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Dynamics of Soil Nutrients, Microbial Community Structure, Enzymatic Activity, and Their Relationships along a Chronosequence of *Pinus massoniana* Plantations

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Abstract: *Pinus massoniana* is the major afforestation and vegetation restoration tree in southern China, and it plays an important role in the sustainable development of plantations. However, long-term single planting of *P. massoniana* has resulted in the decline of soil quality and forest productivity, and a soil fertility assessment is urgently needed. We selected *P. massoniana* plantations of four age stages for plot investigation and sampling to determine the soil physicochemical properties, microbial diversity and composition, and enzyme activities at different soil depths. The results showed that soil total phosphorus (TP) and available phosphorus (AP) decreased with the increase of age, especially low C/N ratio and high C/P and N/P ratio in the 30-year and 36-year stands, leading to P limitation. Meanwhile, the bacterial Shannon index also decreased with the increase of age and was positively correlated with AP, NO₃⁻-N, and pH. However, the fungal Shannon index decreased first and then increased with the increase of age; soil acid phosphatase (S-ACP) and urease activities showed a similar trend. Correlation analysis demonstrated that the increase of total organic carbon (TOC) and total nitrogen (TN) promoted the increase of fungal Shannon index, which was beneficial to the secretion of more enzymes. We found that soil physicochemical properties, microbial diversity, and enzyme activity decreased simultaneously when soil depths increased. Moreover, Acidobacteria and Basidiomycota were the most abundant bacterial and fungal communities, respectively, followed by Proteobacteria and Actinobacteria for bacteria and Ascomycota for fungi, and these microbial taxa were significantly affected by soil water content (SWC), TOC, AP, and C/P. In conclusion, this work reveals the potential correlation among soil physicochemical properties, microbial diversity and composition, and enzyme activities, and revealed potential correlations among them which will help to improve understanding of soil conditions and provide a reference for rational management of soil resources.

Keywords: *Pinus massoniana* plantation; bacteria; fungi; stoichiometric; stand age; soil depth

1. Introduction

Pinus massoniana is the main tree species for afforestation and vegetation rehabilitation in Southern China because of its special advantages such as fast growth, high adaptability to drought and barren soils, and its capacity to retain water and nutrients. According to a national survey, the forest area of the dominant species of *P. massoniana* in China is about 10 million ha, accounting for 6.08% of the total forest area in the country and 20% of the plantation forest areas in Southern China [1,2]. Consequently, *P. massoniana* occupies an extremely important position in promoting wood production and sustainable development

of plantations. However, the long-term monoculture tree plantations leading to significant decline in soil quality and woodland productivity have been often reported by the previous researchers [3–5]. For example, the decline of soil quality leads to the loss of topsoil and associated nutrients, which changes the soil microbial community structure and enzymatic activity, even the instability of the forest ecosystem [3,6]. Therefore, it is of great significance to evaluate the soil quality of the plantation for rational management and utilization of soil resources, maintenance of soil fertility, and improvement of plantation productivity.

The growth and development of plants are strongly affected by soil nutrients, especially the lack of nitrogen (N) and phosphorus (P) in soil which may limit the growth of plants because their nutrients mainly come from soil [7]. Previous studies reported that soil carbon (C), N, and P stoichiometry (C/N, C/P, and N/P ratios) have been used as a comprehensive indicator of soil nutrient status and element balance [8,9]. Ren et al. [10] found that the imbalance of soil C, N, and P elements during afforestation led to P limitation. In addition, changes in soil physical properties affect other soil chemical and biological properties [11], e.g., soil water content (SWC) and pH can affect the absorption of other nutrients [12,13]. As such, they are used as indicators to monitor forest soil quantity.

As a crucial component of biodiversity in terrestrial ecosystems, soil microbial communities play an important role in promoting the decomposition of organic matter, nutrient cycling, and energy flow of forest ecosystems [3,14]. Changes in microbial community composition and diversity can reflect soil fertility status and have been used as an important indicator for assessing soil quality and soil disturbance intensity in many ecosystem processes [15,16]. As previously reported, soil microbes can promote nutrient cycling. In turn, soil physicochemical properties drive microbial community structure, thereby altering the availability, accumulation, and mineralization of nutrients [17]. The evidence has indicated that soil microbial community structure was affected by soil organic matter availability, pH, temperature, water content, and C/N ratio [18–20]. Meanwhile, the unbalanced soil C: N: P stoichiometry leads to soil N and P limitations, which may affect microbial composition and diversity [10]. Besides, the decomposition of litter and dead root by microbes results in a change in the C: N: P ratio and converges toward soil organic matter and microbial biomass. In turn, the changes in the C: N: P ratio affect the decomposition process and nutrient cycling [14,21]. Soil enzymes play a catalytic role in the material cycle and energy flow of the forest ecosystem, e.g., sucrase catalyzes sucrose hydrolysis to produce monosaccharides and increased soil soluble nutrients; urease is involved in nitrogen cycling and catalytic hydrolysis of urea to CO₂ and NH₃; phosphatase can hydrolyze phosphate ester and participate in phosphorus cycling and mineralization [9,22,23]. Thus, enzyme activity can objectively reflect soil fertility to a certain extent, (e.g., a decrease in soil enzyme activity would indicate a decline in soil quality) [6]. Previous studies have shown that the enzymes related to mineralization of C, N, and P are closely related to the C: N: P stoichiometry of soil, e.g., higher C, N, and P nutrient availability can inhibit enzyme activities [24,25]. Simultaneously, soil physicochemical properties such as pH, water content, and temperature also affect enzyme activity [25,26]. Generally, soil enzymes are mainly secreted from roots and microbes, so enzyme activity is also affected by microbial community changes [27]. For example, increasing the number of soil microbes can promote the activities of sucrase and urease [22,26]. Therefore, comprehensive evaluation of soil nutrients, microbial community structure, and soil enzyme activity is of great significance for rational management and utilization of soil resources, maintenance of soil fertility, and improvement of plantation productivity.

Previous evidence revealed the effects of different afforestation years on soil nutrients, microbial communities, and enzyme activities, and revealed the close relationship between them [6,9,28]. Zhang et al. [9] found that soil microbial biomass, enzyme activity, and C/P and N/P ratios increased synergistically with the planting years of the *Pinus tabulaeformis* plantation, and there was a significant correlation between soil organic carbon, total nitrogen, enzyme activity, and microbial biomass. Zhu et al. [6] showed that successive cultivation of *Eucalyptus urophylla* decreased soil bacterial community diversity, microbial

biomass, and enzyme activity, but increased fungal community diversity. Most of the above studies were performed in the topsoil, but the soil properties (physical, chemical, microbes, and enzyme) at different soil depths with afforestation years are still not completely understood, especially in successive cultivated plantations, and further research is needed. Moreover, Shen et al. [29] reported that different soil depths had significant effects on the numbers and activities of soil bacteria and fungi, and considered that the bacteria, fungi and enzyme activities of surface soil (0–20 cm) were more sensitive to nitrogen deposition.

Based on the existing research sites, we selected 10, 20, 30, and 36 years of *P. massoniana* plantations widely distributed in Central Guizhou, China as the research object. Soil samples were collected from different soil depths (0–20, 20–40, and 40–60 cm) to analyze soil physicochemical properties (total organic carbon (TOC), total nitrogen (TN), total phosphorus (TP), and C/N, C/P, and N/P ratios, ammonium nitrogen ($\text{NH}_4^+\text{-N}$), nitrate nitrogen ($\text{NO}_3^-\text{-N}$), available phosphorus (AP), pH, soil water content (SWC), soil temperature (ST), and soil electrical conductivity (SEC)), microbes (bacterial and fungal alpha diversity and composition), and enzyme activity (acid phosphatase (S-ACP), sucrase, and urease). We hypothesized that: (1) the available nutrients in the soil may be consumed with the increase of the chronosequence of *P. massoniana* plantations (especially as it may lead to more severe P limitation in the late stage of plantation under the condition of low P in subtropical regions [30]), the imbalance of element stoichiometry (e.g., C/N, C/P, and N/P ratios), and the variation of other physicochemical properties, which may lead to different trends of bacteria, fungi, and different enzymes; (2) soil nutrient composition and related physicochemical properties decreased with soil deepening, which may lead to a simultaneous decrease in soil bacterial and fungal diversity and enzyme activities, but the differences in the functions of specific bacterial and fungal taxa may cause different trends. The objectives of this study were to (1) elucidate the rules of the changes of soil physicochemical properties, bacterial and fungal alpha diversity and composition, and enzyme activities of *P. massoniana* plantation with planting years; (2) determine whether there are significant differences in soil physicochemical properties, bacterial and fungal alpha diversity and composition, and enzyme activities in different soil depths; (3) reveal the correlation among soil physicochemical properties, bacterial and fungal alpha diversity and composition, and enzyme activities. The related results are expected to provide theoretical reference for soil nutrient monitoring, management, and artificial intervention of *P. massoniana* plantations.

2. Materials and Methods

2.1. Study Region

This study was carried out in the Mengguan Forest Farm (26°21'54.6" N–26°22'30" N, 106°43'20.8" E–106°44'12" E) in Guiyang, Guizhou Province, China (Figure 1). The total area of the forest farm is 1143 ha; it belongs to the mountain geomorphic type in the northern subtropical area, with an altitude of about 1100–1300 m and a relative height difference of 100–200 m. Therefore, it also has a subtropical monsoon climate, with an annual mean temperature of 14.9 °C and annual mean relative humidity of 80%. The annual precipitation reaches 1120 mm and is primarily concentrated from May to August, accounting for about 60% of the annual precipitation. The soil of the forest farm belongs to the siliceous yellow soil developed from sand shale, which is acidic, has a pH value of 3.5–6.5, moderate fertility, and a soil depth of 60–80 cm. The virgin vegetation of the forest farm was *P. massoniana* and broadleaf tree (*Betula platyphylla* and *Myrica rubra*) sparse forest, which was established as a pure forest of *P. massoniana* in 1980s. We conducted a preliminary investigation of the *P. massoniana* plantation on 5–25 August 2017 and determined that four stands aged 10, 20, 30, and 36 years and established in 2007, 1997, 1987, and 1981, respectively, with a plant/row distance of 1 m × 2 m (each stand was consistent). According to the forestry industry standards of the People's Republic of China: Regulations for age-class and age-group division of main tree-species (LY/T 2908-2017), the above four stands represent the young forest, middle-aged forest, near-mature forest, and

mature forest, respectively [31]. After afforestation, the areas were irrigated, pruned, and managed only in the first three years. The first and the second thinning were conducted 10 and 20 years after establishment, respectively, with 40% and 30% of trees being removed, leaving a density of less than 1500 trees ha⁻¹. The main tree species were *P. massoniana*, and the shrub species primarily included *Myrsine africana*, *Viburnum dilatatum*, and *Vaccinium bracteatum*. The herbs were primarily *Dichanopteris dichotoma*, *Odontophoria chinensis*, and *Deyeuxia pyramidalis*.

