

Article

Carbon Stocks across a Fifty Year Chronosequence of Rubber Plantations in Tropical China

Chenggang Liu ^{1,†}, Jiaping Pang ^{1,2,†}, Martin Rudbeck Jepsen ³, Xiaotao Lü ⁴ and Jianwei Tang ^{1,*}

¹ Key Laboratory of Tropical Plant Resources and Sustainable Use, Xishuangbanna Tropical Botanical Garden, Chinese Academy of Sciences, Menglun 666303, China; chenggangliu@hotmail.com (C.L.); pjp_0725@163.com (J.P.)

² Institute of Geographic Sciences and Natural Resources Research, Chinese Academy of Sciences, Beijing 100101, China

³ Department of Geosciences and Natural Resource Management, University of Copenhagen, Øster Voldgade 10, DK-1350 Copenhagen, Denmark; mrj@ign.ku.dk

⁴ State Key Laboratory of Forest and Soil Ecology, Institute of Applied Ecology, Chinese Academy of Sciences, Shenyang 110164, China; lvxiaotao@gmail.com

* Correspondence: tangjw@xtbg.org.cn; Tel.: +86-0691-8715080

† These authors contributed equally to this work.

Academic Editor: Mark E. Harmon

Received: 26 April 2017; Accepted: 7 June 2017; Published: 13 June 2017

Abstract: Transition from forest to rubber (*Hevea brasiliensis* Muell. Arg.) plantation has occurred in tropical China for decades. Rubber has been planted on 1 million ha to provide raw materials to the rubber industry. The role of various-aged rubber plantations in carbon (C) sequestration remains unclear. The biomass C accumulation including latex C and C distribution in soil of five different-aged stands (7, 13, 19, 25 and 47 years old) were examined. The total biomass C stock (TBC) and total net primary productivity (NPP_{total}), whether with or without latex C, had a close quadratic relationship with stand age. Regardless of stand age, around 68% of the C was stored in aboveground biomass, and NPP_{latex} contributed to approximately 18% of C sequestration. Soil organic carbon stock in the 100-cm depth remained relatively stable, but it lost about 16.8 Mg ha⁻¹ with stand age. The total ecosystem C stock (TEC) across stands averaged 159.6, 174.4, 229.6, 238.1 and 291.9 Mg ha⁻¹, respectively, of which more than 45% was stored in the soil. However, biomass would become the major C sink rather than soil over a maximal rubber life expectancy. Regression analysis showed that TEC for rubber plantation at 22 years is comparable to a baseline of 230.4 Mg ha⁻¹ for tropical forest in China, and would reach the maximum value at around 54 years. Therefore, rubber plantation can be considered as alternative land use without affecting net forest ecosystem C storage. In addition to the potential C gains, a full set of ecosystem and economic properties have to be quantified in order to assess the trade-offs associated with forest-to-rubber transition.

Keywords: latex yield; net primary productivity; quadratic regression model; chronosequence; soil organic carbon

1. Introduction

Increasing atmospheric carbon dioxide (CO₂) concentrations and global climate change are of growing concern to humankind. The world's forests are a critical component of the terrestrial ecosystem and play a significant role in regulating the global carbon (C) cycle by serving as C sinks, thereby potentially mitigating future impacts of climate change [1]. Tropical forests represent about half of the global forest area and are believed to be the largest C reservoir of terrestrial biota [2]. They hold approximately 470 Pg C in live biomass, debris, and soil organic matter, which is approximately 55% of the total C stored in the world's forests [3]. Large areas of

naturally-regenerated tropical forests have been lost due to land use changes, which resulted in the emission of around 1.4 Pg C year⁻¹ to the atmosphere during 1990–2010 (~15% of global anthropogenic CO₂ emissions for that time period) [4]. In order to combat global warming and reduce CO₂ emissions from land-use transition, especially due to deforestation and forest degradation, accurate and reliable estimates of C sequestration in various forest ecosystems are needed [5].

Rubber tree (*Hevea brasiliensis* Muell. Arg.) is a cash crop that provides a variety of raw materials to industrial products. During the past decades, state-owned farms and particularly smallholder farmers have gained unprecedented wealth in southern China and elsewhere due to the rising global rubber prices and the sustainability of government subsidies, e.g., the latex return reached approximately 36,000 CNY ha⁻¹ or 13,000 CNY per person in Xishuangbanna (Yunnan Province) in 2010 [6]. Accordingly, large areas of forests have been converted into rubber plantations in the tropics, resulting in major alterations in ecosystem C dynamics due to deforestation and soil erosion [7,8]. Rubber plantations presently cover an area of 10 million ha in southeast Asia, out of which 1 million ha is in tropical China. The rubber-based income from these plantations may be decreased, however, due to the risk of falling latex prices, although it can be increased by the sell of stumpage when rubber trees have reached their latex-tapping lifespan [6]. To evaluate the trade-offs between ecological and economic functions, rubber transitions have been the subject of considerable study [5,9,10]). Rubber plantation C stocks in biomass and soil have been studied [11,12], but information on the C sequestration including its latex C is still lacking, particularly in tropical China. Given the expansion area and the potential expansion rates in the future, a deeper understanding of C sequestration in rubber plantation ecosystems is needed on the part of policymakers, managers, researchers and other groups in the scientific community.

C sequestration in forest ecosystems is affected by stand age, topography, climate, soil type, tree species, and management regime [13,14]. It has been known that the C stored in tree biomass increases with stand age, but the trend may be different for the mineral soil C storage [11,15,16]. Several authors have revealed enhanced biomass C in rubber plantations with stand age, ranging from 1.4 to 6.7 Mg ha⁻¹ year⁻¹ [17–19], but these results excluded an important C component that is stored in latex. Meanwhile, whether net primary productivity (NPP) is related to stand age is still not clear, because Kotowska et al. [10], for example, observed no obvious correlation between total biomass C and NPP when including latex yield. On the other hand, soil C pool might be independent of rubber age sequence. For example, in one study, soil organic C (SOC) stock under rubber plantation increased 1.13 Mg ha⁻¹ year⁻¹ over 11 years [20], whereas in another study, a forest-to-rubber transition led to SOC stock loss by an average of 37.4 Mg ha⁻¹ in 120 cm depth over 46 years [9].

Therefore, the objectives of this study were to: (i) quantify above- and belowground biomass C as well as latex C in five different-aged (7-, 13-, 19-, 25- and 47-year-old) rubber plantations; and (ii) relate SOC storage with increasing stand age. We hypothesized that (i) rubber biomass C and its NPP including latex yield has a close quadratic relationship with stand age; and (ii) SOC stock will not be affected by stand age.

2. Materials and Methods

2.1. Study Site and Plot Establishment

The study site is located in Mengla County of Xishuangbanna, Yunnan Province, southwest China (21°09′–22°23′ N, 101°05′–101°50′ E), which borders Myanmar and Laos at the source of the Mekong River. The region has a typical tropical monsoon climate with a rainy season (May–October) and a dry season (November–April). The mean annual rainfall is 1550 mm, out of which 83% occurs in the rainy season. Mean annual evapotranspiration is 1200 mm. Average relative air humidity ranges from 76% to 89%. Mean annual air temperature is 21.5 °C. Temperatures often exceed 38 °C during March and April when the relative humidity is below 40%. The average annual number of sunshine hours is 1800 h with 170 foggy days a year.

