

# Biogas and Biomethane Production and Usage: Technology Development, Advantages and Challenges in Europe

Josipa Pavičić <sup>1</sup>, Karolina Novak Mavar <sup>2,\*</sup>, Vladislav Brkić <sup>2</sup> and Katarina Simon <sup>2</sup>

<sup>1</sup> Pere Pirkera 38, 10361 Sesvete, Croatia; josipapavicic2003@gmail.com

<sup>2</sup> Faculty of Mining, Geology and Petroleum Engineering, University of Zagreb, 10000 Zagreb, Croatia; vladislav.brkic@rgn.hr (V.B.); katarina.simon@rgn.hr (K.S.)

\* Correspondence: karolina.novak-mavar@rgn.hr

**Abstract:** In line with the low-carbon strategy, the EU is expected to be climate-neutral by 2050, which would require a significant increase in renewable energy production. Produced biogas is directly used to produce electricity and heat, or it can be upgraded to reach the “renewable natural gas”, i.e., biomethane. This paper reviews the applied production technology and current state of biogas and biomethane production in Europe. Germany, UK, Italy and France are the leaders in biogas production in Europe. Biogas from AD processes is most represented in total biogas production (84%). Germany is deserving for the majority (52%) of AD biogas in the EU, while landfill gas production is well represented in the UK (43%). Biogas from sewage sludge is poorly presented by less than 5% in total biogas quantities produced in the EU. Biomethane facilities will reach a production of 32 TWh in 2020 in Europe. There are currently 18 countries producing biomethane (Germany and France with highest share). Most of the European plants use agricultural substrate (28%), while the second position refers to energy crop feedstock (25%). Sewage sludge facilities participate with 14% in the EU, mostly applied in Sweden. Membrane separation is the most used upgrading technology, applied at around 35% of biomethane plants. High energy prices today, and even higher in the future, give space for the wider acceptance of biomethane use.

**Keywords:** biogas; biomethane; gas cleaning and upgrading; anaerobic digestion; circular economy; emission reduction

**Citation:** Pavičić, J.; Novak Mavar, K.; Brkić, V.; Simon, K. Biogas and Biomethane Production and Usage: Technology Development, Advantages and Challenges in Europe. *Energies* **2022**, *15*, 2940. <https://doi.org/10.3390/en15082940>

Academic Editors: Nediljka Gaurina-Međimurec and Borivoje Pašić

Received: 28 March 2022

Accepted: 13 April 2022

Published: 17 April 2022

**Publisher's Note:** MDPI stays neutral with regard to jurisdictional claims in published maps and institutional affiliations.



**Copyright:** © 2022 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (<https://creativecommons.org/licenses/by/4.0/>).

## 1. Introduction

The European Union intention to achieve a sustainable and competitive circular economy by 2050 implies structural changes of energy systems. The first step in transition towards net-zero emissions is a 55% reduction in greenhouse gases (GHG) achieved by 2030 [1]. By the Fit for 55 proposals [2], released in July 2021, the European Commission (EC) is going to require a dramatic technology revamp, through low-carbon processes engagement, comprising energy efficiency and replacing conventional energy sources with green energy sources [1,3]. However, amazing goals from today's point of view affect different sectors. Characterized by huge emissions, the transportation sector is expected to undergo significant changes. The revised Directive (EU) 2018/2001 on the promotion of the use of energy from renewable sources (REDII) [4] sets the targets for both the renewable energy share in transportation and the advanced biofuels sub-target of 14% and 3.5%, respectively, by 2030. While the proposed revision replaces the 14% renewables target with the target of 13% decreasing GHG fossil fuel intensity, it is obvious that both of the goals include a greater use of biofuels. In order to encourage waste-based biofuels production, most of the EU member states have decided to establish a double counting system in the production of biofuels from used cooking oil or animal fat. Directive restrictions cover a limitation of double counting to 1.7%. Further restrictions are focused

on the elimination of biofuels produced from crops which pose a threat of significant expansion of the production area into land with high carbon stock [5].

The use of biofuels varies among the EU member states, depending on their decisions on including renewables in the transportation sector and a reduction of GHG intensity of fuels, primarily under the Directive (EU) 98/70 relating to the quality of petrol and diesel fuels [6], which was later harmonized together with RED (2009/28) [7] into Directive (EU) 2015/1513 [8]. There is, however, a different approach applied among countries. While some countries chose an overall biofuel target, others decided to separate the goals for biofuels in petrol or diesel, or they allowed themselves to rely only on fuel carbon intensity reduction targets (Figure 1). As per the National biofuels policies 2022 [8], and as can be seen in Figure 1, the majority of the countries (18 member states and the UK) set overall biofuel targets. There are 13 member states with differing goals in petrol or/and diesel, as well as six countries having both overall biofuels targets and specific targets in petrol and diesel. Advanced biofuels obligation is recognized in the biofuel policies of 20 countries. The highest biofuel targets are defined in Finland and the Netherlands, setting 19.5% and 17.5%, respectively. On the other hand, Sweden or France opted to set higher targets for reducing the GHG fuel intensity (30.5% for diesel and 7.8% for petrol in Sweden and 10% in France), without overall biofuel related goals. The UK set a 12.6% biofuel target by defining a crop cap at 3.67%, while crop-based biofuels are capped at 4.5% in Estonia. Double counting of advanced targets is not possible in Finland.

However, new strategies developed under the European Green Deal [1], such as the EU Methane Strategy [9], the Energy System Integration Strategy [10], and the From Farm to Fork strategy [11], place biogas in a central position [12]. In light of low carbon strategies, gas market retention requires zero carbon gases and decarbonisation of natural gas (methane). Decarbonized gas options refer to biogas, biomethane, synthetic natural gas, hydrogen from renewable energy, and hydrogen from the reforming or splitting of methane [13]. Biogas is a mixture of gases, primarily consisting of methane ( $\text{CH}_4$ ) and carbon dioxide ( $\text{CO}_2$ ), which is produced by the decomposition of organic compounds derived from different sources [14].

In the past decade, there was an increase of 90% in the biogas industry. The installed capacity globally reached 120 GW in 2019 compared to 65 GW in 2010 where more than 70% of the world biogas generation belongs to Europe [15]. According to the EBA Statistical Report 2021 [16] biogas and biomethane production amounted to 191 TWh in 2020 (18.0 bcm of energy) in Europe. This figure is intended to be doubled in 2030, and to reach over 1000 TWh in 2050, taking a share of up to 40% of the total European gas demand.

In an era of energy crisis, biomass-based bioenergy has been given special attention. Generally, a commonly used feedstock for bioenergy production is lignocellulosic biomass, as the most economical renewable source. Lignocellulosic biomass feedstock refers to municipal solid waste, forest and agricultural products, algae, etc. Lignocellulosic biomass is composed of lignin, cellulose, hemicellulose, ash, and extractives. On the other hand, sewage sludge, animal fat, and animal manure are the major constituents of non-lignocellulosic biomass, rich in protein and fatty acids with less lignin, cellulose, and hemicellulose. Although the term is usually reserved for transport fuels derived from biomass, its wider meaning refers to any fuel derived directly or indirectly from biomass and used for energy purposes that can be sold in liquid and gaseous phases. Various conversion techniques for biofuels are available or under development, reviewed by different authors [17–27].

However, this article reviews the processes deployed for biogas production, cleaning and upgrading, and use. Biogas feedstock can be grouped into agricultural waste, municipal organic waste and industrial waste. Common feedstocks include livestock manure, food-processing waste, and sewage sludge [28–30]. Substrates for biogas production differ significantly by region and commonly depend on the availability of certain raw materials in a certain area [14].

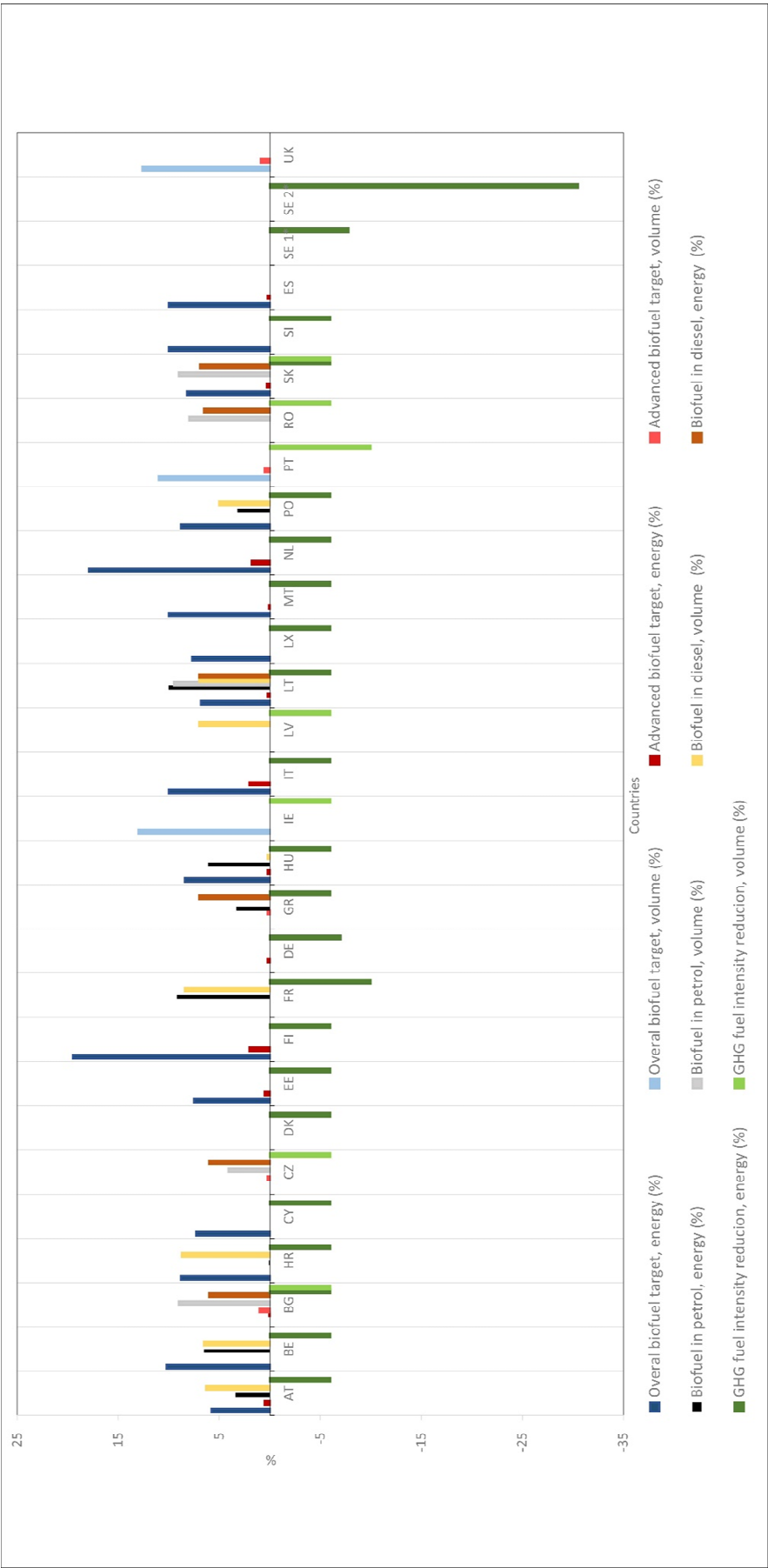
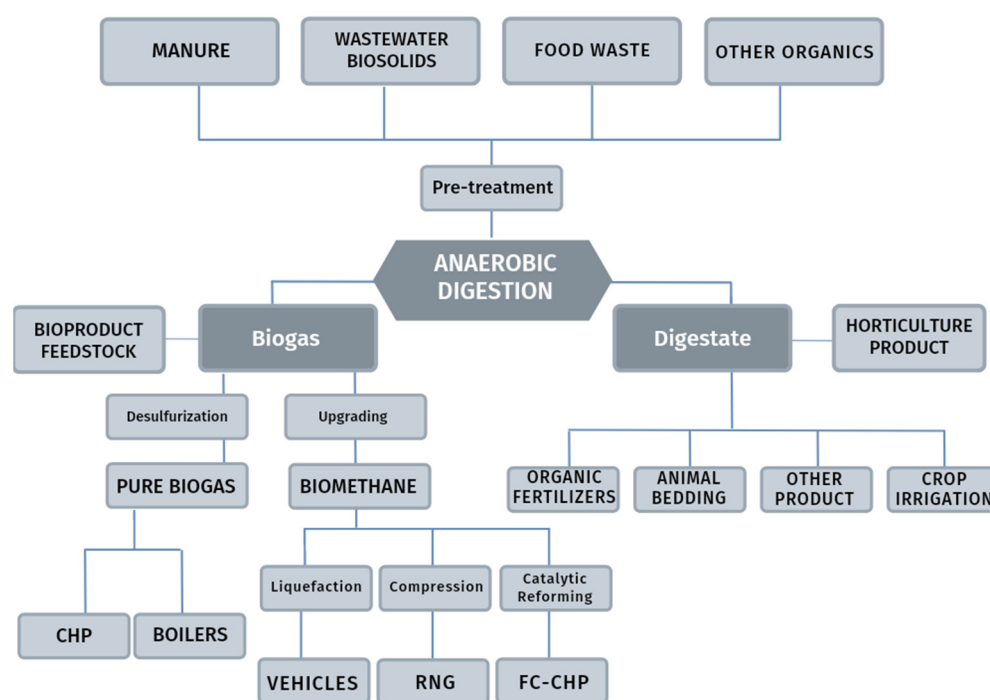


Figure 1. Biofuels targets in the EU (according to [8]).

Biogas usage is mainly connected to electricity and heat generation, and household purposes (cooking and heating). After upgrading to biomethane, it can be injected into a gas grid or used as transport fuel, which is expected to be its main future purpose. Installations for heat and power generation with electrical capacity are in the range from tens of kWe to a few MWe. Different energy applications of biogas are shown in Figure 2. After advanced upgrading, biomethane can be injected into a natural gas pipeline grid, used as a transportation fuel or in fuel cells. Compression and liquefaction are needed to convert biogas into compressed natural gas (bio-CNG) and liquefied biogas (LBG), while catalytic reforming is a commonly applied chemical method to obtain syngas. By Fischer–Tropsch synthesis (FTS) or fermentation process, syngas can be transformed into alcohols, such as methanol or ethanol [15,31–34].

About 90% of globally produced biomethane is generated by the anaerobic digestion process, followed by different upgrading treatments [31]. Production alternatives include thermal and hydrothermal gasification. Engagement of pretreatment results in a wider range of suitable input raw materials and process efficiency improvement [35,36].

Upgrading to biomethane or renewable natural gas (RNG) is being adopted in developed countries, especially in North America and Europe. Such projects are often developed by oil and gas companies in the USA for gaining carbon credits. As a vehicle fuel alternative, biogas has already been applied in the European Union (Sweden and Germany) and the USA, either as pure methane (CBG100) or in a mixture with natural gas (CBG10 and CBG50) [15,37,38].



**Figure 2.** Possibilities of using biogas.

The aim of this paper is to give a representative overview of the overall process of biogas and biomethane production, from feedstock used to the final product in order to get a foundation for further research. In line with that, the paper gives biogas composition, analyses the substrates options that results in satisfactory yield and available technologies for biogas and biomethane production. The analysis is made based on published and publicly available data, with the focus on the most often-used anaerobic digestion process and pretreatment methods. Conventional upgrading techniques are elaborated, but also some

modern, cost effective and environmentally friendly solutions are shown. Biogas and biomethane temporary usage and further streams are given through a literature review on papers published in the last five years.

Another target is to analyse the role of biogas in achieving a climate-neutral Europe, through determination of the current position of biogas in the energy mix and defining its perspective in light of new strategies and geopolitical developments. For that purpose, the EU legal requirements in terms of biogas usage and emission reduction are analysed. Based on data published by different organisations, analyses on biogas presence in Europe, along with leading production countries and distribution of the installed biomethane facilities per both feedstock and upgrading techniques, are elaborated.