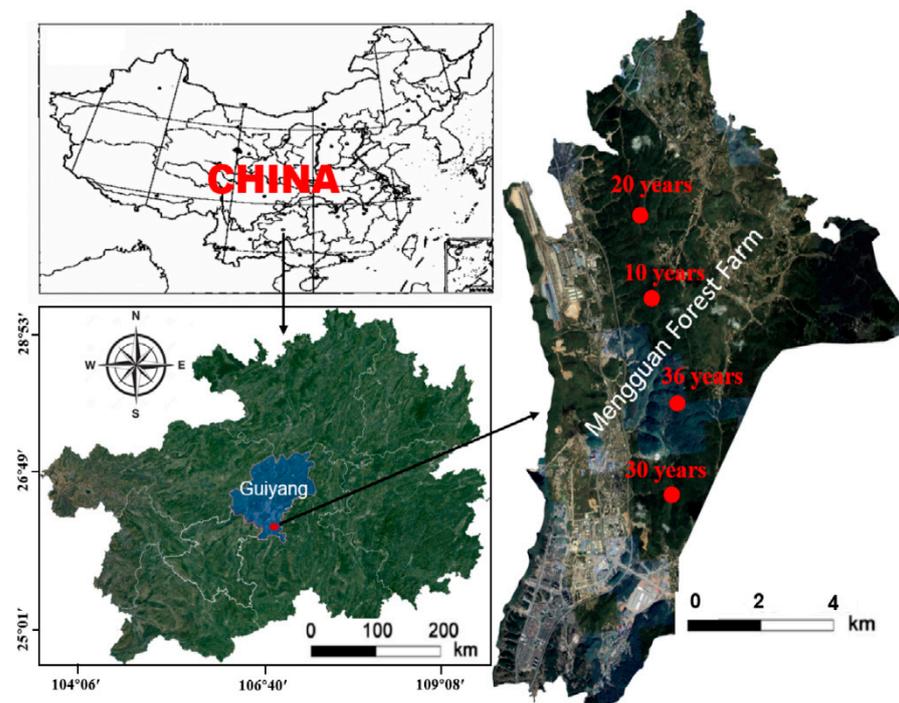


Figure 1. Location of the Mengguan Forest Farm in Guizhou of China and the sampling plots.

2.2. Experimental Design

The sample plot setting was carried out on 20–25 August 2017. The four stands had similar environmental characteristics, such as altitude, slope aspect, slope, slope position, and soil texture. Three 20 m × 20 m typical plots were randomly set in different age stands (12 sample plots in total). The diameter at breast height (DBH), tree height, height under the branch, canopy density, altitude, slope aspect, slope, slope position, and soil parameters of *P. massoniana* in the sample plot were measured (Table 1).

Table 1. The basic situation of different stand ages in *Pinus massoniana* plantations.

Parameter List	10 Years	20 Years	30 Years	36 Years
Altitude (m)	1194	1175	1206	1214
Slope aspect	N/EN	WN	N	EN
Slope (°)	5	6	5	8
Slope position	Downhill	Downhill	Downhill/Middle slope	Middle slope
Soil texture	Sandy loam	Sandy loam	Sandy loam	Sandy loam
Soil parent rock	Sandy shale	Sandy shale	Sandy shale	Sandy shale
Forest stand canopy density	0.90 ± 0.07 a	0.85 ± 0.10 ab	0.75 ± 0.09 b	0.80 ± 0.08 ab
Forest stand density (trees ha ⁻¹)	4675 ± 256 a	2812 ± 135 b	1356 ± 103 c	1083 ± 94 d
Average DBH (cm)	8.53 ± 1.56 d	12.66 ± 2.55 c	18.46 ± 3.01 b	21.26 ± 5.26 a
Average tree height (m)	7.43 ± 1.83 d	15.64 ± 2.74 c	17.94 ± 2.43 b	20.87 ± 3.26 a

N, E, and W mean north, east, and west direction, respectively. DBH: the diameter at breast height. The values show means ± SE. Lowercase letters indicate significant difference among different stand ages ($p < 0.05$).

2.3. Soil Samples Collection

Soil sample collection was carried out on 20 June 2019. Five sampling points were set on the diagonal of each 20 m × 20 m plot to collect soil samples at depths of 0–20, 20–40, and 40–60 cm, respectively. Then, the debris such as stones, litter, and roots were removed, and the five different sampling points with the same soil depth were mixed. Each plot obtains representative samples of three soil depths (using the quartering method, taking 500 g soil for each soil depth). A total of 36 soil samples (4 ages × 3 plots × 3 soil depths) were obtained in this study. Each soil sample was divided into two parts. One-part samples were placed in sterile bags, and these samples were cooled immediately until arrival at the laboratory, and then they were subsampled and frozen at $-80\text{ }^{\circ}\text{C}$ as soon as possible until they were used to extract soil microbial DNA. The other part was used to determine the physicochemical properties and enzyme activities of soil.

2.4. Soil Physicochemical Properties and Enzyme Activity Measurements

The collected soil samples were air dried and crushed in the laboratory to determine the soil physicochemical properties and enzyme activities. The total organic carbon (TOC) concentration of soil was determined using the $\text{K}_2\text{Cr}_2\text{O}_7$ oxidation method. After digestion with $\text{H}_2\text{SO}_4\text{-H}_2\text{O}_2$, the concentrations of total nitrogen (TN) and total phosphorus (TP) were determined by the Kjeldahl and Mo-Sb colorimetric assay methods, respectively [32]. Soil ammonium ($\text{NH}_4^+\text{-N}$) was extracted by 2 mol L^{-1} KCl and determined by the Indigo phenol blue method [33]. Nitrate nitrogen ($\text{NO}_3^-\text{-N}$) was determined by ultraviolet spectrophotometry [34]. Available phosphorus (AP) was extracted by NH_4F , followed by the colorimetric method [35]. The pH value of soil was determined by the glass composite electrode method (water soil ratio of 2.5: 1) [20]. Soil water content (SWC), temperature (ST), and electrical conductivity (SEC) were measured by TDR350 soil moisture, temperature, and electrical conductivity meter. Soil acid phosphatase (S-ACP) was determined by diphenyl sodium phosphate colorimetry, sucrose by 3,5-Dinitrosalicylic acid colorimetry, and urease by indophenol colorimetry [36].

2.5. Determination of Soil Bacteria and Fungi

2.5.1. DNA Extraction, PCR Amplification, and Illumina Sequencing

Soil DNA extraction, PCR amplification, and Illumina sequencing were completed by Hangzhou Kaitai Biotechnology Co., Ltd. (Hangzhou, China). The soil sample weighed 0.25 g. FastDNA[®] Spin Kit (MP Biomedicals, Santa Ana, CA, USA) was used to extract the total DNA in the soil sample according to the manufacturer's instructions. The DNA concentration and purity were detected using the Thermo NanoDrop 2000 UV spectrophotometer and 1% agarose gel electrophoresis. The tested samples were stored at $-80\text{ }^{\circ}\text{C}$ for further analysis. The v3 + v4 region of the bacterial 16S rDNA gene and V4 region of the fungal 18S rDNA gene were amplified by StepOne Plus fluorescent quantitative PCR. The bacterial primers were 341F (5'-CCTACGGGNGGCWGCAG-3') and 805R (5'-GGACTACHVGGGTWTCTAA-3') [37]. The fungal primers were EF4F (5'-GGAAGGGRTGTATTATTAG-3') and EF3R (5'-TCCTCTAAATGACCAAGTTTG-3') [38]. The PCR reaction was performed in a 20.0 μL mixture, which contained 1.0 μL of DNA template (10 ng/ μL), 0.4 μL of forward primer, 0.4 μL of reverse primer, 0.4 μL of ROX Reference Dye, 10.0 μL of 2 × SYBR[®]Premix EX Taq[™], and 7.8 μL of water without RNA enzyme (Rnase-Free Water). The PCR amplification procedure was as follows: pre-denaturation at $95\text{ }^{\circ}\text{C}$, 3 min, 1 cycle; denaturation at $95\text{ }^{\circ}\text{C}$, 30 s, 25 cycles; annealing at $55\text{ }^{\circ}\text{C}$, 30 s; extension at $72\text{ }^{\circ}\text{C}$, 30 s; final extension at $72\text{ }^{\circ}\text{C}$, 5 min, 1 cycle; and $4\text{ }^{\circ}\text{C}$ storage. Library quality inspection was performed by Thermo NanoDrop 2000 UV spectrophotometer and 2% agarose gel electrophoresis. After the quality inspection was qualified, the library was quantified by KAPA Library Quantification Kit Illumina[®] Platforms (KAPA Biosystems, Inc., Wilmington, NC, USA). Finally, an equal amount of PCR product from each sample was put into the single tube, and then high-throughput sequencing was performed using Illumina Miseq platform of Hangzhou Kaitai Biotechnology Co., Ltd. (Hangzhou, China).

2.5.2. Sequencing Data Processing

The two ends of the library were sequenced on the Illumina MiSeq. Each sample generated two files (reads1 and reads2), which were stored in a low-quality fastq format. Conducting quality control and filtering on the original offline data is necessary to obtain high-quality sequencing data and improve the accuracy of biological information analysis [39]. The main steps were as follows: (1) the fastq_mergepairs command of Vsearch software was used to assemble the paired reads from double terminal sequencing into a single sequence; (2) Cutadapt software was used to remove PCR primers from the sequence; (3) the fastq_filter command Vsearch software was used to remove the base sequence containing N (fuzzy base), the same base number of more than 8, and length of less than 100 bp.

2.5.3. Operational Taxonomic Unit (OTU) Clustering and Notes

To reduce the amount of redundant calculation in the process of analysis, we counted the repetitive sequence and extracted the non-repetitive sequence. We used Vsearch software to remove the once-detected sequences (Singleton sequences). The sequences with more than 97% similarity were gathered into one operational taxonomic unit (OTU), the chimeric sequence was removed, and the OTU representative sequence was used for subsequent species annotation. We used the Ribosomal Database Project (RDP) [40] and SLIVA databases [41] to classify and identify bacteria and fungi, respectively. The raw sequencing reads are available at the Sequence Read Archive (SRA) database of the National Center for Biotechnology Information (NCBI) under accession numbers SRP291512: PRJNA675184.

2.6. Statistical Analysis

Based on the OTU information of soil samples, the alpha diversity (Shannon index, H') of bacteria and fungi in soil samples was calculated by using the software of mothur (version v.1.30.1) [10]. The Shannon index was used to estimate the microbial diversity in the soil samples, and the larger the Shannon index, the higher the diversity. Two-way ANOVA was used to examine the effects of stand age and soil depth on soil microbial community composition and diversity and elemental stoichiometry. Significant differences were measured by one-way ANOVA followed by an LSD test and significance was determined at the 0.05 level. Pearson correlation was used to determine the relationship between soil microbial diversity, enzyme activity, and soil physicochemical properties. These statistical analyses were carried out on SPSS 21.0 software (SPSS, Somers, NY, USA) and the figures were mapped on Origin 2020 (Origin Lab Corp., Northampton, MA, USA). Redundancy analysis (RDA) was used to study the effect of soil element stoichiometric variables on soil microbial community composition and implemented on Canoco 5.0 software (Microcomputer Power, Inc., Ithaca, NY, USA). The formula of Shannon index (H') is as follows:

$$H' = - \sum_{i=1}^{S_{obs}} \frac{n_i}{N} \ln \frac{n_i}{N}$$

where S_{obs} = the number of OTU actually measured; n_i = the number of OTU containing i sequences; n = the number of all sequences.

