



Article

## Managing Traffic Flows for Cleaner Cities: The Role of Green Navigation Systems

Fiamma Perez-Prada <sup>1,\*</sup>, Andres Monzon <sup>2</sup> and Cristina Valdes <sup>3</sup>

- Transport Research Centre (TRANSyT), Escuela de Ingenieros de Caminos, Canales y Puertos, Universidad Politécnica de Madrid, Ciudad Universitaria, 28040 Madrid, Spain
- Transport-Civil Eng. Department, Escuela de Ingenieros de Caminos, Canales y Puertos, Universidad Politecnica de Madrid, 28040 Madrid, Spain; andres.monzon@upm.es
- Empresa Municipal de Transportes de Madrid, 28007 Madrid, Spain; cristina.valdes@emtmadrid.es
- \* Correspondence: fiamma.perez@upm.es; Tel.: +34-91-336-6708

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**Abstract:** Cities worldwide suffer from serious air pollution problems and are main contributors to climate change. Green Navigation systems have a great potential to reduce fuel consumption and exhaust emissions from traffic. This research evaluates the impacts of different percentages of green drivers on traffic,  $CO_2$ , and  $NO_x$  over the entire Madrid Region. A macroscopic traffic model was combined with an enhanced macroscopic emissions model and a GIS (Geographic Information Systems) to simulate emissions on the basis of average vehicle speeds and traffic intensity at the link level.  $NO_x$  emissions are evaluated, taking into account not only the exhaust emissions produced by transport activity, but also the amount of the population exposed to these air pollutants. Results show up to 10.4%  $CO_2$  and 13.8%  $NO_x$  reductions in congested traffic conditions for a 90% penetration of green drivers; however, the population's exposure to  $NO_x$  increases up to 20.2%. Moreover, while traffic volumes decrease by 13.5% for the entire region, they increase by up to 16.4% downtown. Travel times also increase by 28.7%. Since green drivers tend to choose shorter routes through downtown areas, eco-routing systems are an effective tool for fighting climate change, but are ineffective to reduce air pollution in dense urban areas.

**Keywords:** eco-routing; green navigation; traffic emissions; climate change; air pollution; co-benefits; trade-offs; ICT;  $CO_2$ ;  $NO_x$ 

#### 1. Introduction

ICT (Information Communication Technology)-based strategies applied to traffic have brought about new possibilities to improve traffic performance in urban environments, all while reducing their externalities. The emissions reduction potential of ICT-based solutions has traditionally been viewed as secondary to impacts on traffic safety and congestion. Nevertheless, recently, different researchers have been studying the effects of these new technologies on fuel consumption and  $CO_2$  emissions [1–4].

Out of the different ICT-based strategies, Green Navigation (GN) systems have shown important fuel consumption and  $CO_2$  reductions when compared to typical travel-time and monetary-cost minimization routing strategies [5–7]. These reported reductions were largely attributed to vehicle travel distance reductions, which, when eco-routing is applied to an entire metropolitan area, lead to an increase in emissions on urban streets, due to transferring traffic to shorter routes [7]. GN systems produce global benefits in terms of  $CO_2$ . However, the impact of GN systems applied at the city-wide level on the reallocation of emissions and its effect on air pollution is not yet well understood.

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This research aims to bridge this gap and evaluate the impacts of GN systems on air pollution, taking into account the varying influence of transport emissions on the urban environment [8]. Our hypothesis is that GN systems provide clear benefits to mitigate greenhouse gases (GHG) emissions, but they could be ineffective at reducing local air pollution in urban environments. To this end, the research evaluates the environmental and traffic performance impacts of GN Systems, with a special focus on the analysis of climate change and air pollution co-benefits and conflicts. Thus, we calculate both  $CO_2$  and  $NO_x$  emissions from the Madrid Metropolitan area assuming different percentages of GN drivers in the Madrid Region. Unlike other papers, we also take into account the local impact of air pollutants; hence, we model the population exposure to  $NO_x$  emissions. This research provides local policy makers and city managers with insights to design and properly implement transport reduction strategies, maximizing climate change and air pollution reduction synergies.

## 2. Climate Change and Air Pollution: Co-Benefits and Conflicts of Transport Emissions Reduction Strategies in Urban Areas

Cities worldwide suffer from serious air pollution problems and are main contributors to climate change, consuming about two-thirds of the world's overall energy and emitting an estimated 70% of the world's anthropogenic GHG emissions [9]. Globally, transportation is responsible for 23% of anthropogenic  $CO_2$  emissions and 54% of  $NO_x$  emissions, of which 75% and 72%, respectively, are attributable to road transport [10]. Road traffic is a major source of GHG emissions, and the largest contributor to air pollutant emissions in urban environments. Moreover, the predicted growth in the urban population and transport activity will further contribute to increasing transport emissions in the near future [11].

The role of sustainable urban mobility strategies in climate change mitigation has drawn increased interest in recent years: a number of authors are examining different transport strategies to reduce GHG emissions [12–16].