Five adjacent rubber plantations with different ages were selected based on similar topography, management practices, and previous vegetation composition (dominated by *Millettia* sp., *Syzygium* sp., *Castanopsis indica* and *Phoebe lanceolata*) in January 2009. The rotation period for rubber plantation is about 40 years in this region due to high tree mortality caused by pests and disease [21]. We identified a chronosequence of rubber stands that were 7-, 13-, 19-, 25-, and 47-years of age. These different-aged stands were first tapped for latex when the trees were 7 years old. All five stands were located within an approximately 10-km radius of each other. Three sampling plots (25 × 40 m) were randomly established in each stand. Within each plot, diameter at breast height (1.3 m, DBH) and height for individual trees were measured using a diameter tape and a Hagl f Vertex clinometer, respectively. Topographical features of each plot, including elevation, slope and aspect, were recorded (Table 1).

Table 1. Stand and soil characteristics of the 7-, 13-, 19-, 25- and 47-year-old rubber plantations at Xishuangbanna, Yunnan Province, China.

Stand Parameter	7-year-old	13-year-old	19-year-old	25-year-old	47-year-old
Longitude	101°14'45.3"	101°14'40.1"	101°14'43.5"	101°16'28.9"	101°16'03.5"
Latitude	21°56'48.8"	21°56'49.6"	21°56'34.8"	21°54'36.8"	21°55'22.9"
Elevation (m)	580–595	600–620	560–570	570–580	550–570
Slope (°)	30–35	25–30	25–30	15–20	0–5
Aspect	Southwest	Southeast	Southeast	Southeast	Southwest
Planting spacing (m)	2.5 × 6	2.5 × 6	2.5 × 8	2.5 × 8	2.5 × 8
Average DBH (cm)	9.3 ± 0.7 ^d	15.5 ± 0.8 ^c	21.4 ± 1.2 ^b	24.9 ± 0.7 ^b	40.4 ± 2.2 ^a
Average height (m)	7.3 ± 0.9 ^c	13.9 ± 1.0 ^b	16.3 ± 1.2 ^{ab}	17.2 ± 0.8 ^a	21.3 ± 1.1 ^a
Density (plants ha ^{−1})	667 ± 31 ^a	637 ± 26 ^a	626 ± 21 ^a	510 ± 23 ^{ab}	203 ± 17 ^b
Basal area (m ² ha ^{−1})	5.1 ± 0.8 ^c	11.3 ± 1.0 ^b	21.9 ± 1.8 ^{ab}	24.6 ± 0.7 ^a	29.5 ± 0.9 ^a
pH	6.5 ± 0.2 ^a	6.4 ± 0.1 ^a	6.0 ± 0.1 ^a	5.7 ± 0.1 ^{ab}	5.4 ± 0.1 ^b
Clay (%)	30.3 ± 3.5 ^a	28.5 ± 2.3 ^a	32.2 ± 2.0 ^a	29.1 ± 3.4 ^a	26.9 ± 2.9 ^a
Silt (%)	34.5 ± 2.2 ^a	33.0 ± 3.7 ^a	33.8 ± 2.7 ^a	31.5 ± 0.5 ^a	28.0 ± 0.7 ^b
Sand (%)	35.2 ± 0.4 ^b	38.5 ± 1.1 ^{ab}	34.0 ± 0.2 ^b	39.4 ± 0.9 ^a	45.1 ± 1.8 ^a

Note: Values are means ± Standard Error (S.E.). Means followed by different lowercase letters within a row indicate significant differences among stand ages according to Tukey's Honest Significant Difference (HSD) test ($P < 0.05$).

2.2. Field Sampling and Measurements

On the basis of the DBH and height measurements in the sampling plots, we harvested six standard trees from different diameter classes in each stand for biomass measurements (30 trees in total). The aboveground portions of the standard trees were divided into 5-cm sections for measurement. We measured the fresh weights of stems, branches and foliage in situ. The belowground portions were obtained by total excavation of the standard trees, extending radially out from the trunk and downwards to bedrock until no more roots were visible. We measured the fresh weights of root collars, stump roots, thick roots (diameter (D) ≥ 2 cm), small roots ($0.5 \leq D < 2$ cm), and fine roots ($D < 0.5$ cm) in situ. Samples (500 g) of each compartment in each standard tree were collected, and then oven-dried at 70 °C to a constant weight to calculate the ratio of fresh weight to dry biomass. We built regression models for the different compartments to evaluate tree biomass ($n = 30$), and the models can be obtained in our previous work [22].

Latex was collected three times a month from the onset of tapping in May 2009 and extended about 7 months. The latex production was recorded through weighing of the fresh material separately for all trees of each plot. The dry weight was measured after oven-drying representative subsamples of latex (six collecting bottles) at 70 °C to a constant weight. Meanwhile, monthly tapping times were also recorded. Then, a model ($y = -0.002x^2 + 0.089x + 0.584$; $P < 0.05$) was fitted using annual yield (y , t ha^{−1} year^{−1}) and stand age (x). The total production during the whole tapping phase was calculated by accumulating consecutive annual yield.

Litter, including leaves, twigs ($D < 2$ cm), flowers/fruits, and miscellaneous litter compartments, were collected every month between January 2009 and December 2009 using 10 litter traps (1 × 1 m)

that were randomly distributed in each sampling plot. Litter samples were oven-dried at 70 °C to a constant weight to determine the ratio of fresh weight to dry biomass. Plant biomass in the shrub and herbaceous layers was negligible or practically missing due to regular removal by local farmers, except for their litter, and has thus not been investigated.

Three soil profiles in each plot were dug to a depth of 100 cm, and the samples were taken from six depths (0–10, 10–20, 20–40, 40–60, 60–80 and 80–100 cm) using a soil corer (5 cm inner diameter). Soil samples from the same depth in the same plot were mixed in equal volume proportions and air-dried at room temperature. Bulk density for each soil depth was determined by collecting samples from a stainless steel cylinder (100 cm³) and oven drying the soil core at 105 °C for constant weight. Bulk density was calculated by dividing the mass of oven-dried soil by the volume of the core. Three soil samples were taken from every depth.

For analysis of C concentration, biomass samples were ground and passed through a 1 mm sieve. Mineral soil samples were sieved through a 0.149 mm sieve before chemical analysis. The C concentrations of tree, latex and soil samples were measured by a dry combustion method with a vario MAX CNS elemental analyzer (Elementar, Langenselbold, Germany). The tree, latex and litter C stocks (Mg ha⁻¹) were calculated by multiplying C concentration by dry biomass (t ha⁻¹). The tree NPP (t ha⁻¹ year⁻¹) was calculated by dividing total C stock by the corresponding stand age. The latex and litter NPP was estimated by total collection amount during a year. The SOC stock (Mg ha⁻¹) was calculated by multiplying C concentration by bulk density and thickness of the soil layer with a correction for stone content.

2.3. Statistical Analysis

One-way ANOVA following Tukey's Honest Significant Difference (HSD) test was used to test the differences in biomass production, and C stocks in different compartments affected by stand age. A quadratic polynomial regression was applied to check the relationship between C stocks in different compartments and stand age. The software SPSS 13.0 (SPSS Inc., Chicago, IL, USA) was employed for all statistical analyses at a significance level of 0.05.