## 2. Methods

A literature review was used as a methodology for conducting research. The intention was to make a comprehensive review of technology applied for biogas production and usage. Therefore, the authors have tried to collect data on the available processes, not only those which are commonly in use, but also those that are currently being under research and development. Additionally, a systematic literature review was initiated to identify relevant data related to biogas cleaning and upgrading techniques, and possible pretreatments used to improve process efficiency. A separate focus was on detecting the biogas current energetic position, as well as its future potential in the EU, keeping in mind the required transition to a low carbon economy.

Turbulent developments of climate-policies, technology progress and a recent upturn in biogas facilities are the reason that we have mostly, but not exclusively, focused on in recent publications. Although most of the considered literature was published in the period from 2015 to 2020, some of the earlier published papers were found convenient due to their information regarding the general technology characteristics.

This review paper covers data obtained from academic literature and professional reports issued by different agencies and organisations. Academic literature covers peer-reviewed papers, books, theses, abstracts and technical reports containing exact figures and have served as an accurate basis in our speculation on technology prospects. The statistics related to the status of biomethane facilities were mostly adopted from reports published by different professional associations and agencies. The figures are additionally analysed, rearranged and compiled in order to analyse actual developments.

Construction of this article ensures a clear review on technology development, divided into separate process phases, i.e., pretreatment, production, purification and application of biogas. Each chapter concludes with an analysis of future technology trends.

## 3. Biogas Production

### 3.1. Biogas Composition

Biogas streams mainly consist of methane ( $\text{CH}_4$ , up to 75%) and carbon dioxide ( $\text{CO}_2$ , up to 50%), while hydrogen ( $\text{H}_2$ ), nitrogen ( $\text{N}_2$ ), water vapour ( $\text{H}_2\text{O}_{(\text{g})}$ ) and hydrogen sulphide ( $\text{H}_2\text{S}$ ) may be present in a smaller proportion. Biogas composition and its properties depend on the substrate type (source of biomass entering the process), the type of facility and the process conditions. Table 1 compares the composition of raw biogas and natural gas. Individual volume shares differ to a lesser extent depending on the literature source.

**Table 1.** Biogas vs. natural gas composition [14,39–41].

Compound	Formula	Volume share (%)			
		Biogas			Natural gas
		Korbag et al. [14]	Persson et al. [39]	Moya et al. [40]	IEA Bioenergy [41]
Methane	$\text{CH}_4$	50–75	53–70	55–70	83–98
Carbon dioxide	$\text{CO}_2$	25–50	30–47	30–45	0–1.4

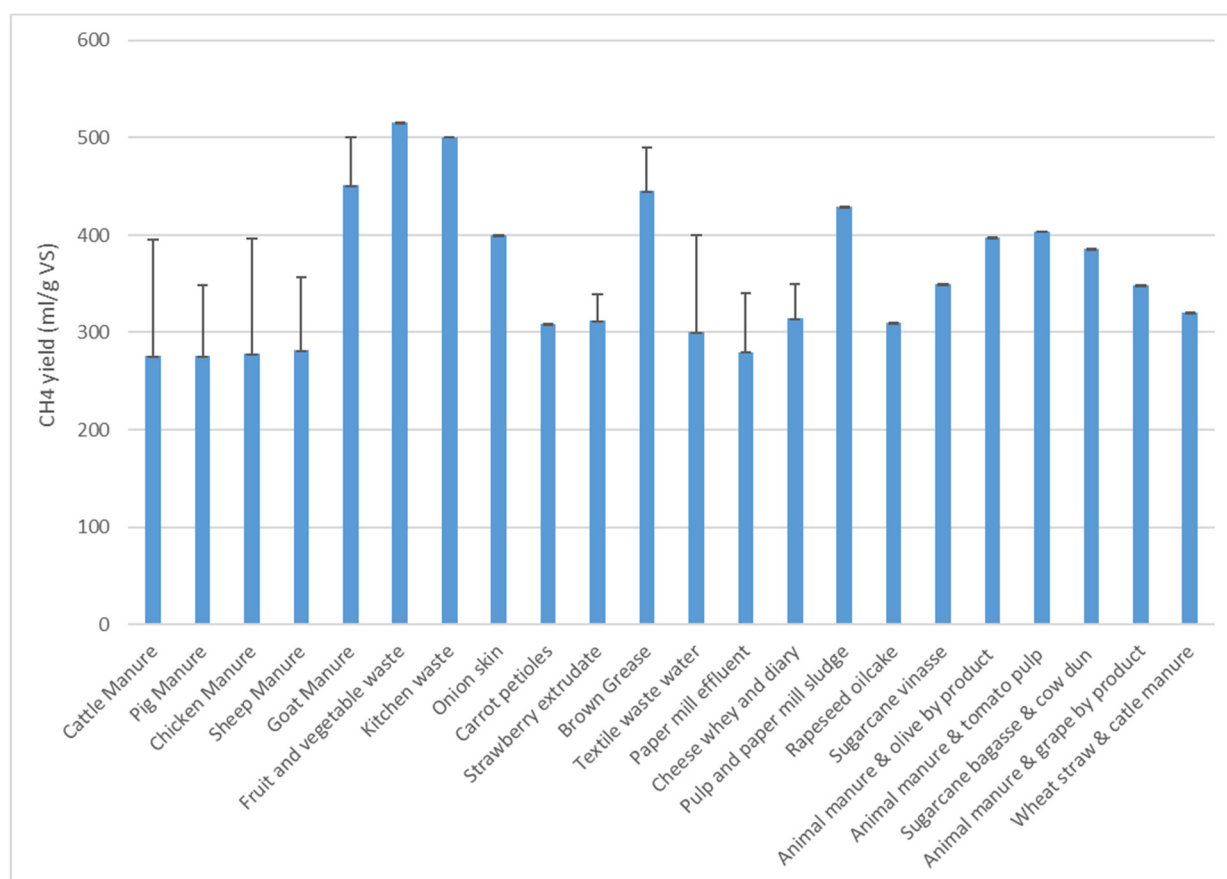
Water	H <sub>2</sub> O	5–10	/	/–5	
Oxygen	O <sub>2</sub>	<2	0	0–3	
Nitrogen	N <sub>2</sub>	<10	0.2	<15	0.6–2.7
Ammonia	NH <sub>3</sub>	<1	<1 (<100 ppm)	<1 (0–100 ppm)	/
Hydrogen	H <sub>2</sub>	<1	0	/	/
Hydrogen sulphide	H <sub>2</sub> S	<3	<1 (0–10,000 ppm)	<1 (0–10,000 ppm)	/
Ethane					<11%
Propane					<3%

In addition to the listed substances, traces of iron, nickel, cobalt, selenium, molybdenum and tungsten can be present [42]. The chemical composition of biogas will also determine the process that needs to be applied to optimize its usage properties. The presence of CO<sub>2</sub> and N<sub>2</sub> negatively affect the gas calorific value, while CO<sub>2</sub>, H<sub>2</sub>S and water vapour form corrosive conditions, which cause equipment damage [14].

### 3.2. The Biogas Production: From Substrate to Biomethane

The chemical composition and biodegradability of the used substrates affect the biogas composition. Generally, substrates rich in lipids and proteins have a greater potential of methane yield compared to those rich in carbohydrates, but high lipid content leads to the anaerobic digestive system failure due to long-chain fatty acid generation. On the other hand, substrates rich in carbohydrates can negatively affect the C/N ratio, resulting in nutrient restriction and quick acidification [30]. Formerly, only one input substrate entered the process, but recent practice prefers a combination of different substrates: the so-called co-digestion. Co-digestion has led to a higher gas production per ton of substrate [43].

Livestock manure is commonly considered a cost-effective and viable feedstock for biogas production, but differences in animal digestive systems, nutrition or manure storage mechanisms result in different and most often lower biomethane production. Although an abundance of lignocellulosic biomass (crop residues, forest residues, energy crops, etc.) could serve as a good basis for sustainable energy production, lignin in lignocellulosic biomass encapsulates hemicellulose and cellulose into a hydrophobic structure of Lignin-Carbohydrate Complexes (LCC), requiring the application of suitable pretreatments. Regarding industrial waste, waste from the food industry has been commonly used in bioreactors due to its chemical properties and biodegradability. Figure 3 represents the selected substrate types used in anaerobic digestion that pose good biomethane yield in mL/g volatile solids (VS). Among a large number of possible substrates, only those with an average biomethane yield over 250 mL/g VS are shown. Livestock manures generally provide a lower methane yield (around 200 mL CH<sub>4</sub>/g VS) due to higher ammonia inhibiting compounds (sheep manure and goat manure substrates show the best results). On the other hand, food by-products pose a significant biomethane yield (over 500 mL/g VS). Food by-products rich in lipid content (kitchen waste) have a higher methane yield. To satisfy the nutritional requirements of microorganisms, a low C/N ratio characterized for livestock manure is combined with carbon rich feedstock, achieving satisfactory results in biomethane yields [44].



**Figure 3.** Biogas yield from different substrates with biomethane yield above 200 mL/g VS (according to [30]).

Recently, production plants have been more oriented towards the use of municipal and industrial waste. However, manure and sludge from wastewater are the most common input (80%) [14]. Although AD is a promising technique for waste sludge management [4], complex microstructure components make sewage sludge difficult to hydrolyze and digest. Components such as extracellular polymeric substances have low volatile solid degradation (30–50%) even with a long retention time, which leads to low methane yield. Biomethane productivity is enhanced by different pretreatment options and by co-digestion with organic wastes [45–47].

Song et al. [48] determined the optimal mixing ratio of 2:1:1 for sewage sludge, food waste, and livestock manure by applying the biochemical methane potential (BMP) test, Design Expert software and continuous reactor operation.

Process input varies significantly by region, being imitated by those materials that are locally available due to transport costs minimization [49,50].

### 3.3. Using Technologies in Biogas and Biomethane Production

Although AD with upgrading treatments is a commonly used technology in biomethane yielding, which means that 90% of the produced biomethane comes from AD, other alternatives are commercially available, and some emerging technologies can be found. The International Energy Agency (IEA) divides biogas based on the feedstock and treatment type into sewage sludge gas, landfill gas and other AD biogas (methanization of non-hazardous waste or raw plant matter) [31].

#### 3.3.1. Anaerobic Digestion

Anaerobic digestion is the biochemical process of decomposition of organic compounds by microbe bacteria under anaerobic conditions. The process itself is actually similar to some natural processes, such as peat formation, or the process in the ruminant digestive system. Methane fermentation requires diverse associations of bacteria, i.e., acetoclastic and hydrogenotrophic methanogens, syntrophic acetogens, fermentative bacteria and homoacetogen [15,51].

Biogas production can be accomplished via two main fermentation processes. In the dry process, the solid concentration in the fermenter is lower than 10%, and in the wet process, it ranges from 15 to 35% [15,52]. However, different factors affect the AD process, such as temperature, pH value, carbon to nitrogen ratio (C/N) and substrate particle size. It can be expected that smaller particles undergo faster decomposition, resulting in faster biogas generation. The process pH ranges from 5.5 to 8.5, with optimal results obtained at higher values. As per Sahota et al. [53], the optimal C/N ratio is in the range from 20/1 up to 30/1 [54,55].

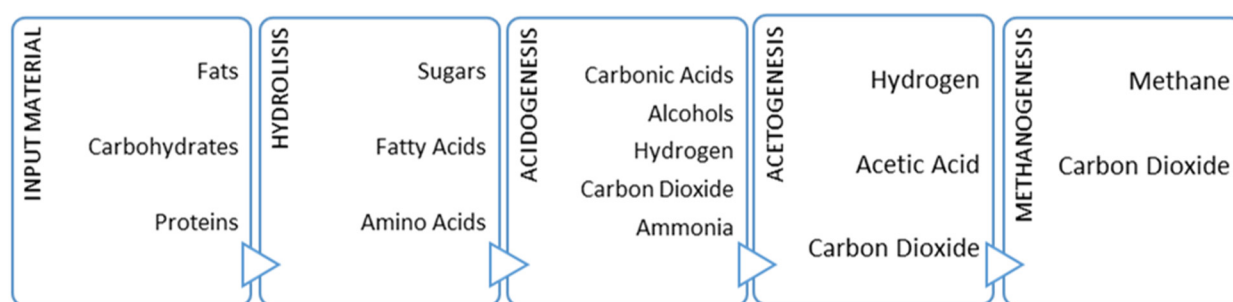
An AD environment can be generally divided into psychrophilic (low temperature), mesophilic (moderate temperature), and thermophilic (warm). Each regimen is suitable for a specific type of microorganism. The psychrophilic process takes place at temperatures less than 30 °C, the mesophilic process occurs in the temperature range of 30 to 42 °C, while the temperature ranges from 43 to 55 °C is needed for the thermophilic process. However, too low or too high temperatures are not suitable due to the fact that they slow down the process [14,42,55].

Mesophilic and thermophilic regimes are more commonly used, and the temperature in these processes is maintained by installed heat exchangers. Some of the advantages of thermophilic biogas production are a higher rate of methanogenic bacteria, shorter digestion time, higher efficiency and better substrate utilization. In general, higher temperatures result in higher biogas yield and shorter retention time. However, higher energy needs require optimal conditions to be defined. Furthermore, the risk of ammonia inhibition and irreversible acidification caused by volatile fatty acid accumulation are important reasons for closer supervision of thermophilic digesters [51,56]. While conventional thermophilic anaerobic digestion optimization focuses primarily on the development of robust microbiome and co-digestion, new approaches consider integration with microbial electrochemical systems and the use of conductive additives [56].

A comprehensive review of reactors for biogas production is given by Banerjee et al. [55]. The authors stated that one-stage systems are usually installed in industry due to their performance simplicity, although two-stage systems are less energy demanding.

As seen in Figure 4, the biogas production procedure includes four important phases: hydrolysis, acidogenesis, acetogenesis, and methanogenesis [57,58]. Principal reactions which occurred in each phase are shown in Table 2. In the first phase, i.e., hydrolysis, complex organic components, such as fats, proteins and carbohydrates, are broken down into simpler molecules, i.e., amino acids, fatty acids and sugars. Equation (1) represents the hydrolysis of cellulose ( $C_6H_{10}O_5$ ) via the addition of water ( $H_2O$ ) to form glucose ( $C_6H_{12}O_6$ ) and  $H_2$ . Chemical reactions take place inside the fermenter (digester) under bacterial activities. Acidogenesis, or the second phase, considers decomposition of chemical compounds formed by hydrolysis into methanogenic compounds, i.e., hydrogen, alcohols, carbon dioxide, carbon acids and ammonia. This is about the fermentation stage, presented by the Equations (2)–(4). Acetogenesis and methanogenesis often take place simultaneously. Carbon dioxide, hydrogen and acetic acid are formed during acetogenesis, as presented by Equations (5)–(7). However, a clear boundary between acidogenesis and acetogenesis is often impossible since both phases result in the generation of  $H_2$  and  $CH_3COO^-$ —which are methanogenic bacteria substrates. Methanogenesis phase results in  $CH_4$  and  $CO_2$  generation. The conversion of  $CH_3COOH$  into  $CH_4$  and  $CO_2$  is shown by Equation (8).  $CO_2$  is further reduced to  $CH_4$  through  $H_2$  (Equation (9)). The production of  $CH_4$  by decarboxylation of  $CH_3CH_2OH$  is presented by Equation (10) [58].