Traffic emissions have different spatial and duration impacts [17]: While CO<sub>2</sub> emissions are a global problem with large-scale impacts—no matter where the emissions are produced, all of them contribute to climate change—short lived GHGs such as ozone have a regional effect. On the other hand, traffic air pollution has important regional and local impacts. Pollutant concentrations decrease in concentration as they are transported by the wind and vehicle movement turbulence from the source [18]. Increased health risks for populations located near major roadways and transportation infrastructures, particularly in densely populated urban areas, have been continuously reported [19]. However, transport planning has usually tackled them in a similar way, considering that most of the strategies designed to reduce GHGs emissions also reduce air pollutants and vice-versa and do not account for the spatial variation of transport emissions,. This can be observed through the gross indicators habitually used to measure the environmental impact of different transport projects: tons of reduced  $CO_2$ , kg of  $NO_x$ , or kg of PM. To obtain these indicators, we only need to take traffic activity attributes into account. However, when analyzing the impacts of the projects on air pollution, it is important to take their greater impact on highly-populated areas into account. Therefore, it is necessary to develop indicators that also take the spatial impacts of transport emissions, particularly at an urban level, into account [20].

Transport policies designed to avert climate change show co-benefits for air pollution mitigation [21–28]; however, they also present conflicts. A well-known case is the shift to diesel cars, which not only delivered less  $CO_2$  reductions than expected [29,30] but also increased air pollution. For example, a study in Ireland showed how a car taxation policy promoting the purchase of lower  $CO_2$  emitting cars could lead to a higher share of fuel-efficient diesel vehicles, with up to 7% in  $CO_2$  reductions, but a 28% increase in  $NO_x$  emissions [31]. The use of biofuels also has potential trade-offs regarding air pollution, particularly regarding  $NO_x$  emissions. The literature illustrates two different explanations; on the one hand, some studies have reported higher levels of  $NO_x$  due to higher

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temperatures of combustion [32,33] while others reported lower emissions due to the lower amount of premixed fuel and peak burning temperature compared to diesel and gasoline fuels [34,35]. A review on the impacts of biofuel exhaust emissions can be found in [36]. Concerning electric vehicles, studies showed that they have a directly positive impact on pollution in urban environments, since exhaust emissions are not produced during the operation of the vehicle [37,38]. Studies also reported the need for further regulation to address the non-exhaust emissions [37] of electric vehicles in the future, and the importance of taking into account the electricity mix of the countries [39].

Measures directly affecting traffic performance were also analyzed, from the combined perspective of climate change and air pollution. For example, traffic-calming measures reported increases in CO<sub>2</sub> and  $NO_x$  [40,41]. Traffic calming can provide emissions reductions only if they generate a reduction in traffic activity, or a shift toward more efficient transport modes. In a study conducted in Madrid City, Garcia-Castro et al. [42] showed that while under low and medium congested traffic conditions eco-driving leads to reductions of up to 2.3% in CO<sub>2</sub> and of up to 4.3% in NO<sub>x</sub> emissions, under highly-congested traffic conditions, emissions increase by up to 1.3% and 0.5%, respectively. In Belgium, a study showed reductions of about 10% in NO<sub>x</sub> and CO<sub>2</sub>, thanks to implementing a green-wave signal coordination scheme along an urban arterial road [43]. Cohen et al. [44] reported global environmental benefits from a speed-reduction strategy in segments along different Lille freeways, and Perez-Prada and Monzon [8] reported a 14.4% and 16.4% reduction in  $CO_2$  and  $NO_x$  emissions, respectively, thanks to implementing a reduced speed limit on an urban ring-road. Regarding road capacity increases, Monzon et al. [45] predicted a reduction in GHGs and air pollutants in Madrid after the renovation of its inner ring-road. An ex-post study [46] confirmed this trend after the renovated beltway had been operating for three years (short-term). However, other studies [47] suggested an increase in emissions in the long run after a road capacity increase.

Green Navigation (GN) systems, or eco-routing systems, provided routing recommendations based on calculation of the environmental impact and the real-time traffic situation, i.e., for a given origin and destination, they provide the route that minimizes fuel consumption. Previous research [1,7] reported up to 8.2% in CO<sub>2</sub> savings under highly congested conditions, and up to 9.5% in CO<sub>2</sub> savings in free-flow or low-congested conditions. These studies also reported increasing CO<sub>2</sub> savings as the level of "green drivers" increases, and the vehicle-km travelled decreases, due to the selection of shorter routes. In a later work, Valdes et al. [7] indicated that those benefits were mostly concentrated on freeways and highways, while urban streets and extra-urban roads experience an increase in emissions, due to the transfer of traffic to shorter routes. This research suggested that, since length has an important effect on fuel consumption, green drivers choose routes similar to the minimum length, although this may imply crossing the downtown area or selecting a road with lower speed than a highway. Those eco-routing studies show the effectiveness of that measure to reduce fuel consumption, and therefore CO<sub>2</sub> emissions.