3. Results

3.1. Biomass C Concentration and Stock

The C concentration varied substantially among different compartments of rubber trees (Figure 1). The C concentration of all tree compartments in the 19-year-old stand was lower in range than in other stands (although not significantly) (Figure 1a). Across all age groups, foliage had the highest C concentration ($51.2 \pm 0.4\%$), whereas roots had the lowest value ($46.7 \pm 0.4\%$). The latex C concentration ($84.7 \pm 0.8\%$) was not affected by stand age (Figure 1b). A significant difference in litter C concentration was only observed between the 7- and 13-year-old stands (Figure 1c). For root compartments, average C concentration declined in the following order: root collar > stump root > thick root > small root > fine root (Figure 1d).

The total biomass (TB) including above- (AGB) and belowground (BGB) and latex yield increased with stand age (Table 2). AGB was the largest compartment of C storage in biomass ($68.3 \pm 1.0\%$), followed by latex yield ($17.3 \pm 2.7\%$) and BGB ($14.3 \pm 2.0\%$) across all age groups. A quadratic regression analysis showed that the proportion of aboveground biomass C (AGBC)/total biomass C (TBC) was not related to stand age ($R^2 = 0.10$, $P = 0.35$), but a high correlation in latex C/TBC ($R^2 = 0.62$, $P = 0.003$), belowground biomass C (BGBC)/TBC ($R^2 = 0.93$, $P < 0.001$), and BGBC/AGBC ($R^2 = 0.97$, $P < 0.001$) could be observed, respectively.

3.2. Biomass NPP and Associated C Sequestration

The biomass NPP and associated C sequestration of various compartments increased first and then decreased with stand age, with the higher values between the 19- and 25-year-old stands, except NPP_{litter} (Table 3). Across all age groups, more than 75% of C was sequestered in aboveground tree and litter. Moreover, around 18% of NPP was attributed to latex yield, whereas a

smaller percentage of NPP was allocated to root. NPP_{total} , whether including latex C or not, showed a close quadratic relationship with stand age ($R^2 = 0.80$, $P < 0.001$; or $R^2 = 0.79$, $P < 0.001$).

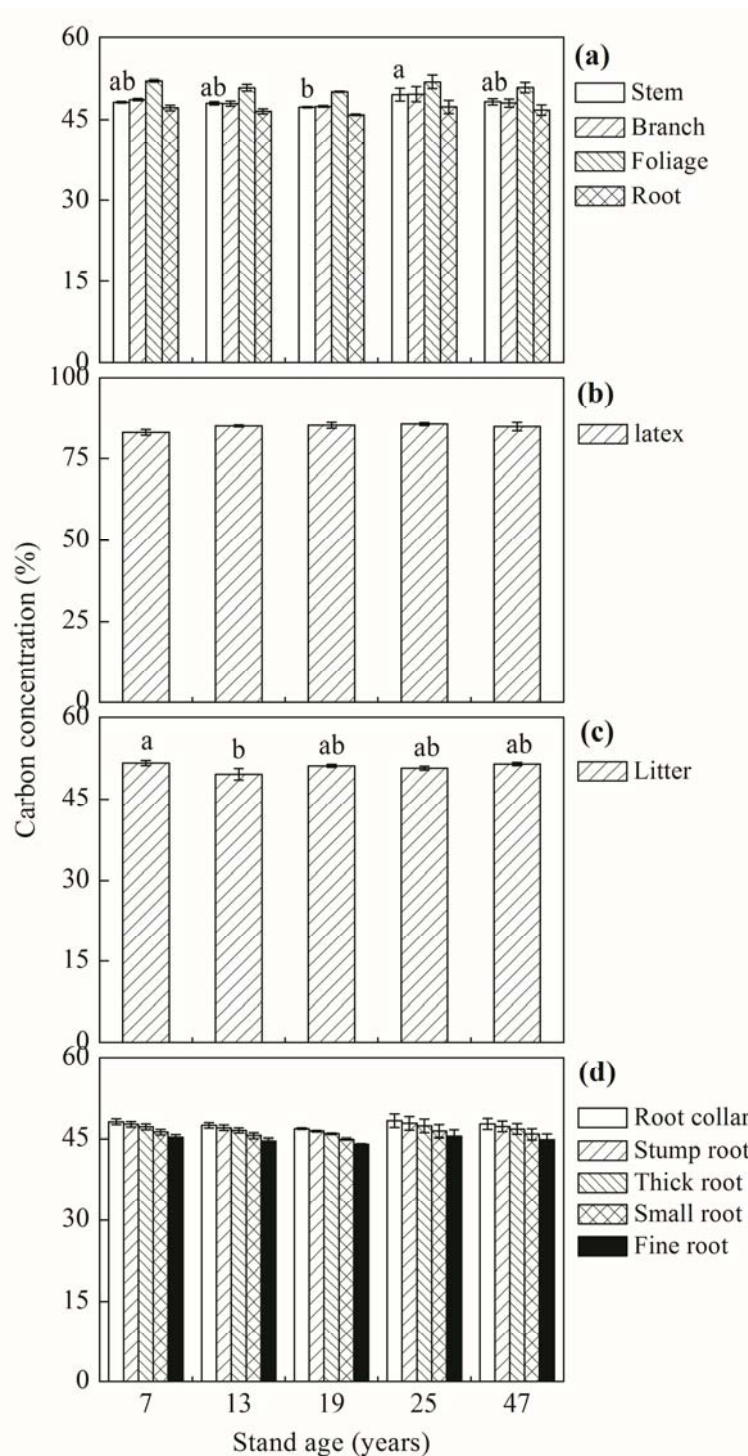


Figure 1. Carbon concentrations (as % of biomass) in different biomass compartments: tree (a), latex (b), litter (c) and specific root (d), of the 7-, 13-, 19-, 25- and 47-year-old rubber plantations at Xishuangbanna, Yunnan Province, China. Presence of different lowercase letters indicates significant differences among means of stand ages according to Tukey's HSD test ($P < 0.05$). Vertical bars show $\pm$ S.E.

Table 2. Above- and belowground biomass and latex yield (t ha^{-1}) and carbon stock (in brackets Mg C ha^{-1}) in different compartments of the 7-, 13-, 19-, 25- and 47-year-old rubber plantations at Xishuangbanna, Yunnan Province, China.