**Figure 4.** Anaerobic digestion process phases (modified to [51]).

**Table 2.** Principal reactions in the AD process (according to [58]).

AD process stage	Chemical reactions
Hydrolysis	$(C_6H_{10}O_5)_n + n H_2O \rightarrow n C_6H_{12}O_6 + n H_2$ (1)
	$C_6H_{12}O_6 \leftrightarrow 2 CH_3CH_2OH + 2 CO_2$ (2)
Acidogenesis	$C_6H_{12}O_6 + 2 H_2 \leftrightarrow 2 CH_3CH_2COOH + 2 H_2O$ (3)
	$C_6H_{12}O_6 \rightarrow 3 CH_3COOH$ (4)
	$CH_3CH_2COO^- + 3 H_2O \leftrightarrow CH_3COO^- + H^+HCO_3^- + 3 H_2$ (5)
Acetogenesis	$C_6H_{12}O_6 + 2 H_2O \leftrightarrow 2 CH_3COOH + 2 CO_2 + 4 H_2$ (6)
	$CH_3CH_2OH + 2 H_2O \leftrightarrow CH_3COO^- + 3 H_2 + H^+$ (7)
	$CH_3COOH \rightarrow CH_4 + CO_2$ (8)
Methanogenesis	$CO_2 + 4 H_2 \rightarrow CH_4 + 2 H_2O$ (9)
	$2 CH_3CH_2OH + CO_2 \rightarrow CH_4 + 2 CH_3COOH$ (10)

Mao et al. [41,59] investigated the latest trends in AD and anaerobic membrane bioreactors (AnMBRs). Even though the intention of breaking down the comprehensive AD process into two groups, i.e., acetogenesis–methanation and hydrolysis–acidogenesis separated reactors, enhances the rate of the conversion process, this would result in higher costs [60].

### 3.3.2. Thermal and Hydrothermal Gasification

Thermal gasification is passing through its early commercialization phase. Such a process uses dry woody and lignocellulose biomass. Under controlled oxygen and steam, thermal breakdown occurs, resulting in syngas and biochar. Conversion into biomethane is done in a catalytic reaction or through biological processes [36,43,61]. Another option is hydrothermal gasification. It uses dry or wet raw material, which is then heated and compressed to supercritical conditions, favourable for methane rich biogas generation. Hydrothermal gasification demonstration projects in France and the Netherlands represent the forerunners of industrial application, which can be expected in forthcoming years [35,36].

### 3.3.3. Process Pretreatments

Pretreatment is used to overcome the obstruction of digestion. Complex organic matter needs to be solubilized and hydrolyzed into simple components, i.e., to be prepared for microbial activities. Cellulose and hemicellulose are polysaccharides that can be hydrolyzed to simple sugars. Lignin serves as a support to the cell structure and thus impedes the proneness to microbial attack during the hydrolysis process. Pretreatment processes break the lignin layer, making the biomass prone to digestion. The other role is in decreasing the crystallinity of cellulose and increasing porosity [62]. There are different

advanced pretreatments, which enable a broader range of input raw materials, and improve the AD process efficiency by faster digestion and yield increasing. A decrease in the retention period leads to economically reasonable small-scale reactors.

Various in situ and ex situ pretreatment techniques are used in order to achieve better results [15,44,62–67]. Pretreatment methods can be grouped into physical, chemical, physico-chemical and biological methods. (Table 3).

**Table 3.** Types of pretreatments most used to improve biogas production [62–66].

Pretreatment type	Method	Mode of action
Physical	Milling	Biomass is altered by cells breaking when physical force is used.
	Extrusion	
	Cavitation	
	Microwave Irradiation	
Physico-chemical	Liquid Hot Water	A combination of mechanical forces and chemical effects applied to biomass.
	Steam Explosion Ammonia Fiber Expansion	
	CO <sub>2</sub> Explosion	
Chemical	Alkaline Hydrolysis	Use of organic or inorganic compounds to disrupt the structural recalcitrance of lignocellulosic biomass.
	Acid Hydrolysis	
	Ionic Liquids	
	Organosolv	
	Wet Oxidation	
Biological		Enzyme-producing fungi are used to alter biomass.

Physical pretreatment methods reduce the particle size resulting in an increase in surface area and a decrease in polymerization degree and crystallinity. Increased surface area improves accessibility to microbial and enzymatic degradation activities. The main disadvantages of physical pretreatments are in their inability to remove the lignin content of lignocellulosic biomass and their high energy request [58,62–68]. Chemical pretreatment of biomass considers different chemical compounds to be applied in order to accomplish a disruption of biomass structure through interaction with intra and inter polymer bonds. The lignocellulosic biomass structure disruption results in cellulose crystallinity reduction, depolymerisation and degradation of cellulose and lignin breakdown. Alkali Hydrolysis is based on the solubilization of lignin in the alkali solution (NaOH, Ca(OH)<sub>2</sub>, NH<sub>3</sub>). Acid Hydrolysis takes advantage of the sensitivity of glucosidic bonds between hemicellulose and cellulose to acid (H<sub>2</sub>SO<sub>4</sub>, HCl, CH<sub>3</sub>COOH, HNO<sub>3</sub>). The organosolv process uses organic solvents for breaking down the bonds between lignin and hemicellulose. Wet Oxidation considers applying oxygen in the temperature range of 140–200 °C. It enhances contact between molecular oxygen and organic matter and thus achieves the complete degradation of organic compounds into CO<sub>2</sub>, H<sub>2</sub>O and simpler, more oxidized organic compounds into low molecular weight carboxylic acids [58,62–68]. Physico-chemical methods represent both mechanical forces and chemical effects applied to biomass. In the Steam Explosion, bioreactors are saturated with steam at 160–200 °C and 0.69–4.83 MPa. Steam condensation and its penetration into biomass causes sudden depressurization, breaking of glycosidic hemicellulose bonds and solubilization. Liquid Hot Water is similar to Steam Explosion, using water at elevated temperatures (170–230 °C) and pressure (up to 5 MPa). The pressure application is due to maintaining the liquid phase. The process hydrolyses hemicellulose by liberating its acetyl groups and removes lignin. Ammonia Fiber Explosion uses NH<sub>3</sub> at relatively high temperatures (90–100 °C) and moderate pressure for approximately 30 min, before its rapid release. The rapid release of pressure

disrupts the fibrous structure of biomass and decrystallizes the cellulose, hydrolyses hemicellulose, removes and depolymerises lignin, and increases the size and number of micropores and improves the accessibility of enzyme. The CO<sub>2</sub> Explosion uses supercritical CO<sub>2</sub> as a “greener” alternative compared to ammonia [69]. CO<sub>2</sub> acts as a solvent in the pretreated biomass, transforming it into glucose through the enzymatic hydrolysis of cellulose. The pretreatment process is not satisfactory to biomass with no moisture content, since CO<sub>2</sub> dissolved in water forms carbonic acid, which catalyses the hydrolysis of hemicellulose. Pressurized gas release is responsible for breaking the compact matrix structure of biomass [70]. Low temperatures are required, and no flammable or corrosive products generation are the main process advantages [58,62–68]. Biological pretreatments use microorganisms, i.e., fungi and bacteria (brown-, white- and soft-rot fungi, actinomycetes and bacteria) to degrade lignin and hemicellulose, while leaving the cellulose undamaged. Enzymes secreted by the fungi are responsible for lignin degradation. The main advantages of biological pretreatments refer to the absence of toxic substance generation and the low amount of required energy. However, the long duration of the process, the need for vigilant control of conditions and large space occupancy are the main disadvantages [58,62–68].

Thermal, thermal-alkaline, alkaline and electrochemical pretreatments are found to be the most effective for sewage sludge solubilizing. However, other pretreatment methods (such as ultrasonication, microwave and high-pressure homogenization) are being considered. When it comes to food waste, a hybrid biological-physiochemical treatment is suggested. Since the manure consists of lignocellulose fibres; its pretreatment is similar to that for energy crops/plant residues [71–73].

By applying different chemical additives, such as NaOH, Ca(OH)<sub>2</sub>, NH<sub>4</sub>OH, or H<sub>3</sub>PO<sub>4</sub>, process stability can be improved and up to 40% higher methane yield can be achieved [15]. According to Abraham et al. [74], enhanced biogas yield of 1200% can be achieved if ionic liquid is used in pretreatment process.

Rodriguez et al. [75] investigated mechanical, microwave, thermal, chemical and biological pretreatment methods applicable for grass structure, while one of the latest studies [76] includes convenient pretreatments for improving biogas production of macroalgae. *Fucus vesiculosus* and *Fucus serratus*, as the dominant species of brown seaweed growing in the Baltic Sea, were used there. The research included mechanical, microwave (600 W, 2 min), ultrasonic (110 V, 15 min), and microwave combined with ultrasonic (600 W, 2 min; 110 V, 15 min) treatment applied to the seaweed in order to track methane yields from AD. The results showed that the ultrasonic, ultrasonic combined with microwave, and microwave pretreatments increase cumulative methane yields of 167%, 185%, and 156%, respectively. The combined pretreatment showed a maximum methane yield of 260 mL/g of volatile solids after 20 days of digestion. The ultrasonic combined with microwave pretreatment showed a significant improvement in methane yield when compared with the mechanical pretreatment.

### 3.4. Biogas Purification and Upgrading

More review reports on biogas purification and upgrading were issued in the last decade, covering a comparison of different commercial upgrading techniques by their efficiency and barriers [77–79]. In a detailed review, Sun et al. [80] elaborated biogas purity, CH<sub>4</sub> recovery and loss, upgrading efficiency, investment and operating costs.

However, conventional biogas upgrading is based on physical/chemical technologies, which result in high CH<sub>4</sub> concentration (88–98%) and high removal efficiency for H<sub>2</sub>S, halocarbons, and siloxanes (min. 99%). Unfortunately, typical high energy and chemical demands affect environmental and economic sustainability [81]. Conventional biogas purification and upgrading can be divided into sorption and separation processes; they are shown in Tables 4 and 5.

**Table 4.** Biogas purification methods (modified to Adnan et al. [82]).

	Method	Working principle
Absorption	Water scrubbing	Different solubilities of H <sub>2</sub> S and CO <sub>2</sub> .
Adsorption	In situ chemical precipitation (iron salts)	Chemicals used dissolves sulphides into either insoluble metallic sulphide compounds or elemental S.
	Metal oxides (Al, Fe, Mn, Co, Cu, Zn oxides)	H <sub>2</sub> S adsorbs metal oxides by the sulphur binding as metal sulphide; Efficient in mercaptans removal.
	Activated carbon (impregnated, virgine)	Catalyze H <sub>2</sub> S oxidation into elemental S.
Biological		Microorganisms are used to convert H <sub>2</sub> S into S.
Membranes		Selective selectivity characteristics of CH <sub>4</sub> and H <sub>2</sub> S.
In-situ microaeration		Adding O <sub>2</sub> or air directly in digester.

**Table 5.** Biogas upgrading methods.

	Method	Working principle
Absorption	Water scrubbing	Different solubilities of H <sub>2</sub> S and CO <sub>2</sub> .
	Physical scrubbing	
	Chemical scrubbing (amines)	Chemical reactions of amine mixture with CO <sub>2</sub> .
Adsorption	Pressure swing adsorption (PSA)	Pressure changes in different process stages. VSA adsorption occur near atmospheric pressure.
	Vacuum swing adsorption (VSA)	
	Temperature swing adsorption	Temperature changes in different process stages.
Membranes (gas–gas separation, gas-liquid separation)		Selective selectivity characteristics of CH <sub>4</sub> and H <sub>2</sub> S.
Cryogenic separation		Different boiling temperatures of CO <sub>2</sub> and CH <sub>4</sub> .

Physical scrubbing means that CO<sub>2</sub> is dissolved in a liquid under pressure (various commercially available physical solvents), while chemical scrubbing means that CO<sub>2</sub> reacts with chemical solvents (amines) to separate it from CH<sub>4</sub>. The operation of scrubbing begins with removing hydrogen sulphide and other impurities. Water scrubbing is widely used for biogas upgrading. Water is used as a solvent for CO<sub>2</sub> because within it, there is a higher solubility of CO<sub>2</sub> than solubility of CH<sub>4</sub> [83,84].

Depending on the type of amine solution used, different reactions may occur. However, the most commonly used amine mixture is aMDEA (methyl diethanolamine + piperazine), as CO<sub>2</sub> absorption rate is quite high compared to other available amines.

Lee et al. [85] made a performance comparison of the removal efficiency of CO<sub>2</sub> and H<sub>2</sub>S from biogas by blending amine absorbents. The research, which included twelve different absorbents, showed that CO<sub>2</sub> absorption and regeneration rates were excellent in MDEA/piperazine, while H<sub>2</sub>S absorption and regeneration rates were excellent in MDEA/2-amino-2-methyl-1-propanol. MDEA-based blending absorbent showed better absorption and regeneration performance than MDEA, and MDEA/PZ showed good performance for CO<sub>2</sub> but very poor performance for H<sub>2</sub>S.

Adsorption is a mass transfer process, a phenomenon of sorption of gases or solutes by solid or liquid surface. It is often preferred in gas purification due to its simplicity, relatively low-cost, flexibility, high energy-efficiency (>85%), high selectivity and regeneration possibility [86–88]. Adsorption technologies and material commonly used are shown in Tables 4 and 5. Among such processes, pressure swing adsorption appears to be the most frequent. Separation of CO<sub>2</sub> and CH<sub>4</sub> is obtained based on the difference in their attraction to an adsorbent surface. A pressure increase results in adsorbed gas. Activated carbon is commonly applied due to low costs, high specific area and porous space, and H<sub>2</sub>S and CO<sub>2</sub> selectivity [87,89]. Furthermore, carbon can be produced from biomass residues and modified with metal oxides representing additional benefit. Sawalha et al. [90]

published the study on removing H<sub>2</sub>S from biogas by adsorption using synthesized activated carbon prepared of locally available biomass. The best results in removal efficiency were achieved with eucalyptus (690 mg H<sub>2</sub>S/g adsorbent), almond shells (230 mg H<sub>2</sub>S/g adsorbent) and coffee grains (22 mg H<sub>2</sub>S/g adsorbent).

Besides carbon, zeolites are also preferable due to their high selectivity, regenerability, thermal resistance and separation of H<sub>2</sub>S through chemisorption at very low partial pressures and physisorption at low to moderate partial pressures [91,92].

Furthermore, research studies on metal-organic-frameworks (MOFs) have shown their great adsorption and selection properties toward H<sub>2</sub>S and CO<sub>2</sub> [93,94]. Nowadays, the MOFs are subject to various research on their application in different gas separation processes. Taddei & Petit [95] made a review on technical information on MOF design, synthesis, characterisation and sorption testing. According to Alonso et al. [96], biogas purification and CH<sub>4</sub> storage would become a new motivation for the development of nanomaterials as sorbents. Their review on the current use of novel nanomaterials as adsorbents for CO<sub>2</sub> and CH<sub>4</sub> shows that materials modified with amine or metals are most perspective. The Fe<sub>3</sub>O<sub>4</sub>-graphene and the MOF-117 based nanoparticles show the greatest CO<sub>2</sub> sorption capacities, due to their high thermal stability and high porosity. IRMOF-6, MOF-177 and MOF-5 showed the highest adsorption capacities for CH<sub>4</sub>.

Separation methods consider different membrane technologies and cryogenic separation. Membrane separation is based on a permeation process. This technology is well known and used in natural gas processing [97,98]. Comparison analysis made for organic, inorganic and hybrid materials, shows advantages of inorganic materials in terms of temperature stability, permeation rates and selectivity [99].

Cryogenic (low-temperature) purification of biogas uses very low temperature (−70 °C) and high pressure (approx. 40 bar) to remove the various undesirable components, including CO<sub>2</sub>, H<sub>2</sub>S, water vapour and siloxane. Process principles are based on the differences in the constituents' boiling temperature. Different gases condense at different temperature-pressure domains. Usually, the main gas stream is cooled down in a multi-stage refrigeration sequence, below a certain cryogenic temperature according to operating pressure. Acid components are concentrated in the liquid effluent at the bottom of distillation column. Numerous and expensive process equipment, such as compressors, turbines, heat exchangers and coolers, represent the main disadvantages of this method [100–102].