Based on the results of previous works, the present research argues that GN systems present clear benefits for combating climate change, but that they could be ineffective for reducing air pollution in urban environments. To this end, we will analyze the impacts of GN systems on  $CO_2$  and  $NO_x$  emissions in the city of Madrid, taking the different spatial impacts of GHG and air pollutant emissions into account in this research.

### 3. Method

This research has analyzed the environmental and traffic performance impacts of different percentages of GN drivers in the region of Madrid (Spain). The emissions assessment methodology is based on the interaction of different types of models: a macroscopic transport model, a macroscopic emissions model, and a GIS. These models will be used to generate three types of indicators for the evaluation of impacts: traffic performance (TPIs), traffic emissions (TEIs), and population exposure (PEIs) indicators (see Section 3.2).

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The interaction of the models takes place in two phases, summarized in Figure 1. The first stage follows the ICT-Emissions project methodology [48]. It assesses CO<sub>2</sub> emissions of ICT-based strategies at the macro-scale level, linking the widespread macroscopic emissions model, COPERT, to a macroscopic traffic model. The traffic demand model fulfills three functions at this stage. First, it is used to generate the different scenarios under analysis. Base-case (BC) scenarios represent the current traffic situation for each of the traffic periods analyzed, while Green Navigation (GN) and Shortest-Route (SR) scenarios are obtained by varying the traffic model impedance function. SR scenarios will only be used to understand to what extent the GN impedance function, which is designed to minimize fuel consumption, takes changes in speed and traffic intensity into account, and not only changes in itinerary lengths. Second, it provides inputs for the emission model (links lengths, average speeds, and traffic intensity). Finally, it generates the TPIs used in the evaluation. On the other hand, the emissions model generates TEIs based on EFs, fleet composition, and the inputs of the traffic model. TEIs only take traffic activity and fleet composition into account.

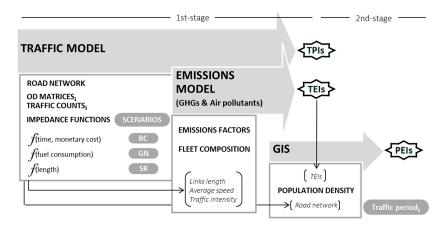


Figure 1. Methodology flow chart.

The second stage of the model evaluates traffic emissions, taking their effect on the population into account. We use a GIS to calculate PEIs, weighting TEIs by population concentrations near each roadway. The higher the PEI values, the worse the impact of the air pollutant emissions. The GIS needs inputs from the traffic model (road network) and the emissions model (TEIs), and information regarding population densities.

## 3.1. Traffic Scenarios

For the assessment, three different scenarios have been considered, varying the traffic model impedance function. The model impedance function, also known as the model objective function, is the key factor in driver's route choices. It assigns the effort connected to each route or path. Usually, this effort is measured in terms of travel time. Additional properties, such as travel expenses, can be converted into travel time units, and considered within the travel impedance function [49]. The traffic assignment selects the routes that minimize the impedance function.

The scenarios considered are:

- The BC represented the existing traffic situation. It will assign the typical impedance function for conventional drivers, which is based on time and cost.
- The GN scenario assigned an impedance function defined in terms of fuel consumption, which is directly related to CO<sub>2</sub> emissions, i.e., the assignment will seek results that minimize fuel consumption, and therefore CO<sub>2</sub> emissions, as well. The ICT-Emissions project tested five different fuel consumption functions from different studies, and the one that best performed under congested traffic conditions was selected, i.e., the default fuel consumption function of COPERT (g/km), expressed as a function of time (g/s) [1,7]:

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$$I_{GN} = FC(v_{Cur})x t_{Cur} \tag{1}$$

where:  $I_{GN}$  is the GN impedance function expressed in seconds, FC is the COPERT default fuel consumption function expressed in g/s as:

$$FC = av_{Cur}^5 + bv_{Cur}^4 + cv_{Cur}^3 + dv_{Cur}^2 + ev_{Cur} + f$$
 (2)

 $a=-3.29612\times 10^{-1}$ ,  $b=1.097\times 10^{-7}$ ,  $c=-1.1893\times 10^{-5}$ ,  $d=5.30345\times 10^{-4}$ ,  $e=-1.56253\times 10^{-3}$ , f=0.256344,  $v_{Cur}$  is the link average speed calculated in the traffic model in km/h, and  $t_{Cur}$  is the link travel time calculated by the traffic model in seconds.

• A SR scenario has also been considered. This last scenario assigned an impedance function defined exclusively in terms of trip length.