3. Results

3.1. The Variation of Soil C: N: P Stoichiometry and Physicochemical Properties with Stand Age and Soil Depth

The two-way ANOVA showed that different stand ages and soil depths had significant effects on soil TOC, TN, and TP concentrations and their stoichiometry (Table 2, $p < 0.05$ or $p < 0.01$). TOC and TN concentrations in different soil depths showed different trends with the increase of stand age. That is, the 0–20 and 20–40 cm soil depths showed that 30-year stand was significantly higher than other stands, while the 40–60 cm soil depth showed that 20-year and 30-year stands were significantly lower than the other stands (Figure 2, $p < 0.05$). TP concentrations of 0–20, 20–40, 40–60 cm soil depths increased first, then

decreased with the increase of stand age and peaked in 20-year stand ($0.62 \pm 0.04 \text{ g kg}^{-1}$, $0.45 \pm 0.02 \text{ g kg}^{-1}$, and $0.40 \pm 0.03 \text{ g kg}^{-1}$, respectively). C/N ratio of 0–20 and 40–60 cm soil depths first increased, then decreased with the increase of stand age and peaked in 30-year stand (6.32 ± 0.14 and 7.65 ± 0.49 , respectively), while C/N ratio of 20–40 cm soil depth tended to decrease (Figure 2). The changing trend of C/P and N/P ratios in the three soil depths were opposite to TP concentration with the increase of stand age. Except for the C/N ratio, the soil nutrient concentration and their stoichiometry of 0–20 cm soil depth in four stands were significantly higher than other soil depths (Figure 2, $p < 0.05$).

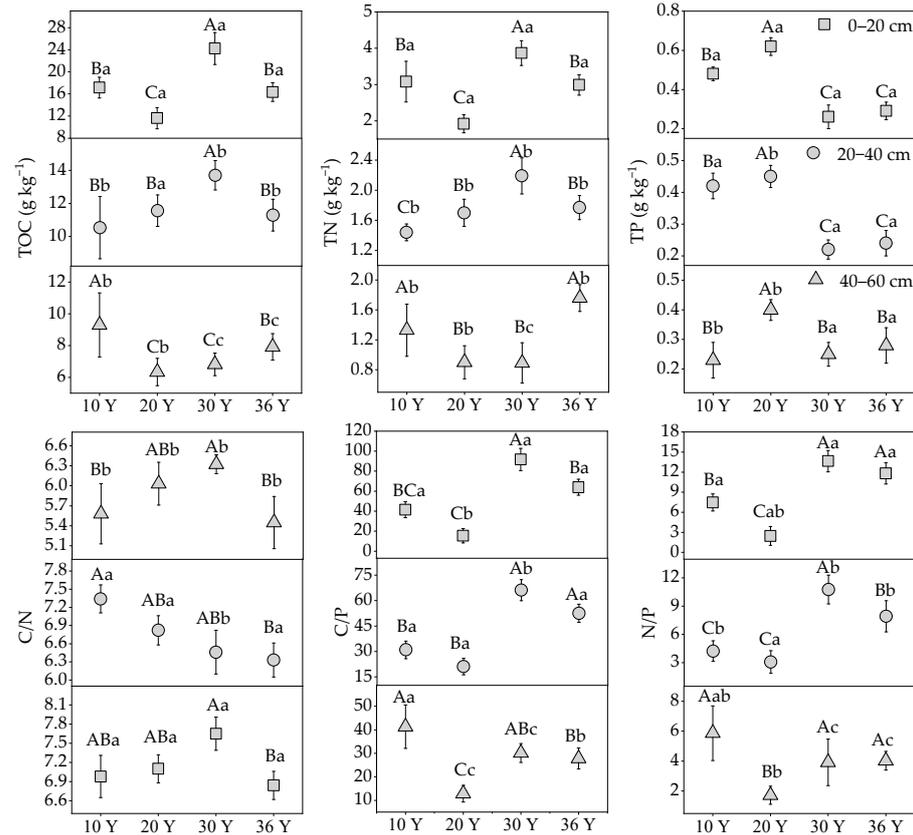


Figure 2. Soil total organic carbon (TOC), total nitrogen (TN), and total phosphorus (TP) concentrations and their stoichiometry of different stand ages and soil depths. 10 Y: 10-year stand; 20 Y: 20-year stand; 30 Y: 30-year stand; 36 Y: 36-year stand; 0–20 cm: 0–20 cm soil depth; 20–40 cm: 20–40 cm soil depth; 40–60 cm: 40–60 cm soil depth. The values show means \pm SE. Different capital and lowercase letters represent differences among different stand ages and soil depths, respectively ($p < 0.05$).

Different stand ages had significant effects on soil physicochemical properties ($p < 0.05$ or $p < 0.01$). Different soil depths had no significant effects on $\text{NH}_4^+\text{-N}$, $\text{NO}_3^-\text{-N}$, pH, and SEC ($p > 0.05$), but had significant effects on AP, SWC, and ST (Table 2, $p < 0.01$). $\text{NH}_4^+\text{-N}$ and ST of three soil depths tended to rise with the increase of stand age and reached the highest in 36-year stand (the mean values were $12.59 \pm 0.58 \text{ mg kg}^{-1}$ and $20.01 \pm 0.37 \text{ }^\circ\text{C}$, respectively). On the contrary, except for 40–60 cm soil depth, $\text{NO}_3^-\text{-N}$, AP, and pH tended to decrease with the increase of stand age, and the maximum appeared in 10-year stand (the mean values were 3.51 ± 1.56 and $7.15 \pm 2.06 \text{ mg kg}^{-1}$, and 4.60 ± 0.05 , respectively) (Figure 3). SWC and SEC showed irregular changes, and they decreased first and then increased in 0–20 cm soil depth with the increase of stand age, which was the lowest in 20 ($20.00 \pm 2.46\%$) and 30 years ($32.00 \pm 2.00 \text{ } \mu\text{s cm}^{-1}$), respectively. AP and ST of four stands decreased significantly with the increase of soil depth, while SWC increased significantly (Figure 3, $p < 0.05$).

Table 2. Effects of stand age (A), soil depth (D), and their interactions on soil physicochemical properties, enzyme activities, and the microbial diversity indicated by a two-way repeated measures ANOVA.

Variables	F (p) Value		
	A	D	A × D
TOC	22.857 (<0.01)	158.197 (<0.01)	14.171 (<0.01)
TN	8.122 (<0.01)	68.933 (<0.01)	5.745 (<0.01)
TP	10.625 (<0.01)	4.283 (0.026)	1.520 (0.214)
C/N	3.916 (0.021)	34.720 (<0.01)	1.525 (0.213)
C/P	81.256 (<0.01)	46.031 (<0.01)	17.106 (<0.01)
N/P	83.850 (<0.01)	74.682 (<0.01)	18.641 (<0.01)
pH	5.162 (<0.01)	2.950 (0.072)	2.548 (0.047)
AP	481.016 (<0.01)	223.540 (<0.01)	154.060 (<0.01)
NH ₄ ⁺ -N	7.864 (<0.01)	0.333 (0.720)	0.148 (0.988)
NO ₃ ⁻ -N	3.852 (0.022)	0.220 (0.804)	2.237 (0.074)
SWC	18.574 (<0.01)	138.262 (<0.01)	12.449 (<0.01)
SEC	38.255 (<0.01)	2.493 (0.104)	28.426 (<0.01)
ST	7.196 (<0.01)	12.471 (<0.01)	0.843 (0.550)
S-ACP	14.595 (<0.01)	111.858 (<0.01)	7.782 (<0.01)
Sucrase	5.564 (<0.01)	1.523 (0.238)	4.751 (<0.01)
Urease	18.137 (<0.01)	116.335 (<0.01)	22.932 (<0.01)
B-OTU	5.960 (<0.01)	0.661 (0.526)	1.667 (0.173)
B-Shannon	27.456 (<0.01)	11.752 (<0.01)	1.448 (0.238)
F-OTU	5.745 (<0.01)	7.696 (<0.01)	3.575 (0.011)
F-Shannon	28.070 (<0.01)	9.711 (<0.01)	8.495 (<0.01)

NH₄⁺-N: ammonium nitrogen; NO₃⁻-N: nitrate nitrogen; AP: available phosphorus; SWC: soil water content; SEC: soil electrical conductivity; ST: soil temperature; S-ACP: soil acid phosphatase; B-OTU: the bacterial operational taxonomic unit (OTU); B-Shannon: the bacterial Shannon index; F-OTU: the fungal OTU; F-Shannon: the fungal Shannon index. Bold fonts indicate significant (*p* < 0.05) or extreme significance (*p* < 0.01).

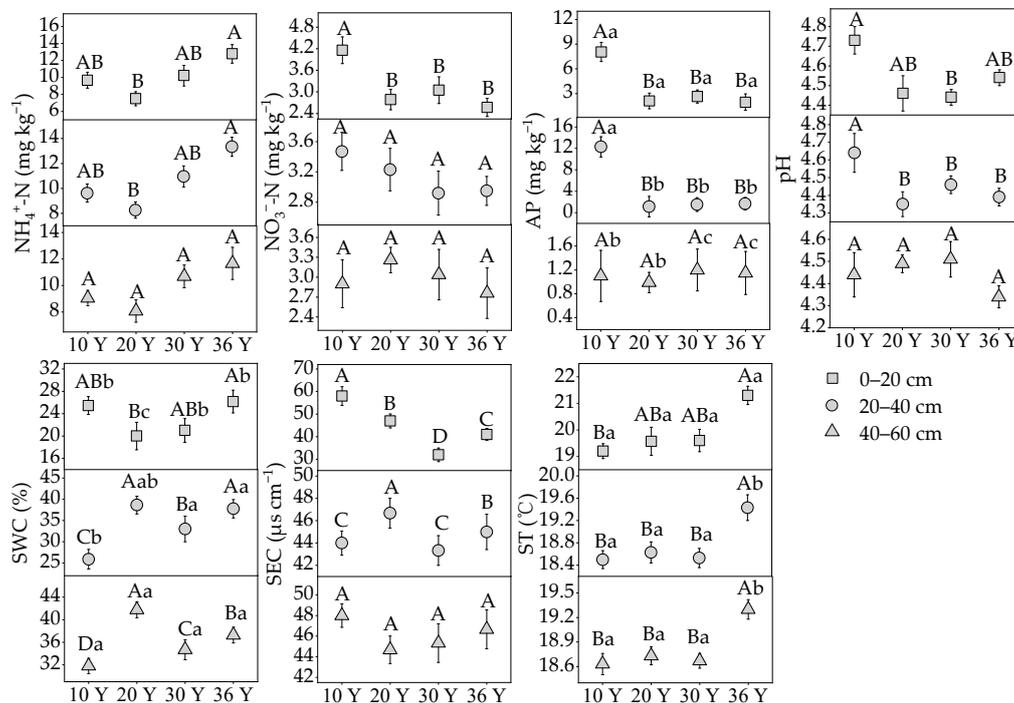


Figure 3. Soil physicochemical properties of different ages and soil depths. NH₄⁺-N: ammonium nitrogen; NO₃⁻-N: nitrate nitrogen; AP: available phosphorus; SWC: soil water content; SEC: soil electrical conductivity; ST: soil temperature. 10 Y: 10-year stand; 20 Y: 20-year stand; 30 Y: 30-year stand; 36 Y: 36-year stand; 0–20 cm: 0–20 cm soil depth; 20–40 cm: 20–40 cm soil depth; 40–60 cm: 40–60 cm soil depth. The values show means ± SE. Different capital and lowercase letters represent differences among different stand ages and soil depths, respectively (*p* < 0.05).