Modern developments in biogas purification and upgrading techniques have been reviewed by different authors [14,81,103]. Rodero et al. [81] stated that technologies which use aerobic and anoxic biotrickling filters, algal-bacterial photobioreactors and digesters under microaerobic conditions, successfully remove H<sub>2</sub>S (>99%). They also made a comparison of various modern technologies for the biogas purification of H<sub>2</sub>S, halocarbons, and siloxanes. In the same article the authors stated that biological CO<sub>2</sub> removal processes, including microalgae-based CO<sub>2</sub> fixation, H<sub>2</sub>-assisted lithoautotrophic CO<sub>2</sub> bioconversion to CH<sub>4</sub>, enzymatic CO<sub>2</sub> dissolution or fermentative CO<sub>2</sub> reduction, have recently appeared as cost effective and environmentally friendly solutions, resulting in CO<sub>2</sub> removal of 80–100% and CH<sub>4</sub> purities of 88–100%.

### 3.5. Biogas Usage

The very first examples of biogas usage date back to the Assyrian empire, about 3000 years ago. The biogas, obtained by decomposing organic material, had been initially used for water heating. Later, in the 17th century, more intensive research of biogas and its properties began. The first modern digestion plant was recorded in India in the second half of the 19th century [104,105]. A literature review on papers published over the last five years with the topic of biogas usage possibilities and the related issues are shown in Table 6.

**Table 6.** Biogas usage.

Biogas usage	Topics elaborated in recent publications on biogas related issues
Production of heat or steam	Advantages of biogas cook stoves in wood scarcity and agricultural regions [106]; Life cycle energy and cost analysis of small-scale biogas facilities in rural regions [107]; Design features, construction material, feedstock, and operation parameters that made anaerobic digestion in small digesters [108]; Optimisation of techno-economic design of biogas digester [109].
Electricity generation or Combined Heat and Power (CHP)	Modelling and simulations of demand-orientated biogas-based power production [110]; Modelling of rural multi-energy complementary system with biogas cogeneration and electric vehicle considering carbon emission and satisfaction [111]; Load response of biogas CHP systems in a power grid [112]; Comparative analysis of different CHP systems using biogas for the cassava starch plants [113].
Vehicle fuel	Perspectives of biogas conversion into Bio-CNG for automobile fuel, Bangladesh case study [114]; Comparative analyses of the impact of biogas and biogas with hydrogen fuels on vehicle emissions and performance [115]; Sugarcane bagasse and straw as a biofuel to propel light vehicles [116]; System perspective on biogas use for transport and electricity production. Energies [117].
Fuel cell	Potential and constraints of solid Oxide Fuel Cells fuelled with biogas [33]; Biogas production and usage with legislations framework across the globe [15]; Biogas fed-fuel cell-based electricity generation: a life cycle assessment approach [118].

As per the IEA Report 2020 [32], about two-thirds of biogas produced globally goes for electricity and heat purposes, with an almost equal share of electricity and co-generation facilities. About 30% of biogas was consumed by the residential sector globally, while the residual share refers to biogas upgraded to biomethane, injected into a gas grid or used as fuel.

Small biogas systems are conducive in developing countries, where the rural population commonly depends on traditional biomass for cooking and heating. As the energy strategies commonly include small bioreactors, different governments and non-governmental institutions have supported their applications in Asia, South America and Africa. Nowadays, there are about 50 million small biogas systems [106,119,120]. Pilloni & Hamed [108] gave a review on design features, construction material, feedstock and operation parameters that make anaerobic digestion in small digesters attractive.

There is a special benefit in the possibility of the adjustment of conventional gas burners for biogas by changing the air-to-gas ratio. The only requirements of gas burners are a gas pressure of 8 to 25 mbar and a concentration of H<sub>2</sub>S below 100 ppm to achieve a dew point of 150 °C [121].

The current increase in renewable electricity production requires power balancing, where demand-orientated combined heat and power production from biogas can be a suitable option. Grim et al. [110] analysed agricultural facilities, using cattle manure and sugar beet for biogas and CHP production. A developed dynamic production model showed that feeding management reduced the storage requirement due to the fast biogas production response to feeding. Furthermore, Ishikawa et al. [112] addressed the issue of load response with biogas CHP systems in power grid. They made an analysis of the systems operating in Germany and Japan that included load adjustment, start-up and shut-down behaviour, as well as exhaust gas emissions measurements during supply-demand adjustment. As concluded, biogas CHP systems have the potential to contribute to a long-term power equivalent adjustment, such as tertiary control reserve (TCR).

CHP systems use biogas and waste heat, resulting in much higher efficiency than using fuel to produce only power or heat. Gas turbines (in the range from micro to >100 kW) or internal combustion engines are commonly engaged. Such systems require removal of water vapour and H<sub>2</sub>S (<100 ppm) from biogas [121]. Yin et al. [113] analysed

energy demands for cassava starch plant. Results showed the ability to meet energy needs and economic feasibility for both the CPH system with reciprocating internal-combustion engine (ICE-CHP) and micro turbine (MT-CHP). Due to the lowest environmental impact, and the lowest and short payback period (4.57 year) ICE-CHP is suggested as first choice.

Although being intermittent, renewables are abundant in rural areas, and therefore a system that combines different energy sources can be recognized as a convenient option. Fan et al. [111] proposed a rural multi-energy optimisation model combining different technologies, where biogas cogeneration units improve system scheduling flexibility. The proposed model poses economic, environmental, and service benefits.

Future biomethane application relies on fuel cells as small-scale power plants of high efficiency exceeding 60% and low emissions. One of the largest digester/fuel cell units can be found at the South Treatment Plant in Renton, Washington, USA. About 4400 m<sup>3</sup>/day of biogas produces 1 MW of electricity for site operation purposes [122]. Biogas can be converted into different renewable transport fuels, i.e., bio-CNG, liquefied biogas, syngas, and gasoline. A whole range of merits can be connected to the production of biogas as biofuels. However, the most emphasized are environmental benefits related to the reduction of GHG emission in the transport sector, generally from 60% to 80% compared to gasoline [106,114,123].

Shah et al. [114] gave an overview of available liquid and gaseous fuel commonly used as transportation fuel in Bangladesh. If the methane share in biogas is greater than 97%, biogas can be compressed at a pressure of 20–25 MPa to biomethane (Bio-CNG). Like regular CNG, bio-CNG shows high thermal efficiency due to a high octane number [124]. Special pertinence is that both petrol and diesel vehicle engines can be converted to bio-CNG through retrofitting with a gas tank and a gas supply system, in addition to the conventional petrol fuel system [125].

Biomethane can be liquefied (LBM). Achieved fuel has similar characteristics to LNG and is convenient for conventional vehicle engines. However, lower energy content of LBM (70% of gasoline) and required energy intensive methods, such as cryogenic technology, liquefaction, and pressure letdown, negatively impact the economics [114].

Syngas, as a mixture of hydrogen, carbon monoxide and a small amount of carbon dioxide, can be produced from pure biogas through reforming processes, such as dry reforming, steam reforming, and partial oxidative reforming, or any of their combination. Steam reforming is a widespread commercial process used to produce high purity hydrogen as clean transportation fuel [126,127].

Syngas can be converted into methanol and dimethyl ether and further upgraded into transport fuel like diesel, jet fuel or gasoline through the Fischer–Tropsch synthesis method [128]. Gasoline is produced by upgrading methanol through the methanol-to-gasoline (MTG) process. Biogasoline has a higher energy content and higher vapour pressure compared to biomethanol. Biogasoline can be directly used as transport fuel in a vehicle, or it can be mixed with biomethanol. Recently, partial oxidation has emerged as the most attractive and commonly used technology for biogasoline production

Karagöz [115] used a numeric engine model empirically validated by the global standard procedures for determining the levels of emissions and fuel consumption, LTC and NEDC. The research showed positive results when a low ratio of hydrogen (5% as molar) was added to biogas fuel. With the addition of H<sub>2</sub>, the increase in fuel consumption dropped to 12.1% (compared to 16.1% in the case where pure biomethane is used). On the other hand, the increase in CO emissions decreased from 21.6% to 11.7% during the NEDC cycle.

Hydrogen derived from biogas can directly feed fuel cells. The reforming process can be succeeded by either internally employing fuel cells or externally with a catalytic pre-reformer. Abanades et al. [15] sees fuel cells as the cutting-edge application of biogas. Fuel cells utilization ranges from large-scale power plants to small scale portable power supply microelectronic equipment [129]. Although fuel cell-based power plants using biogas are not common in industrial application, more pilot projects are being initiated [33]. Their

integration with other power generation systems like gas turbines or microgas turbines improve their performance. A biogas fuelled integrated solid oxide fuel cell (SOFC)-CHP represents a new efficient system that could surpass the issues of heat and power demands for decentralized grids [130,131].

#### 4. Biogas Production in EUROPE

The EU has been recording an increase in the use of bioenergy, which is mainly due to the higher use of biogas rather than solid biomass. Biogas for electricity generation, heating and cooling took 8.3% (394.5 PJ) of the total bioenergy used in 2017, which means a five-time increase compared to 2005. Almost 60% of biogas is used for electricity production [132,133].

As for the size of the installed biogas plants capacity in 2017, the figure has tripled since 2005. Table 7 shows the total number of biogas and biomethane plants in the EU in the last decade. As seen, the number of biogas plants manifestly increased from less than 10,500 in 2010 to almost 19,000 in 2020 [134]. The increase in the number of facilities is accompanied by an increase in biogas production. As per the 19th EU Observ's Report [135], the European Union produced 16.8 Mtoe of biogas in 2018. In the meantime, since the UK withdrew from the EU, biogas production in the EU 27 was 13.8 Mtoe in 2018.

**Table 7.** Total number of biogas and biomethane facilities in the EU in the period 2010–2020 (according to [132,135]).

Year	2010	2012	2014	2016	2018	2020
Biogas facilities	10,508	13,812	16,834	17,432	18,943	Approx. 19,000
Biomethane	187	232	367	477	627	880

The Biogas barometer 2020 published the biogas production data by the EU member states in 2018. Analysis of published data is made and represented by Figures 5–7. Germany, UK, France and Italy are the leaders in biogas production in Europe (Figures 5 and 6). Analysis of the report data shows that Germany stands out as the most important producer, producing about 44% of the total biogas quantities in 2018. The United Kingdom participated with 17%, while Italy and France took shares of 12% and 4%, respectively. Among other countries, the Czech Republic and Denmark show a notable shift in production. Biogas from AD processes is most represented in the EU, participating with a high 84% of total biogas production. Germany, France and Italy produce a significant part of biogas by other AD processes, amounting to 98%, 79% and 62%, respectively, of their national biogas production (Figure 7). AD gas production in Germany far surpasses the production of other member states, amounting 52% of total biogas production by AD process in the EU. Landfill gas production is well represented in the UK, participating with 43% of total national biogas production. With regards to the EU, more significant landfill production can be found in Italy, France and Spain. Italy produces 27% of total landfill biogas in the EU, which is 18% of the country production. France participates with 24% of biogas from landfills, which is 34% of the country production. A high presence of landfill biogas locations can be found in Spain, where landfill production participates with 56%. Spain produces about 12% of landfill biogas in the EU. Biogas from sewage sludge is presented by less than 5% in total biogas quantities produced in the EU. Poland, Spain and the Netherlands take advantages of sewage sludge for biogas production. Sewage sludge participates with 40%, 25% and 18%, respectively, in their domestic production.



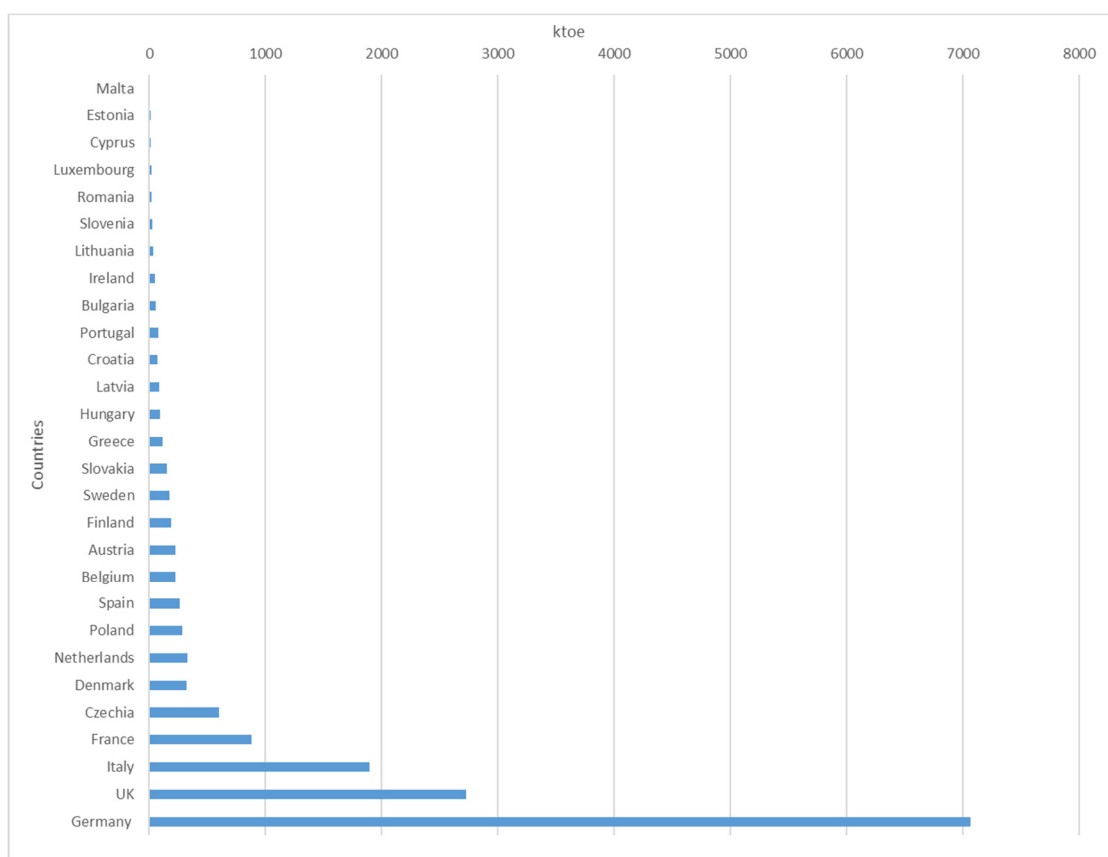


Figure 5. Biogas production in ktoe by countries (according to [135]).