It is not realistic to assume that all the vehicles in the network would be equipped with eco-routing navigation systems; therefore, different levels of penetration will be considered in this research. Five GN scenarios representing different percentages of eco-routing drivers will be compared to the BC. The penetration rates selected for study are 10% (GN10 scenario), 25% (GN25), 50% (GN50), 75% (GN75), and 90% (GN90). GN scenarios will consist of a combination of GN drivers and conventional drivers; the origin-destination (OD) matrix has been divided into two, one for GN drivers and the second for conventional drivers. For example, in the GN25 scenario, 75% of the OD matrix is assigned to conventional drivers, and the other 25% to green drivers according to Equation (1). The same reasoning applied to the SR scenarios (five penetration rates and assignment of OD matrices between the shortest route drivers and the conventional ones).

Navigation systems provide eco-routing recommendations based on a calculation of the environmental impact and the real-time traffic situation. The GN-scenario assignment process was designed to continuously collect new traffic conditions and then, to allow for different route alternatives, taking this new traffic situation into account. First, heavy vehicles and conventional car drivers were assigned to the network, and subsequently, the impedance function of green drivers is calculated for the new traffic levels and average speeds. Next, green drivers were assigned to the network in 10 steps, and after each of the stages, the impedance function was calculated again to show the new traffic conditions. This implies that, in every step of the GN demand segment assignment, the impedance function is updated with the new traffic situation, and henceforth, routing recommendations will be modified consequently. In each stage, 10% of the OD matrix corresponding to green drivers is assigned.

## 3.2. Assessment Indicators

The model outputs are three different types of indicators designed to understand the impact of eco-routing systems on traffic, climate change, and air pollution: traffic performance indicators (TPIs), traffic emissions indicators (TEIs), and population exposure indicators (PEIs).

Traffic performance was evaluated in terms of volume (VKM) and travel times (VEH). In view of the fact that GN systems do not affect the geometry of the network (link length and capacity), changes in traffic volume also represent changes in saturation rates (traffic intensity divided by section capacity). The TPIs used for the assessment are:

• VKM (Vehicle KiloMeters travelled) represents traffic volume as a product of traffic intensity (number of vehicles crossing a specific section in an hour) and section length. See Equation (1), where *I* is the traffic intensity (vehicles/hour) in a specific link of the network for an average hour in the time period considered; *l* is the length (km) of the link; *i* represents the links of the network; and *p* represents an hour under the traffic condition considered (congested and free-flow). *N* represents the different aggregation levels used to show the results, e.g., by road type (highways, urban roads, extra-urban). Traffic volume is often described as an indicator of traffic demand. It characterizes the traffic flow over a road link on an average hour in a day. Saturation

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(traffic intensity divided by section capacity) and traffic volume will be used as synonyms for traffic intensity when used in relative terms, since the measurement under study does not modify the length nor the capacity of the links.

$$VKM_{i,p} = \sum_{i}^{N} I_{i,p} l_i \tag{3}$$

• VEH (VEhicles per Hour) represents the total travel time of all vehicles on a link or segment of the road for an average hour of the three defined periods. It is calculated as a product of the traffic intensity and travel time. See Equation (2), where *t* is the travel time in hours on a specific link of the network for an average hour in the time period considered, and the rest of the terms are the same as those used in VKM.

$$VEH_{i,p} = \sum_{i}^{N} I_{i,p} t_{i,p}$$

$$\tag{4}$$

Traffic emissions were evaluated in terms of  $CO_2$  and  $NO_x$ .  $CO_2$  is the largest GHG anthropogenic emission, and  $NO_x$  has commonly been used as a proxy for traffic-related air pollution [45,46].

TEIs were calculated by the link as a product of traffic volume and the EF corresponding to the average speed of the link for the Madrid fleet composition. See Equation (3), where m represents vehicle classes (fleet composition), depending on vehicle fuel technology, engine type and technology, and k is the type of emission considered ( $CO_2$  or  $NO_x$ ). The rest of the terms are the same as those used before.

$$TEI_{i,p,k} = \sum_{i}^{N} EF(v)_{k,m,i,p} I_{i,p} l_{i,p}$$
 (5)

In order to show the population exposure to air pollution, we calculated PEIs, weighting transport exhaust emissions with population density, following an approach similar to the one used by Perez-Prada and Monzon [20]. See Equation (4), where *D* is the population density of the area that the link crosses over. The rest of the terms are the same as those used before.

$$PEI_{i,p,k} = \sum TEI_{i,p,k} D_{i,p}$$
 (6)

Very complex dispersion models and detailed chemistry models are needed to develop accurate population exposure indicators that take all the variables affecting concentration levels into account, e.g., precipitation, wind speed and direction, atmospheric pressure, temperature, topography, and urban street disposition, as well as the type and height of buildings [50–53].

As explained by [20], the use of population density as a proxy for exposure levels is motivated by the fact that (i) exposure to local emissions from transport is largely a function of both the amount of traffic activity, and the number of people who regularly live or work along roadsides [50,53]; that (ii) numerous studies analyze air pollutant exposure levels at a certain distance from a high-traffic road [54,55]; and that (iii) health effects and residential proximity to high traffic roads show a correlation [51].

