



David Dušek¹, Jiří Novák¹ and Jakub Černý^{1,2,*}

- ¹ Department of Silviculture, Forestry and Game Management Research Institute, Na Olivě 550, 517 73 Opočno, Czech Republic; dusek@vulhmop.cz (D.D.); novak@vulhmop.cz (J.N.)
- ² Department of Silviculture, Faculty of Forestry and Wood Technology, Mendel University in Brno, Zemědělská 3, 613 00 Brno, Czech Republic
- * Correspondence: cerny@vulhmop.cz; Tel.: +420-736-124-662

Abstract: Norway spruce stands are established and managed along various site conditions in central Europe. Currently, spruce often grows at locations outside of its ecological optimum, resulting in extensive damage elicited by harmful abiotic and biotic factors, which relatively shortens the time to change this adverse status in the adaptation frame by foresters. Except for the rapid change in species composition through clear-cuts, another way is possible, i.e., stabilising current (especially young) spruce stands to extend the time required to implement adaptation measures. The assumption that different site conditions will have to be respected as part of this adaptation was confirmed by our study based on NFI data of the Czech Republic. A semiparametric generalized linear model (GAM) was used to model the relationship between the height-to-diameter ratio and forest stand age, differentially considering particular forest vegetation zones. Spruce stands with lower elevations attain a lower stability (expressed by their height-to-diameter ratio; HDR) than those in the mountains. The HDR culminated in lower and middle altitudes in the first half of the rotation period, representing the most critical timing and effectivity of silvicultural measures. Contrary to previous findings, we found higher HDR values at nutrient-rich sites than those at acid ones, especially up to 50-60 years old. Therefore, more research should be devoted to the issue concerning the same thinning regime under different site conditions.



Citation: Dušek, D.; Novák, J.; Černý, J. The Mechanical Stability of Pure Norway Spruce Stands along an Altitudinal Gradient in the Czech Republic. *Forests* **2023**, *14*, 1558. https://doi.org/10.3390/f14081558

Academic Editor: W. Keith Moser

Received: 16 June 2023 Revised: 23 July 2023 Accepted: 28 July 2023 Published: 30 July 2023



Copyright: © 2023 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (https:// creativecommons.org/licenses/by/ 4.0/). **Keywords:** *Picea abies;* National Forest Inventory (NFI); height-to-diameter ratio (HDR); crown ratio (CR); Forest Vegetation Zone (FVZ); wind; snow; central Europe

1. Introduction

European forest ecosystems are threatened by abiotic (storm, snow, frost, drought, and fire, e.g., [1–4]) and biotic factors (pest insects, diseases, and ungulates, e.g., [3,5–7]) under ongoing global climate change conditions, characterised by increasing mean annual air temperatures and changes in the distribution of precipitation within the growing season [3,8], resulting in a decrease in the provision of multiple ecosystem services [9]. Therefore, the mechanical stability of forest ecosystems, as one of the most productive components in the planetary biosphere, plays a pivotal role [10–12].

Natural disorders significantly affect forest structure and dynamics [13,14]. Forest vulnerability to natural disturbances has recently increased, with around 58% of all European forests being at risk of biomass loss [15]. This represents 39 million hectares of anthropogenically or naturally disturbed forests (in the 1986–2016 period), corresponding to 17% of the total European forest area [16], with an evident increasing trend in recent decades [17]. Specifically, in the Czech Republic, salvage cuttings removing damaged timber accounted for 86.9% of the total felling volume (from 28.72 million m³) in 2020 [18]. Windstorms elicit significant disturbances in all European forest ecosystems [11,19], resulting in severe economic and ecological consequences [20,21] and representing more

than 50% of the volume of all the damaged forests in Europe [22]. These disturbances significantly affect forest productivity, carbon storage, and biodiversity [23–26]. On the contrary, snow damage routinely occurs in northern European spruce stands, mainly in mountainous areas [27]; however, it is complicated to differentiate snow damage from disturbances caused by wind, because both usually coincide [2].

European conifers (gymnosperms), mainly Norway spruce [28], are less stable than angiosperms [29]. Nonetheless, comparison is tricky, because conifers occupy sites that are usually inappropriate for many broad-leaved tree species (wind exposure, high altitudes, shallow soils, and high water tables; [12]). In addition, forest management can significantly enhance mechanical stability and persistence against forest stand disturbances [11] through silvicultural treatments [17,30–33]. The tending of young stands positively affects their resistance to wind and snow at the tree and stand levels [30,34–41]. This has also been confirmed for other conifers (e.g., [42-44]) or broad-leaved tree species [45,46]. Generally, the mechanical stability of forest stands against abiotic factors and stem radial growth is related to crown development [47–49]. In Norway spruce stands, radial increment and resistance to abiotic and biotic factors substantially increase after reductions in stand density by thinning (e.g., [50]), especially in pole stage stands [30]. However, heavy silvicultural measures applied in pole or older stands are connected to the risk of stand disintegration, especially if the treatments are the first in this growing phase [51, 52], which also applies to pine stands [31]. Since the main risk factors affecting mechanical stability are tree species, tree height, and the diameter at the breast height (i.e., at 1.3 m above ground level, DBH [51,53]), the slenderness ratio, defined as the height-to-DBH ratio (HDR), is a crucial parameter for evaluating the mechanical stability at the tree and stand levels [5,54–56]. Generally, intensive stand tending can amplify a favourable HDR value. However, the damage risk may be greater several years after the interventions (before the HDR value decreases [57,58]). Therefore, in-time stand tending (no later than when a 7 m stand dominant height is reached [59]) is necessary to attenuate HDR culmination and lower the tree's centre of gravity in a young growing phase [57]. Trees with lower HDR values withstand natural hazards, such as wind and snow damage, because of their more extended live crown and lower located centre of gravity [60]. Heavy thinning is more effective than mild thinning, but the timing of this intervention is crucial. The earlier the thinning application, the more efficiently it affects the HDR. Once the dominant heights exceed about 20 m, the thinning effect is comparatively slight [61].

On the contrary, trees with higher HDR values are more slender and less stable, resulting in a higher vulnerability to abiotic damage [62]. Generally, HDR values below 80 indicate a sufficient tree mechanical stability against external abiotic factors. In contrast, HDR values exceeding 100 signify specimens that are significantly vulnerable to natural hazards [63].

Although Norway spruce (*Picea abies* L. Karst.) is immensely affected by abiotic damage [64], followed by bark beetle (*Ips typographus* L.) outbreaks in the Czech Republic nowadays [65], its representation will be relatively high with significant economic value in the ongoing century in central Europe [20]. It still represents an ecologically and economically important tree species in central Europe (e.g., [30,66,67]). However, only a few papers have been devoted to the mechanical stability of Norway spruce stands along an altitudinal (ecological) gradient with various site conditions (i.e., acid vs. nutrient-rich sites). Thus, the presented paper aims to find the most suitable sites for spruce growing related to its mechanical stability (HDR; Crown ratio, CR), based on the assessment of an extensive dataset from the National Forest Inventory (NFI) performed in the Czech Republic. The results will also help to formulate silviculture recommendations that could contribute to better mechanical stability parameters in existing spruce stands.