Biomass Fraction	7-year-old	13-year-old	19-year-old	25-year-old	47-year-old
Aboveground tree					
Stem	12.86 ± 1.57 (6.20) ^c	38.66 ± 1.88 (18.56) ^c	92.17 ± 6.77 (43.57) ^b	106.56 ± 9.52 (52.88) ^b	169.50 ± 16.19 (81.84) ^a
Branch	3.63 ± 0.44 (1.77) ^c	10.74 ± 0.53 (5.15) ^c	25.29 ± 1.86 (12.00) ^b	29.15 ± 2.60 (14.48) ^b	45.54 ± 4.32 (21.87) ^a
Foliage	0.75 ± 0.09 (0.39) ^c	1.87 ± 0.10 (0.95) ^c	3.88 ± 0.30 (1.94) ^b	4.32 ± 0.37 (2.24) ^{ab}	5.62 ± 0.51 (2.86) ^a
Litter					
Leaf	2.50 ± 0.09 (1.29) ^b	3.82 ± 0.20 (1.90) ^a	4.24 ± 0.30 (2.17) ^a	3.77 ± 0.06 (1.92) ^a	4.12 ± 0.05 (2.12) ^a
Twig	0.02 ± 0.00 (0.01) ^c	0.40 ± 0.08 (0.20) ^b	0.59 ± 0.14 (0.30) ^{ab}	0.82 ± 0.02 (0.42) ^a	0.40 ± 0.04 (0.21) ^b
Flower/fruit	0.10 ± 0.01 (0.06) ^c	0.35 ± 0.07 (0.18) ^b	0.26 ± 0.07 (0.14) ^b	0.92 ± 0.01 (0.47) ^a	0.76 ± 0.10 (0.39) ^a
Miscellaneous litter	0.12 ± 0.01 (0.06) ^c	0.15 ± 0.01 (0.08) ^c	0.16 ± 0.01 (0.08) ^{bc}	0.23 ± 0.01 (0.12) ^b	0.42 ± 0.03 (0.22) ^a
Total litter	2.74 ± 0.12 (1.42) ^b	4.72 ± 0.36 (2.35) ^a	5.25 ± 0.53 (2.69) ^a	5.74 ± 0.08 (2.93) ^a	5.70 ± 0.21 (2.94) ^a
Total AGB	19.98 ± 2.11 (9.78) ^c	55.99 ± 2.87 (27.01) ^c	126.59 ± 9.46 (60.20) ^b	145.77 ± 12.57 (72.53) ^b	226.36 ± 21.23 (109.51) ^a
Root					
Root collar	2.41 ± 0.26 (1.16) ^c	6.13 ± 0.38 (2.91) ^b	7.21 ± 0.61 (3.38) ^{ab}	9.29 ± 0.79 (4.49) ^a	9.59 ± 0.83 (4.57) ^a
Stump root	2.34 ± 0.25 (1.11) ^c	4.35 ± 0.27 (2.04) ^b	6.20 ± 0.52 (2.87) ^a	4.59 ± 0.39 (2.19) ^b	4.42 ± 0.38 (2.09) ^b
Thick root	0.46 ± 0.05 (0.22) ^c	1.32 ± 0.08 (0.62) ^c	5.15 ± 0.43 (2.37) ^b	9.41 ± 0.80 (4.45) ^a	8.17 ± 0.71 (3.82) ^a
Small root	0.44 ± 0.05 (0.21) ^c	0.99 ± 0.06 (0.45) ^c	4.43 ± 0.37 (1.99) ^a	2.54 ± 0.22 (1.18) ^b	5.58 ± 0.48 (2.56) ^a
Fine root	0.47 ± 0.05 (0.21) ^c	0.75 ± 0.05 (0.33) ^c	2.46 ± 0.21 (1.08) ^b	1.69 ± 0.14 (0.77) ^b	3.42 ± 0.30 (1.54) ^a
Total BGB	6.12 ± 0.66 (2.91) ^c	13.53 ± 0.84 (6.35) ^b	25.45 ± 2.14 (11.69) ^a	27.52 ± 2.34 (13.08) ^a	31.19 ± 2.69 (14.58) ^a
BGB/AGB	0.31 ± 0.00 (0.30) ^a	0.24 ± 0.00 (0.24) ^b	0.20 ± 0.00 (0.19) ^c	0.19 ± 0.00 (0.18) ^d	0.14 ± 0.00 (0.13) ^e
Latex	1.10 ± 0.00 (0.91) ^e	8.74 ± 0.10 (7.43) ^d	17.51 ± 0.28 (14.92) ^c	26.47 ± 0.50 (22.66) ^b	46.15 ± 0.84 (38.63) ^a
TB	27.20 ± 2.88 (13.61) ^c	78.26 ± 3.72 (40.79) ^c	169.55 ± 11.62 (86.82) ^b	199.76 ± 14.93 (108.27) ^b	303.70 ± 23.71 (162.72) ^a

Note: Values are means $\pm$ S.E. Means followed by different lowercase letters in the same row indicate significant differences among stand ages according to Tukey's HSD test ($P < 0.05$). AGB: aboveground biomass; BGB: belowground biomass; TB: total biomass.

Table 3. Net primary productivity (NPP) ($\text{t ha}^{-1} \text{ year}^{-1}$) and carbon stock (in brackets $\text{Mg C ha}^{-1} \text{ year}^{-1}$) in different compartments of the 7-, 13-, 19-, 25- and 47-year-old rubber plantations at Xishuangbanna, Yunnan Province, China.

NPP Fraction	7-year-old	13-year-old	19-year-old	25-year-old	47-year-old
NPP _{aboveground tree}	2.46 ± 0.30 (1.19) ^c	3.94 ± 0.19 (1.90) ^{b,c}	6.39 ± 0.47 (3.03) ^a	5.60 ± 0.50 (2.78) ^{a,b}	4.69 ± 0.45 (2.27) ^{a,b}
NPP _{litter}	2.74 ± 0.12 (1.42) ^b	4.72 ± 0.36 (2.35) ^a	5.25 ± 0.53 (2.69) ^a	5.74 ± 0.08 (2.93) ^a	5.70 ± 0.21 (2.94) ^a
NPP _{root}	0.87 ± 0.09 (0.42) ^{b,c}	1.04 ± 0.06 (0.49) ^{a,b}	1.33 ± 0.11 (0.62) ^a	1.10 ± 0.09 (0.52) ^{a,b}	0.66 ± 0.05 (0.31) ^c
NPP _{latex}	1.10 ± 0.00 (0.91) ^c	1.25 ± 0.01 (1.06) ^b	1.35 ± 0.02 (1.14) ^a	1.39 ± 0.03 (1.19) ^a	1.12 ± 0.02 (0.94) ^c
NPP _{total}	7.17 ± 0.51 (3.94) ^b	10.95 ± 0.62 (5.79) ^a	14.32 ± 1.11 (7.48) ^a	13.83 ± 0.67 (7.42) ^a	12.18 ± 0.71 (6.46) ^a

Note: Values are means $\pm$ S.E. Means followed by different lowercase letters in the same row indicate significant differences among stand ages according to Tukey's HSD test ($P < 0.05$).

3.3. SOC Concentration and Stock

The SOC concentration decreased from 1.2% to 1.0% in 100 cm depth with stand age. Regardless of stand age, the average SOC concentration declined significantly with depth, ranging from 1.6% at 0–20 cm to 0.6% at 80–100 cm. By contrast, bulk density increased from 1.32 g cm⁻³ to 1.41 g cm⁻³ with stand age. Irrespective of stand age, the average bulk density increased with depth, ranging from 1.31 g cm⁻³ at 0–20 cm to 1.39 g cm⁻³ at 80–100 cm.

The SOC stock in 100 cm depth slightly decreased from 146.0 Mg ha⁻¹ in the 7-year-old stand to 129.1 Mg ha⁻¹ in the 47-year-old stand (Figure 2). Across all age groups, SOC stock decreased rapidly with depth, ranging from 41.9 Mg ha⁻¹ at 0–20 cm to 17.5 Mg ha⁻¹ at 80–100 cm. With the exception of the 7-year-old stand, more than 52.5% of SOC was accumulated within the 40 cm depth for each stand.