## 3.3. Attributes of Model Components in the Madrid Case Study

The Madrid Region is the capital of Spain and the country's most populated region. The population of the Madrid Region amounts to about 6.5 million inhabitants, half of which live in the main city. The region is spread over an 8030 km² surface, and the population density is high in the central city, 5232 inhab./km², and rather low in the whole region, 605 hab./km²) [56]. Figure 2 shows the distribution of the population within the Madrid Region.

Madrid's road network has a radio-centric structure, with eight main free-radial highways linking central Madrid to all of Spain's coastal regions (A1, A2, A3, A4, A5, A6) and four ring-roads (two of

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them unfinished) encircling it: the first one is the M-30, which divides the city into two parts, the inner part—Central Area—and the peripheral districts, and it acts as a local distributor road; the second one is the M-40, which encompasses most of the city of Madrid, and is the most congested road in the Madrid metropolitan area; and the M-45 and M-50, both yet incomplete rings. Four tollways run in parallel to four of the main highways.

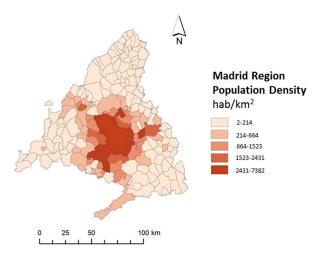


Figure 2. Population density by municipality in the Madrid Region.

## 3.3.1. Traffic Model Attributes

The traffic model was based on the one developed by Valdes [57] and enhanced for the ICT-Emissions project. Information regarding the model characterization and calibration can be found in [58]. The regional network consists of 15,712 road links, with a higher level of detail in Madrid City, and in particular, for the city center, i.e., inside the M30.

The road network was classified into three types (see Figure 3), depending on the type of road they represent.

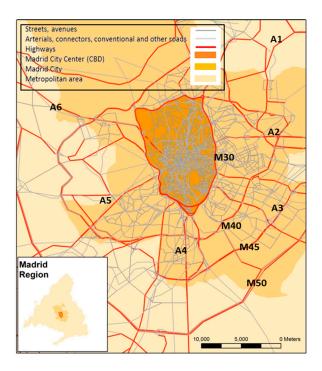


Figure 3. Madrid road network.

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CBD NETWORK represents 1670 km of streets and avenues inside the M30 and accounts for 28% of the whole network.

- HIGHWAYS represent 1647 km of urban ring-roads, highways, and motorways and also account for 28% of the whole road network.
- OTHER ROADS aggregate 2560 km of conventional roads, junctions, and service lanes belonging to the peripheral, metropolitan, and regional areas, and account for 44% of the entire network.

#### 3.3.2. Emissions Model Attributes

COPERT provides a straightforward expression for hot EF, based on two or three speed ranges by vehicle category. Each vehicle category is also classified according to the fuel (gasoline, diesel) and engine type, and the associated emissions reduction technology (from Conventional to Euro IV) [58].

As an innovative approach to customized urban emission assessment, the emissions model used the circulating fleet, instead of the existing fleet, to estimate emissions more accurately. The circulating fleet considers a different fleet composition by the three road-types, providing more accurate results, while the existing fleet considers that all the vehicles are equally distributed through all type of roads. However there is no data on the fleet used at different traffic periods: peak vs. off-peak. Table 1 shows the final distribution of the fleet throughout the network, adapted from [59].

Vehicle Category and Fuel	CBD Network	Ring-Roads	Other Roads and Rest of the Highways		
Passenger Cars	83.4%	88.0%	86.2%		
Gasoline	25.2%	25.8%	25.1%		
Diesel	55.8%	61.4%	60.1%		
Others	2.4%	0.8%	1.0%		
LDV	4.7%	8.5%	8.6%		
Gasoline	0.1%	0.3%	0.2%		
Diesel	4.6%	8.2%	8.4%		
HDV	1.0%	2.0%	1.8%		
Bus	2.0%	1.0%	1.6%		
Diesel	1.1%	1.0%	1.3%		
Others	0.8%	0.0%	0.3%		
Mopeds and Motorcycles	8.9%	0.5%	1.8%		

**Table 1.** Madrid's circulating fleet by category and fuel and zone distribution.

Notes: LDV stands for Light Duty Vehicles, HDV stands for Heavy Duty Vehicles, CBD stands for Central Business District.

Passenger cars represented more than the 80% of all the vehicles in the considered zones. The share of light-duty vehicles (LDV) on urban roads was almost half that on highways and other roads. Other roads included the extra-urban network of the model, and all the highways except ring-roads. The share of two-wheel vehicles was only relevant at an urban level, where it represented almost 9%, while on ring-roads its share is almost negligible (0.5%). Finally, heavy-duty vehicles (HDV) in urban areas only accounted for 1%, while this figure was double on ring-roads (2.0%) and almost double on other roads (1.8%). The share of buses at an urban level (2.0%) doubled the bus portion on ring-roads (1.0%), and was also higher than the bus segment on other roads (1.6%).