We hypothesise:

H1. *Height-to-diameter ratio* (HDR) *values culminate in 40-year-old and younger pure Norway spruce stands.*

H2. HDR increases and CR decreases with lower altitudinal vegetation zones.

H3. A higher HDR will be observed at nutrient-rich sites, promoting intensive height growth.

2. Materials and Methods

2.1. Data Source

All the biometric data were obtained within the first NFI in the Czech Republic and performed by the Forest Management Institute (government organisation) in 2001–2004 [68]. Circular inventory plots with an area of 500 m² were selected randomly within inventory squares with a spacing of 2 km \times 2 km (two plots per square). For this study, 14,220 inventory plots were established and measured across broad forest stand conditions countrywide—the field measurements involved all trees with a DBH above 12 cm occurring in the inventory plots.

We selected 3969 inventory plots with dominating Norway spruce and a minimum of 15 spruces per plot, using information about stand age, Forest Vegetation Zones (FVZ), and Ecological Series (ES) at the plot level. In the Czech typological system, FVZs are defined by the relationship between the climate (elevation is a proxy for climate conditions) and the tree layer composition of forest stands along an altitudinal gradient [69]. For modelling purposes, six FVZs (from second to seventh) with a sufficient data pool from the NFI were selected (Table 1).

Table 1. Basic characteristics of forest vegetation zones (FVZs) in the Czech Republic.

Forest Vegetation Zone (FVZ)	Mean Annual Air Temperature (°C)	Mean Sum of Annual Precipitation (mm)	Growing Season Length (Days)
2nd Beech—Oak zone	7.5-8.0	600–650	160–165
3rd Oak—Beech zone	6.5–7.5	650-700	150-160
4th—Beech zone	6.0-6.5	700-800	140-150
5th Fir—Beech zone	5.5-6.0	800-900	130–140
6th Spruce—Beech zone	4.5-5.5	900-1050	115–130
7th Beech—Spruce zone	4.0-4.5	1050-1200	100–115

In the typological system of the Czech Republic, forest site units are organised by edaphic categories, which are further divided into ecological series (ES) based on the similarity of the soil conditions [69]. Within the statistical software environment R 4.1.0 [70], the two most frequent ESs were analysed in the study:

- (1) nutrient-rich—this represents mesotrophic soils without a significant soil water influence.
- (2) acidic—this represents oligotrophic soils with no significant soil water effect.

2.2. Aggregation of Tree-Level Data

We used the DBH, total height, and crown base height data at the individual tree level. The diameter–height relationship was modelled individually per plot according to the Levakovic function [71]:

$$h = a \times \left(\frac{d}{d+1}\right)^b + 1.3\tag{1}$$

where h is the tree height (m), d is the DBH (cm), and a and b are regression coefficients.

The dominant stem diameter was calculated as the quadratic mean diameter for the 100 largest trees per hectare, i.e., the five thickest trees per plot (500 m^2).

$$d_{dom} = \sqrt{\frac{\sum d_i^2}{5}} \tag{2}$$

The height-to-diameter ratio (HDR) of the dominant stem at the plot level was calculated as follows:

$$HDR = \frac{h_{dom}}{d_{dom}} \times 100 \tag{3}$$

where *HDR* is the height-to-diameter ratio, h_{dom} is the dominant tree height (m), and d_{dom} is the dominant stem diameter (cm).

The mean crown ratio (*CR*) at the plot level was calculated as follows:

$$CR = \frac{\sum_{i=1}^{n} \frac{l_i}{h_i}}{n} \tag{4}$$

where *CR* is the crown ratio, l_i is the crown length of a tree (m), h_i is a tree's total height (m), and *n* is the number of trees per plot.

2.3. HDR-Age Relationship Modelling

A semiparametric generalised additive model (GAM) from the R mgcv library [72] was used to model the HDR–age relationships. A GAM can be viewed as a GLM (generalised linear model), in which the linear predictor partly and linearly depends on some unknown smooth function *f*. GAMs are estimated by a penalized version of the method used to fit GLMs. In general, a GAM has the form [73]:

$$E(Y) = g^{-1}(\beta_0 + \sum_{j=1}^{J} f_j(x_j))$$
(5)

where E(Y) is the mean value of the response Y (with an appropriately defined distribution and link function g), f_j is a smooth function of the covariate x_j , β_0 is the intercept term, and g^{-1} is the inverse link function. Each smoother f_j is represented by a sum of K basis functions ($b_{i,k}$) multiplied by coefficients ($\beta_{i,k}$):

$$f_j(x_j) = \sum_{k=1}^{K} \beta_{j,k} \, b_{j,k} \, (x_j)$$
(6)

where *K* determines the maximum complexity of each smoother and the large basis size *K* could lead to overfitting, which is compensated for by a smoothing penalty [73]. In the mgcv library, thin plate regression spline (TPRS) [72,74] uses a penalty matrix based on the integral of the squared derivates of the basis functions [72,73,75] in the form:

$$\lambda \int \left[f_j^m(x) \right]^2 dx \tag{7}$$

where f_j^m is an *m*-th order derivation (usually second-order) and λ is a penalty constant (also called a smoothing parameter). The integral (7) is a roughness penalty [76]. The integral (7) can be rewritten into linear algebra form:

$$\lambda \beta^T S \beta \tag{8}$$

where *S* is the penalty matrix multiplied by the parameter vector β (T denotes matrix transposition), and the penalty term (8) is then subtracted from the model log-likelihood [73]. The restricted maximum likelihood method (REML) was chosen for searching the optimal λ value, because the default method in the mgcv library (GCV—generalized cross-validation) tends to overfit smaller or noisy datasets. Overfitting tends to look like spline fits that are too wiggly [73].

The GAM allows us to model the nonlinear relationship between the HDR and stand age (covariate) for each level of factor variable (*FVZ*). As a smoother, we specifically

used thin-plate regression spline [72,74] in the mgcv R library, which is appropriate for models with many independent variables, both continuous and categorical. The dependent variable (*HDR*) was modelled as it comes from a gamma distribution (with a logarithmic link transform function), because the model with a normal distribution showed skewness and heteroscedasticity in its residuals. The model can be informally written as follows:

$$\log(HDR_{ij}) = \alpha + f_j(Age_i) + FVZ_j + f(Age_i) : FVZ_j$$
(9)

$$HDR_{ij} \sim Gamma(\mu_{ij}, \varphi)$$
 (10)

where α is the intercept term, *f* is the *j*-th smoothing term for the *i*-th age, *Age* is a smoothed covariate, *FVZ* is the factor variable with the *j*-th level of forest vegetation zones, and *Age: FVZ* is the interaction. Including an interaction term in the equation guarantees the computation of a smooth term for each factor variable (*FVZ*) level with its penalties. The term log(*HDR*) indicates that the dependent variable was calculated on a logarithmic scale. The notation: $HDR_{ij} \sim Gamma(\mu_{ij}, \varphi)$ indicates that the dependent variable came from a gamma distribution with a mean of μ for the *i*-th age and *j*-th FVZ, and φ is the dispersion parameter. The gamma distribution is a continuous, two-parameter distribution belonging to the exponential family and is used for positively skewed and strictly positive real data [77].