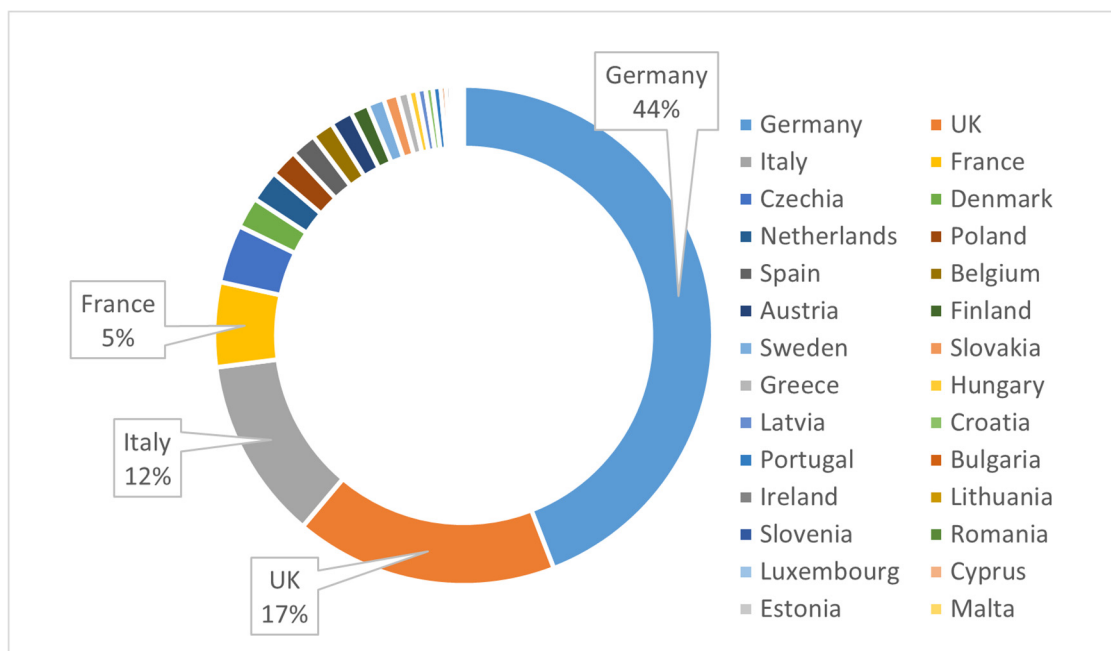
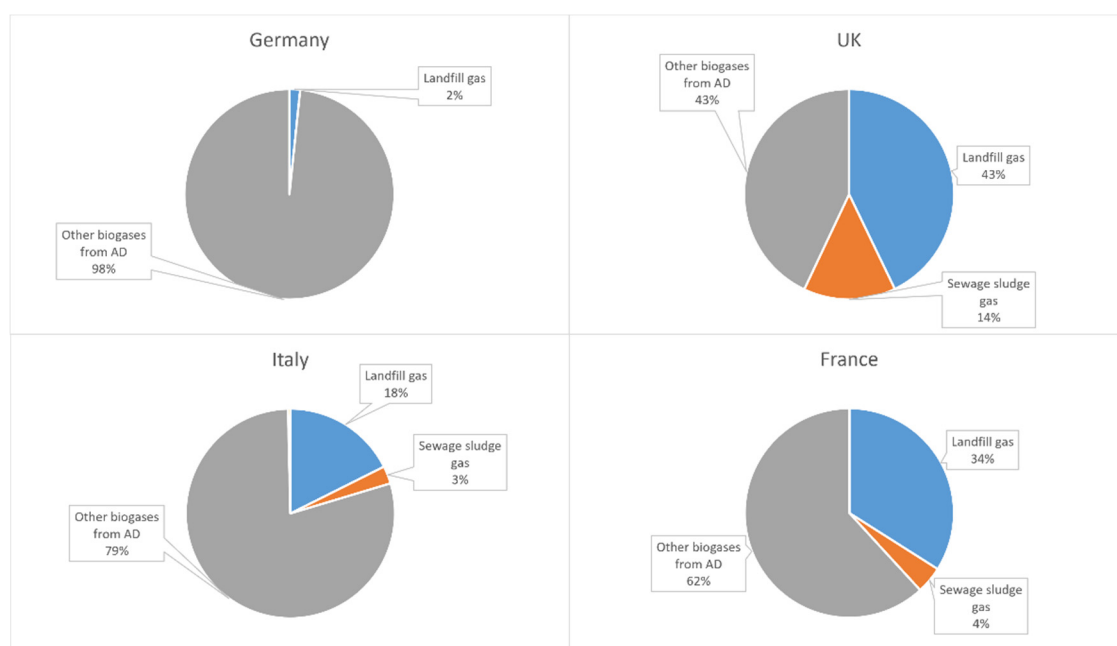


Figure 6. Related shares in biogas production (according to [135]).



**Figure 7.** Leading countries in biogas production (according to [135]).

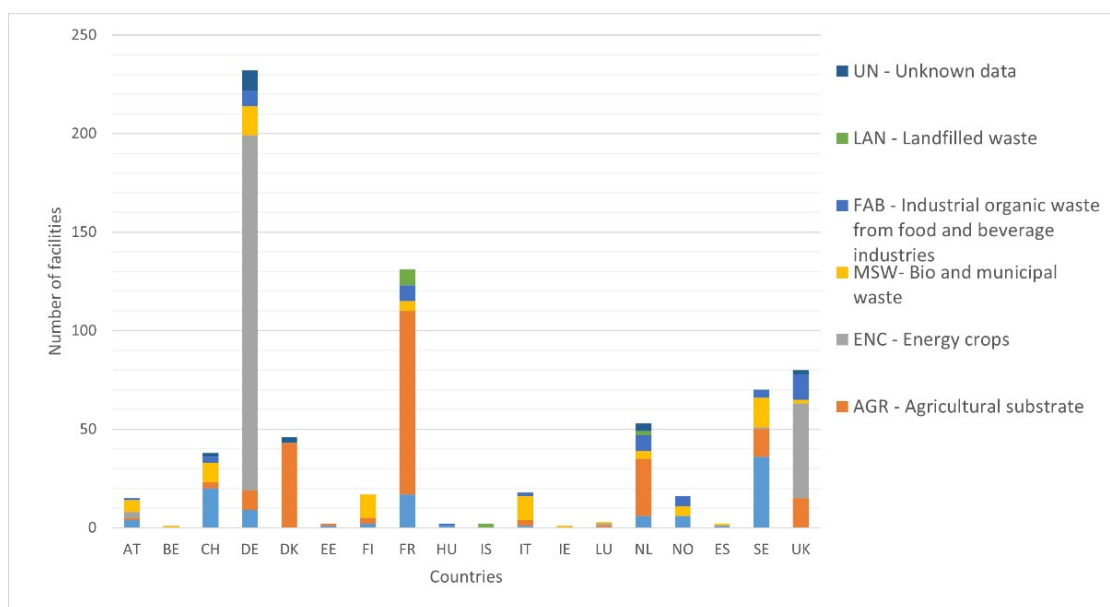
The increase in biomethane facilities is even steeper. In the last decade, the number of biomethane facilities in Europe have advanced from 483 plants in 2010 to 880 plants operating in 2020, reaching a total of 32 TWh of biomethane in Europe. Particularly significant progress can be seen in an only two-year period, from 2018 to 2020, when the number of biomethane facilities increased rapidly by more than 50%. There are currently 18 countries producing biomethane in Europe. Germany has the highest share of biomethane plants (232), followed by France (131) [136].

The Biomethane Map 2020 [136] published by the European Biogas Association (EBA) and the Gas Infrastructure Europe (GIE) shows data on infrastructure for biomethane production. Data analysis on substrate use was made and presented in Figure 8. It can be concluded that most of the plants installed in the EU use agricultural substrate as feedstock. Agricultural substrate (AGR), which includes agricultural residues, manure and plant residues represents the base for 28% of facilities. Energy crops (ENC) are used in 25%, while sewage sludge and waste (SWW), and bio and municipal waste (MWW) are feedstock for respectively 14% and 12% of the total biomethane facilities operating in the EU.

Production in Germany is highly dependent on energy crops. There, about 78% of crop-based plants are installed, while bio and municipal waste substrates have the second position (6% of installed facilities in Germany), slightly higher than other substrate alternatives. About 71% of biogas plants installed in France base their production on substrates coming from agricultural production except for energy crops, which includes manure, straw, cover/catch crops and crop residues. Sewage sludge facilities are represented by 14% in the EU, mostly applied in Sweden. Sweden has about 37% of the total number of such facilities installed in the European Union. Industrial organic waste from food and beverage industries (FAB) is the base for 5% of biomethane facilities, while landfilled waste based biomethane facilities (LAN) participate with only 2%. For about 2% of biomethane facilities there was no data (marked as UN in Figure 8) on the substrate presented on the Biomethane Map 2020 that was used.

Among other member states, growth in biomethane plants can be seen in Denmark and the Netherlands, having respectively 7% and 6% of the total number of installed plants in the EU. The Czech Republic has 38 facilities installed, which represent 5% of the total installed plants.

Substantial biomethane production is seen in non-EU member states, UK and Norway. In the UK, there were 80 biomethane facilities, mostly relying production on energy crops (60%).

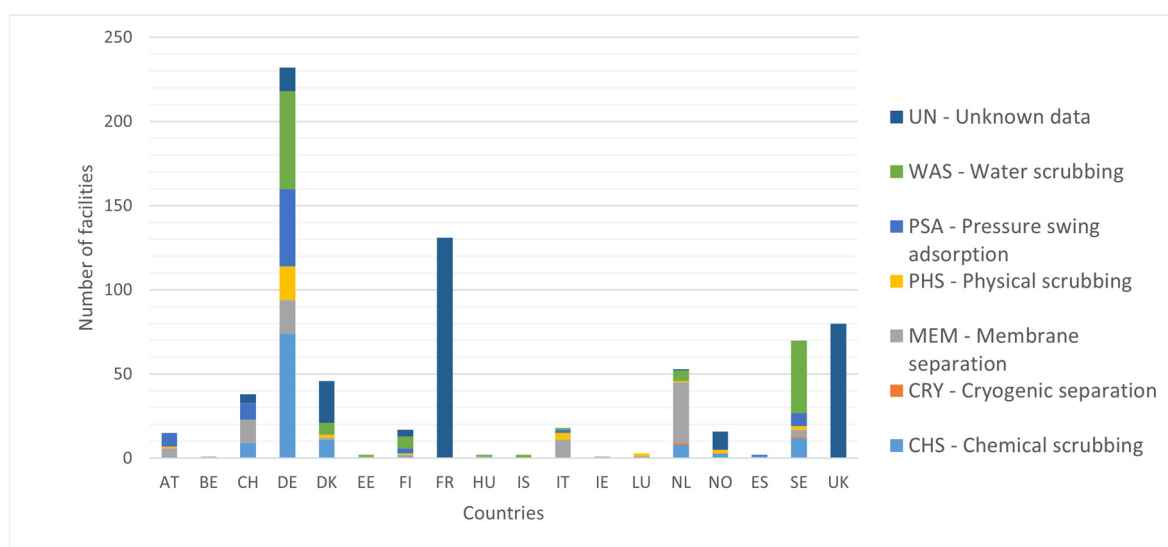


**Figure 8.** Distribution of facilities per feedstock in EU member states, UK and Norway (according to [136]).

The Biomethane Map 2020 [136] also shows data on infrastructure for biomethane production. Data analysis on the upgrading process used in 18 countries—biomethane producers in Europe, which means 16 EU member states, the UK and Norway were included and are presented in Figure 9. The facilities are grouped by upgrading the process used into chemical scrubbing (CHS), cryogenic separation (CRY), membrane separation (MEM), physical scrubbing (PHS), pressure swing adsorption (PSA) and water scrubbing (WAS). However, there is an evident lack of published data for some countries, which means that about one third of the data is missing (marked as UN data on diagram), mostly related to the UK and France. Since France and the UK belong to the European leaders in biomethane production, general conclusions cannot be made. Nevertheless, it can be seen that in other member states (data for France is not included) CHS and WAS processes are almost equally presented, with a light higher contribution of WAS (almost 20% compared to 19% of CHS). On the other hand, Germany, as the largest producer relies on CHS (34%), followed by WAS (27%) and PSA (21%). WAS is also highly present in Sweden with 61% of upgrading.

However, the same order in representation of individual technologies in biomethane purification processes in Europe is given by Niesner et al. [83], who stated that WAS was the most prevailing technology, applied at almost 40% of all biomethane plants.

MS appears as the main basis for biomethane production in the Netherlands, amounting to more than 69% or 42% in the Czech Republic. With regards to CRY technology, there are 2 facilities operating in the Netherlands and Sweden.



**Figure 9.** Distribution of facilities per upgrading process in EU member states, UK and Norway (according to [136]).

## 5. Discussion

In Europe, AD technology for biogas production was applied during the Second World War, when the lack of energy-initiated thoughts of switching to alternatives. Later, being less competitive than fossil fuels, biofuels were marginalized. A non-systematic approach to the biogas issue has resulted in its further neglect. Biogas solutions were just sporadically mentioned in the “Waste Directive” [137] and the “EU Bioeconomy Strategy” [138]. Recently, the situation has started to change. The EU targets adapted in the “REDII Directive” [4] affect biogas indirectly by requiring the share of energy from renewables of at least 32% and 14% of renewables in transport fuel to be achieved in 2030. Today, biogas production is a part of circular economy [139–141], with a central position in new European Green Deal strategies, such as the EU Methane Strategy [9], The Energy System Integration Strategy [10] and the “From Farm to Fork-Strategy” [11]. Furthermore, the recently proposed framework to decarbonize EU gas markets set the rules for the easier integration of biomethane in the gas grid with a legislative proposal to reduce methane emissions in the energy sector.

As per the EBA Statistical Report [15], membrane separation is still the leading upgrading technology, applied at more than 35% of biomethane facilities. The use of cryogenic separation is in its early development phase with 11 new plants in 2020. Currently, thermal gasification is experiencing an early commercialization phase, and hydrothermal gasification is being demonstrated with pilot projects located in France and the Netherlands [35,36,142]. Most conventional technologies applied to natural gas processing are efficient in biogas cleaning and upgrading. However, if separated CO<sub>2</sub> is vented into the atmosphere, the environmental benefit of the applied technology is questionable.

Due to that, biological CO<sub>2</sub> removal processes (e.g., microalgae-based CO<sub>2</sub> fixation, enzymatic CO<sub>2</sub> dissolution or fermentative CO<sub>2</sub> reduction) have to be the focus of future research [14,81,103].

Same “green trend” should be applied in each stage of biogas production. Process pretreatments, such as mechanical, microwave, thermal, chemical or biological, are available for both extending feedstock options and improving AD process efficiency. Today, due to the use of chemicals and waste liquid generation, some commonly applied pretreatment methods (hydrothermal, acid, alkali pretreatments) are not considered as green options. The feedstock for biogas production may contain crystallized cellulose and hemicellulose polymer matrix that is encrusted by the highly polymerized phenolic lignin causing difficulties in the conversion processes. Therefore, there is a need for other inno-

vative pretreatment methods in order to facilitate biological digestion. Ecologically acceptable options are needed. Ultimate research showed that great biogas yield can be achieved if ionic liquid is used in the pretreatment process [74]. While high cost and immature recovery technologies limit its usage, electrochemical pretreatment of biomass appears as a promising tool for the utilization of lignocellulose.

While biogas production from locally available feedstock represents a generally sustainable energy model, special attention must be paid to the occupation of a huge surface of arable land with the large-scale production of crops intended for energy purposes. Therefore, new regulations are focused on incentives that favour the use of organic and farming waste. The first generation of biofuels with a high risk of indirect land use change (ILUC) when land purpose is converted from grasslands and forests to crop production, is going to be phased out by 2030.

Generally, biogas production and utilization are present in Europe and North America. There were about 1000 biomethane facilities operating worldwide in 2021. Besides the markets developed in Europe and the United States, the biomethane market is under development in Brazil, Canada, China and India [142].

Combined biogas and biomethane production in Europe exceeded 191 TWh of energy in 2020, and it is expected to be doubled by 2030, while by 2050, production can achieve over 1000 TWh). The great availability of biodegradable substrate ensures biomethane potential in Europe at the level of 1350 TWh [36]. Over the past decade, there was an increase of 90% in biogas industry (120 GW installed in 2019), and Europe contributes with over 70% of the world biogas generation. Developed countries have installed advanced large-scale facilities for utilizing biogas [15]. About two-thirds of biogas produced globally goes for electricity and heat purposes, approximately 30% is consumed by the residential sector, while a small residual share refers to biogas upgraded to biomethane, injected into the gas grid or used as fuel [31]. Biogas has usually been used in the production location vicinity for electricity and heat generation purposes. Produced electricity is usually consumed locally, or it can be handed over to the grid. Even the most promising forecast of biomethane upgrading considers consuming 60% of biogas in this way by 2030 [13], because high capital and operating costs were the main obstacles for more significant developments, especially when compared to the relatively low natural gas price. The natural gas price had stayed relatively low and stable for a decade. The current situation could be changed by the latest Ukraine crisis that dramatically affected the energy market. Although the need for renewables is well recognized in the EU energy strategy and action plans, strong dependence on the gas import from Russia requires faster actions in supply source diversification. This will certainly result in higher gas prices. In such circumstances, biomethane production can become a cost-effective option.