Regarding fuel types, the highest share of gasoline cars appeared at the urban level (34.2%), which was more than 25% higher than the gasoline vehicle share on orbital highways (26.6%) and extra-urban roads (27.1%). Therefore, the share of diesel vehicles was higher on orbital motorways and other roads. It accounted for 72.5% and 71.5%, versus a 62.4% share in urban areas. The share of other types of fuels was trivial for other roads and highways, and very small for the urban network (3.3%).

Newest passenger vehicle technologies (EURO 4 and 5) had a higher share on urban roads, while the incidence of pre-EURO 2 vehicles (both diesel and gasoline) was higher on other roads. The share of gasoline pre-EURO 2 vehicles was more than four times the share of diesel pre-EURO 2 vehicles, higher on urban and orbital-highways: 17.5% and 26.5%, versus 4% and 6.6%, respectively. On other roads, its share was also high, at 23.5%, versus 7.1%.

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#### 3.3.3. GIS Attributes

Finally, using a GIS tool, population maps provided by the National Statistical Institute (http://www.ine.es/) were combined with the model network to assign a new link characteristic: population density. We have used information from 4341 census sections in the Madrid Region.

These sections are an official geographical reference of statistical character. Since they have basically an operational character and therefore always have to be defined by more or less fixed sizes: to be accessible to an interviewing agent for the purpose of population counts or statistical surveys, it is recommended that the size of a section does not exceed 2500 legal residents [60].

For links crossing more than one census section, the density was calculated as follows:

$$Link\ Density = \frac{\sum_{i}^{N} A_{i} P_{i}}{\sum A_{i}}$$
 (7)

where A is the area of the census section, P is the population in the census section, and N is the total number of census sections crossed by the link.

## 3.4. Model Assumptions and Limitations

Some authors questioned the appropriateness of average speed models to assess emissions at a city level [61,62] since similar average speeds can be reached with different speed patterns in each link. Notwithstanding, this model was further developed for calculating emissions at a link level in urban environments [63], including (i) algorithms suitable for working with hourly time periods and for links that are hundreds of meters long; (ii) a subroutine to distribute the data among the hundreds of types of vehicles, subdivided by fuel type, age, etc. Limitations may appear for high saturation levels (e.g., >80%) and/or for very short road segments (e.g., <400 m). Moreover, as is customary, start emissions and evaporation emissions are not considered.

Due to the use of population density, PEIs are mainly residentially focused and only partially reflect exposure to air pollution, as they exclude the effects on working areas and on tourist and commercial areas that frequently exhibit higher numbers of pedestrians. Besides, PEIs do not completely capture the fact that pollution decreases quickly when moving away from the roadside. The model assumes that the population is equally distributed throughout each census section.

## 4. Results and Discussion

In general, eco-routing studies show the effectiveness of the system to reduce fuel consumption, and therefore reduce  $CO_2$  emissions, e.g., a study in Lund showed that for 46% of trips, drivers do not choose the most fuel-efficient route, and when they do, they save up to 8.2% on fuel consumption [5]. Ahn and Rakha [64] reported a fuel consumption reduction of 14–18% and  $CO_2$  emissions reduction of 17–25% through field trials and micro emissions simulation models for a shorter but slowly arterial route compared to a longer by faster highway corridor. Minett et al. [65] reported a 17% reduction in fuel consumption between the cities of Delft and Zoetermeer using the shortest local road instead of a faster motorway or a provincial roads. Regarding  $NO_x$  emissions, Ahn and Rakha [64] showed up to 45% reductions in the arterial route. The focus of these studies was at the corridor level, lacking interaction with other corridors in the context of a complete road network.

In a latter work, Ahn and Rakha [6] quantified the impacts of implementing dynamic eco-routing in two American cities (Cleveland and Columbus) considering various levels of market penetration and congestion. For a full level of market penetration results showed a reduction of travel distance by 5.11% and 5.46%, of  $NO_x$  by 8% and 9.5%, and  $CO_2$  by 3.3% and 5.2% in Cleveland and Columbus, respectively. Results also showed an increase in travel times by 4.54% in Cleveland and 3.21% in Columbus. Their results led to the conclusion that eco-routing impacts are sensitive to the road network configuration. The eco-routing market penetration analysis showed further reduction as the percentage of eco-routing drivers increased; in particular, for Cleveland 0.23% and 3.85% of fuel

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savings were reported for 10% and 90% eco-routing market penetration while these figures were 4.97% and 6.12% in Columbus. The study also reported increased fuel consumption and emission savings with increased congestion levels.