In *R* code, the model is specified as:

model_9_10 <- gam (HDR ~ s(Age, by = FVZ, bs = 'tp') + FVZ, method = 'REML', family = 'Gamma'(link = log)).

2.4. Modelling HDR–Age Relationship by Ecological Series (ES)

Inventory plots in two ES series were selected for modelling the HDR–age relationships in terms of ES: nutrient-rich (1492 plots) and acidic (1598 plots). As mentioned above, we added the factor ES to the *HDR*–age relationship model.

$$\log(HDR_{ijk}) = \alpha + f_j(Age_i) + FVZ_j + ES_k + f(Age_i) : FVZ_j$$
(11)

$$HDR_{iik} \sim Gamma(\mu_{iik}, \varphi)$$
 (12)

In *R* code, the model is specified as:

,

model_11_12 <- gam (HDR ~ s(Age, by = FVZ, bs = 'tp') + FVZ + ES, method = 'REML', family = 'Gamma'(link = log)).

2.5. CR–Age Relationship Modelling

The CR–age relationship model was built analogically, the same as the HDR–age model, but the distribution function was switched from Gamma to Beta with a logit-link function. The beta distribution is a two-parameter distribution beyond an exponential family distribution and is appropriate for response variables ranging between 0 and 1 [77]. The logistic function is a canonical link function for beta distribution models. The model can be written as follows:

$$\log\left(\frac{\pi_{ij}}{1-\pi_{ij}}\right) = \alpha + f_j(Age_i) + FVZ_j + f(Age_i) : FVZ_j$$
(13)

$$CR_{ij} \sim Beta(\mu_{ij}, \varphi)$$
 (14)

where $\log(\pi_{ij}/(1 - \pi_{ij}))$ is the logistic transformation of the response variable *CR*, which is assumed to come from a beta distribution with the parameters μ (mean) and φ (dispersion parameter).

In *R* code, the model is specified as:

model_13_14 <- gam (CR ~ s(Age, by = FVZ, bs = 'tp') + FVZ, method = 'REML', family = 'betar'(link = logit)).

3. Results

3.1. Development of Dominant Trees' HDR by Forest Vegetation Zones (FVZ)

The HDR development of the dominant spruce trees differed in terms of the FVZ (Figure 1), although the trend of an initial increase and a later HDR decrease was noted for almost all the analysed FVZs. Generally, the highest FVZ (i.e., highest elevation) corresponded to the lowest HDR of the dominant trees. The lowest HDR values (below 65 throughout the observed period, i.e., for 20–100-year-old stands) were found for "montane" FVZ 7. In the downwardly following FVZ 6, the HDR continually increased from 62 to 70 in the analysed age period.



Figure 1. Model (GAM) of HDR development by forest vegetation zones (FVZs 2–7). Shadow areas depict 95% confidence intervals of means.

An initial steeper HDR value increase was found in FVZs 2–5. The HDR in FVZ 5 culminated before the age of 40 years at the level below 75 and slowly decreased in older stands. A similar pattern was shown in the results from FVZs 4 and 3. However, the HDR culminated at approximately 45 years and at higher values of 77 and 79 in FVZs 4 and 3, respectively. The highest HDR values, indicating the lowest mechanical stability, were found in the lowest-located FVZ 2. In this "lowland" zone, the HDR of the spruce dominant trees increased to 83 in 55-year-old stands. Subsequently, it decreased to level 77.

The above-mentioned order of the HDR values of the dominant trees in the studied FVZs is also evident from Figure 2. Empirical cumulative density curves (ECDF) showed a similar pattern, while the most favourable HDR distribution was evident in FVZ 7, and the most adverse was observed in FVZ 2.



Figure 2. Empirical cumulative density curves (ECDF) of dominant trees' HDR by forest vegetation zones (FVZs 2–7).

3.2. Development of Dominant Trees' HDR in Acidic and Nutrient-Rich Ecological Series (ES)

The differences between the HDR values of the dominant trees at acidic and nutrientrich sites were found mainly for the first half of the rotation period, i.e., up to the age of 50–60 years (Figure 3). The HDR culminated in lower values at the nutrient-rich sites compared to the acidic ones for all the analysed FVZs. Therefore, more propitious conditions at the nutrient-rich sites significantly contributed to higher radial and lower height increments, leading to lesser HDR values.



Figure 3. Model (GAM) of HDR development at two ecological series by forest vegetation zones (FVZs 2–7). Shadow areas depict 95% confidence intervals of means. Red and blue lines represent acid and nutrient-rich sites, respectively.

The results revealed that the CR naturally decreased with an increasing age in all the analysed FVZs (Figure 4). The highest CR values (0.7–0.8) were detected in 20-year-old stands and the lowest (below 0.5) in 100-year-old ones. This age-related decrease trend was more gradual in the lowest (2–5) compared to the highest (6–7) FVZs. Thus, the most favourable CR values (i.e., the highest) were detected in "montane" (FVZ 7) throughout the studied stand age range.



Figure 4. Model (GAM) of mean CR development by forest vegetation zones (FVZs 2–7). Shadow areas depict 95% confidence intervals of means.

The empirical cumulative density curves (ECDF) showed a similar pattern (Figure 5) as well, while the CR distribution in FVZ 7 was the most favourable (approximately 75% of all the analysed inventory plots showed a mean CR exceeding 0.5). The percentage decreased with lower FVZs (only around 30% of all the inventory plots showed a CR of > 0.5 in FVZ 2).



Figure 5. Empirical cumulative density curves (ECDF) of mean CR along an altitudinal gradient (FVZs).

4. Discussion

NFI data are routinely used to monitor forest stands' productivity (e.g., [33,50,78]) or evaluate ecosystem services and biodiversity (e.g., [79–81]). The presented study used NFI data from the Czech Republic to assess the mechanical stability of Norway spruce. Recently, a nonlinear mixed-effects HDR model was also developed on the basis of these Czech NFI data by Sharma et al. [55] for several tree species. Additionally, the predictive performance of this model was evaluated using data from research sample plots. The dominant height and diameter, relative spacing index, and DBH-to-quadratic mean DBH ratio were identified as the most important predictors of HDR variations. Our approach to the NFI data processing differed from the above, because we used the HDR–age and CR–age relationships for the modelling.

Our results confirmed that the HDR changed naturally with age, culminating in the first half of the standard rotation period for spruce in lower and middle elevations, supporting hypothesis H1 at these altitudes. This was also consistent with the findings of Cremer et al. [57], where increases in the HDR values were most pronounced up to dominant heights of 20 m or 25 m, and for greater heights, the HDR values levelled off and even declined with further growth.