Process input varies significantly by region, since it is imitated to locally available materials that have acceptable transportation costs. Germany, UK, Italy and France are the leaders in biogas production in Europe. Biogas from AD processes is most represented in the European Union, participating with a high 84% of total biogas production, and agricultural-based biogas and biomethane facilities dominate production. Germany is deserving for the majority (52%) of AD biogas in the EU. Landfill gas production is well represented in the UK (43% of total national biogas production). More significant landfill production can also be found in the EU member states Italy, France and Spain. Biogas from sewage sludge is poorly presented by less than 5% in total biogas quantities produced in the EU. Sewage sludge practice is applied in Poland Spain and the Netherlands.

Biomethane facilities reached a production of 32 TWh in 2020 in Europe. There are currently 18 countries producing biomethane. Germany has the highest share of such plants (232), followed by France (131). Most of the European plants use agricultural substrate (28%), while the second position refers to energy crop feedstock (25%). Germany production relies on energy crops (78%) and biomethane production in France is highly dependent on substrates coming from manure, straw, cover/catch crops and crop residues (71%). Sewage sludge facilities participate with 14% in the EU, mostly applied in Sweden

[15,136]. While Germany, Great Britain, France and Italy stand out as the largest producers of biogas in Europe, most of the countries are in the early development stage. One of them is the Republic of Croatia, where biogas production and use is in its early phase, while biomethane production realization is planned for the future, as per the new national strategy for development of energy sector for the period 2030, with a view to 2050.

The first biogas plant in Croatia with installed power of 1 MW started to operate in 2009. According to the Report 2020 issued by Croatian Energy Market Operator (HROTE), in Croatia there are 41 biogas power plant with overall installed power capacity of 45.9 MW. Power plants with installed capacity of 1 MW are predominant. Commonly used feedstocks are manure (5–60%) and corn or grass silage (25–35%). The facilities are mostly located in the continental part of Croatia.

The analysis confirmed that biogas production and utilization differ significantly by region. Non-aggressive progress in the development of the biogas industry has been driven for a long time by specific states that recognized the value of biogas, primarily in Europe and North America.

Nevertheless, the implementation of adequate incentive is essential for the further development of the biomethane market. Irrespective of significant initiatives and the evident development of common biogas policy, it is obvious that specific instruments for biogas development stimulation may not be directly transferred to all countries since they are strongly dependent on regional and local conditions, mainly driven by energy system differences, but also the available substrate options.

## 6. Conclusions

The main advantages of using biogas and biomethane are its contribution to baseload power generation and reducing energy dependence. Biogas is suitable as a substitute for fossil fuels in power generation (both small and large scale), transportation and household sectors. Although the production of biogas is constantly increasing in European countries, the energy is mostly used for heat and electricity. The profit from biogas would be better if the biogas was purified and sold as fuel. In this way, less greenhouse gases are released into the atmosphere. Emission reduction can be considered from different points of view: Fossil fuel replacement, avoidance of methane emissions from animal or plant manure, production of green manure, etc. Overall, the benefits lead to a negative carbon footprint. However, the multiple benefits of biogas are not usually recognised as a value, resulting in the price of the final product being much higher than that of fossil natural gas. For this to change, the technologies for upgrading biogas to biomethane must be continuously improved. Recent geopolitical relations have led to instability in the natural gas market, resulting in the highest natural gas prices ever. This could also be the trigger for faster and greater inclusion of biogas and biomethane in the future energy mix.

Decarbonized Europe considers renewable electricity and renewable and low-carbon gases. Strong policy and regulatory support for the promotion of biomethane production and clarification of market rules are needed to achieve these ambitious goals.

**Author Contributions:** J.P.—created the idea for the paper, provided biogas and biomethane technology review, participated in the discussion and conclusion section; K.N.M.—wrote discussion section, participated in analyses and prepared the figures and tables, V.B.—participated in analyses and reviewed the whole paper, K.S.—provided biogas purification section and reviewed the whole paper. All authors have read and agreed to the published version of the manuscript.

**Funding:** The dissemination process is financially supported by the University of Zagreb within the project “Circular economy in petroleum engineering” (In Croatian: Kružna ekonomija u naftnom inženjerstvu—KENI).

**Institutional Review Board Statement:** Not applicable.

**Informed Consent Statement:** Not applicable.

**Data Availability Statement:** Data available in a publicly accessible repository.

**Conflicts of Interest:** The authors declare no conflict of interest.

## References

1. European Commission (EC). *The European Green Deal. European*; Commission: Brussels, Belgium, 2019. Available online: [https://ec.europa.eu/info/publications/communication-european-green-deal\\_en](https://ec.europa.eu/info/publications/communication-european-green-deal_en) (accessed on 15 April 2022).
2. European Commission (EC). *Fit for 55: Delivering the EU's 2030 Climate Target on the Way to Climate Neutrality*; European Commission: Brussels, Belgium, 2019. Available online: <https://www.eesc.europa.eu/en/our-work/opinions-information-reports/opinions/fit-55-delivering-eus-2030-climate-target-way-climate-neutrality> (accessed on 15 April 2022).
3. Catuti, M.; Egenhofer, C.; Elkerbout, M. *The Future of Gas in Europe: Review of Recent Studies on the Future of Gas*; CEPS Research Report No. 2019/03; CEPS Energy Climate House: Brussels, Belgium, 2019.
4. European Commission. Directive (EU) 2018/2001 of the European Parliament and of the Council of 11 December 2018 on the Promotion of the Use of Energy from Renewable Sources (Text with EEA Relevance). 2018, p. 328. Available online: [https://eur-lex.europa.eu/legal-content/EN/TXT/?uri=uriserv:OJ.L\\_.2018.328.01.0082.01.ENG](https://eur-lex.europa.eu/legal-content/EN/TXT/?uri=uriserv:OJ.L_.2018.328.01.0082.01.ENG) (accessed on 13 April 2022).
5. Panoutsou, C.; Germer, S.; Karka, P.; Papadokostantakis, S.; Kroyan, Y.; Wojcieszyn, M.; Maniatis, K.; Marchand, P.; Landalv, I. Advanced biofuels to decarbonise European transport by 2030: Markets, challenges, and policies that impact their successful market uptake. *Energy Strat. Rev.* **2021**, *34*, 100633. <https://doi.org/10.1016/j.esr.2021.100633>.
6. Directive 98/70/EC of the European Parliament and of the Council of 13 October 1998 Relating to the Quality of Petrol and Diesel Fuels and Amending Council Directive 93/12/EEC. Available online: <https://eur-lex.europa.eu/legal-content/EN/ALL/?uri=CELEX%3A31998L0070> (accessed on 23 March 2022).
7. Directive 2009/28/EC of the European Parliament and of the Council of 23 April 2009 on the Promotion of the Use of Energy from Renewable Sources and Amending and Subsequently Repealing Directives 2001/77/EC and 2003/30/EC. Available online: <https://eur-lex.europa.eu/legal-content/EN/ALL/?uri=CELEX%3A32009L0028> (accessed on 23 March 2022).
8. ePURE. *2022 National Biofuels Policies*; ePURE: Brussels, Belgium, 2022. Available online: <https://www.epure.org/news/updated-for-2022-national-biofuels-policies-across-the-eu/> (accessed on 23 March 2022).
9. European Commission (EC). *Communication on an EU Strategy to Reduce Methane Emissions*; European Commission: Brussels, Belgium, 2020. Available online: <https://eur-lex.europa.eu/legal-content/EN/TXT/?uri=CELEX%3A52020DC0663> (accessed on 15 April 2022).
10. European Commission (EC). *Powering a Climate-Neutral Economy: An EU Strategy for Energy System Integration*; European Commission: Brussels, Belgium, 2020. Available online: <https://eur-lex.europa.eu/legal-content/EN/ALL/?uri=COM:2020:299:FIN> (accessed on 23 March 2022).
11. European Commission (EC). *A Farm to Fork Strategy for a Fair, Healthy and Environmentally-Friendly food System*; EC: Brussels, Belgium, 2020. Available online: <https://eur-lex.europa.eu/legal-content/EN/TXT/?uri=CELEX%3A52020DC0381> (accessed on 23 March 2022).
12. Gustafsson, M.; Anderberg, S. Biogas policies and production development in Europe: A comparative analysis of eight countries. *Biofuels* **2022**, 1–14. <https://doi.org/10.1080/17597269.2022.2034380>.
13. Stern, J. *Narratives for Natural Gas in Decarbonising European Energy Markets*; The Oxford Institute for Energy Studies, Oxford, United Kingdom, 2019. Available online: <https://www.oxfordenergy.org/publications/narratives-natural-gas-decarbonising-european-energy-markets/> (accessed on 23 March 2022).
14. Korbag, I.; Omer, S.M.S.; Boghazala, H.; Abusasiyah, M.A.A. Recent Advances of Biogas Production and Future Perspective. In *Biogas—Recent Advances and Integrated Approaches*; Abomohra, A.E., Elsayed, M., Qin, Z., Ji, H., Liu, Z., Eds.; IntechOpen: London, UK, 2020.
15. Abanades, S.; Abbaspour, H.; Ahmadi, A.; Das, B.; Ehyaei, M.A.; Esmaeilion, F.; Assad, M.E.H.; Hajilounezhad, T.; Jamali, D.H.; Hmida, A.; et al. A critical review of biogas production and usage with legislations framework across the globe. *Int. J. Environ. Sci. Technol.* **2021**, *19*, 3377–3400. <https://doi.org/10.1007/s13762-021-03301-6>.
16. European Biogas Association (EBA). *EBA Statistical Report 2021*; EBA: Brussels, Belgium, 2022. Available online: <https://www.europeanbiogas.eu/eba-statistical-report-2021/> (accessed on 23 March 2022).
17. Fang, Z. (Ed.) *Liquid, Gaseous and Solid Biofuels—Conversion Techniques*; IntechOpen: London, UK, 2013.
18. Olugbade, T.; Ojo, O.; Mohammed, T. Influence of Binders on Combustion Properties of Biomass Briquettes: A Recent Review. *BioEnergy Res.* **2019**, *12*, 241–259. <https://doi.org/10.1007/s12155-019-09973-w>.
19. Olugbade, T.O.; Ojo, O.T. Biomass Torrefaction for the Production of High-Grade Solid Biofuels: A Review. *BioEnergy Res.* **2020**, *13*, 999–1015. <https://doi.org/10.1007/s12155-020-10138-3>.
20. Olugbade, T.O.; Ojo, O.T. Binderless briquetting technology for lignite briquettes: A review. *Energy Ecol. Environ.* **2021**, *6*, 69–79. <https://doi.org/10.1007/s40974-020-00165-3>.
21. Zhu, P.; Abdelaziz, O.; Hultberg, C.P.; Riisager, A. New synthetic approaches to biofuels from lignocellulosic biomass. *Curr. Opin. Green Sustain. Chem.* **2020**, *21*, 16–21. <https://doi.org/10.1016/j.cogsc.2019.08.005>.
22. Awogbemi, O.; Von Kallon, D.V.; Onuh, E.I.; Aigbodion, V.S. An Overview of the Classification, Production and Utilization of Biofuels for Internal Combustion Engine Applications. *Energies* **2021**, *14*, 5687. <https://doi.org/10.3390/en14185687>.

23. Mahapatra, S.; Kumar, D.; Singh, B.; Sachan, P.K. Biofuels and their sources of production: A review on cleaner sustainable alternative against conventional fuel, in the framework of the food and energy nexus. *Energy Nexus* **2021**, *4*, 100036. <https://doi.org/10.1016/j.nexus.2021.100036>.
24. Malode, S.J.; Prabhu, K.K.; Mascarenhas, R.J.; Shetti, N.P.; Aminabhavi, T.M. Recent advances and viability in biofuel production. *Energy Convers. Manag.* **2021**, *10*, 100070.
25. Mizik, T.; Gyarmati, G. Economic and Sustainability of Biodiesel Production—A Systematic Literature Review. *Clean Technol.* **2021**, *3*, 19–36. <https://doi.org/10.3390/cleantechnol3010002>.
26. Zabermaawi, N.; Alsulaimany, F.A.; El-Saadony, M.T.; El-Tarabily, K.A. New eco-friendly trends to produce biofuel and bioenergy from microorganisms: An updated review. *Saudi J. Biol. Sci.* **2022**, *in press*.
27. Oke, D.; Dunn, J.B.; Troy, R.; Hawkins, T.R. The contribution of biomass and waste resources to decarbonizing transportation and related energy and environmental effects. *Sustain. Energy Fuels* **2022**, *6*, 721–735.
28. Jiang, B.; Zhang, Y.; Guo, T.; Zhao, H.; Jin, Y. Structural Characterization of Lignin and Lignin-Carbohydrate Complex (LCC) from Ginkgo Shells (*Ginkgo biloba* L.) by Comprehensive NMR Spectroscopy. *Polymers* **2018**, *10*, 736. <https://doi.org/10.3390/polym10070736>.
29. Hernández-Beltrán, J.U.; Lira, I.O.H.-D.; Cruz-Santos, M.M.; Saucedo-Luevanos, A.; Hernández-Terán, F.; Balagurusamy, N. Insight into Pretreatment Methods of Lignocellulosic Biomass to Increase Biogas Yield: Current State, Challenges, and Opportunities. *Appl. Sci.* **2019**, *9*, 3721. <https://doi.org/10.3390/app9183721>.
30. Nwokolo, N.; Mukumba, P.; Obileke, K.; Eneme, M. Waste to energy: A focus on the impact of substrate type in biogas production. *Processes* **2020**, *8*, 1224.
31. International Energy Agency (IEA). *Outlook for Biogas and Biomethane: Prospects for Organic Growth*; IEA Paris, France, 2022. Available online: [https://iea.blob.core.windows.net/assets/03aeb10c-c38c-4d10-bcec-de92e9ab815f/Outlook\\_for\\_biogas\\_and\\_biomethane.pdf](https://iea.blob.core.windows.net/assets/03aeb10c-c38c-4d10-bcec-de92e9ab815f/Outlook_for_biogas_and_biomethane.pdf) (accessed on 15 April 2022).
32. Scarlat, N.; Dallemand, J.-F.; Fahl, F. Biogas: Developments and perspectives in Europe. *Renew. Energy* **2018**, *129*, 457–472. <https://doi.org/10.1016/j.renene.2018.03.006>.
33. Saadabadi, S.A.; Thattai, A.T.; Fan, L.; Lindeboom, R.E.; Spanjers, H.; Aravind, P. Solid Oxide Fuel Cells fuelled with biogas: Potential and constraints. *Renew. Energy* **2019**, *134*, 194–214. <https://doi.org/10.1016/j.renene.2018.11.028>.
34. Stern, J. The role of gases in the European energy transition. *Russ. J. Econ.* **2020**, *6*, 390–405. <https://doi.org/10.32609/j.ruje.6.55105>.
35. Jafri, Y.; Waldheim, L.; Lundgren, J. *Emerging Gasification Technologies for Waste & Biomass*; IEA Bioenergy, Dublin, Ireland: 2020. Available online: [https://www.ieabioenergy.com/wp-content/uploads/2021/02/Emerging-Gasification-Technologies\\_final.pdf](https://www.ieabioenergy.com/wp-content/uploads/2021/02/Emerging-Gasification-Technologies_final.pdf) (accessed on 23 March 2022).
36. European Biogas Association (EBA). *Market State and Trends in Renewable and Low-Carbon Gases in Europe*; EBA Brussels Belgium: 2020. Available online: <https://www.europeanbiogas.eu/market-state-and-trends-in-renewable-and-low-carbon-gases-in-europe/> (accessed on 23 March 2022).
37. Yang, L.; Ge, X.; Wan, C.; Yu, F.; Li, Y. Progress and perspectives in converting biogas to transportation fuels. *Renew. Sustain. Energy Rev.* **2014**, *40*, 1133–1152. <https://doi.org/10.1016/j.rser.2014.08.008>.
38. Damyanova, S.; Beschkov, V. Biogas as a source of energy and chemicals. In *Biorefinery Concepts*; Beschkov, V. Ed.; IntechOpen: London, UK, 2019.
39. Persson, M.; Jonsson, O.; Wellinger, A. Biogas Upgrading to Vehicle Fuel Standards and grid Injection; IEA Bioenergy: 2006; p. 32. Available online: [https://www.ieabioenergy.com/wp-content/uploads/2007/12/upgrading\\_report\\_final.pdf](https://www.ieabioenergy.com/wp-content/uploads/2007/12/upgrading_report_final.pdf) (accessed on 15 April 2022).
40. Moya, C.; Santiago, R.; Hospital-Benito, D.; Lemus, J.; Palomar, J. Design of biogas upgrading processes based on ionic liquids. *Chem. Eng. J.* **2021**, *428*, 132103. <https://doi.org/10.1016/j.cej.2021.132103>.
41. Thrän, D.; Persson, T.; Daniel-Gromke, J.; Ponitka, J.; Seiffert, M.; Baldwin, J.; Kranzl, L.; Schipfer, F.; Matzenberger, J.; Devriendt, N.; et al. *Biomethane Status and Factors Affecting Market Development and Trade*; IEA Bioenergy, Dublin Ireland: 2014. Available online: [https://publik.tuwien.ac.at/files/PubDat\\_234171.pdf](https://publik.tuwien.ac.at/files/PubDat_234171.pdf) (accessed on 23 March 2022).
42. Energy Institute Hrvoje Požar (EIHP). Available online: [http://www.eihp.hr/wp-content/uploads/2018/03/BiogasAction-Viro-Expo-prezentacija\\_EIHP.pdf](http://www.eihp.hr/wp-content/uploads/2018/03/BiogasAction-Viro-Expo-prezentacija_EIHP.pdf) (accessed on 23 March 2022).
43. Kovačić, Đ. Development of the Lignocellulose Pretreatment by Heat and Electric Field for Biogas Production by Anaerobic Co-digestion with Cow Manure. Ph.D. Thesis, University of Osijek, Osijek, Croatia, 2017.
44. Garcia, N.H.; Mattioli, A.; Gil, A.; Frison, N.; Battista, F.; Bolzonella, D. Evaluation of the methane potential of different agricultural and food processing substrates for improved biogas production in rural areas. *Renew. Sustain. Energy Rev.* **2019**, *112*, 1–10. <https://doi.org/10.1016/j.rser.2019.05.040>.
45. Pereira, A.J.M.; Elvira, P.S.I.; Oneto, S.J.; Cruz, D.L.R.; Portela, J.R.; Nebot, E. Enhancement of methane production in mesophilic anaerobic digestion of secondary sewage sludge by advanced thermal hydrolysis pretreatment. *Water Res.* **2015**, *71*, 330–340.
46. Farida, H.; Lee Chang, Y.; Hirotsugu, K.; Abdul, A.H.; Yoichi, A.; Takeshi, Y.; Hiroyuki, D. Treatment of sewage sludge using anaerobic digestion in Malaysia: Current state and challenges. *Front. Energy Res.* **2019**, *7*, 19.
47. Kaur, G.; Luo, L.; Chen, G.; Wong, J.W. Integrated food waste and sewage treatment—A better approach than conventional food waste-sludge co-digestion for higher energy recovery via anaerobic digestion. *Bioresour. Technol.* **2019**, *289*, 121698.
48. Song, Y.-J.; Oh, K.-S.; Lee, B.; Pak, D.-W.; Cha, J.-H.; Park, J.-G. Characteristics of Biogas Production from Organic Wastes Mixed at Optimal Ratios in an Anaerobic Co-Digestion Reactor. *Energies* **2021**, *14*, 6812. <https://doi.org/10.3390/en14206812>.