The ICT-Emissions project estimated fuel and  $CO_2$  impacts of different percentages of green drivers at different traffic periods [66]. The reported fuel and  $CO_2$  reductions were consistent with the results of previous studies. In congested conditions, the results showed reduction by 2.22%, 4.89%, 7.07%, 7.92%, and 9.16% for 10%, 25%, 50%, 75%, and 90% of eco-routing drivers. For off-peak hours these figures ranged from 1.13% to 4.66% and for free-flow conditions from 5.91% to 9.47%. These reductions were attributed to a reduction of vehicle-km travelled (from 3.65% to 13.51% for congested conditions, from 1.61% to 8.53 for off-peak hours, and 5.85% to 9.48% for free-flow conditions) and the relocation of traffic to shorter routes. They also reported an important increase of travel times, that were higher as the level of green drivers increased. Our results are consistent with the ICT-Emissions results; minor differences in reductions are due to the use of the circulating fleet, instead of the existing fleet for the Madrid Region.

## 4.1. Whole Region

Eco-routing systems generated important reductions at a regional level in terms of CO<sub>2</sub> emissions, up to 10.4% in reductions under highly-congested traffic conditions, and up to 7.1% under free-flow traffic conditions (see Table 2). When considering gross  $NO_x$  emissions (TEI\_ $NO_x$  indicator), similar benefits were also shown. Reductions in traffic volumes were in line with the reduction in traffic emissions. However, two indicators considerably increased with the GN penetration rate. On the one hand, travel times increased for all GN penetration rates, except for the GN10 under free-flow conditions. This increase was greater for lower penetration rates (from 6.8% to 14.1% and to 21.3% for GN10, GN25 and GN50, respectively) and it was moderated at the highest penetration levels (GN75 and GN90). Analyzing the PEI for  $NO_x$ , we find the first evidence that GN systems could be ineffective regarding air pollution in dense urban environments. At high levels of congestion, even the lower penetration rates produced an increase of 6.4% in the PEI\_ $NO_x$  indicator, reaching 20.2% in the GN90 scenario. For free-flow conditions, this increase was more moderate, ranging from 1.2% to 16.3% for GN10 and GN90, respectively.

**Table 2.** Variation of traffic and environmental indicators for the whole regional network for each GN penetration rate.

Penetration Rates	VKM (veh × km)	VEH (veh × h)	TEI_CO <sub>2</sub> (kg)	TEI_NO <sub>x</sub> (kg)	$PEI\_NO_x  (kg \times hab./km^2)$
BC_cong	5,884,915	96,541	904,802	3009	10,277,277
GN10	-3.7%	6.8%	-3.3%	-3.3%	6.4%
GN25	-7.7%	14.1%	-6.3%	-7.5%	9.4%
GN50	-11.3%	21.3%	-8.9%	-11.2%	14.0%
GN75	-12.9%	25.7%	-9.9%	-12.9%	17.9%
GN90	-13.5%	28.7%	-10.4%	-13.8%	20.2%
BC_free	2,170,829	25,623	333,790	1164	3,779,809
GN10	-5.8%	-3.6%	-1.7%	-1.6%	1.2%
GN25	-6.8%	0.2%	-3.0%	-3.8%	4.5%
GN50	-7.8%	7.2%	-4.8%	-7.2%	9.6%
GN75	-8.9%	13.2%	-6.4%	-10.2%	14.0%
GN90	-9.5%	16.8%	-7.1%	-1.8%	16.3%

Notes: BC\_cong stands for Base case scenario under congested conditions, BC\_free stands for Base case scenario under free-flow conditions, VKM stands for vehicles-km travelled, i.e., traffic volume, VEH stands for vehicles-hour travelled, i.e., travel times,  $TEI\_CO_2$  stands for the  $CO_2$  traffic emission indicator,  $TEI\_NO_x$  stand for the  $NO_x$  traffic emission indicator,  $PEI\_NO_x$  stands for the population-weighted  $NO_x$  indicator.

Table 3 confirms that the fuel-consumption impedance function used for the simulation of the eco-routing measure took average speeds and traffic intensities into account, and not only the trip distance.

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riigiliy-corigested	scenarios.					
		Traffic Volum	ie (veh × km)	Average Speed (veh $ imes$ km/veh $ imes$ hour)		
		SR	GN	SR	GN	
	ВС	5,884,915	5,884,915	61.0	61.0	

-3.7%

-7.7%

-11.3%

-12.9%

-13.5%

7.40%

6.80%

-14.50%

-43.40%

-56.20%

-9.8%

\_19 1%

-26.9% -30.7%

-32.8%

**Table 3.** Changes in traffic volumes and averages speeds between the SR and the GN highly-congested scenarios.

Notes: BC is the Base case scenario, SR is the Shortest Route scenario, GN is the Green Navigation Scenario.

-6.9%

\_17.7%

-26.0%

-22.8%

-15.6%

## 4.2. Results by Road Type

Penetration rate (%)

10

25

50

75

90

Figure 4 shows the share by road type of each of the indicators assessed for the base scenarios for congested and free-flow traffic conditions. While the CBD network only carried 14.8% of vehicle-km driven in Madrid and generated 15.7% of the  $CO_2$  emissions, it represented 63.8% of the air pollution exposure levels in the highly-congested scenario. This is due to the fact that population density in the city center is higher than in the peripheral areas. For the free-flow scenario, similar levels of traffic flow (14.5%) led to a lower share of the  $PEI_NO_x$  indicator, 54.6%. This is due to the fact that the gross traffic  $NO_x$  emission share was higher under congested conditions than under free-flow conditions, 13.3% vs. 10.7%, respectively, in CBD.