3.2. The Variation of Soil Enzyme Activities with Stand Age and Soil Depth

The two-way ANOVA showed that different stand ages had significant effects on S-ACP, sucrase, and urease ($p < 0.01$). Different soil depths had significant effects on soil acid phosphatase and urease ($p < 0.01$), while they had no significant effects on sucrase (Table 2, $p > 0.05$). On the whole, except for urease of 30-year stand at 40–60 cm, soil depth was significantly higher than other stands ($p < 0.05$); S-ACP and urease decreased first and then increased with the increase of stand age. S-ACP and urease of 30-year stand at 0–20 cm soil depth was significantly lower than other stands, and S-ACP of 20-year stand at 20–40 and 40–60 cm soil depths and urease of 20-year stand at 20–40 cm soil depth was significantly lower than other stands (Figure 4, $p < 0.05$). The sucrase of 0–20 and 20–40 cm soil depth tended to rise with the increase of stand age, but there was no significant difference between different stands at the 20–40 cm soil depth ($p > 0.05$). The sucrase of 20-year stand at 40–60 cm soil depth was significantly lower than other stands (Figure 4, $p < 0.05$). Moreover, S-ACP and urease of 40–60 cm soil depth in four stands were significantly higher than other depths ($p < 0.05$).

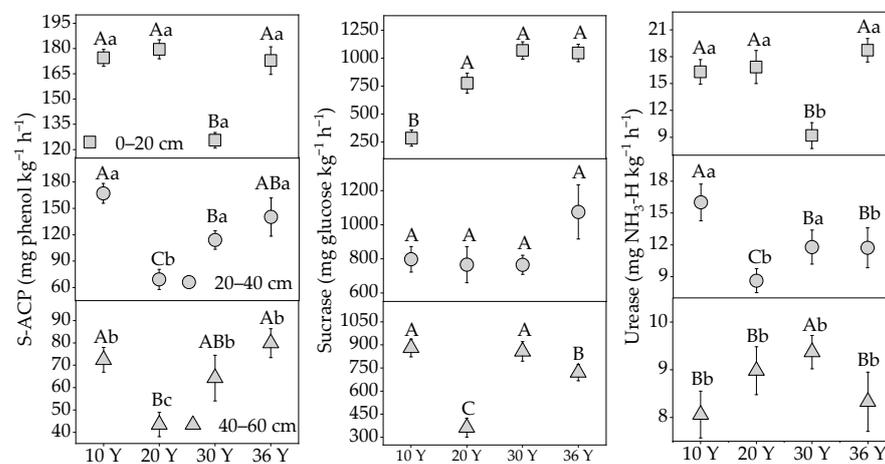


Figure 4. Soil enzyme activities of different ages and soil depths. S-ACP: soil acid phosphatase. 10 Y: 10-year stand; 20 Y: 20-year stand; 30 Y: 30-year stand; 36 Y: 36-year stand; 0–20 cm: 0–20 cm soil depth; 20–40 cm: 20–40 cm soil depth; 40–60 cm: 40–60 cm soil depth. The values show means \pm SE. Different capital and lowercase letters represent differences among different stand ages and soil depths, respectively ($p < 0.05$).

3.3. Alpha Diversity of Soil Bacteria and Fungi

The total quantity of bacterial and fungal OTU obtained from 36 samples was 78,132 and 13,230, respectively, and the number of bacterial and fungal OTU in each sample was 1780 to 2461 and 151 to 528, respectively (Figure 5). Based on 97% similarity, the rarefaction curves for bacteria and fungi tended to saturate for each sample (Figure S1), which showed that the sequencing data were reasonable, and the majority of sequence types could be detected. The two-way ANOVA showed that different stand ages had significant effects on bacterial and fungal OTU and Shannon index ($p < 0.05$). Different soil depths had significant effects on bacterial Shannon index and fungal OTU and Shannon index ($p < 0.05$), while they had no significant effects on bacterial OTU (Table 2, $p > 0.05$). The bacterial and fungal Shannon index had the same trend with OTU, that is, the bacterial alpha diversity decreased with the increase of stand age, while the fungal alpha diversity first decreased and then increased (Figure 5). The bacterial Shannon index of 0–20, 20–40, and 40–60 cm soil depths were the largest in 10-year stand, which were 6.32 ± 0.08 , 6.20 ± 0.10 , and 6.14 ± 0.03 , respectively. The fungal Shannon index of 20–40 and 40–60 cm depths were the smallest in the 30-year stand, which were 1.14 ± 0.18 and 0.95 ± 0.28 , respectively (Figure 5). Furthermore, the bacterial and fungal alpha diversity of 0–20 cm soil depth was significantly higher than that of the other two soil depths (Figure 5).

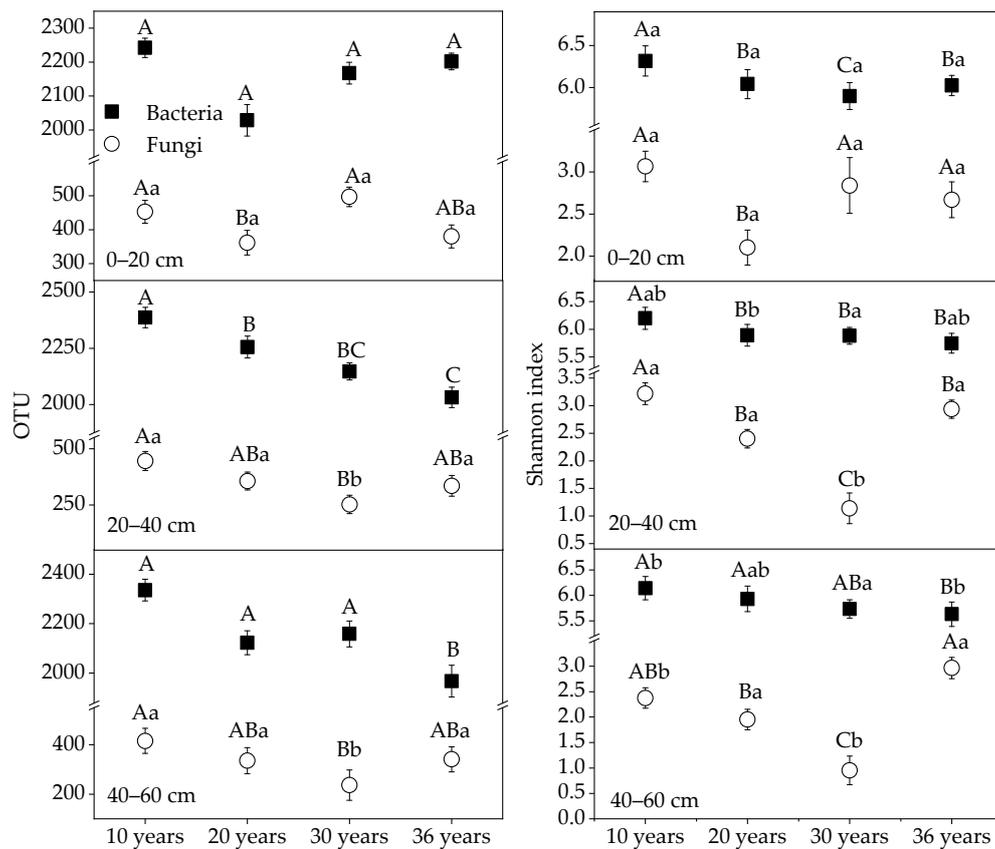


Figure 5. Alpha diversity of soil bacterial and fungal communities at different stand ages and soil depths. 0–20 cm: 0–20 cm soil depth; 20–40 cm: 20–40 cm soil depth; 40–60 cm: 40–60 cm soil depth. The values show means \pm SE. Different capital and lowercase letters represent differences among different stand ages and soil depths, respectively ($p < 0.05$).

3.4. Community Composition of Bacteria and Fungi

A total of 9 and 3 phyla of bacteria and fungi with a relative abundance of more than 1% were identified (phyla with more than 1% average abundance in at least one sample were considered as predominant) (Figure 6a,b). The average relative abundances of Acidobacteria (31.12%), Proteobacteria (16.33%), Actinobacteria (17.06%), Bacteria-unclassified (16.02%), and Chloroflexi (11.11%) were all higher than 10%, which were dominant in the soil bacterial community. The relative abundances of Planctomycetes, Verrucomicrobia, candidate division WPS-2, and candidate division WPS-1 were all higher than 1% (Figure 6a), and they were also important members of soil bacteria. The trend of different bacterial relative abundance with the increase of stand age and soil depth was different. Acidobacteria and Proteobacteria first decreased and then increased with the increase of stand age, while Actinobacteria and Bacteria-unclassified were the opposite. Acidobacteria showed no obvious rule with the increase of soil depth, Proteobacteria and Actinobacteria decreased, while Bacteria unclassified and Chloroflexi increased (Figure 6a). The average relative abundance of Basidiomycota (83.34%) in the soil fungal community was the highest, followed by Ascomycota (8.96%) and Eukaryota-unclassified (6.71%). The relative abundance of Basidiomycota increased with the increase of stand age and soil depth, but Ascomycota and Eukaryota-unclassified were the opposite (Figure 6b).

The heat map showed that bacterial and fungal communities as 20-year, 30-year, and 36-year stands were clustered together, clearly separated from the 10-year stand. The 20–40 cm and 40–60 cm soil depths were clustered together, clearly separated from 0–20 cm soil depth (Figure 7).

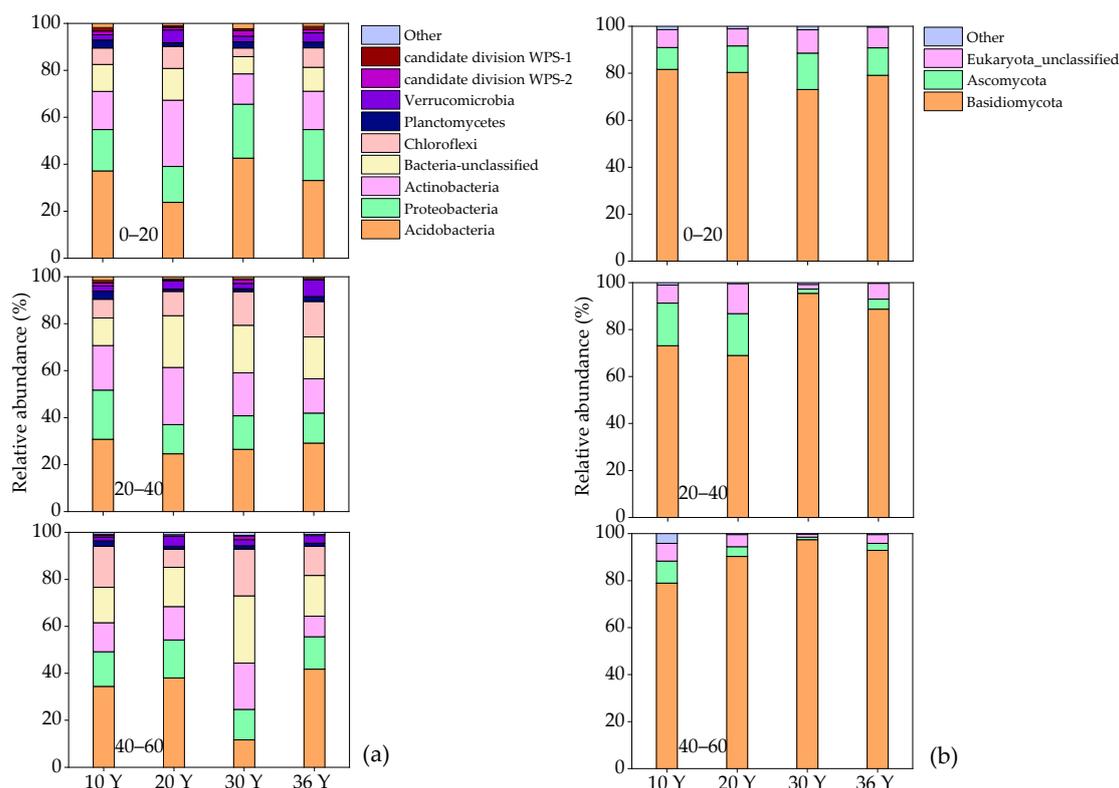


Figure 6. Relative abundance distribution of bacterial and fungal communities at the phylum level at different soil depths and stand ages. Phyla with more than 1% average abundance in at least one sample were considered as predominant. (a,b) show the variation of relative abundance of bacteria and fungi, respectively. Other: the taxa of bacteria or fungi with relative abundance less than 1%; 10 Y: 10-year stand; 20 Y: 20-year stand; 30 Y: 30-year stand; 36 Y: 36-year stand; 0–20 cm: 0–20 cm soil depth; 20–40 cm: 20–40 cm soil depth; 40–60 cm: 40–60 cm soil depth.