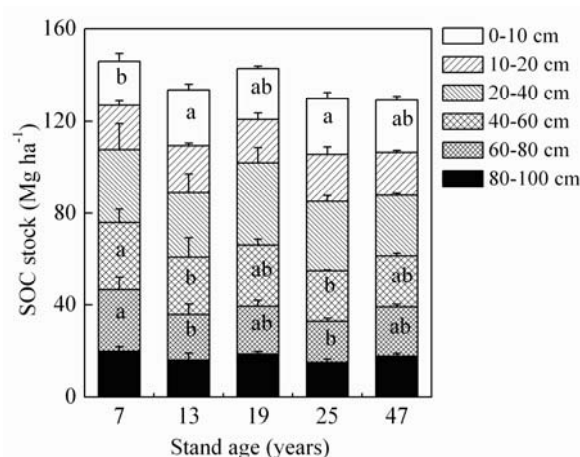


Figure 2. Soil organic carbon (SOC) stock in different soil layers of the 7-, 13-, 19-, 25- and 47-year-old rubber plantations at Xishuangbanna, Yunnan Province, China. Presence of different lowercase letters across a respective soil layer indicates significant differences among means of stand ages according to Tukey's HSD test ($P < 0.05$). Vertical bars show $\pm$ S.E.

3.4. Ecosystem C Stock and Its Relationship between Main Compartments

The total ecosystem C stock (TEC) increased with stand age, ranging from 159.6 Mg ha⁻¹ in the 7-year-old stand to 291.9 Mg ha⁻¹ in the 47-year-old stand. The contribution from TBC to TEC increased from 8.6% for the 7-year-old stand to 55.1% for the 47-year-old stand; therein, latex C contributed 0.6–13.4% to TEC with stand age (Figure 3). Most of TEC was occupied by SOC rather than TBC. Indeed, 91.4% of TEC in the 7-year-old stand was sequestered in 100 cm soil depth. This proportion decreased significantly with stand age, and fell to 44.9% in the 47-year-old stand. SOC stock was not correlated with age classes ($P = 0.30$) (Table 4).

A quadratic regression analysis showed that C stored in main biomass compartments was highly correlated with stand age ($P < 0.001$), whereas SOC stock was not correlated with age classes ($P = 0.30$). The TEC, whether excluding latex C or not, also exhibited a close correlation with stand age ($P < 0.001$). The quadratic model predicted that the maximum TEC will occur at around 54 years, which is much larger than comprehensive prediction (38 years) using previous data excluding latex C.

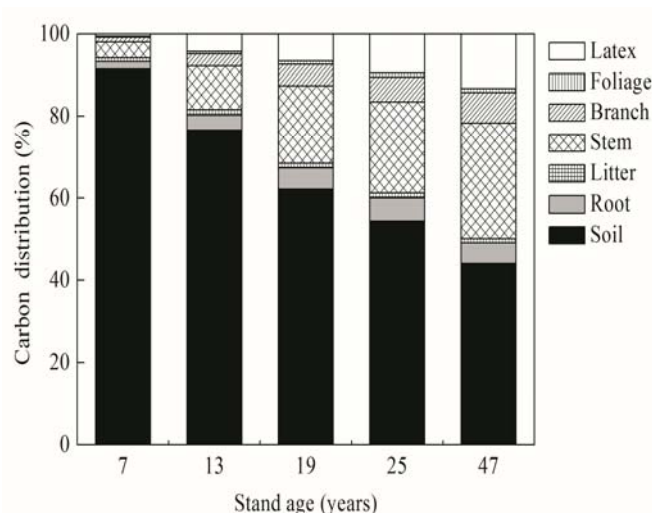


Figure 3. Percentage distribution of carbon stock in various compartments of rubber plantation ecosystem of five different ages.

Table 4. The relationship between carbon stored in different compartments of rubber plantation ecosystem and stand age.

Compartments	Regression Equation	R^2	SEE	F-Value	P
Aboveground tree	$y = -0.049x^2 + 5.150x - 27.724$	0.802	19.028	24.398	< 0.001
Litter	$y = -0.002x^2 + 0.154x + 0.547$	0.841	0.271	31.703	< 0.001
Root	$y = -0.014x^2 + 1.034x - 3.934$	0.711	3.105	14.746	< 0.001
Latex	$y = -0.011x^2 + 1.533x - 9.878$	0.995	1.090	1076.113	< 0.001
Soil	$y = 0.012x^2 - 1.015x + 150.766$	0.181	12.590	1.329	0.300
Total excluding latex C (our study)	$y = -0.053x^2 + 5.323x + 119.655$	0.774	20.412	20.529	< 0.001
Total excluding latex C (comprehensive) *	$y = -0.098x^2 + 7.508x + 104.044$	0.328	58.128	6.349	0.006
Total including latex C (our study)	$y = -0.063x^2 + 6.856x + 109.78$	0.869	20.343	39.612	< 0.001

* Data of total ecosystem C stock (comprehensive) were collected from previous studies irrespective of management practices, soil texture and environmental conditions [17–21,23–34]. Note: a, b, R^2 , SEE denotes slope, intercept, coefficient of determination and model's standard error, respectively.

4. Discussion

4.1. Biomass C Accumulation

Accurate estimates of C stocks in different development stages of a plantation are important for assessing their role in regional and global C management. The C concentration varies with type of tree compartments (stems, branches, foliage and roots), type of wood (normal, tension or compression), soil conditions, geographical location, and climate [35]. Thus, the compartment-specific C concentration is often used when converting biomass to C stock instead of a standard factor of 50% [36]. The average C concentration of rubber tree (48.4%) (Figure 1) was higher than that of taller plants (43.6%) [37]. The use of 50% therefore could overestimate biomass C stock by as much as 4% in this region. Moreover, the lowest C concentration in the 19-year-old tree is attributed to a highest increment rate, so maybe the fast growth rate will result in less dense biomass. We suggest that a general C concentration of 48% can be applied as an alternative to empirical assessments of site-specific values.

The TBC including AGBC, BGBC, and latex C of rubber plantation increased and showed a quadratic trend with stand age (Table 2). The BGBC/AGBC ratio decreased with stand age, which is different from another study reporting larger C accumulation in roots over a long time period

(105-year-old trees) [38]. The highest C increment was $6.4 \text{ Mg ha}^{-1} \text{ year}^{-1}$ between the 13–19 years stands (Table 2), which is associated with the fast growth rates. Rubber plantations grown in potentially dissimilar conditions in Brazil [18,20] and Asia [17,19,23] have revealed comparable C increments, varying from 1.4 to $6.7 \text{ Mg ha}^{-1} \text{ year}^{-1}$, with the highest value occurring between the 10–20 years stands. Moreover, the largest TBC in the 47-year-old stand was higher than the value of 110.9 Mg ha^{-1} measured for tropical forest in China [39]. These comparisons indicate that rubber plantation can serve as an effective C sink by a rapid rate of biomass C accumulation in the early stages, although the rate will slow down with stand age.

The C stock of fine roots was age-dependent (Table 2). The majority of fine root C was located in the upper 40 cm soil depth (data not shown) due to abundant water and nutrients, which is consistent with recent findings in other tree species [40,41]. Furthermore, litter C stock increased with stand age. This trend is in line with some chronosequence studies of broad-leaved plantations [42,43], but not with some others [13,44], indicating large between-stand variability. The discrepancies result from input and decomposition rates of litter that are highly susceptible to disturbances and spatial variation [36,45]. We also found that the C stock of litter in each stand was greater than that of foliage, suggesting that litter should be considered an essential component in TBC assessment.