Dominant trees are naturally more stable, i.e., the HDR values are the lowest for dominant and highest for suppressed trees [82], but especially at lower elevations (FVZ 2), where we found the HDR culmination of the dominant spruce trees to be over the value of 80, where the risk of abiotic damage (firstly from snow and later from wind) is increased [63].

In this study, higher altitudes (represented by higher FVZs) generally represented lower HDR values throughout all the analysed age periods (see hypothesis H2). Although spruce stands in the Czech Republic do not always correspond to their original populations (due to the historical use of spruce outside of its natural distribution area), the results confirmed that these stands are more stable in higher locations [83,84], primarily due to their lower height growth, which is the result of adaptation to the given environment [62,85–88], and the higher radiation use efficiency of mountainous spruce stands [89], which probably allocate carbon to the stem (i.e., to a radial increment resulting in a lower HDR value and higher stand stability). In contrast, silvicultural measures promoting crown morphology and the growth of young spruce stands are less effective in the mountains than lower-located sites [90].

A lower elevation also generally represents more favourable site conditions (mainly through deeper soils with higher available nutrients). These conditions lead to larger diameters, but especially to height growth (expressed, for instance, by site index), i.e., a higher site index corresponds with a lower mechanical stability—higher HDR values [91,92]. On the contrary, we found in our analyses that the spruce trees attained more favourable HDRs at nutrient-rich sites compared to acid ones at the same age (hypothesis H3 was rebutted). The differences were not high, but they were confirmed for all the analysed FVZs in the first half of the rotation period. More favourable growth conditions promoted that height growth did not substantially overwhelm radial growth, resulting in a nottoo-intense increase in the HDR value in youth (thickets and small pole stands). The increased growth of spruce and beech recorded in Europe [93] could also play a specific role, especially at nutrient-rich sites. On the other hand, Dobbertin [94] in Switzerland and Bouchard et al. [95] in Canada observed that windthrow occurred more frequently at fertile sites than poor sites. Therefore, our finding leads to the need for further research, focusing on the characteristics of stands in different conditions (site indexes), but under identical thinning management.

The analysed NFI plots are managed "in a standard way", and the detailed thinning regimes for individual locations are unknown. Stand density and management applied through thinning treatments are essential from a stability point of view, expressed in terms of HDR or CR. The competition was recorded in ponderosa pine [87], loblolly pine [96], or Engelmann spruce [88] dense stands as a significant factor affecting the stem and crown

10 of 14

parameters in connection with the risk of damage from snow or wind. Thus, windstorms often suffer well-stocked mature Norway-spruce-dominated stands [97] or dense coniferous stands [94,95].

The HDR or CR values are also influenced by a tree's distance from the stand edge [98]. Trees growing at the stand's edge with longer live crowns are naturally more stable (a lower HDR and higher CR) than trees inside the stand.

These results should be, together with the knowledge of the thinning effects on stability [51,99], the basis of practical recommendations for forest practices to reduce the risk of forest damage. Of course, besides thinning applications as a mitigation measure, it is necessary to focus on forest regeneration, which is a principal measure enabling species composition change [100,101] in order to reduce the proportion of spruce (especially in lowlands and middle elevations) and increase the share of mixtures, characterised by a higher annual volume increment [102] and having a more positive effect on productivity [103] and increased temporal stability than monocultures [104].

5. Conclusions

Norway spruce stands are damaged by abiotic and biotic harmful factors in many European localities, often due to their historical extension to sites outside of the ecological optimum for this tree species. In addition to a gradual change in species composition towards mixtures, it is essential to manage spruce stands to minimise the risk of their large-scale disintegration and allow enough time for the abovementioned conversion. In forestry management, it is advisable to adapt procedures to the different site conditions where spruce stands grow. The presented study confirmed the facts based on the NFI data from the Czech Republic. Spruce stands in lower elevations show a lower stability (expressed in terms of HDR) than those in the mountains. The HDR culminated in lower and middle elevations in the first half of the rotation period, representing the time with potentially the highest silviculture measures (thinning) effectiveness. The study did not confirm predicted higher HDR values at nutrient-rich sites than at acid ones, especially for the first half of the rotation period. Therefore, this topic should be the subject of further research, mainly concerning the same thinning regime at different site conditions.

Author Contributions: Conceptualization D.D., J.N. and J.Č.; methodology D.D.; software D.D.; validation J.Č.; formal analysis D.D.; investigation D.D., J.N. and J.Č.; resources D.D., J.N. and J.Č.; data curation D.D., J.N. and J.Č.; writing—original draft preparation D.D., J.N. and J.Č.; writing—review and editing J.Č.; visualization D.D.; supervision J.Č.; project administration J.Č.; funding acquisition J.N. and J.Č. All authors have read and agreed to the published version of the manuscript.

Funding: The research was financially supported by the Ministry of Agriculture of the Czech Republic, institutional support MZE-RO0123, the National Agency of Agricultural Research (Project No. QK21020307 "Optimising silvicultural practices for adaptation of forest ecosystems to climate change") and Internal Grant Agency, Faculty of Forestry and Wood Technology, MENDELU Brno, the Czech Republic (Project No. IGA-LDF22TP2-102 "Influence of climatic factors on growth dynamics of stands with different structures and its effect on the soil").

Data Availability Statement: The data presented in this study are available on request from the corresponding author.

Acknowledgments: We sincerely thank the Forestry Management Institute (FMI) for sharing raw data from the National Forest Inventory (NFI) analysed in the presented study. Authors are indebted to anonymous reviewers for their valuable comments, significantly improving the manuscript.

Conflicts of Interest: The authors declare no conflict of interest. The funders had no role in the study's design, data collection, analysis and interpretation, manuscript writing, or publication.