49. McCabe, B.K.; Schmidt, T. *Integrated Biogas Systems: Local Applications of Anaerobic Digestion towards Integrated Sustainable Solutions; Technical Report*; IEA Bioenergy. 2008. Available on: [https://www.ieabioenergy.com/wp-content/uploads/2018/06/Integrated-biogas-systems\\_WEB.pdf](https://www.ieabioenergy.com/wp-content/uploads/2018/06/Integrated-biogas-systems_WEB.pdf) (accessed on 15 April 2022).
50. Jain, S. *Global Potential of Biogas*; World Biogas Association, London, UK: 2019. Available online: [https://www.worldbiogasassociation.org/wp-content/uploads/2019/09/WBA-globalreport-56ppa4\\_digital-Sept-2019.pdf](https://www.worldbiogasassociation.org/wp-content/uploads/2019/09/WBA-globalreport-56ppa4_digital-Sept-2019.pdf) (accessed on 18 March 2022).
51. Al Seadi, T.; Rutz, D.; Prassl, H.; Kottner, M.; Finsterwalder, T.; Volk, S.; Janssen, R. *Handbook of Biogas*; University of Southern Denmark: Esbjerg, Denmark, 2008.
52. Weiland, P. Biogas production: Current state and perspectives. *Appl. Microbiol. Biotechnol.* **2010**, *85*, 849–860. <https://doi.org/10.1007/s00253-009-2246-7>.
53. Sahota, S.; Shah, G.; Ghosh, P.; Kapoor, R.; Sengupta, S.; Singh, P.; Vijay, V.; Sahay, A.; Vijay, V.K.; Thakur, I.S. Review of trends in biogas upgradation technologies and future perspectives. *Bioresour. Technol. Rep.* **2018**, *1*, 79–88. <https://doi.org/10.1016/j.biteb.2018.01.002>.
54. Sriramajayam, S.; Ramesh, D. Studies on biomethanation of kitchen wastes. *Trends in biosciences* **2006**, *8/16*, 4332–4335.
55. Banerjee, S.; Prasad, N.; Selvaraju, S. Reactor Design for Biogas Production-A Short Review. *J. Energy Power Technol.* **2021**, *4/1*, p.14. <https://doi.org/10.21926/jept.2201004>.
56. Ryue, J.; Lin, L.; Kakar, F.L.; El beshbishy, E.; Al-Mamun, A.; Dhar, B.R. A critical review of conventional and emerging methods for improving process stability in thermophilic anaerobic digestion. *Energy Sustain. Dev.* **2020**, *54*, 72–84. <https://doi.org/10.1016/j.esd.2019.11.001>.
57. Kumar RNeerudu, U.K.; Ragini Gothwal, R.; Mohapatra, S.; Deshpande, P.K.; Mukunda Vani, M.; Merugu, R. Bioprocess Parameters for Thermophilic and Mesophilic Biogas Production: Recent Trends and Challenges. In *Bioenergy Research: Basic and Advanced Concepts. Clean Energy Production Technologies*; Srivastava, M., Srivastava, N., Singh, R., Eds.; Springer: Singapore, 2021; pp. 225–256.
58. Anukam, A.; Berghel, J. Biomass Pretreatment and Characterization: A Review. In *Biotechnological Applications of Biomass*; Basso, T.P., Basso, T.O., Basso, L.C., Eds.; IntechOpen: London, UK, 2020.
59. Mao, C.; Feng, Y.; Wang, X.; Ren, G. Review on research achievements of biogas from anaerobic digestion. *Renew. Sustain. Energy Rev.* **2015**, *45*, 540–555. <https://doi.org/10.1016/j.rser.2015.02.032>.
60. Yu, L.; Ma, J.; Frear, C.; Zaher, U.; Chen, S. Two-Stage Anaerobic Digestion Systems Wherein One of the Stages Comprises a two-Phase System. United States Patent Application Publication, 2017. Available online: <https://patents.google.com/patent/US20130309740A1/en> (accessed on 23 March 2022).
61. Hamad, M.A.F.; Radwan, A.M.; Amin, A. Review of Biomass Thermal Gasification. In *Biomass Volume Estimation and Valorization for Energy*; InTech, London, UK: 2017.
62. Karuppiyah, T.; Azariah, V.E. Biomass Pretreatment for Enhancement of Biogas Production. In *Anaerobic Digestion*; Banu, J.R., Ed.; IntechOpen: London, UK, 2019.
63. Wagner, A.O.; Lackner, N.; Mutschlechner, M.; Prem, E.M.; Markt, R.; Illmer, P. Biological pretreatment strategies for second-generation lignocellulosic resources to enhance biogas production. *Energies* **2018**, *11*, 1797.
64. Zieliński, M.; Kisielińska, M.; Dudek, M.; Rusanowska, P.; Nowicka, A.; Krzemieniewski, M.; Kazimierowicz, J.; Dębski, M. Comparison of microwave thermohydrolysis and liquid hot water pretreatment of energy crop *Sida hermaphrodita* for enhanced methane production. *Biomass Bioenergy* **2019**, *128*, 105324.
65. Marta, K.; Paulina, R.; Magda, D.; Nowicka, A.; Krzywik, A.; Dębski, M.; Kazimierowicz, J.; Zieliński, M. Evaluation of ultrasound pretreatment for enhanced anaerobic digestion of *Sida hermaphrodita*. *Bioenerg. Res.* **2020**, *13*, 824–832.
66. Orlando, M.-Q.; Borja, V.-M. Pretreatment of animal manure biomass to improve biogas production: A review. *Energies* **2020**, *13*, 3573.
67. Achinas, S.; Achinas, V.; Euverink, G.J.W. A technological overview of biogas production from biowaste. *Eng. J.* **2017**, *3*, 299–230.
68. Baruah, J.; Nath Bikash Kar, N.B.; Ritika, S.; Sachin, K.; Chandra, D.R.; Chandra, B.D.; Eeshan, K. Recent trends in the pretreatment of lignocellulosic biomass for value-added products. *Front. Energy Res.* **2018**, *6*, 141.
69. Bharathiraja, B.; Jayamuthunagai, J.; Chakravarthy, M.; Kumar, R.P. Bioprocess of biofuels for Green and Clean Environment. In *Bioprocess Engineering for a Green Environment*, 1st ed.; Sivasubramanian, V., Ed.; CRC Press: Boca Raton, FL, USA, 2018; p. 13.
70. Capolupo, L.; Faraco, V. Green methods of lignocellulose pretreatment for biorefinery development. *Appl. Microbiol. Biotechnol.* **2016**, *100*, 9451–9467. <https://doi.org/10.1007/s00253-016-7884-y>.
71. Yu, H.; Zhang, F.; Li, L.; Wang, H.; Sun, Y.; Jiang, E.; Xu, X. Boosting levoglucosan and furfural production from corn stalks pyrolysis via electro-assisted seawater pretreatment. *Bioresour. Technol.* **2021**, *346*, 126478. <https://doi.org/10.1016/j.biortech.2021.126478>.
72. Panigrahi, S.; Sharma, H.B.; Tiwari, B.R.; Krishna, N.V.; Ghangrekar, M.; Dubey, B.K. Insight into understanding the performance of electrochemical pretreatment on improving anaerobic biodegradability of yard waste. *Renew. Energy* **2021**, *180*, 1166–1178. <https://doi.org/10.1016/j.renene.2021.08.123>.
73. Kasinath, A.; Fudala-Ksiazek, S.; Szopinska, M.; Bylinski, H.; Artichowicz, W.; Remiszewska-Skwarek, A.; Luczkiewicz, A. Biomass in biogas production: Pretreatment and codigestion. *Renew. Sustain. Energy Rev.* **2021**, *150*, 111509. <https://doi.org/10.1016/j.rser.2021.111509>.

74. Abraham, A.; Mathew, A.K.; Park, H.; Choi, O.; Sindhu, R.; Parameswaran, B.; Pandey, A.; Park, J.H.; Sang, B.I. Pre-treatment strategies for enhanced biogas production from lignocellulosic biomass. *Bioresour. Technol.* **2020**, *301*, 122725.
75. Rodriguez, C.; Alaswad, A.; Benyounis, K.Y.; Olabi, A.G. Pretreatment techniques used in biogas production from grass. *Renew. Sustain. Energy Rev.* **2017**, *68*, 1193–1204.
76. Wu, Y.N.; Mattsson, M.; Ding, M.W.; Wu, M.T.; Mei, J.; Shen, Y.L. Effects of Different Pretreatments on Improving Biogas Production of Macroalgae *Fucus vesiculosus* and *Fucus serratus* in Baltic Sea. *Energy Fuels* **2019**, *33*, 2278–2284. <https://doi.org/10.1021/acs.energyfuels.8b04224>.
77. Ryckebosch, E.; Drouillon, M.; Vervaeren, H. Techniques for transformation of biogas to biomethane. *Biomass Bioenergy* **2011**, *35*, 1633–1645. <https://doi.org/10.1016/j.biombioe.2011.02.033>.
78. Bauer, F.; Persson, T.; Hultberg, C.; Tamm, D. Biogas upgrading—Technology overview, comparison and perspectives for the future. *Biofuels Bioprod. Biorefining* **2013**, *7*, 499–511. <https://doi.org/10.1002/bbb.1423>.
79. Bora, D.; Barbora, L.; Borah, A.J.; Mahanta, P. A Comparative assessment of biogas upgradation techniques and its utilization as an alternative fuel in internal combustion engines. In *Alternative Fuels and Advanced Combustion Techniques as Sustainable Solutions for Internal Combustion Engines*; Singh, A.P., Kumar, D., Agarwal, A.K., Eds.; Springer: Singapore, 2021; pp. 95–11.
80. Sun, Q.; Li, H.; Yan, J.; Liu, L.; Yu, Z.; Yu, X. Selection of appropriate biogas upgrading technology—a review of biogas cleaning, upgrading and utilisation. *Renew. Sustain. Energy Rev.* **2015**, *51*, 521–532. <https://doi.org/10.1016/j.rser.2015.06.029>.
81. Rodero, M.R.; Angeles, R.; Marin, D.; Diaz, I.; Colzi, A.; Posadas, E.; Lebrero, R.; Munoz, R. Biogas Purification and Upgrading Technologies. In *Biogas: Fundamentals, Process, and Operation*; Tabatabaei, M., Ghanavati, H., Eds; Springer International Publishing: Cham, Switzerland, 2018.
82. Adnan, A.I.; Ong, M.Y.; Nomanbhay, S.; Chew, K.W.; Show, P.L. Technologies for Biogas Upgrading to Biomethane: A Review. *Bioengineering* **2019**, *6*, 92. <https://doi.org/10.3390/bioengineering6040092>.
83. Niesner, J.; Jeca, D.; Stehlik, P. Biogas upgrading technologies: State of art review in European region. *Chem. Eng. Trans.* **2013**, *35*, 517–522.
84. De Oliveira, L.H.; Meneguín, J.G.; Pereira, M.V.; Nascimento, J.F.D.; Arroyo, P. Adsorption of hydrogen sulfide, carbon dioxide, methane, and their mixtures on activated carbon. *Chem. Eng. Commun.* **2019**, *206*, 1533–1553. <https://doi.org/10.1080/00986445.2019.1601627>.
85. Lee, W.Y.; Park, S.Y.; Lee, K.B.; Nam, S.C. Simultaneous removal of CO<sub>2</sub> and H<sub>2</sub>S from biogas by blending amine absorbents: A performance comparison Study. *Energy Fuels* **2020**, *34*, 1992–2000.
86. Ramli, A.; Ahmed, S.; Yusup, S. Adsorption behavior of Si-MCM-41 for CO<sub>2</sub>: Effect of pressure and temperature on adsorption. *Chem. Eng. Trans.* **2014**, *39*, 271–276.
87. He, X.; Zhu, J.; Wang, H.; Zhou, M.; Zhang, S. Surface Functionalization of Activated Carbon with Phosphonium Ionic Liquid for CO<sub>2</sub> Adsorption. *Coatings* **2019**, *9*, 590. <https://doi.org/10.3390/coatings9090590>.
88. Ziobrowski, Z.; Rotkegel, A. Comparison of CO<sub>2</sub> Separation Efficiency from Flue Gases Based on Commonly Used Methods and Materials. *Materials* **2022**, *15*, 460. <https://doi.org/10.3390/ma15020460>.
89. Ou, H.; Fang, M.; Chou, M.; Chang, H.; Shiao, T. Long-term evaluation of activated carbon as an adsorbent for biogas desulfurization. *J. Air Waste Manag. Assoc.* **2020**, *70*, 641–648. <https://doi.org/10.1080/10962247.2020.1754305>.
90. Sawalha, H.; Maghalseh, M.; Qutaina, J.; Junaidi, K.; Rene, E.R. Removal of hydrogen sulfide from biogas using activated carbon synthesized from different locally available biomass wastes—A case study from Palestine. *Bioengineered* **2020**, *11*, 607–618. <https://doi.org/10.1080/21655979.2020.1768736>.
91. Narang, K.; Akhtar, F. Freeze Granulated Zeolites X and A for Biogas Upgrading. *Molecules* **2020**, *25*, 1378. <https://doi.org/10.3390/molecules25061378>.
92. Raganati, F.; Miscio, F.; Ammendola, P. Adsorption of carbon dioxide for post-combustion capture: A review. *Energy Fuels* **2021**, *35*, 12845–12868.
93. Bhatt, P.M.; Belmabkhout, Y.; Assen, A.H.; Weseliński, Ł.J.; Jiang, H.; Cadiau, A.; Xue, D.X.; Eddaoudi, M. Isorecticular rare Earth fcu-MOFs for the selective removal of H<sub>2</sub>S from CO<sub>2</sub> containing gases. *Chem. Eng. J.* **2017**, *324*, 392–396. <https://doi.org/10.1016/j.cej.2017.05.008>.
94. Yu, J.; Xie, L.H.; Li, J.R.; Ma, Y.; Seminario, J.M.; Balbuena, P.B. CO<sub>2</sub> Capture and separations using MOFs: Computational and experimental studies. *Chem. Rev.* **2017**, *117*, 674–9754.
95. Taddei, M.; Petit, C. Engineering metal–organic frameworks for adsorption-based gas separations: From process to atomic scale. *Mol. Syst. Des. Eng.* **2021**, *6*, 841–875. <https://doi.org/10.1039/d1me00085c>.
96. Alonso, A.; Vico, J.M.; Markeb, A.A.; Busquets-Fité, M.; Komilis, D.; Puentes, V.; Sánchez, A.; Font, X. Critical review of existing nanomaterial adsorbents to capture carbon dioxide and methane. *Sci. Total Environ.* **2017**, *595*, 51–62. <https://doi.org/10.1016/j.scitotenv.2017.03.229>.
97. Sanaeepur, H.; Amooghin, A.E.; Bandehali, S.; Moghadassi, A.; Matsuura, T.; Van der Bruggen, B. Polyimides in membrane gas separation: Monomer’s molecular design and structural engineering. *Prog. Polym. Sci.* **2019**, *91*, 80–125. <https://doi.org/10.1016/j.progpolymsci.2019.02.001>.
98. Norahim, N.; Yaisanga, P.; Faungnawakij, K.; Charinpanitkul, T.; Klayson, C. Recent Membrane Developments for CO<sub>2</sub> Separation and Capture. *Chem. Eng. Technol.* **2018**, *41*, 211–223. <https://doi.org/10.1002/ceat.201700406>.