3.5. Relationship between Microbes, Enzyme Activity, and Soil Physicochemical Properties

Pearson correlation analysis showed that soil enzyme activities and the bacterial and fungal Shannon index were significantly correlated with soil physicochemical properties (Figure 8, $p < 0.05$ or $p < 0.01$). S-ACP was positively correlated with TOC, TN and N/P ratios, pH, AP, and ST, as well as negatively correlated with C/N ratio and SWC; sucrase was positively correlated with C/P and N/P ratios, ST, and negatively correlated with SC; urease was positively correlated with TN, pH, AP, and ST, and negatively correlated with C/N ratio (Figure 8). Bacterial Shannon index was positively correlated with pH, NO_3^- -N, SC, and negatively correlated with NH_4^+ -N, and SWC. Fungal Shannon index was positively correlated with TOC, TN, AP, and ST. Besides, bacterial and fungal Shannon index were significantly positively correlated with S-ACP and urease ($p < 0.01$), but not significantly correlated with sucrase ($p > 0.05$) (Figure 8). In conclusion, S-ACP, urease, bacterial, and fungal Shannon index were significantly positively correlated with AP ($p < 0.01$), but negatively correlated with SWC.

Redundancy analysis was conducted with the abundance of soil bacteria and fungi as response variables and soil physicochemical properties as explanatory variables (environmental variables). The results showed that the first two axes of the RDA plot explained 40.56% and 9.36% (49.92% in total) of the total microbial variation (Figure 9). A Monte Carlo permutation test showed that the SWC, TOC, AP, and C/P ratio were significantly correlated ($p < 0.05$) with soil microbial community composition in all environmental variables considered. In particular, the SWC was the first factor that contributed to 28.2% of total variation (Table 3). Figure 9 show that SWC was positively correlated with Planctomycetes, Chloroflexi, Verrucomicrobia, and Basidiomycota, while negatively correlated with Acidobacteria, Actinobacteria, Proteobacteria, Bacteria-unclassified, candidate division WPS-1,

candidate division WPS-2, Ascomycota, and Eukaryota-unclassified. Conversely, TOC, AP, and C/P ratio were negatively correlated with Planctomycetes, Chloroflexi, Verrucomicrobia, and Basidiomycota, while positively correlated with Acidobacteria, Bacteria-unclassified, Actinobacteria, Proteobacteria, candidate division WPS-1, candidate division WPS-2, Ascomycota, and Eukaryota-unclassified.

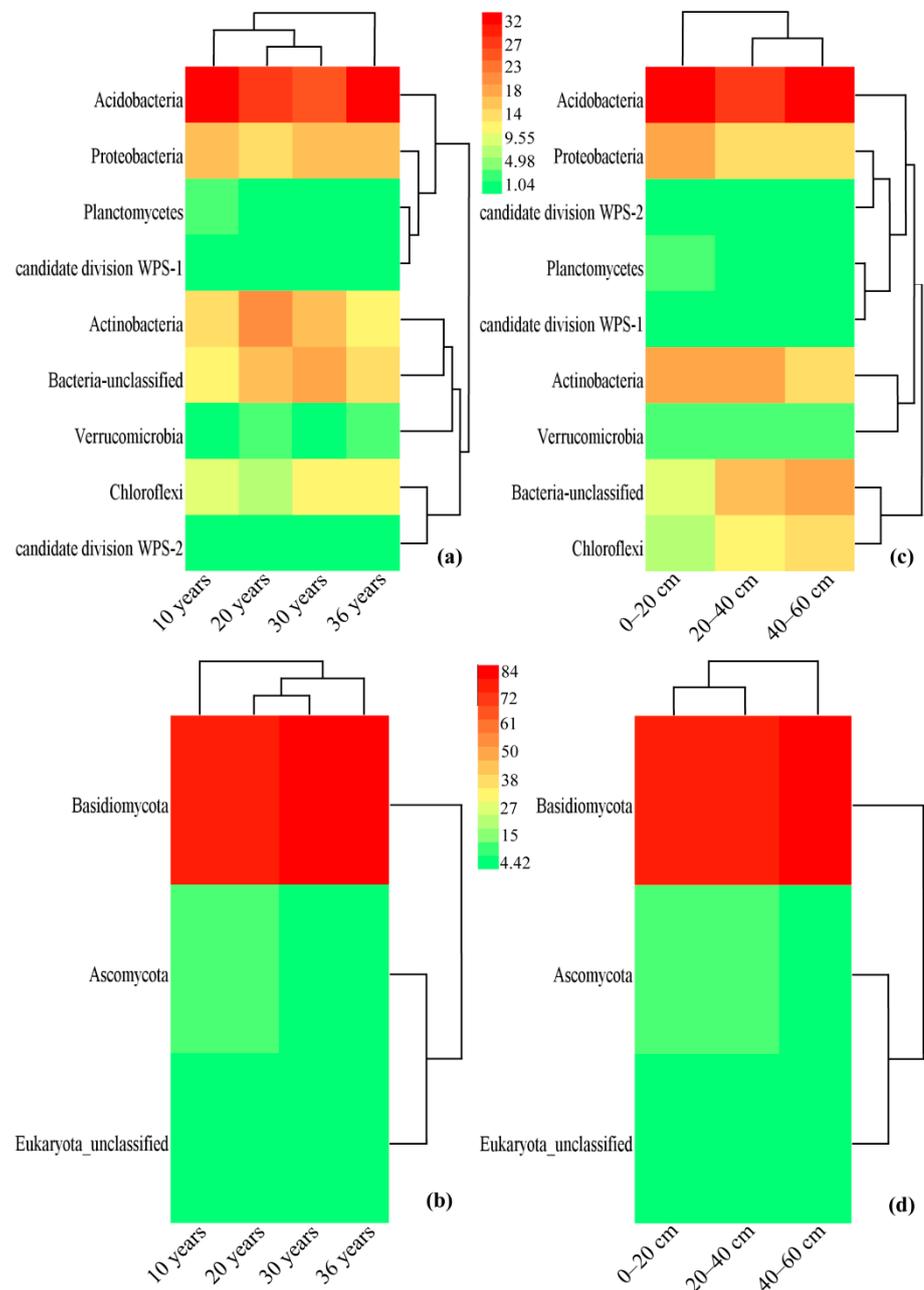


Figure 7. Heat map depicting the hierarchy cluster results for the abundance of bacteria and fungi at the phylum level. Phyla with more than 1% average abundance in at least one sample were considered as predominant. (a,b) show the clustering of bacteria and fungi at different ages, respectively. (c,d) show the clustering of bacteria and fungi at different soil depths, respectively. Red indicates high relative abundance and green indicates low relative abundance. 0–20 cm: 0–20 cm soil depth; 20–40 cm: 20–40 cm soil depth; 40–60 cm: 40–60 cm soil depth.

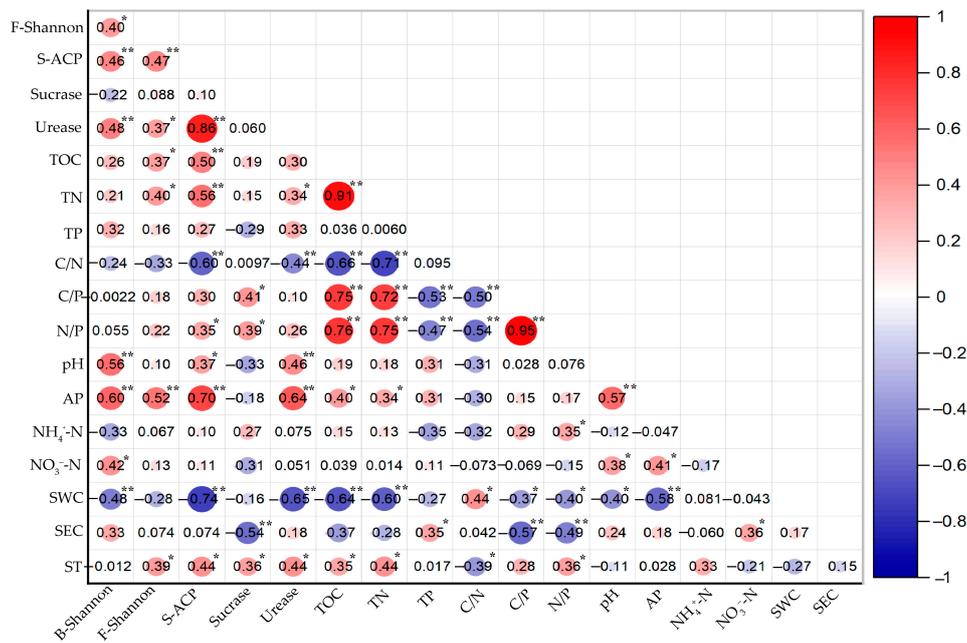


Figure 8. Pearson correlation coefficients between the soil physicochemical properties, the enzyme activities, and the microbial alpha diversity. B-Shannon index: the bacterial Shannon index; F-Shannon index: the fungal Shannon index; NH₄⁺-N: ammonium nitrogen; NO₃⁻-N: nitrate nitrogen; AP: available phosphorus. The value shows the correlation coefficient, positive values represent positive correlation and negative values represent negative correlation. * $p < 0.05$, ** $p < 0.01$.