References

- 1. Kohler, M.; Sohn, J.; Nägele, G.; Bauhus, J. Can drought tolerance of Norway spruce (*Picea abies* (L.) Karst.) be increased through thinning. *Eur. J. For. Res.* 2010, 129, 1109–1118. [CrossRef]
- Hlásny, T.; Křístek, Š.; Holuša, J.; Trombik, J.; Urbaňcová, N. Snow disturbances in secondary Norway spruce forests in Central Europe: Regression modeling and its implications for forest management. *For. Ecol. Manag.* 2011, 262, 2151–2161. [CrossRef]
- 3. Seidl, R.; Thom, D.; Kautz, M.; Martin-Benito, D.; Peltoniemi, M.; Vacchiano, G.; Wild, J.; Ascoli, D.; Petr, M.; Honkaniemi, J.; et al. Forest disturbances under climate change. *Nat. Clim. Change* **2017**, *7*, 395–402. [CrossRef]
- Steckel, M.; del Río, M.; Heym, M.; Aldea, J.; Bielak, K.; Brazaitis, G.; Černý, J.; Coll, L.; Collet, C.; Ehbrecht, M.; et al. Species mixing reduces drought susceptibility of Scots pine (*Pinus sylvestris* L.) and oak (*Quercus robur* L.; *Quercus petraea* (Matt.) Liebl.)—Site water supply and fertility modify the mixing effect. For. Ecol. Manag. 2020, 461, 117908. [CrossRef]
- 5. Hanewinkel, M.; Hummel, S.; Albrecht, A. Assessing natural hazards in forestry for risk management: A review. *Eur. J. For. Res.* **2011**, *130*, 329–351. [CrossRef]
- 6. Cukor, J.; Vacek, Z.; Linda, R.; Sharma, R.P.; Vacek, S. Afforested farmland vs. forestland: Effects of bark stripping by *Cervus elaphus* and climate on production potential and structure of *Picea abies* forests. *PLoS ONE* **2019**, *14*, e0221082. [CrossRef]
- Vacek, Z.; Cukor, J.; Linda, R.; Vacek, S.; Šimůnek, V.; Brichta, J.; Gallo, J.; Prokůpková, A. Bark stripping, the crucial factor affecting stem rot development and timber production of Norway spruce forests in Central Europe. *For. Ecol. Manag.* 2020, 474, 118360. [CrossRef]
- Cavin, L.; Mountford, E.P.; Peterken, G.F.; Jump, A.S. Extreme drought alters competitive dominance within and between tree species in a mixed forest stand. *Funct. Ecol.* 2013, 27, 1424–1435. [CrossRef]
- Bottero, A.; Forrester, D.I.; Cailleret, M.; Kohnle, U.; Gessler, A.; Michel, D.; Bose, A.K.; Bauhus, J.; Bugmann, H.; Cuntz, M.; et al. Growth resistance and resilience of mixed silver fir and Norway spruce forests in central Europe: Contrasting responses to mild and severe droughts. *Glob. Change Biol.* 2021, 27, 4403–4419. [CrossRef]
- Pan, Y.; Birdsey, R.A.; Fang, J.; Houghton, R.; Kauppi, P.E.; Kurz, W.A.; Phillips, O.L.; Shvidenko, A.; Lewis, S.L.; Canadell, J.G.; et al. A large and persistant carbon sink in the world's forests. *Science* 2011, 333, 988–993. [CrossRef] [PubMed]
- 11. Suvanto, S.; Peltoniemi, M.; Tuominen, S.; Strandström, M.; Lehtonen, A. High-resolution mapping of forest vulnerability to wind for disturbance-aware forestry. *For. Ecol. Manag.* **2019**, *453*, 117619. [CrossRef]
- 12. Gardiner, B. Wind damage to forests and trees: A review with an emphasis on planted and managed forests. *J. For. Res.* 2021, 26, 248–266. [CrossRef]
- 13. Łaska, G. The disturbance and vegetation dynamics: A review and an alternative framework. *Plant Ecol.* **2001**, *157*, 77–99. [CrossRef]
- 14. Tsvetanov, M.; Dountchev, A.; Panayotov, M.; Zhelev, P.; Bebi, P.; Yurukov, S. Short- and long-term natural regeneration after windthrow disturbances in Norway spruce forests in Bulgaria. *iForest* **2018**, *11*, 675–684. [CrossRef]
- Forzieri, G.; Girardello, M.; Ceccherini, G.; Spinoni, J.; Feyen, L.; Hartmann, H.; Beck, P.S.A.; Camps-Valls, G.; Chirici, G.; Mauri, A.; et al. Emergent vulnerability to climate-driven disturbances in European forests. *Nat. Commun.* 2021, *12*, 1081. [CrossRef]
- 16. Senf, C.; Seidl, R. Mapping the forest disturbance regimes of Europe. Nat. Sustain. 2021, 4, 63–70. [CrossRef]
- 17. Schuck, A.; Schelhaas, M.-J. Storm damage in Europe—An overview. In *Living with Storm Damage to Forests*; Gardiner, B., Schuck, A., Schelhaas, M.-J., Orazio, C., Blennow, K., Nicoll, B., Eds.; European Forest Institute: Joensu, Finland, 2013; pp. 15–23.
- 18. MZe. Report about Stage of Forests and Forest Management in the Czech Republic in 2021; Ministry of Agriculture of the Czech Republic: Praha, Czech Republic, 2021; p. 47. (In Czech)
- Forzieri, G.; Pecchi, M.; Girardello, M.; Mauri, A.; Klaus, M.; Nikolov, C.; Rüetschi, M.; Gardiner, B.; Tomaštík, J.; Small, D.; et al. A spatially explicit database of wind disturbances in European forests over the period 2000–2018. *Earth Syst. Sci. Data* 2020, 12, 257–276. [CrossRef]
- Hanewinkel, M.; Cullmann, D.A.; Schelhaas, M.-J.; Nabuurs, G.-J.; Zimmermann, N.E. Climate change may cause severe loss in the economic value of European forest land. *Nat. Clim. Change* 2012, *3*, 203–207. [CrossRef]
- Konôpka, B.; Zach, P.; Kulfan, J. Wind—An important ecological factor and destructive agent in forests. For. J. 2016, 62, 123–130. [CrossRef]
- 22. Schelhaas, M.J.; Nabuurs, G.J.; Schuck, A. Natural disturbances in the European forests in the 19th and 20th centuries. *Glob. Change Biol.* **2003**, *9*, 1620–1633. [CrossRef]
- 23. Seidl, R.; Schelhaas, M.-J.; Rammer, W.; Verkerk, P.J. Increasing forest disturbances in Europe and their impact on carbon storage. *Nat. Clim. Change* **2014**, *4*, 806–810. [CrossRef]
- Zell, J.; Hanewinkel, M. How treatment, storm events and changed climate affect productivity of temperate forests in SW Germany. *Reg. Environ. Change* 2015, 15, 1531–1542. [CrossRef]
- 25. Thom, D.; Seidl, R. Natural disturbance impacts on ecosystem services and biodiversity in temperate and boreal forests. *Biol. Rev. Camb. Phylos. Soc.* 2016, *91*, 760–781. [CrossRef]
- Reyer, C.P.O.; Bathgate, S.; Blennow, K.; Borges, J.G.; Bugmann, H.; Delzon, S.; Faias, S.P.; Garcia-Gonzalo, J.; Gardiner, B.; Gonzalez-Olabarria, J.R.; et al. Are forest disturbances amplifying or canceling out climate change-induced productivity changes in European forests? *Environ. Res. Lett.* 2017, 12, 034027. [CrossRef]