99. Galizia, M.; Chi, W.S.; Smith, Z.P.; Merkel, T.C.; Baker, R.W.; Freeman, B.D. 50th Anniversary Perspective: Polymers and Mixed Matrix Membranes for Gas and Vapor Separation: A Review and Prospective Opportunities. *Macromolecules* **2017**, *50*, 7809–7843. <https://doi.org/10.1021/acs.macromol.7b01718>.
100. Baccanelli, M.; Langè, S.; Rocco, M.V.; Pellegrini, L.A.; Colombo, E. Low temperature techniques for natural gas purification and LNG production: An energy and exergy analysis. *Appl. Energy* **2016**, *180*, 546–559. <https://doi.org/10.1016/j.apenergy.2016.07.119>.
101. Font-Palma, C.; Cann, D.; Udemu, C. Review of cryogenic carbon capture innovations and their potential applications. *C* **2021**, *7*, 58.
102. Babar, M.; Bustam, M.A.; Ali, A.; Maulud, A.S.; Shafiq, U.; Mukhtar, A.; Shah, S.N.; Maqsood, K.; Mellon, N.; Shariff, A.M. Thermodynamic data for cryogenic carbon dioxide capture from natural gas: A review. *Cryogenics* **2019**, *102*, 85–104. <https://doi.org/10.1016/j.cryogenics.2019.07.004>.
103. Kapoor, R.; Ghosh, P.; Kumar, M.; Vijay, V.K. Evaluation of biogas upgrading technologies and future perspectives: A review. *Environ. Sci. Pollut. Res.* **2019**, *26*, 11631–11661. <https://doi.org/10.1007/s11356-019-04767-1>.
104. Edison, S.M. Bio-methane generation from organic waste: A review. In Proceedings of the World Congress on Engineering and Computer Science (WCECS), San Francisco, CA, USA, 22–24 October 2014.
105. Bond, T.; Templeton, M. History and future of domestic biogas plants in the developing world. *Energy Sustain. Dev.* **2011**, *15*, 347–354.
106. Aurelia Figueroa, A.; Boshell, F.; van Velzen, L.; Anisie, A. *Biogas for Domestic Cooking Technology Brief*; International Renewable Energy Agency: Abu Dhabi, United Arab Emirates, 2017. Available online: [https://www.irena.org/-/media/Files/IRENA/Agency/Publication/2017/Dec/IRENA\\_Biogas\\_for\\_domestic\\_cooking\\_2017.pdf](https://www.irena.org/-/media/Files/IRENA/Agency/Publication/2017/Dec/IRENA_Biogas_for_domestic_cooking_2017.pdf) (accessed on 25 May 2022).
107. Ali, M.Y.; Hassan, M.; Rahman, M.A.; Kafy, A.A.; Ara, I.; Javed, A.; Rahman, M.R. Life cycle energy and cost analysis of small scale biogas plant and solar PV system in rural areas of Bangladesh. *Energy Procedia* **2019**, *160*, 277–284.
108. Pilloni, M.; Hamed, T.A. Small-Size Biogas Technology Applications for Rural Areas in the Context of Developing Countries. In *Anaerobic Digestion in Built Environments*; Sikora, A., Ed.; IntechOpen: London, UK, 2021.
109. Sawyerr, N.; Trois, C.; Workneh, T.S.; Oyeboode, O.; Babatunde, O.M. Design of a household biogas digester using co-digested cassava, vegetable and fruit waste. *Energy Rep.* **2020**, *6*, 1476–1482. <https://doi.org/10.1016/j.egy.2020.10.067>.
110. Grim, J.; Nilsson, D.; Hansson, P.-A.; Nordberg, Å. Demand-Orientated Power Production from Biogas: Modeling and Simulations under Swedish Conditions. *Energy Fuels* **2015**, *29*, 4066–4075. <https://doi.org/10.1021/ef502778u>.
111. Fan, W.; Huang, L.; Tan, Z.; Xue, F.; De, G.; Song, X.; Cong, B. Multi-objective Optimal Model of Rural Multi-energy Complementary System with Biogas Cogeneration and Electric Vehicle Considering Carbon Emission and Satisfaction. *Sustain. Cities Soc.* **2021**, *74*, 103225. <https://doi.org/10.1016/j.scs.2021.103225>.
112. Ishikawa, S.; Connell, N.O.; Lechner, R.; Hara, R.; Kita, H.; Brautsch, M. Load response of biogas CHP systems in a power grid. *Renew. Energy* **2021**, *170*, 12–26. <https://doi.org/10.1016/j.renene.2021.01.120>.
113. Yin, Y.; Chen, S.; Li, X.; Jiang, B.; Zhao, J.R.; Nong, G. Comparative analysis of different CHP systems using biogas for the cassava starch plants. *Energy* **2021**, *232*, 121028. <https://doi.org/10.1016/j.energy.2021.121028>.
114. Shah, M.S.; Halder, P.K.; Shamsuzzaman AS, M.; Hossain, M.S.; Pal, S.K.; Sarker, E. Perspectives of biogas conversion into bio-CNG for automobile fuel in Bangladesh. *J. Renew. Energy* **2017**, *2017*, 4385295.
115. Karagöz, Y. Analysis of the impact of gasoline, biogas and biogas + hydrogen fuels on emissions and vehicle performance in the WLTC and NEDC. *Int. J. Hydrog. Energy* **2019**, *44*, 31621–31632. <https://doi.org/10.1016/j.ijhydene.2019.10.019>.
116. de Castro, R.E.N.; Alves, R.M.B.; Nascimento, C.A.O. Assessing the sugarcane bagasse and straw as a biofuel to propel light vehicles. *Sustain. Energy Fuels* **2021**, *5*, 2563–2577. <https://doi.org/10.1039/d1se00129a>.
117. Rosén, T.; Ödlund, L. System Perspective on Biogas Use for Transport and Electricity Production. *Energies* **2019**, *12*, 4159. <https://doi.org/10.3390/en12214159>.
118. Shafie, S.M.; Othman, Z.; Hani, N.; Omar, S.; Nu'man, A.H.; Yusuf, N.N.A.N.; Shah, A. Biogas fed-fuel cell based electricity generation: A life cycle assessment approach. *Int. J. Energy Econ. Policy* **2020**, *10*, 498–502.
119. Roubík, H.; Mazancová, J. Suitability of small-scale biogas systems based on livestock manure for the rural areas of Sumatra. *Environ. Dev.* **2020**, *33*, 100505.
120. Putti, V.R.; Tsan, M.; Mehta, S.; Kammila, S. *The State of the Global Clean and Improved Cooking Sector*; Technical Paper No. 007/15, Energy Sector Management Assistance Program ESMAP; World Bank: Washington, DC, USA, 2015.
121. Livestock and Poultry Environmental Learning Community (LPELC). Available online: <https://lpec.org/> (accessed on 25 March 2022).
122. Frazier, S.; Hamilton, D.; Ndegwa, P.M. *Anaerobic Digestion: Biogas Utilization and Cleanup*. Oklahoma State University, 2017. Available online: <https://extension.okstate.edu/fact-sheets/anaerobic-digestion-biogas-utilization-and-cleanup.html> (accessed on 25 March 2022).
123. Patterson, T.; Dinsdale, R.; Esteves, S. Review of Energy Balances and Emissions Associated with Biomass-Based Transport Fuels Relevant to the United Kingdom Context. *Energy Fuels* **2008**, *22*, 3506–3512. <https://doi.org/10.1021/ef800237q>.
124. Chandra, R.; Vijay, V.; Subbarao, P.; Khura, T. Performance evaluation of a constant speed IC engine on CNG, methane enriched biogas and biogas. *Appl. Energy* **2011**, *88*, 3969–3977. <https://doi.org/10.1016/j.apenergy.2011.04.032>.
125. Clarke, S. *Vehicle Conversion to Natural Gas or Biogas*; Ministry of Agriculture, Food and Rural Affairs: Ontario, CA, USA, 2012.

126. Boyano, A.; Morosuk, T.; Blanco-Marigorta, A.M.; Tsatsaronis, G. Conventional and advanced exergo environmental analysis of a steam methane reforming reactor for hydrogen production. *J. Clean. Prod.* **2012**, *20*, 152–160.
127. Serrano-Lotina, A.; Daza, L. Influence of the operating parameters over dry reforming of methane to syngas. *Int. J. Hydrog. Energy* **2014**, *39*, 4089–4094. <https://doi.org/10.1016/j.ijhydene.2013.05.135>.
128. Li, Y.; Yang, L. *Converting Biogas to Transportation Fuels*; Ohio State University Extension, Columbus, Ohio, USA, 2014. Available online: <https://ohioline.osu.edu/factsheet/AEX-653.2> (accessed on 25 March 2022).
129. Alves, H.J.; Junior, C.B.; Niklevicz, R.R.; Frigo, E.P.; Frigo, M.S.; Coimbra-Araújo, C.H. Overview of hydrogen production technologies from biogas and the applications in fuel cells. *J. Hydrog. Energy* **2013**, *38*, 5215–5225.
130. Safari, S.; Ghasedi, A.H.; Ozgoli, H.A. Integration of solar dryer with a hybrid system of gasifier-solid oxide fuel cell/micro gas turbine: Energy, economy, and environmental analysis. *Environ. Prog. Sustain. Energy* **2021**, *40*, 13569. <https://doi.org/10.1002/ep.13569>.
131. Safari, S.; Hajilounezhad, T.; Ehyaei, M.A. Multi-objective optimization of solid oxide fuel cell/gas turbine combined heat and power system: A comparison between particle swarm and genetic algorithms. *Int. J. Energy Res.* **2020**, *44*, 9001–9020. <https://doi.org/10.1002/er.5610>.
132. Banja, M.; Jégard, M.; Motola, V.; Sikkema, R. Support for biogas in the EU electricity sector—A comparative analysis. *Biomass Bioenergy* **2019**, *128*, 105313. <https://doi.org/10.1016/j.biombioe.2019.105313>.
133. Ivanković, M. Biogas in circular economy of the European Union. In Proceedings of the 35th International Scientific & Expert Meeting of Gas Professionals, Opatija, Croatia, 21–23 October 2020.
134. European Biogas Association (EBA). *EBA Statistical report 2020*; EBA: Brussels, Belgium, 2021. Available online: <https://www.europeanbiogas.eu/eba-statistical-report-2020/> (accessed on 25 March 2022).
135. EurObserv'ER. The State of Renewable Energies in Europe, 2019. Available online: <https://www.eurobserv-er.org/19th-annual-overview-barometer/> (15 April 2022).
136. European Biogas Association (EBA) & Gas Infrastructure Europe (GIE). European Biomethane Map 2020. Available online: [https://www.europeanbiogas.eu/wp-content/uploads/2020/06/GIE\\_EBA\\_BIO\\_2020\\_A0\\_FULL\\_FINAL.pdf](https://www.europeanbiogas.eu/wp-content/uploads/2020/06/GIE_EBA_BIO_2020_A0_FULL_FINAL.pdf) (accessed on 25 March 2022).
137. Directive 2008/98/EC on Waste and Repealing Certain Directives. Available online: <https://eur-lex.europa.eu/legal-content/EN/TXT/?uri=CELEX%3A02008L0098-20180705> (accessed on 25 March 2022).
138. European Commission (EC). *A Sustainable Bioeconomy for Europe: Strengthening the Connection between Economy, Society and the Environment: Updated Bioeconomy Strategy*; EC: Brussels, Belgium, 2018. Available online: <https://data.europa.eu/doi/10.2777/478385> (accessed on 25 March 2022).
139. Bedeković, G.; Grčić, I.; Anić Vučinić, A.; Premur, V. Recovery of waste expanded polystyrene in lightweight concrete production. *Min.-Geol.-Pet. Eng. Bull.* **2019**, *34*, 73–80.
140. Friant, M.C.; Vermeulen, W.J.; Salomone, R. Analysing European Union circular economy policies: Words versus actions. *Sustain. Prod. Consum.* **2021**, *27*, 337–353. <https://doi.org/10.1016/j.spc.2020.11.001>.
141. Mazzi, A.; Ren, J. Circular Economy in Low-Carbon Transition. *Energies* **2021**, *14*, 8061. <https://doi.org/10.3390/en14238061>.
142. CEDIGAZ. Global Biomethane Market 2021 Assessment. Available online: <https://www.cedigaz.org/global-biomethane-market-2021/> (accessed on 25 March 2022).