Our estimate of $\text{NPP}_{\text{total}}$ for rubber plantation, whether excluding latex C or not, had a close quadratic relationship with stand age. The average $\text{NPP}_{\text{total}}$ amounting to $6.2 \text{ Mg ha}^{-1} \text{ year}^{-1}$ during the 47 years was a little lower than the estimated rates ($\sim 7.8 \text{ Mg ha}^{-1} \text{ year}^{-1}$) reported in previous studies [10,46]. This discrepancy is mainly attributed to single age measurement (e.g., 16- or 33-year-old) in their studies. However, a rubber plantation can be considered to have a considerable C sequestration capacity, although its estimated NPP is lower than the global NPP of $10.9 \text{ Mg ha}^{-1} \text{ year}^{-1}$ measured for tropical forests [47]. Moreover, latex C was more prominent in its contribution to NPP relative to roots (Table 3), which is consistent with the estimate by Kotowska et al. [10], who reported about 20.8–31.3% of NPP was allocated to latex, while a smaller percentage was attributed to leaf and root in rubber monoculture of Indonesia. Eventually, latex C will be stored in industrial products, such as tires. It is worth noting that latex yield measured in a given year or estimated from our model does not accurately reflect the rubber productivity of the whole economic lifespan. Therefore, a continuous monitoring in latex yield is necessary to estimate rubber plantation $\text{NPP}_{\text{total}}$ accurately.

4.2. Soil C Stock

Although soil C generally is expected to decrease due to soil disturbance and aeration in the early decades after plantation establishment [38,44], some studies suggest that SOC stock is neutral, i.e., neither increases nor decreases with stand age [1,16,48]. This phenomenon may be ascribed to numerous factors, such as soil properties, tree species, pre-planting disturbance, and site management, all of which may individually or jointly overshadow the effect of stand age on SOC stock [13,36]. We found no significant correlation between SOC stock and stand age, indicating fast organic C turnover in rubber soils. However, about 22.3 Mg ha^{-1} SOC stock has been lost over a period of 47 years compared with its amount before rubber establishment (151.4 Mg ha^{-1}). Similarly, de Blécourt et al. [9] found that forest-to-rubber transition resulted in SOC stock loss of 120 cm (37.4 Mg ha^{-1}) over a period of 46 years, which was equal to 19.3% of the initial SOC stock in the secondary forests. By contrast, Maggionto et al. [20] found a small, but significant, increase in SOC stock (12.4 Mg ha^{-1}) over 11 years of rubber plantation. Over the long run, however, SOC stock trajectory with stand age will tend to be relatively stable and remains an essential part of the C budget of the overall plantation ecosystem.

In addition, SOC in the topsoil did not increase with stand age, whereas its stock estimated using a modelling method showed a positive trend [48]. Approximately 48.0–57.8% of SOC was found in the top 40 cm soil in all stands (Figure 2), despite the topsoil being vulnerable to human disturbance and natural erosion [16]. Thus, focus on the vertical variability of SOC stock, and protection of the topsoil from loss is essential for C sequestration [14]. The average SOC stock of

136.0 Mg ha⁻¹ in 100 cm depth (Figure 2) was similar to the mean value of 146.2 Mg ha⁻¹ obtained from different-aged rubber plantations [18,23], but higher than that (89.5–96.9 Mg ha⁻¹) in Thailand [19] and Hainan Island [29]. These results indicate that SOC stock is largely determined by the site-specific C concentration, and the soil with the low initial concentration will easily increase C sink following transition to rubber, and vice versa [8]. Moreover, SOC stock that we observed was higher than the 116.5 Mg ha⁻¹ measured for tropical forest soil in China [39]. Thus, we speculate that organic matter decomposition and subsequent accumulation in soil C in our study region happens faster than that in the whole Chinese tropical forest due to precipitation surplus or intensive management.

4.3. Ecosystem C Stock

The TEC increased by nearly 82.9% with stand age and would reach maximal value in a 54-year-old stand according to quadratic model including latex C, indicating a considerable C sequestration potential in rubber plantation. By contrast, Nizami et al. [21] predicted rubber plantation to be a smaller C sink (186.7 Mg ha⁻¹) in a 40-year rotation using the CO₂FIX Model. Another important finding was that SOC stock was relatively stable over age sequence, whereas its contribution to TEC decreased rapidly with stand age (Figure 3), similar to the results from *Pinus strobus* [36] and *Cyclobalanopsis glauca* [41] plantations. This trend did not occur in some other plantations due to the differences in the rates of biomass and soil C accumulation among tree species during stand development [40,44]. Uri et al. [13] found that AGBC will ultimately surpass SOC stock of the older stand. We suggest that when considering a maximal life expectancy, biomass including latex yield will become the major C stock of rubber plantation. Therefore, further evidence for ecosystem C pool dependence on stand age will be required, in particular from studies with a very long time perspective (>50 years).

4.4. The Implications of Rubber Plantation on C Sequestration

Although forest-to-rubber transition can increase farmer income in the short term, more attention should be paid to the implications for C sequestration of the transition, especially transitions with baselines in the two main forest types: degraded and natural (maybe even undisturbed) forests. Firstly, large areas of degraded forest lands with low C stocks could benefit from the establishment of rubber plantation [5]. For example, biomass C in five different-aged rubber plantations (Table 2) was always higher than that in degraded forest in the same region as reported by Tang et al. [49], indicating a rapid biomass accumulation rate in rubber plantation. Secondly, the consequence of converting natural forest with high C stock, old-growth forest in particular, to rubber has been shown to result in large C debt [9], because rubber plantation C payback time is about 40 years [21,31]. Under these two land-use scenarios, biomass C has been predicted to increase by about 11.8 Tg or drop by about 4.13 Tg in Xishuangbanna during the next 20 years [7]. Moreover, rubber plantations could induce increased C sequestration at around 22 years of growth, according to a quadratic regression model (Table 4), because the stock at that time is projected to exceed the average TEC (230.4 Mg ha⁻¹) for Chinese tropical forest. As a result, there is no doubt that rubber plantation can be considered as alternative land use without affecting forest ecosystem C storage.

There are big drawbacks, however, to the large-scale rubber monoculture in the form of reduction in biodiversity, loss of soil and water, and an increase in vulnerability to diseases and pests [5]. For example, as little as 129.4 Mg C ha⁻¹ in soil is left once a rubber plantation is harvested for rotation (40 years). By contrast, rubber agroforestry systems are found to be environmentally-friendly due to their diversity and complexity. For example, leguminous cover crops planted between rubber trees can substantially increase C stock in addition to improving soil fertility [50]. Further studies on ecosystem C stock of rubber plantations should include not only potential ecological issues associated with monocultures but also comparisons to mixed agroforestry systems in general.

5. Conclusions

In a chronosequence of rubber plantations in tropical China, the TBC and NPP_{total}, whether including latex C or not, had a close quadratic relationship with stand age. Regardless of stand age, around 68% of the C was stored in aboveground biomass, and NPP_{latex} contributed to approximately 18% of C sequestration. However, SOC stock in the 100 cm depth was age-independent, and decreased with soil depth for each stand. The TEC across stands averaged 159.6, 174.4, 229.6, 238.1 and 291.9 Mg ha⁻¹, respectively, of which more than 45% was stored in the soil. When considering a maximal life expectancy of rubber trees, the C accumulated in biomass would exceed soil over age sequence. Regression analysis showed that rubber plantations have a potential role in improving the regional C budget in the long term (achieving maximum at about 54 years), and thus they can be considered as alternative land use without affecting forest ecosystem C storage. We also suggest that further research on ecosystem C pool dependence on stand age in rubber agroforestry systems is needed to fully achieve a “win-win” balance between environmental and economic benefits in tropical China.