- 27. Díaz-Yáñez, O.; Mola-Yudego, B.; Eriksen, R.; González-Olabarria, J.R.; Carcaillet, C. Assessment of the main natural disturbances on Norwegian forest based on 20 years of national inventory. *PLoS ONE* **2016**, *11*, e0161361. [CrossRef]
- Dodan, M.; Perić, S. Windthrow resistance of Norway spruce (*Picea abies* L. Karst.) forest cultures—preliminary results. *Seefor-South-East Eur. For.* 2019, 10, 77–88. [CrossRef]
- Gardiner, B.; Blennow, K.; Carnus, J.-M.; Fleischer, P.; Ingemarson, F.; Landmann, G.; Linder, M.; Marzano, M.; Nicoll, B.; Orazio, C.; et al. *Destructive Storms in European Forests: Past and Forthcoming Impacts*; Final report to European Commission—DG Environment: Brussels, Belgium, 2010; p. 138. [CrossRef]
- Slodičák, M.; Novák, J. Silvicultural measures to increase the mechanical stability of pure secondary Norway spruce stands before conversion. For. Ecol. Manag. 2006, 224, 252–257. [CrossRef]
- 31. del Río, M.; Bravo-Oviedo, A.; Pretzsch, H.; Löf, M.; Ruiz-Peinado, R. A review of thinning effects on Scots pine stands: From growth and yield to new challenges under global change. *For. Syst.* **2017**, *26*, eR03S. [CrossRef]
- Vacek, Z.; Prokůpková, A.; Vacek, S.; Cukor, J.; Bílek, L.; Gallo, J.; Bulušek, D. Silviculture as a tool to support stability and diversity of forests under climate change: Study from Krkonoše Mountains. *Cent. Eur. For. J.* 2020, 66, 116–129. [CrossRef]
- Romeiro, J.M.N.; Eid, T.; Anton-Fernandez, C.; Kangas, A.; Tromborg, E. Natural disturbances risks in European boreal and temperate forests and their links to climate change—A review of modelling approaches. *For. Ecol. Manag.* 2022, 509, 120071. [CrossRef]
- Pařez, J. Škody Sněhem a Větrem ve Smrkových Porostech Probírkových Pokusných Ploch v Období 1959–1968 [Damages Caused by Snow and Wind in Pure Norway Spruce Stands at Silvicultural Experiments in the Period of 1959–1968]; VÚLHM: Jíloviště-Strnady, Czech Republic, 1972; p. 97. (In Czech)
- 35. Chroust, L. Tvar kmene a velikost korun při výchově smrkových porostů ve vztahu ke škodám působeným sněhem a větrem [Stem shape and crown size in during Norway spruce tending related to damages elicited by snow and wind]. *Práce VÚLHM* 1980, 56, 31–52. (In Czech)
- Slodičák, M. Resistance of young spruce stands to snow and wind damage in dependence on thinning. *Commun. Inst. For. Cech.* 1987, 15, 75–86.
- Nykänen, M.L.; Peltola, H.; Quine, C.; Kellomäki, S.; Broadgate, M. Factors affecting snow damage of trees with particular reference to European conditions. *Silva Fenn.* 1997, *31*, 193–213.
- Mason, B.; Valinger, E. Managing forests to reduce storm damage. In *Living with Storm Damage to Forests. What Science Can Tell Us*; Gardiner, B., Schuck, A., Schelhaas, M.-J., Orazio, C., Blennow, K., Nicoll, B., Eds.; European Forest Institute: Joensuu, Finland, 2013; pp. 87–96.
- 39. Vacchiano, G.; Derose, R.J.; Shaw, J.D.; Svoboda, M.; Motta, R. A density management diagram for Norway spruce in the temperate European montane region. *Eur. J. For. Res.* **2013**, *132*, 535–549. [CrossRef]
- 40. Gauthier, M.-M.; Tremblay, S. Precommercial thinning as a silvicultural option for treating very dense conifer stands. *Scand. J. For. Res.* **2018**, *33*, 446–454. [CrossRef]
- Dušek, D.; Novák, J.; Kacálek, D.; Slodičák, M. Norway spruce production and static stability in IUFRO thinning experiments in the Czech Republic. J. For. Sci. 2021, 67, 185–194. [CrossRef]
- 42. Bragg, D.C.; Shelton, M.G.; Zeide, B. Impacts and management implications of ice storms on forests in the southern United States. *For. Ecol. Manag.* 2003, *186*, 99–123. [CrossRef]
- 43. Weiskittel, A.R.; Kenefic, L.S.; Seymour, R.S.; Phillips, L.M. Long-term effects of precommercial thinning on the stem dimensions, form and branch characteristics of red spruce and balsam fir crop trees in Maine, USA. *Silva Fenn.* **2009**, *43*, 397–409.
- 44. Zhang, X.; Wang, H.; Cchin, S.; Zhang, J. Effects of competition, age and climate on tree slenderness of Chinese fir plantations in southern China. *For. Ecol. Manag.* 2020, *458*, 117815. [CrossRef]
- 45. Swift, D.E.; Knight, W.; Béland, M.; Boureima, I.; Bourque, C.P.-A.; Meng, F.R. Stand dynamics and tree quality response to precommercial thinning in a northern hardwood forest of the Acadian forest region: 23 years of intermediate results. *Scand. J. For. Res.* **2017**, *32*, 45–59. [CrossRef]
- 46. Klinar, B.; Klopčič, M.; Bončina, A. Individual tree damage due to abiotic natural disturbances on European beech sites in Slovenia with the main focus on snow damage. *Acta Silvae Et. Ligni* 2020, 122, 53–69, (in Slovenian with English abstract and summary). [CrossRef]
- 47. Brüchert, F.; Becker, G.; Speck, T. The mechanics of Norway spruce (*Picea abies* L. Karst) mechanical properties of standing trees from different thinning regimes. *For. Ecol. Manag.* **2000**, *135*, 45–62. [CrossRef]
- Seifert, T.; Pretzsch, H.; Bücking, M. Coppice with spruce from high forest? Part II: Year ring width, stem taper and branchiness of long crowned Norway spruce. *Forst. Holz.* 2003, 59, 63–68.
- 49. Pretzsch, H. Tree growth as affected by stem and crown structure. *Trees* 2021, 35, 947–960. [CrossRef]
- 50. Katrevičs, J.; Džeriņa, B.; Neimane, U.; Desaine, I.; Bigača, Z.; Jansons, A. Production and profitability of low density Norway spruce (*Picea abies* L. Karst.) plantation at 50 years age: Case study from eastern Latvia. *Agron. Res.* 2018, *16*, 113–121. [CrossRef]
- Albrecht, A.; Hanewinkel, M.; Bauhus, J.; Kohnle, U. How does silviculture affect storm damage in forests of south-western Germany? Results from empirical modelling based on long-term observations. *Eur. J. For. Res.* 2012, 131, 229–247. [CrossRef]

