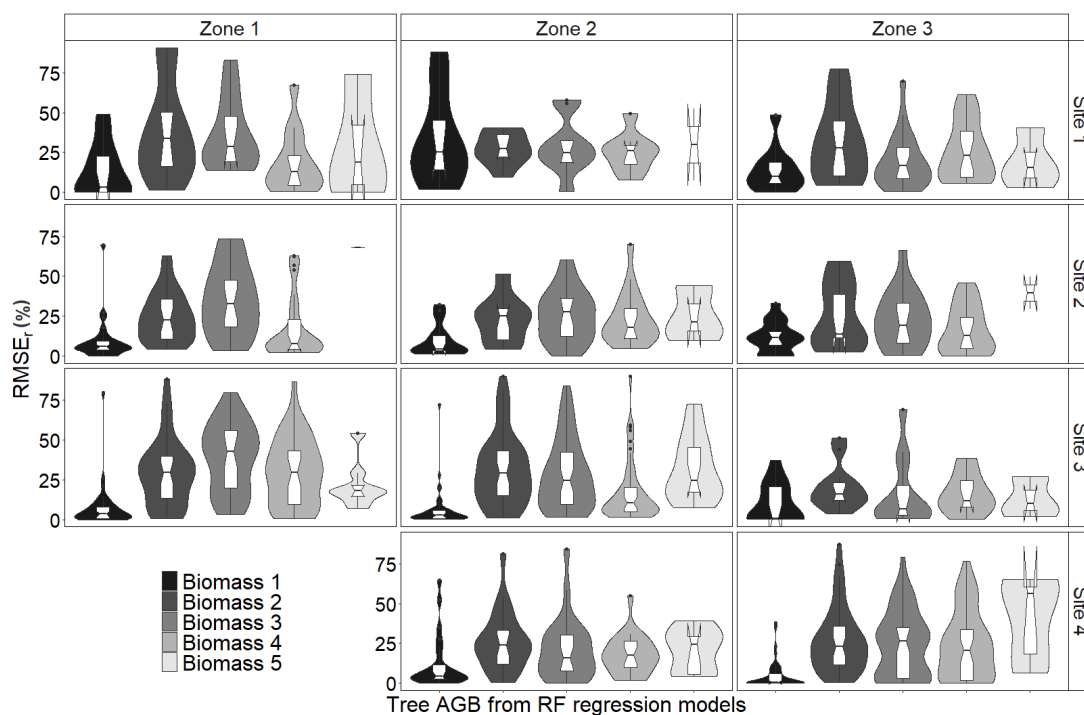


**Figure A8.** Site-specific AGB values obtained from allometric equations on variables measured by UAV (vertical axis) and reference values (horizontal axis) in coppice stands in Zone 2.

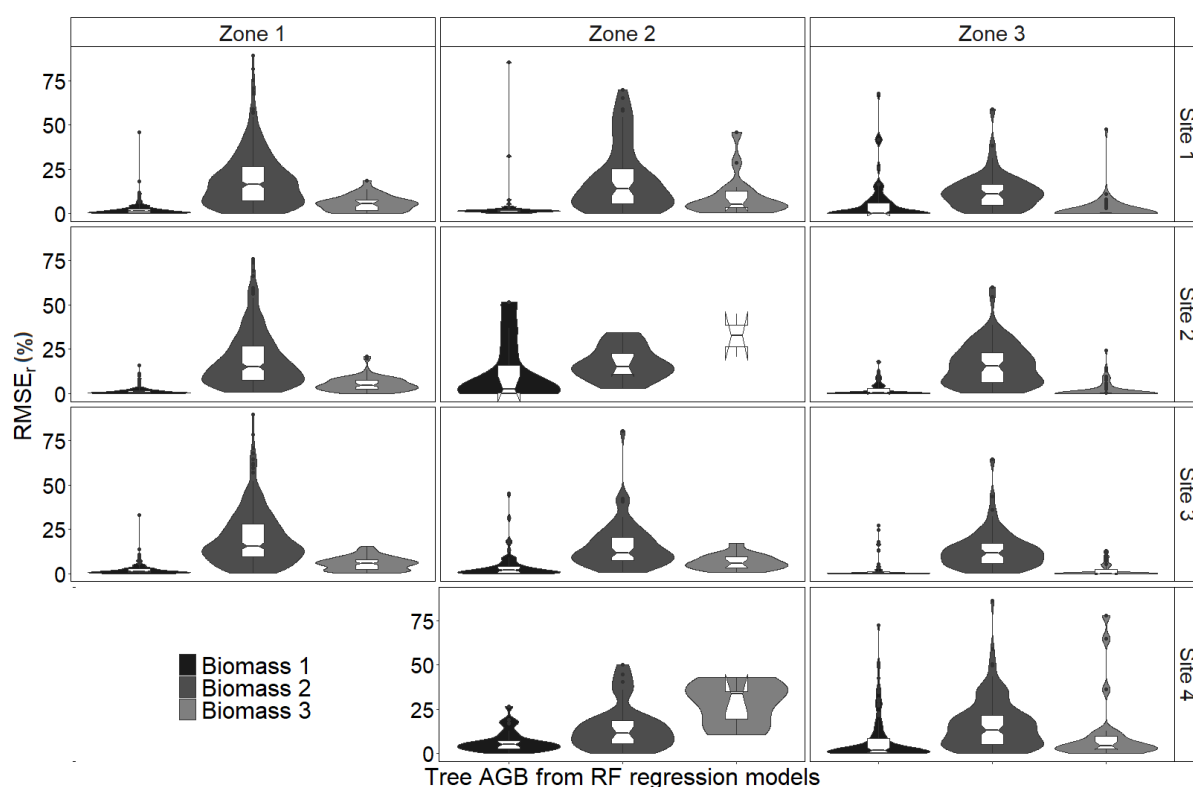


**Figure A9.** Site-specific AGB values obtained from allometric equations on variables measured by UAV (vertical axis) and reference values (horizontal axis) in coppice stands in Zone 3.

## Appendix D



**Figure A10.** Site-specific relative RSME values of RFR models for AGB estimation in high stands.



**Figure A11.** Site-specific relative RSME values of RFR models for AGB estimation in coppice stands.

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