Acknowledgments: We thank Yun Lu and Defu Chen for helping with field work and the members in the Biogeochemistry Laboratory of the Xishuangbanna Tropical Botanical Garden for C analysis. This study was supported by the National Natural Science Foundation of China (31600507), the Chinese Academy of Sciences (CAS) “Light of West China” Program, the Natural Science Foundation of Yunnan Province (2016FA047), and the key program of CAS (KFZD-SW-312).

Author Contributions: J.T. and X.L. conceived and designed the experiments; J.P. performed the experiments; C.L. and J.P. analyzed the data; M.J. contributed reagents/materials/analysis tools; C.L. wrote the paper.

Conflicts of Interest: The authors declare no conflict of interest.

References

- Cheng, J.Z.; Lee, X.Q.; Theng, B.K.G.; Zhang, L.K.; Fang, B.; Li, F.S. Biomass accumulation and carbon sequestration in an age-sequence of *Zanthoxylum bungeanum* plantations under the Grain for Green Program in karst regions, Guizhou province. *Agric. For. Meteorol.* **2015**, *203*, 88–95.
- Nogueira, E.M.; Yanai, A.M.; Fonseca, F.O.; Fearnside, P.M. Carbon stock loss from deforestation through 2013 in Brazilian Amazonia. *Glob. Chang. Biol.* **2015**, *21*, 1271–1292.
- Pan, Y.D.; Birdsey, R.A.; Fang, J.Y.; Houghton, R.; Kauppi, P.E.; Kurz, W.A.; Phillips, O.L.; Shvidenko, A.; Lewis, S.L.; Canadell, J.G.; et al. A large and persistent carbon sink in the world’s forests. *Science* **2011**, *333*, 988–993.
- Hansen, M.C.; Potapov, P.V.; Moore, R.; Hancher, M.; Turubanova, S.A.; Tyukavina, A.; Thau, D.; Stehman, S.V.; Goetz, S.J.; Loveland, T.R. High-resolution global maps of 21st-century forest cover change. *Science* **2013**, *342*, 850–853.
- Fox, J.; Castella, J.C.; Ziegler, A.D.; Westley, S.B. Rubber plantations expand in mountainous Southeast Asia: What are the consequences for the environment? *AsiaPacific Issues* **2014**, *114*, 1–8.
- Zhang, L.; Kono, Y.; Kobayashi, S.; Hu, H.B.; Zhou, R.; et al. The expansion of smallholder rubber farming in Xishuangbanna, China: A case study of two Dai villages. *Land Use Policy* **2015**, *42*, 628–634.
- Li, H.M.; Ma, Y.X.; Aide, T.M.; Liu, W.J. Past, present and future land-use in Xishuangbanna, China and the implications for carbon dynamics. *For. Ecol. Manag.* **2008**, *255*, 16–22.
- van Straaten, O.; Corre, M.D.; Wolf, K.; Tchienkoua, M.; Cuellar, E.; Matthews, R.B.; Veldkamp, E. Conversion of lowland tropical forests to tree cash crop plantations loses up to one-half of stored soil organic carbon. *Proc. Natl. Acad. Sci. USA* **2015**, *112*, 9956–9960.
- de Blécourt, M.; Brumme, R.; Xu, J.C.; Corre, M.D.; Veldkamp, E. Soil carbon stocks decrease following conversion of secondary forests to rubber (*Hevea brasiliensis*) plantations. *PloS ONE* **2013**, *8*, e69357.
- Kotowska, M.M.; Leuschner, C.; Triadiati, T.; Meriem, S.; Hertel, D. Quantifying above- and belowground biomass carbon loss with forest conversion in tropical lowlands of Sumatra (Indonesia). *Glob. Chang. Biol.* **2015**, *21*, 3620–3634.
- de Blécourt, M.; Hänsel, V.M.; Brumme, R.; Corre, M.D.; Veldkamp, E. Soil redistribution by terracing alleviates soil organic carbon losses caused by forest conversion to rubber plantation. *For. Ecol. Manag.* **2014**, *313*, 26–33.

12. Kobayashi, N.; Kumagai, T.O.; Miyazawa, Y.; Matsumoto, K.; Tateishi, M.; Lim, T.K.; Mudd, R.G.; Ziegler, A.D.; Giambelluca, T.W.; Yin, S. Transpiration characteristics of a rubber plantation in central Cambodia. *Tree Physiol.* **2014**, *34*, 285–301.
13. Uri, V.; Varik, M.; Aosaar, J.; Kanal, A.; Kukumägi, M.; Lõhmus, K. Biomass production and carbon sequestration in a fertile silver birch (*Betula pendula* Roth) forest chronosequence. *For. Ecol. Manag.* **2012**, *267*, 117–126.
14. Li, T.J.; Ren, B.W.; Wang, D.H.; Liu, G.B. Spatial variation in the storages and age-related dynamics of forest carbon sequestration in different climate zones—evidence from *Black Locust* plantations on the Loess Plateau of China. *PLoS ONE* **2015**, *10*, e0121862.
15. Yang, Y.H.; Luo, Y.Q.; Finzi, A.C. Carbon and nitrogen dynamics during forest stand development: A global synthesis. *New Phytol.* **2011**, *190*, 977–989.
16. Khasanah, N.; van Noordwijk, M.; Ningsih, H.; Rahayu, S. Carbon neutral? No change in mineral soil carbon stock under oil palm plantations derived from forest or non-forest in Indonesia. *Agric. Ecosyst. Environ.* **2015**, *211*, 195–206.
17. Dey, S.K. A preliminary estimation of carbon stock sequestered through rubber (*Hevea brasiliensis*) plantation in north eastern region of India. *Indian For.* **2005**, *131*, 1429–1436.
18. Wauters, J.; Coudert, S.; Grallien, E.; Jonard, M.; Ponette, Q. Carbon stock in rubber tree plantations in Western Ghana and Mato Grosso (Brazil). *For. Ecol. Manag.* **2008**, *255*, 2347–2361.
19. Saengruksawong, C.; Khamyong, S.; Anongrak, N.; Pinthong, J. Growths and carbon stocks of para rubber plantations on Phonpisai Soil Series in Northeastern Thailand. *Rubber Thai J.* **2012**, *1*, 1–18.
20. Maggiotto, S.R.; Oliveira, D.D.; Marur, C.J.; Stivari, S.M.S.; Leclerc, M.; Wagner-Riddle, C. Potential carbon sequestration in rubber tree plantations in the northwestern region of the Paraná State, Brazil. *Acta Sci. Agron.* **2014**, *36*, 239–245.
21. Nizami, S.M.; Zhang, Y.P.; Sha, L.Q.; Zhao, W.; Zhang, X. Managing carbon sinks in rubber (*Hevea brasiliensis*) plantation by changing rotation length in SW China. *PLoS ONE* **2014**, *9*, e115234.
22. Tang, J.W.; Pang, J.P.; Chen, M.Y.; Guo, X.M.; Zeng, R. Biomass and its estimation model of rubber plantations in Xishuangbanna, Southwest China. *Chin. J. Ecol.* **2009**, *28*, 1942–1948.
23. Yang, J.C.; Huang, J.H.; Tang, J.W.; Pan, Q.M.; Han, X.G. Carbon sequestration in rubber tree plantations established on former arable lands in Xishuangbanna, SW China. *Chin. J. Plant Ecol.* **2005**, *29*, 296–303.
24. Cotta, M.K.; Jacovine, L.A.G.; Valverde, S.R.; Paiva, H.N.D.; Virgens Filho, A.D.C.; Silva, M.L.D. Economic analysis of the rubber-cocoa intercropping for generation of certified emission reduction. *Revista Árvore* **2006**, *30*, 969–979. (In Portuguese)
25. Cheng, C.M.; Wang, R.S.; Jiang, J.S. Variation of soil fertility and carbon sequestration by planting *Hevea brasiliensis* in Hainan Island, China. *J. Environ. Sci.* **2007**, *19*, 348–352.
26. Fernandes, T.J.G.; Soares, C.P.B.; Jacovine, L.A.G.; Alvarenga, A.D.P. Quantification of carbon stored in the aerial part and roots of 12-year-old *Hevea* sp., in the “Zona da Mata Mineira”. *Revista Árvore* **2007**, *31*, 657–665. (In Portuguese)
27. Sha, L.Q. Carbon Storage and Soil CO₂ Efflux of Tropical Seasonal Rain Forest, Rubber Tree Plantation and Paddy Soil in Xishuangbanna Ph.D. Thesis, Xishuangbanna Tropical Botanical Garden, University of Chinese Academy of Sciences, Beijing, China, 1 June 2008.
28. Peng, Y. Researches on Carbon Sequestration Function of Plant Subsystem in Rubber (*Hevea brasiliensis*) Plantation Ecosystem in Western Hainan. Master’s Thesis, Hainan University, Haikou, China, 1 July 2010.
29. Zheng, Z.B.; Wang, X.; Zhou, Z.D.; Yang, H. Carbon storage and distribution law of the rubber forest ecological system in Danzhou of Hainan. *Trop. Agric. Eng.* **2010**, *34*, 45–50.
30. Egbe, A.E.; Tabot, P.T.; Fonge, B.A.; Bechem, E. Simulation of the impacts of three management regimes on carbon sinks in rubber and oil palm plantation ecosystems of South-Western Cameroon. *J. Ecol. Nat. Environ.* **2012**, *46*, 154–162.
31. Kongsager, R.; Napier, J.; Mertz, O. The carbon sequestration potential of tree crop plantations. *Mitig. Adapt. Strateg. Glob. Chang.* **2013**, *18*, 1197–1213.
32. Miao, X.L. Study on the Potential for Carbon Sequestration Fartificial PR107 Rubber Forest Ecosystem in Rubber Tree Planting Area of Qiongzong of Hainan. Master’s Thesis, Hainan University, Haikou, China, 1 July 2013.