- 52. Piticar, M.A. The effects of thinning on a spruce stand, over a period of 40 years of monitoring. Adv. Agric. Bot. 2016, 8, 77–86.
- 53. Peltola, H.; Kellomaki, S.; Hassinen, A.; Granander, M. Mechanical stability of Scots pine, Norway spruce and birch: An analysis of tree-pulling experiments in Finland. *For. Ecol. Manag.* 2000, *135*, 143–153. [CrossRef]
- 54. Rust, S. Analysis of regional variation of height growth and slenderness in populations of six urban tree species using a quantile regression approach. *Urban For. Urban Green.* **2014**, *13*, 336–343. [CrossRef]
- 55. Sharma, R.P.; Vacek, Z.; Vacek, S.; Kučera, M. A nonlinear mixed-effects height-to-diameter ratio model or several tree species based on Czech National Forest Inventory data. *Forests* **2019**, *10*, 70. [CrossRef]
- 56. Tian, D.; Jiang, L.; Wang, J. The influence of climate, soil physicochemical properties and tree size inequality on tree slenderness in mixed forests of Northeastern China. *For. Ecol. Manag.* **2023**, *529*, 120719. [CrossRef]
- 57. Cremer, K.W.; Borough, C.J.; McKinnell, F.H.; Carter, P.R. Effects of stocking and thinning on wind damage in plantations. *New Zealand J. For. Sci.* **1982**, *12*, 224–268.
- Bayar, E.; Deligoz, A. Effects of precommercial thinning on growth parameters in *Pinus nigra* Arn. subsp. *Pallasiana* (*Lamb.*) Holmboe Stand. J. For. Res. 2020, 31, 1803–1811. [CrossRef]
- Slodičák, M.; Novák, J.; Dušek, D. Management of Norway spruce stands in the Western Carpathians. In *The Carpathians:* Integrating Nature and Society towards Sustainability. Environmental Science and Engineering; Kozak, J., Ed.; Springer: Berlin, Heidelberg, 2013. [CrossRef]
- 60. Jelonek, T.; Walkowiak, R.; Jakubowski, M.; Tomczak, A. Indices of tree stability in Scots pine stands damaged by wind. *Sylwan* **2013**, *157*, 323–329.
- Slodičák, M.; Novák, J. Růst, Struktura a Statická Stabilita Smrkových Porostů s Různým Režimem Výchovy. [Growth, Structure and Static Stability of Norway Spruce Stands with Different Thinning Regimes]; Folia Forestalia Bohemica 3; Lesnická Práce: Kostelec nad Černými lesy, Czech, 2006; p. 128. ISBN 978-80-86386-91-1. (In Czech)
- 62. Vospernik, S.; Monserud, R.A.; Sterba, H. Do individual-tree growth models correctly represent height:diameter ratios of Norway spruce and Scots pine? *For. Ecol. Manag.* 2010, 260, 1735–1753. [CrossRef]
- Wonn, H.T.; O'Hara, K.L. Height:diameter ratios and stability relationships for four Northern Rocky Mountain tree species. West. J. Appl. For. 2001, 16, 87–94. [CrossRef]
- Bošeľa, M.; Konôpka, B.; Šebeň, V.; Vladovič, J.; Tobin, B. Modelling height to diameter ratio—An opportunity to increase Norway spruce stand stability in the Western Carpathians. For. J. 2014, 60, 71–80. [CrossRef]
- Hlásny, T.; Zimová, S.; Merganičová, K.; Štěpánek, P.; Modlinger, R.; Turčáni, M. Devasting outbreak of bark beetle in the Czech Republic: Drivers, impacts, and management implications. For. Ecol. Manag. 2021, 490, 119075. [CrossRef]
- 66. Albrecht, A.; Kohnle, U.; Hanewinkel, M.; Bauhus, J. Storm damage of Douglas-fir unexpectedly high compared to Norway spruce. *Ann. For. Sci.* 2013, 70, 195–207. [CrossRef]
- Hlásny, T.; König, L.; Kroneke, P.; Lindner, M.; Montagné-Huck, C.; Müller, J.; Qin, H.; Raffa, K.F.; Schelhaas, M.-J.; Svoboda, M.; et al. Bark beetle outbreaks in Europe: State of knowledge and ways forward for management. *Curr. For. Rep.* 2021, 7, 138–165. [CrossRef]
- FMI. National Forest Inventory in the Czech Republic 2001–2004; Forest Management Institute: Brandýs nad Labem, Czech Republic, 2007; p. 224. (In Czech)
- 69. Viewegh, J.; Kusbach, A.; Mikeska, M. Czech forest ecosystem classification. J. For. Sci. 2003, 49, 74–82. [CrossRef]
- R Core Team. R: A Language and Environment for Statistical Computing; R Foundation for Statistical Computing: Vienna, Austria, 2021. Available online: https://www.R-project.org/ (accessed on 12 April 2023).
- 71. Levakovic, A. Analytical form of growth laws. Glas. Za Sumske Pokuse 1935, 4, 189–282.
- 72. Wood, S. Generalized Additive Models, 2nd ed.; Chapman & Hall/CRC: New York, NY, USA, 2017; p. 476. [CrossRef]
- 73. Pedersen, E.J.; Miller, D.L.; Simpson, G.L.; Ross, N. Hierarchical generalized additive models in ecology: An introduction with mgcv. *PeerJ* 2019, 7, e6876. [CrossRef] [PubMed]
- 74. Wood, S. Thin plate regression splines. J. R. Stat. Society. Ser. B 2003, 65, 95–114. [CrossRef]
- 75. Faraway, J. Extending the Linear Model with R, 2nd ed.; Chapman & Hall/CRC: New York, NY, USA, 2016; p. 413. [CrossRef]
- 76. Harezlak, J.; Ruppert, D.; Wand, M.P. Semiparametric Regression with R; Springer: New York, NY, USA, 2018; p. 331. [CrossRef]
- 77. Rigby, R.A.; Stasinopoulos, M.D.; Heller, G.Z.; Bastiani, F.D. *Distributions for Modeling Location, Scale, and Shape Using GAMLSS in R*; Chapman & Hall/CRC: New York, NY, USA, 2020; p. 588. [CrossRef]
- Gschwantner, T.; Lanz, A.; Vidal, C.; Bošel'a, M.; Cosmo, L.; Fridman, J.; Gasparini, P.; Kuliesis, A.; Tomter, S.; Schadauer, K. Comparison of methods used in European National Forest Inventories for the estimation of volume increment: Towards harmonisation. *Ann. For. Sci.* 2016, 73, 807–821. [CrossRef]
- 79. Chirici, G.; McRoberts, R.E.; Winter, S.; Bertini, R.; Brandli, U.-B.; Asensio, I.A.; Bastrup-Birk, A.; Rondeaux, J.; Barsoum, N.; Marchetti, M. National Forest Inventory contributions to forest biodiversity monitoring. *For. Sci.* 2012, *58*, 257–268. [CrossRef]
- Alberdi, I.; Nunes, L.; Kovac, M.; Bonheme, I.; Cañellas, I.; Rego, F.C.; Dias, S.; Duarte, I.; Notarangelo, M.; Rizzo, M.; et al. The conservation status assessment of Natura 2000 forest habitats in Europe: Capabilities, potentials and challenges of national forest inventories data. *Ann. For. Sci.* 2019, *76*, 34. [CrossRef]