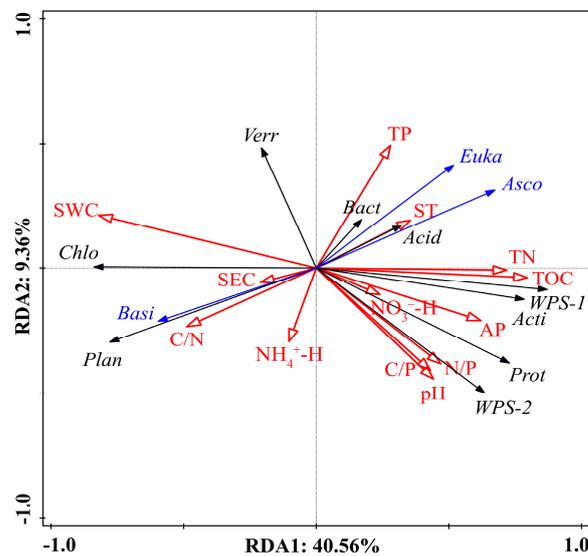


Figure 9. Ordination plots of the redundancy analysis (RDA) to determine the relationship between bacterial (black arrow) and fungal (blue arrow) phylum with relative abundance greater than 1% and soil physicochemical properties (red arrow). NH₄⁺-N: ammonium nitrogen; NO₃⁻-N: nitrate nitrogen; AP: available phosphorus; SWC: soil water content; SEC: soil electrical conductivity; ST: soil temperature; S-ACP: soil acid phosphatase; Acidobacteria (*Acid*), Proteobacteria (*Prot*), Actinobacteria (*Acti*), Bacteria-unclassified (*Bact*), Chloroflexi (*Chlo*), Planctomycetes (*Plan*), Verrucomicrobia (*Verr*), candidate division WPS-1 (*WPS-1*), candidate division WPS-2 (*WPS-2*), Basidiomycota (*Basi*), Ascomycota (*Asco*), Eukaryota-unclassified (*Euka*).

Table 3. Conditional effects of soil properties on soil microbial community composition obtained from the summary of forward selection in the redundancy analysis (RDA).

Variables	Explains (%)	F-Value	p-Value
SWC	28.2	13.4	0.002 **
TOC	7	3.6	0.012 *
AP	6.2	3.4	0.016 *
C/P	5.7	3.4	0.014 *
ST	3.1	1.9	0.082
SEC	3	1.8	0.11
TN	2	1.2	0.288
NH ₄ ⁺ -N	1.9	1.2	0.338
C/N	1.7	1.1	0.378
pH	1.8	1.1	0.344
TP	1.2	0.8	0.518
NO ₃ ⁻ -N	0.7	0.5	0.798
N/P	0.7	0.4	0.834

Explains(%): the explanation rate of single variable to total variable in soil. NH₄⁺-N: ammonium nitrogen; NO₃⁻-N: nitrate nitrogen; AP: available phosphorus; AK: available potassium. * $p < 0.05$, ** $p < 0.01$.

4. Discussion

4.1. Effects of Soil Physicochemical Properties on Microbial Alpha Diversity

The results were in line with our first expectation that soil bacterial and fungal alpha diversity (Shannon index) showed different trends with the increase of the chronosequence of *P. massoniana* plantations (Figure 5), and the bacterial Shannon index was significantly correlated with soil AP, NO₃⁻-N, pH, and SWC ($p < 0.05$), while fungal Shannon index was significantly correlated with TOC, TN, AP, and ST ($p < 0.05$) (Figure 8). Previous studies have shown that soil physicochemical properties, root biomass, and chemical composition change constantly with the continuous development of forests, which inevitably leads to changes in the diversity and composition of the soil microbial community [6,21,42]. Our research revealed that the bacterial Shannon index tended to decline with the increasing age of the *P. massoniana* plantation (Figure 5), and a similar result has been reported in other studies [6,42]. In our study, the decline of bacterial Shannon index could be attributed to the gradual decrease of soil nutrient availability, especially AP and NO₃⁻-N during the aging process of the *P. massoniana* plantation. As N and P were essential nutrients for microbial growth, the change of their effectiveness affected microbial composition and diversity [43]. As we knew, P was a necessary component in the production of the ribosome, ATP, DNA, and RNA to maintain rapid growth [44]. In 0–20 cm and 20–40 cm soil depths, the TP concentration in near-mature and mature forests was significantly lower than that in young and middle forests, while C/P and N/P ratios were just the opposite, which may lead to P limitation and thus inhibit bacterial growth (Figure 2). Moreover, fungi are more likely to thrive at a lower pH, while bacteria occupy the dominant position at a higher pH [45]. Our results showed that soil pH was significantly positively correlated with bacterial Shannon index (Figure 9), indicating that the decrease of pH in the near-mature and the mature forest was also the reason for the decrease of bacterial diversity. Generally, excessive water reduces soil porosity and restricts oxygen input and CO₂ emission in the process of aerobic respiration, and thus affects aerobic microbial activities [46]. Hence, SWC increases with the age of the *P. massoniana* plantation, which may inhibit aerobic bacterial growth and affect its diversity. However, we found that the fungal Shannon index decreased first and then increased with the increase of the *P. massoniana* plantation age (Figure 5). This result indicates that the soil condition may be more helpful to the growth and reproduction of fungi with the increase of *P. massoniana* plantation stand age. It is generally believed that the availability and quality of soil C and nutrients regulate microbial biomass and activity, and the C/nutrients ratio can reflect the quality of soil microbial growth environment to some extent [20]. Correlation analysis showed that fungal Shannon index was significantly positively correlated with soil TOC and TN (Figure 8, $p < 0.05$), which confirmed the above conclusion. In the present study, the increase of TOC and TN in the near-mature and mature forests provided carbon sources and nutrients for fungi and

promoted the increase of fungal diversity. On the other hand, the needles of *P. massoniana* contain refractory components (cellulose, lignin, tannin, etc.) and fungi can decompose the lignocellulosic matrix which other microbes cannot use to obtain more nutrients [47,48]. Thus, the increase of fungal alpha diversity may be related to the increase of litters of *P. massoniana*. Kubartova et al. [49] studied the fungal diversity of litter decomposition in temperate forests of southern France and found that the decomposition rate and fungal richness of the mixed litter were significantly higher than that of beech and spruce litters in the late stage of decomposition (24 months). This indicates that the mixed litter can provide diverse niches for microbes. Consequently, *P. massoniana* coniferous broadleaved mixed forest should be more conducive to the microbial decomposition of litters and improve soil fertility.

Different soil depths had significant effects on bacterial and fungal Shannon index (Table 2), and they decreased significantly with the increase of soil depth (Figure 5, $p < 0.05$), which agreed with our second hypothesis. Du et al. [50] studied variations in bacterial and fungal communities through soil depth profiles in a *Betula albosinensis* forest and similar results were obtained. The main reason is that the soil organic C, N, and AP decreased significantly with the increase of soil depth (Figures 2 and 3), and the direct carbon sources and nutrients required by soil bacteria and fungi were limited, resulting in the decrease of their diversity. Secondly, SWC increases with the increase of soil depth, which may change soil structure, affect soil respiration, and inhibit aerobic microbial activities, causing their diversity to decrease [46]. In addition, the input of litter and roots on the soil surface can provide more abundant carbon and nitrogen for the surface soil to increase the microbial diversity [14,51]. However, the nutrients of the deep soil depth are derived from the combinations of allocation of root secretions, translocations of mineral nutrients from top surface by leaching, and weathering of minerals [52]. We also found that the soil C/P and N/P ratios significantly decreased with the increase of soil depth, while the C/N ratio significantly increased (Figure 2), indicating that the deep soil microbial diversity may be limited by N and the surface soil (0–20 cm) microbial diversity may be limited by P [29]. With no P fertilizer application, the P source in the 0–20 cm soil is primarily from weathering soil containing P minerals. Even though the P content in soil minerals of the *P. massoniana* forest in this study area is high, the rate of P released from the weathering process is less than the rate of P demand by both trees and the microbial. In addition, phosphorus can easily be fixed by Al and Fe when soil pH is as low as in our studied sites (<4.8 , Figure 3). Thus, more N fertilizer should be applied to the deep soil and more P fertilizers should be applied to the topsoil to promote the absorption of nutrients by plants and maintain microbial diversity.

4.2. Effects of Soil Physicochemical Properties on Microbial Composition

Cluster analysis showed that the bacteria and fungi were clearly separated at 10-year stand from the others, and they were clearly separated at 0–20 cm soil depth from the other soil depths (Figure 7). It indicates that the composition of soil microbial communities varied greatly with the change of stand age and soil depth. Our two hypotheses have been confirmed once again. RDA analysis indicated that SWC, TOC, AP, and C/P ratio were the main factors affecting the microbial community (Figure 9 and Table 3), and different bacterial and fungal taxa had different responses to these soil physicochemical properties, which supported our hypothesis. Specifically, the abundance of Acidobacteria was the highest among all soil depths and stands, followed by Proteobacteria and Actinobacteria (Figure 6a). The same results were found in *B. albosinensis* forest by Du et al. [50]. However, Acidobacteria were positively correlated with soil TOC, AP, and pH in our results (Figure 9), which were different from most previous studies. Zhang et al. [53] studied the forest soil in Shennongjia and found that Acidobacteria_GP1, Acidobacteria_GP2, and Acidobacteria_GP3 were significantly positively correlated with soil pH, and other subgroups (excluding Acidobacteria_GP7) and were significantly negatively correlated with pH. Fierer et al. [54] found that the relative abundance of Acidobacteria and soil TOC and

C/N ratio were significantly positively correlated. The above showed that different sub-populations of Acidobacteria had different adaptability to soil physicochemical properties, and not all Acidobacteria were acidophilic and oligotrophic. Proteobacteria and Actinobacteria decreased with soil deepening and were significantly positively correlated with soil TOC and AP (Figures 7a and 9), which was consistent with the changes in soil bacteria in the *B. platyphylla* forest [50]. In general, the distribution of Proteobacteria is affected by the content of organic carbon, which is related to the high utilization rate of carbon [54]. More importantly, Proteobacteria preferred the high P, because its rapid growth requires a large number of RNA to meet the rapid cell division [55]. Previous studies showed that Actinobacteria is a key litter decomposer in soil, and part species of this phylum can secrete extracellular enzymes to promote the degradation of lignin, polysaccharides, and phenolics [17,56]. Consequently, Actinobacteria in topsoil can promote nutrient cycling by decomposing litter. SWC was negatively correlated with Acidobacteria, Proteobacteria, and Actinobacteria, and was the main limiting factor (the explanation rate was 28.2%) (Figure 9 and Table 3). As mentioned earlier, SWC mediates microbial activity by affecting soil aeration and nutrient availability, so high SWC may not be beneficial for microbial growth [13,46].

For fungal communities, Basidiomycota was the most abundant in all soil layers and stands, followed by Ascomycota and Eukaryota-unclassified (Figure 6b). We observed that the abundance of Basidiomycota increased as the soil depths increased, and similar results were reported in previous studies [50,57]. In this study, soil nutrients (e.g., TOC, TN, AP) tended to decrease with the deepening of soil, and they were negatively correlated with the abundance of Basidiomycota (Figures 2, 3 and 9). Therefore, the change of Basidiomycota abundance may possibly be attributed to other causes. In most cases, the majority genera of Basidiomycota formed a symbiotic relationship with plant roots, enabling them to obtain carbon sources directly from the roots [57]. Cheng et al. [58] found that the *P. massoniana* roots of 2–10 mm were mainly distributed in the 20–60 cm depth range. Thus, the Basidiomycota in deep soil (40–60 cm) can form mycorrhiza with *P. massoniana* roots, providing carbon and nutrients, which makes the abundance of Basidiomycota higher. Usually, Ascomycota is mostly saprophytes, which decompose litter to obtain nutrients, while only a few of them can form mycorrhiza with plant roots, so Ascomycota is mainly distributed in surface soil [59]. Moreover, the soil nutrient content decreased and SWC increased with the deepening of the soil (Figures 2 and 3), which may not be conducive to the survival of Ascomycota, resulting in the decline of its abundance. We also noticed that the abundance of Ascomycota in the topsoil (0–20 cm) increased with the age of *P. massoniana* plantation (Figure 6b). This may be related to the increase of undergrowth shrubs and herbs in the near mature forest and mature forest stage after thinning, which makes litter easier to decompose. However, the abundance of Ascomycota in 20–40 cm and 40–60 cm soil depths tended to decrease with age (Figure 6b), which could be explained by the increase of nitrogen concentration (especially $\text{NH}_4^+\text{-N}$) in these two soil depths with age, because high nitrogen was harmful to the phylum [6].