33. Sun, Y. Simulation of Carbon Dynamic Change and Accumulation Process in Rubber Plantation of Xishuangbanna. Ph.D. Thesis, Xishuangbanna Tropical Botanical Garden, University of Chinese Academy of Sciences, Beijing, China, 1 June 2013.
34. Xiao, Z.W.; Wang, X.H.; Zheng, L.; Wang, X.L.; Gao, L.H.; et al. Biomass and its allocation pattern of monoculture and mixed rubber-tree plantations in Xishuangbanna. *J. Central South Univ. For. Technol.* **2014**, *34*, 108–116.
35. Bert, D.; Danjon, F. Carbon concentration variations in the roots, stem and crown of mature *Pinus pinaster* (Ait.). *Forest Ecol. Manag.* **2006**, *222*, 279–295.
36. Peichl, M.; Arain, M.A. Above-and belowground ecosystem biomass and carbon pools in an age-sequence of temperate pine plantation forests. *Agr. Forest Meteorol.* **2006**, *140*, 51–63.
37. Zheng, W.J.; Bao, W.K.; Gu, B.; He, X.; Leng, L. Carbon concentration and its characteristics in terrestrial higher plants. *Chin. J. Ecol.* **2007**, *26*, 307–313.
38. Cao, J.X.; Wang, X.P.; Tian, Y.; Wen, Z.Y.; Zha, T.S. Pattern of carbon allocation across three different stages of stand development of a Chinese pine (*Pinus tabulaeformis*) forest. *Ecol. Res.* **2012**, *27*, 883–892.
39. Zhou, Y.R.; Yu, Z.L.; Zhao, S.D. Carbon storage and budget of major Chinese forest types. *Acta Phytoecol. Sin.* **2000**, *24*, 518–522.
40. Wang, F.M.; Xu, X.; Zou, B.; Guo, Z.H.; Li, Z.A.; Zhu, W.X. Biomass accumulation and carbon sequestration in four different aged *Casuarina equisetifolia* coastal shelterbelt plantations in South China. *PLoS ONE* **2013**, *8*, e77449.
41. Zhang, H.; Song, T.Q.; Wang, K.L.; Du, H.; Yue, Y.M.; Wang, G.X.; Zeng, F.P. Biomass and carbon storage in an age-sequence of *Cyclobalanopsis glauca* plantations in southwest China. *Ecol. Eng.* **2014**, *73*, 184–191.
42. Mao, R.; Zeng, D.H.; Hu, Y.L.; Li, L.J.; Yang, D. Soil organic carbon and nitrogen stocks in an age-sequence of poplar stands planted on marginal agricultural land in Northeast China. *Plant Soil* **2010**, *332*, 277–287.
43. Fonseca, W.; Alice, F.E.; Rey-Benayas, J.M. Carbon accumulation in aboveground and belowground biomass and soil of different age native forest plantations in the humid tropical lowlands of Costa Rica. *New For.* **2012**, *43*, 197–211.
44. Ming, A.; Jia, H.; Zhao, J.; Tao, Y.; Li, Y. Above-and below-ground carbon stocks in an indigenous tree (*Mytilaria laosensis*) plantation chronosequence in subtropical China. *PLoS ONE* **2014**, *9*, e109730.
45. Pregitzer, K.S.; Euskirchen, E.S. Carbon cycling and storage in world forests: Biome patterns related to forest age. *Glob. Chang. Biol.* **2004**, *10*, 2052–2077.
46. Song, Q.H.; Tan, Z.H.; Zhang, Y.P.; Sha, L.Q.; Deng, X.B.; et al. Do the rubber plantations in tropical China act as large carbon sinks? *Forest* **2014**, *7*, 42–47.
47. Cleveland, C.C.; Taylor, P.; Chadwick, K.D.; Dahlin, K.; Doughty, C.E.; Malhi, Y.; Smith, W.K.; Sullivan, B.W.; Wieder, W.R.; Townsend, A.R. A comparison of plot-based satellite and Earth system model estimates of tropical forest net primary production. *Glob. Biogeochem. Cycle* **2015**, *29*, 626–644.
48. Peltoniemi, M.; Mäkipää, R.; Liski, J.; Tamminen, P. Changes in soil carbon with stand age—an evaluation of a modelling method with empirical data. *Glob. Chang. Biol.* **2004**, *10*, 2078–2091.
49. Tang, J.W.; Zhang, J.H.; Song, Q.S.; Cao, M.; Feng, Z.L.; Dang, C.L.; Wu, Z.L. A preliminary study on the biomass of secondary tropical forest in Xishuangbanna. *Acta Phytoecol. Sin.* **1998**, *22*, 489–498.
50. Ziegler, A.D.; Phelps, J.; Yuen, J.Q.; Webb, E.L.; Lawrence, D.; Fox, J.M.; Bruun, T.B.; Leisz, S.J.; Ryan, C.M.; Dressler, W.; et al. Carbon outcomes of major land-cover transitions in SE Asia: great uncertainties and REDD+ policy implications. *Glob. Chang. Biol.* **2012**, *18*, 3087–3099.