- Atkinson, M.A.; Edwards, D.M.; Jensen, F.S.; van der Jagt, A.P.N.; Ditchburn, B.R.; Sievanen, T.; Gasparini, P. Harmonising, improving and using social and recreational data in National Forest Inventories across Europe. *Ann. For. Sci.* 2020, 77, 76. [CrossRef]
- Šebeň, V.; Bošeľa, M.; Konôpka, B.; Pajtík, J. Indices of tree competition in dense spruces stand originated from natural regeneration. *Lesn. Časopis For. J.* 2013, 59, 172–179.
- Konôpka, B.; Konôpka, J. Static stability of forest stands in the seventh altitudinal vegetation zone in Slovakia. J. For. Sci. 2003, 49, 474–481.
- 84. Homeier, J.; Breckle, S.W.; Günter, S.; Rollenbeck, R.T.; Leuschner, C. Tree diversity, forest structure and productivity along altitudinal and topographical gradients in a species-rich Ecuadorian Montane rain forest. *Biotropica* 2010, 42, 140–148. [CrossRef]
- 85. Lines, E.R.; Zavala, M.A.; Purves, D.W.; Coomes, D.A. Predictable changes in aboveground allometry of trees along gradients of temperature, aridity and competition. *Glob. Ecol. Biogeogr.* **2012**, *21*, 1017–1028. [CrossRef]
- 86. Mitchell, S.J. Wind as a natural disturbance agent in forests: A synthesis. Forestry 2013, 86, 147–157. [CrossRef]
- 87. Qiu, H.; Liu, S.; Zhang, Y.; Li, J. Variation in height-diameter allometry of ponderosa pine along competition, climate, and species diversity gradients in the western United States. *For. Ecol. Manag.* **2021**, 497, 119477. [CrossRef]
- Liu, S.; Liu, Y.; Xia, R. Using random forest to disentangle the effects of environmental conditions on height-to-diameter ratio of Engelmann spruce. New For. 2023. (online first). [CrossRef]
- Černý, J.; Pokorný, R.; Vejpustková, M.; Šrámek, V.; Bednář, P. Air temperature is the main driving factor of radiation use efficiency and carbon storage of mature Norway spruce stands under global climate change. *Int. J. Biometeorol.* 2020, 64, 1599–1611. [CrossRef] [PubMed]
- 90. Orman, O.; Adamus, M.; Foremník, K. Norway spruce sapling plasticity in their response of architecture and growth to light gradient decreases with altitude in subalpine stands. *For. Ecol. Manag.* **2023**, *536*, 120898. [CrossRef]
- Bošel'a, M.; Máliš, F.; Kulla, L.; Šebeň, V.; Deckmyn, G. Ecologically based height growth model and derived raster maps of Norway spruce site index in the Western Carpathians. *Eur. J. For. Res.* 2013, 132, 691–705. [CrossRef]
- 92. Konôpka, J.; Konôpka, B. Static stability of spruce stands subjected to target trees method tending. Rep. For. Res. 2020, 65, 82–92.
- Pretzsch, H.; Biber, P.; Schütze, G.; Uhl, E.; Rötzer, T. Forest stand growth dynamics in Central Europe have accelerated since 1870. *Nat. Commun.* 2014, 5, 4967. [CrossRef] [PubMed]
- 94. Dobbertin, M. Influence of stand structure and site factors on wind damage comparing the storms Vivian and Lothar. *For. Snow Landsc. Res.* **2002**, *77*, 187–205.
- 95. Bouchard, M.; Pothier, D.; Ruel, J.-C. Stand-replacing windthrow in the boreal forests of eastern Quebec. *Can. J. For. Res.* 2009, 39, 481–487. [CrossRef]
- Zeide, B.; Vanderschaaf, C. The effect of density on the height-diameter relationship. In *Proceedings of the Eleventh Biennial Southern* Silvicultural Research Conference; Outcalt, K.W., Ed.; General Technical Report SRS-48; Department of Agriculture, Forest Service, Southern Research Station: Asheville, NC, USA, 2002; pp. 463–466.
- 97. Valinger, E.; Fridman, J. Factors affecting the probability of windthrow at stand level as a result of Gudrun winter storm in southern Sweden. *For. Ecol. Manag.* 2011, 262, 398–403. [CrossRef]
- 98. Brüchert, F.; Gardiner, B. The effect of wind exposure on the tree aerial architecture and biomechanics of Sitka spruce (*Picea sitchensis, Pinaceae*). *Am. J. Bot.* **2006**, *93*, 1512–1521. [CrossRef] [PubMed]
- 99. Bianchi, S.; Huuskonen, S.; Hynynen, J.; Siipilehto, J.; Niemistö, P. Tree-level differences in Norway spruce and Scots pine growth after extreme thinning treatments. *Scand. J. For. Res.* 2022, *31*, 109–118. [CrossRef]
- Yousefpour, R.; Hanewinkel, M.; Le Moguédec, G. Evaluating the stability of management strategies of pure Norway spruce forests in the Black Forest area of southwest Germany for adaptation to or mitigation of climate change. *Environ. Manag.* 2010, 45, 387–402. [CrossRef]
- 101. Pasztor, F.; Matulla, C.; Zuvela-Aloise, M.; Rammer, W.; Lexer, M.J. Developing predictive models of wind damage in Austrian forests. *Ann. For. Sci.* 2014, 72, 289–301. [CrossRef]
- 102. Pretzsch, H.; Steckel, M.; Heym, M.; Biber, P.; Ammer, C.; Ehbrecht, M.; Bielak, K.; Bravo, F.; Ordóñez, C.; Collet, C.; et al. Stand growth and structure of mixed-species and monospecific stands of Scots pine (*Pinus sylvestris* L.) and oak (*Q. robur* L.; *Quercus petraea* (Matt.) Liebl.) analysed along a productivity gradient through Europe. Eur. J. For. Res. 2020, 139, 349–367. [CrossRef]
- 103. Vospernik, S.; Heym, M.; Pretzsch, H.; Pach, M.; Steckel, M.; Aldea, J.; Brazaitis, G.; Bravo-Oviedo, A.; del Río, M.; Löf, M.; et al. Tree species growth response to climate in mixtures of *Quercus robur/Quercus petraea* and *Pinus sylvestris* across Europe—A dynamic, sensitive equilibrium. *For. Ecol. Manag.* 2023, 530, 120753. [CrossRef]
- 104. del Río, M.; Pretzsch, H.; Ruiz-Peinado, R.; Jactel, H.; Coll, L.; Löf, M.; Aldea, J.; Ammer, C.; Avdagić, A.; Barbeito, I.; et al. Emerging stability of forest productivity by mixing two species buffers temperature destabilizing effect. *J. Appl. Ecol.* 2022, 59, 2730–2741. [CrossRef]

Disclaimer/Publisher's Note: The statements, opinions and data contained in all publications are solely those of the individual author(s) and contributor(s) and not of MDPI and/or the editor(s). MDPI and/or the editor(s) disclaim responsibility for any injury to people or property resulting from any ideas, methods, instructions or products referred to in the content.