4.3. Relationship among Soil Physicochemical Properties, Microbes, and Enzyme Activity

The activities of several enzymes have different trends with the increase of the chronosequence of *P. massoniana* plantations, but they decreased synchronously with the soil depths, and were significantly correlated with soil physicochemical properties and microbial alpha diversity. These results are in accordance with our hypothesis. Generally, a high S-ACP level is essential for improving soil P availability [6]. Our study observed that the topsoil has higher P and S-ACP, and there was a strong positive correlation between them, which was consistent with the function of S-ACP. Strangely, S-ACP first decreased and then increased with increasing stand age while AP continued to decrease, which indicated that S-ACP was not only related to phosphorus but may also be influenced by other factors. Previous studies demonstrated that TOC and TN could provide sufficient substrates for enzyme synthesis [28], so the lower soil TOC and TN in the middle-aged

forest stage and deep soil led to the decrease of enzyme activity. S-ACP and urease were positively correlated with soil TOC and TN, which confirmed the above results. At the same time, the improvement of soil pH and the increase of ST can also create a favorable condition for the enzyme to exist stably in the soil [9]. Correlation analysis showed that S-ACP and urease were significantly negatively correlated with soil C/N ratio and SWC; sucrose was significantly negatively correlated with SEC, which indicated that higher C/N, SWC, and SEC might not be conducive to enzyme activity. In addition, soil enzyme activity was not only related to soil physicochemical properties, but also positively correlated with bacterial and fungal diversity, because most exoenzymes and endoenzymes are mainly from microbes [27,29]. In particular, S-ACP was significantly negatively correlated with C/N ratio, positively correlated with N/P ratio, and positively correlated with bacterial and fungal Shannon index (Figure 8). Consequently, microbes may consume more C and N to secrete enzymes related to phosphorus metabolism under the condition of phosphorus deficiency [9]. In conclusion, fungi may have more competitive advantages than bacteria on TOC and TN in *P. massoniana* plantations under low phosphorus conditions. Therefore, the recovery of TOC and TN leads to an increase in fungal diversity, thus improving the enzyme activity. In turn, increased fungal diversity and enzyme activity promoted nutrient cycling.

5. Conclusions

Our results showed that the nutrient availability of woodland decreased significantly with the increase of the chronosequence of *P. massoniana* plantations, especially AP. Meanwhile, soil C/N ratio decreased while C/P and N/P ratios increased, indicating that there was a more serious P limitation in the near-mature forest and mature forest stage. On this basis, coupled with the decrease of pH and NO_3^- -N, these resulted in the decrease of soil bacterial alpha diversity. However, fungi can increase their diversity by consuming more TOC and TN under the condition of P limitation, which leads to the increase of enzyme activity (especially S-ACP). From different soil depths, soil physicochemical properties, microbial diversity, and enzyme activities decreased synchronously, and there was a close correlation between them. We also analyzed the microbial composition at different ages and soil depths and found that bacteria and fungi in the 10-year stand were significantly separated from other stands, and bacteria and fungi in the 0–20 cm soil depth were significantly separated from the other two soil layers. In all stands and soil depths, Acidobacteria was the most abundant bacterial community, followed by Proteobacteria and Actinobacteria; Basidiomycota was the most abundant fungal community, followed by Ascomycota. Furthermore, SWC, TOC, AP, and C/P ratio were the main factors affecting the microbial community; in particular, SWC had the highest contribution rate, and excessive SWC might inhibit the survival of most microbes. In short, these results reflect the potential relationship between soil physicochemical properties, microbial diversity and composition, and enzyme activities, which enhances understanding of soil fertility in different soil depths during the development of *P. massoniana* plantations and provides a reference for manual intervention.

Supplementary Materials: The following are available online at <https://www.mdpi.com/1999-4907/12/3/376/s1>, Figure S1: Rarefaction curves of bacteria (a) and fungi (b).

Author Contributions: Q.G. planned and supervised the project, conceived and designed the experiments, reviewed drafts of the paper. J.P. performed the experiments, analyzed the data and led manuscript production. H.L., S.L., Y.Z., S.Y., X.S., and Y.Q. collected and analyzed the data. X.F. analyzed the data and made charts. All authors have read and agreed to the published version of the manuscript.

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References

1. Yang, H.; Wang, S.; Zhang, J.; Fan, B.; Zhang, W. Biomass and nutrients of *Pinus massoniana* plantations in southern China: Simulations for different management practices. *J. Food Agric. Environ.* **2011**, *9*, 689–693.
2. SFA (State Forestry Administration P.R. China). *National Forest Resources Statistics (2009–2014)*. NFI 8th; State Forestry Administration P.R.: Beijing, China, 2014.
3. Wen, L.; Lei, P.; Xiang, W.; Yan, W.; Liu, S. Soil microbial biomass carbon and nitrogen in pure and mixed stands of *Pinus massoniana* and *Cinnamomum camphora* differing in stand age. *For. Ecol. Manag.* **2014**, *328*, 150–158. [[CrossRef](#)]
4. Chen, F.; Zheng, H.; Zhang, K.; Ouyang, Z.; Lan, J.; Li, H.; Shi, Q. Changes in soil microbial community structure and metabolic activity following conversion from native *Pinus massoniana* plantations to exotic *Eucalyptus* plantations. *For. Ecol. Manag.* **2013**, *291*, 65–72. [[CrossRef](#)]
5. Wang, W.; Su, D.; Qiu, L.; Wang, H.; An, J.; Zheng, G.; Zu, Y. Concurrent changes in soil inorganic and organic carbon during the development of larch, *Larix gmelinii*, plantations and their effects on soil physicochemical properties. *Environ. Earth Sci.* **2013**, *69*, 1559–1570. [[CrossRef](#)]
6. Zhu, L.; Wang, X.; Chen, F.; Li, C.; Wu, L. Effects of the successive planting of *Eucalyptus urophylla* on soil bacterial and fungal community structure, diversity, microbial biomass, and enzyme activity. *Land Degrad. Dev.* **2019**, 1–11. [[CrossRef](#)]
7. Koerselman, W.; Meuleman, A.F.M. The vegetation N:P ratio: A new tool to detect the nature of nutrient limitation. *J. Appl. Ecol.* **1996**, *33*, 1441–1450. [[CrossRef](#)]
8. Mooshammer, M.; Wanek, W.; Zechmeister-Boltenstern, S.; Richter, A. Stoichiometric imbalances between terrestrial decomposer communities and their resources: Mechanisms and implications of microbial adaptations to their resources. *Front. Microbiol.* **2014**, *5*, 22. [[CrossRef](#)]
9. Zhang, W.; Qiao, W.; Gao, D.; Dai, Y.; Deng, J.; Yang, G.; Han, X.; Ren, G. Relationship between soil nutrient properties and biological activities along a restoration chronosequence of *Pinus tabulaeformis* plantation forests in the Ziwuling Mountains, China. *Catena* **2018**, *161*, 85–95. [[CrossRef](#)]
10. Ren, C.; Zhao, F.; Kang, D.; Yang, G.; Han, X.; Tong, X.; Feng, Y.; Ren, G. Linkages of C:N:P stoichiometry and bacterial community in soil following afforestation of former farmland. *For. Ecol. Manag.* **2016**, *376*, 59–66. [[CrossRef](#)]
11. Fageria, N.K. Soil quality vs. environmentally-based agricultural management practices. *Commun. Soil Sci. Plant Anal.* **2002**, *33*, 2301–2329. [[CrossRef](#)]
12. Minasny, B.; Hong, S.Y.; Hartemink, A.E. Soil pH increase under paddy in South Korea between 2000 and 2012. *Agric. Ecosyst. Environ.* **2016**, *221*, 205–213. [[CrossRef](#)]
13. Mayakaduwa, S.; Mosley, L.M.; Marschner, P. Phosphorus pools in acid sulfate soil are influenced by soil water content and form in which P is added. *Geoderma* **2021**, *381*, 114692. [[CrossRef](#)]
14. Zechmeister-Boltenstern, S.; Keiblinger, K.M.; Mooshammer, M.; Peñuelas, J.; Richter, A.; Sardans, J.; Wanek, W. The application of ecological stoichiometry to plant–microbial–soil organic matter transformations. *Ecol. Monogr.* **2015**, *85*, 133–155. [[CrossRef](#)]
15. Powlson, D.S.; Jenkinson, D.S. A comparison of the organic matter, biomass, adenosine triphosphate and mineralizable nitrogen contents of ploughed and direct-drilled soils. *J. Agric. Sci.* **1981**, *97*, 713–721. [[CrossRef](#)]
16. Van, D.H.M.G.; Bardgett, R.D.; Van Straalen, N.M. The unseen majority: Soil microbes as drivers of plant diversity and productivity in terrestrial ecosystems. *Ecol. Lett.* **2008**, *11*, 296–310.
17. Zhang, C.; Liu, G.B.; Xue, S.; Wang, G.L. Soil bacterial community dynamics reflect changes in plant community and soil properties during the secondary succession of abandoned farmland in the Loess Plateau. *Soil Biol. Biochem.* **2016**, *97*, 40–49. [[CrossRef](#)]
18. Tian, J.; He, N.; Hale, L.; Niu, S.; Yu, G.; Liu, Y.; Blagodatskaya, E.; Kuzyakov, Y.; Gao, Q.; Zhou, J. Soil organic matter availability and climate drive latitudinal patterns in bacterial diversity from tropical to cold-temperate forests. *Funct. Ecol.* **2018**, *32*, 61–70. [[CrossRef](#)]
19. Lauber, C.L.; Ramirez, K.S.; Aanderud, Z.; Lennon, J.; Fierer, N. Temporal variability in soil microbial communities across land-use types. *ISME J.* **2013**, *7*, 1641–1650. [[CrossRef](#)] [[PubMed](#)]
20. Wan, X.; Huang, Z.; He, Z.; Yu, Z.; Wang, M.; Davis, M.R.; Yang, Y. Soil C:N ratio is the major determinant of soil microbial community structure in subtropical coniferous and broadleaf forest plantations. *Plant Soil.* **2015**, *387*, 103–116. [[CrossRef](#)]
21. Ren, C.; Chen, J.; Deng, J.; Zhao, F.; Han, X.; Yang, G.; Tong, X.; Feng, Y.; Shelton, S.; Ren, G. Response of microbial diversity to C:N:P stoichiometry in fine root and microbial biomass following afforestation. *Biol. Fertil. Soils* **2017**, *53*, 457–468. [[CrossRef](#)]
22. Fang, F.; Wu, C.Z.; Hong, W.; Fan, H.L.; Song, P. Study on the relationship between rhizospheric or non-rhizospheric soil enzyme and microbe in different plants. *Subtrop. Agric. Resear.* **2007**, *3*, 209–215.

23. Adamczyk, B.; Kilpeläinen, P.; Kitunen, V.H.; Smolander, A. Potential activities of enzymes involved in N, C, P and S cycling in boreal forest soil under different tree species. *Pedobiologia* **2014**, *57*, 97–102. [CrossRef]
24. Xu, Z.; Yu, G.; Zhang, X.; He, N.; Wang, Q.; Wang, S.; Wang, R.; Zhao, N.; Jia, Y.; Wang, C. Soil enzyme activity and stoichiometry in forest ecosystems along the North-South Transect in eastern China (NSTEC). *Soil Biol. Biochem.* **2017**, *104*, 152–163. [CrossRef]
25. Stock, S.C.; Kster, M.; Dippold, M.A.; Nájera, F.; Kuzyakov, Y. Environmental drivers and stoichiometric constraints on enzyme activities in soils from rhizosphere to continental scale. *Geoderma* **2019**, *337*, 973–982. [CrossRef]
26. Zhang, C.B.; Jin, Z.X.; Ke, S.X. Study on the relationship among enzymes activities, microorganism, respiration rate and physico-chemical properties of soil under different forests of Tiantai mountain. *J. Plant Nutr. Fertil.* **2004**, *10*, 51–56.
27. Burns, R.G.; Deforest, J.L.; Marxsen, J.R.; Sinsabaugh, R.L.; Stromberger, M.E.; Wallenstein, M.D.; Weintraub, M.N.; Zoppini, A. Soil enzymes in a changing environment: Current knowledge and future directions. *Soil Biol. Biochem.* **2013**, *58*, 216–234. [CrossRef]
28. Lucas-Borja, M.E.; Hedo, J.; Cerda, A.; Candel-Perez, D.; Vinegla, B. Unravelling the importance of forest age stand and forest structure driving microbiological soil properties, enzymatic activities and soil nutrients content in Mediterranean Spanish black pine (*Pinus nigra* Ar. ssp. *salzmannii*) *Forest. Sci. Total Environ.* **2016**, *562*, 145–154. [CrossRef]
29. Shen, F.; Wu, J.; Fan, H.; Liu, W.; Guo, X.; Duan, H.; Hu, L.; Lei, X.; Wei, X. Soil N/P and C/P ratio regulate the responses of soil microbial community composition and enzyme activities in a long-term nitrogen loaded Chinese fir forest. *Plant Soil.* **2019**, *436*, 91–107. [CrossRef]
30. Li, B.; Ge, T.; Xiao, H.; Zhu, Z.; Li, Y.; Shibistova, O.; Liu, S.; Wu, J.; Inubushi, K.; Guggenberger, G. Phosphorus content as a function of soil aggregate size and paddy cultivation in highly weathered soils. *Environ. Sci. Pollut. Res. Int.* **2016**, *23*, 7494–7503. [CrossRef]
31. NFA (National Forestry Administration). *Regulations for Age-Class and Age-Group Division of Main Tree-Species*; Wang, H.C., Li, G.C., Luan, G.J., Wang, R.N., Wang, H.Z., Bian, F., Shi, J., Zhou, J., Wang, Q., Gao, Z.F., et al., Eds.; National Forestry Administration: Beijing, China, 2017. Available online: <http://www.doc88.com/p-5671706807327.html> (accessed on 27 October 2017).
32. Bao, S. *Soil and Agricultural Chemistry Analysis*; China Agriculture Press: Beijing, China, 2000.
33. Shi, Y.; Shen, Q.R.; Mao, Z.S.; Xu, G.H. Time and Horizontal Spatial Variations of NH_4^+ -N and NO_3^- -N of Rhizospheric Soil with Rice Cultivation on Upland Condition Mulched with Half-decomposed Rice Straw. *Sci. Agric. Sin.* **2002**, *5*, 520–524.
34. Norman, R.J.; Stucki, J.W. The Determination of Nitrate and Nitrite in Soil Extracts by Ultraviolet Spectrophotometry. *Soil Sci. Soc. Am. J.* **1981**, *45*, 347–353. [CrossRef]
35. Pfahler, V.; Bielnicka, A.; Smith, A.C.; Granger, S.J.; Turner, B.L. A rapid ammonium fluoride method to determine the oxygen isotope ratio of available phosphorus in tropical soils. *Rapid Commun. Mass Spectrom.* **2020**, *34*, e8647. [CrossRef]
36. Guan, S.Y. *Soil Enzyme and Its Research Methods*; China Agricultural press: Beijing, China, 1986.
37. Hugerth, L.W.; Wefer, H.A.; Lundin, S.; Jakobsson, H.E.; Lindberg, M.; Rodin, S.; Engstrand, L.; Andersson, A.F. DegePrime, a Program for Degenerate Primer Design for Broad-Taxonomic-Range PCR in Microbial Ecology Studies. *Appl. Environ. Microbiol.* **2014**, *80*, 5116–5123. [CrossRef] [PubMed]
38. Smit, E.; Leeflang, P.; Glandorf, B.; Van Elsas, J.D.; Wernars, K. Analysis of fungal diversity in the wheat rhizosphere by sequencing of cloned PCR-amplified genes encoding 18S rRNA and temperature gradient gel electrophoresis. *Appl. Environ. Microbiol.* **1999**, *65*, 2614–2621. [CrossRef]
39. Yang, C.; Liu, N.; Zhang, Y. Soil aggregates regulate the impact of soil bacterial and fungal communities on soil respiration. *Geoderma* **2019**, *337*, 444–452. [CrossRef]
40. Wang, Q.; Garrity, G.M.; Tiedje, J.M.; Cole, J.R. Naïve Bayesian Classifier for Rapid Assignment of rRNA Sequences into the New Bacterial Taxonomy. *Appl. Environ. Microbiol.* **2007**, *73*, 5261–5267. [CrossRef] [PubMed]
41. Quast, C.; Pruesse, E.; Yilmaz, P.; Gerken, J.; Schweer, T.; Yarza, P.; Peplies, J. The SILVA ribosomal RNA gene database project: Improved data processing and web-based tools. *Nucleic Acids. Res.* **2013**, *41*, D590–D596. [CrossRef] [PubMed]
42. Wu, Z.; Elizabeth, H.S.; Wenxiong, L.; Bailian, L.; Linkun, W.; Changxun, F.; Zhixing, Z.; Zhen, J. Soil Microbial Community Structure and Metabolic Activity of *Pinus elliottii* Plantations across Different Stand Ages in a Subtropical Area. *PLoS ONE.* **2015**, *10*, e0135354. [CrossRef] [PubMed]
43. Liu, Y.; Zang, H.; Tida, G.; Bai, J.; Guggenberger, G. Intensive fertilization (N, P, K, Ca, and S) decreases organic matter decomposition in paddy soil. *Appl. Soil Ecol.* **2018**, *127*, 51–57. [CrossRef]
44. Delgado-Baquerizo, M.; Reich, P.B.; Khachane, A.N.; Campbell, C.D.; Thomas, N.; Freitag, T.E.; Al-Soud, W.A.; Sørensen, S.; Bardgett, R.D.; Singh, B.K. It is elemental: Soil nutrient stoichiometry drives bacterial diversity. *Environ. Microbiol.* **2017**, *19*, 1176–1188. [CrossRef] [PubMed]
45. Högberg, M.N.; Högberg, P.; Myrold, D.D. Is microbial community composition in boreal forest soils determined by pH, C-to-N ratio, the trees, or all three? *Oecologia* **2007**, *150*, 590–601. [CrossRef]
46. Shi, P.L.; Zhang, X.Z. Some advance in the main factors controlling soil respiration. *Adv. Earth Sci.* **2005**, *20*, 778–785.
47. Kjäller, A.; Struwe, S. Microfungi in ecosystems: Fungal occurrence and activity in litter and soil. *Oikos* **1982**, *39*, 389–442. [CrossRef]
48. Guo, P.; Jiang, H.; Yu, S.; Ma, Y.; Dou, R.; Song, X. Comparison of Litter Decomposition of Six Species of Coniferous and Broad-leaved Trees in Subtropical China. *Chin. J. Appl. Environ. Biol.* **2009**, *15*, 655–659.

49. Kubartová, A.; Ranger, J.; Berthelin, J.; Beguiristain, T. Diversity and Decomposing Ability of Saprophytic Fungi from Temperate Forest Litter. *Microb. Ecol.* **2009**, *58*, 98–107. [[CrossRef](#)]
50. Du, C.; Geng, Z.; Wang, Q.; Zhang, T.; Wang, Y. Variations in bacterial and fungal communities through soil depth profiles in a *Betula albosinensis* forest. *J. Microbiol.* **2017**, *55*, 684–693. [[CrossRef](#)]
51. Izumi, H.; Cairney, J.W.G.; Killham, K.; Moore, E.; Alexander, I.J.; Anderson, I.C. Bacteria associated with ectomycorrhizas of slash pine (*Pinus elliotti*) in south-eastern Queensland, Australia. *FEMS Microbiol. Lett.* **2008**, *282*, 196–204. [[CrossRef](#)] [[PubMed](#)]
52. Clemmensen, K.E.; Bahr, A.; Ovasainen, O.; Dahlberg, A.; Ekblad, A.; Wallander, H.; Stenlid, J.; Finlay, R.D.; Wardle, D.A.; Lindahl, B.D. Roots and Associated Fungi Drive Long-Term Carbon Sequestration in Boreal Forest. *Science* **2013**, *339*, 1615–1618. [[CrossRef](#)]
53. Zhang, Y.; Cong, J.; Lu, H.; Li, G.; Qu, Y.; Su, X.; Zhou, J.; Li, D. Community structure and elevational diversity patterns of soil Acidobacteria. *J. Environ. Sci.* **2014**, *26*, 1717–1724. [[CrossRef](#)]
54. Fierer, N.; Bradford, M.A.; Jackson, R.B. Toward an ecological classification of soil bacteria. *Ecology* **2007**, *88*, 1354–1364. [[CrossRef](#)] [[PubMed](#)]
55. Elser, J.J. Growth rate–stoichiometry couplings in diverse biota. *Ecol. Lett.* **2010**, *6*, 936–943. [[CrossRef](#)]
56. Kopecky, J.; Kyselkova, M.; Omelka, M.; Cermak, L.; Novotna, J.; Grundmann, G.L.; Moënne-Loccoz, Y.; Sagova-Mareckova, M. Actinobacterial community dominated by a distinct clade in acidic soil of a waterlogged deciduous forest. *FEMS Microbiol. Ecol.* **2011**, *78*, 386–394. [[CrossRef](#)] [[PubMed](#)]
57. Santalahti, M.; Sun, H.; Jumpponen, A.; Pennanen, T.; Heinonsalo, J. Vertical and seasonal dynamics of fungal communities in boreal Scots pine forest soil. *FEMS Microbiol. Ecol.* **2016**, *92*, fiw170. [[CrossRef](#)] [[PubMed](#)]
58. Cheng, R.M.; Wang, R.L.; Xiao, W.F.; Feng, X.H.; Liu, Z.B.; Ge, X.G.; Wang, X.R.; Zhang, W.Y. Spatial distribution of root biomass of *Pinus massoniana* plantation in Three Gorges Reservoir area, China. *Acta Ecol. Sin.* **2012**, *32*, 823–832. [[CrossRef](#)]
59. He, Y.H.; Zhou, G.Y.; Wang, S.J.; Li, H. Fungal diversity in *Cunninghamia lanceolata* plantation soil. *Acta Ecol. Sin.* **2014**, *34*, 2725–2736.