#### Road network share by indicator Base scenarios ☐ CBD NET. ☐ HIGHWAY. ☐ OTHER R. 5.2% 19.9% 19.6% 20.5% 24.9% 24.5% 24.6% 44.0% 63.8% 54.6% 28.0% 15.7% 13.3% 13.3% 10.7% Length VKM TEI CO2 TEI NOX PEL NOX VKM TEI CO2 TEI NOX PEL NOX CONGESTED FREE-FLOW

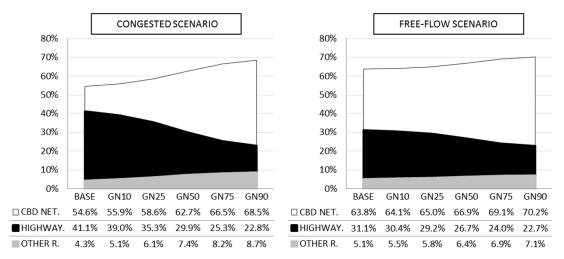
**Figure 4.** Indicators' share of the total road network by road type in the base congested and free-flow scenarios. Notes: CBD NET. stands for Central Business District Network, HIGHWAY. stands for Highways, and OTHER R. stands for Other Roads. Length is the percentage of km of road for each type of road considered, VKM stands for vehicles-km travelled i.e., traffic volume, TEI\_CO<sub>2</sub> stands for CO<sub>2</sub> traffic emission indicator, TEI\_NO<sub>x</sub> stands for NO<sub>x</sub> traffic emission indicator, PEI\_NO<sub>x</sub> stands for population-weighted NO<sub>x</sub> indicator.

Figure 5 shows the evolution of the PEI\_NO $_x$  share by road type for the different GN penetration levels considered (rows in the tables below, the sum of the columns add up to 100%). Although for the free-flow base scenario the CBD share of PEI\_NO $_x$  was lower than for the congested scenario, it increased with the penetration rate. Hence, for the highest GN penetration rates (GN75 and GN90), the share of air pollution exposure was similar. This noticeable evolution under free-flow condition was due to the fact that, while NO $_x$  gross traffic emissions increased in similar proportion to the congested condition, traffic flows in CBD grew faster as the GN penetration rate increased. Traffic flows rose

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from 14.5% for the free base-case scenario to 22.5%, while  $NO_x$  gross traffic emissions increased from 11% to 18%. For the congested situation, this growth ranged from 15% to 20% for traffic volumes, and from 13% to 20% for  $NO_x$  emissions.

# Evolution of the network share of the population-weighted NO<sub>x</sub> exposure indicator increasing GN penetration rate



**Figure 5.** Evolution of the population exposure to the  $NO_x$  indicator for different GN penetration rates and road types.

Table 4 shows the variation of traffic performance and environmental indicators by road type (i.e., urban, highway, and extra-urban) for each of the GN penetration rates considered. The impedance function looked for the traffic assignment that minimizes fuel consumption for the entire network. Traffic volumes increased considerably, especially under low-congestion conditions. Travel distances, speed, and congestion had a great influence on traffic emissions [62]. Although the impedance function used also takes the influence of traffic intensity and speed in fuel consumption into account, travel distance was still a major driver of fuel consumption reduction. At the CBD level, all of the indicators worsened with the adoption of eco-routing choices, regardless of the traffic condition considered. Under free-flow conditions, shorter urban routes became even more attractive for green drivers, since road saturation levels were low. Traffic volumes also increased on OTHER roads. However, these growths were compensated by reductions on the highways to obtain a global decrease (Table 2) in the total VKM driven.

Travel times increased noticeably on CBD and OTHER roads. While on CBD roads, there was an increase under highly-congested traffic conditions, and on OTHER peripheral roads, there was a greater increase under free-flow traffic conditions. Nevertheless, travel times on highways significantly decreased as GN penetration rates increases.

In general, GN systems reduced congestion on faster roads (highways) by distributing traffic to shorter, but slower, routes, where saturation rates increase significantly. Therefore, global travel times (Table 2) for the whole region increased. Then GN systems were ineffective at improving traffic performance since congestion was transferred to other areas in the city.

As expected,  $CO_2$  emissions also increased at a CBD level and OTHER roads, although this increase was compensated by the  $CO_2$  emission reduction on HIGHWAYS. The greatest impact of  $CO_2$  emissions lies in their contribution to climate change, i.e., in general, no matter where they are emitted, any emission contributes to global figures, and hence to global warming [18]; therefore, we can consider GN systems as an effective measure to reduce  $CO_2$  emissions, especially when applied during highly-congested traffic periods.

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Table 4. Variation of traffic and environmental indicators by road type for each GN penetration rate.

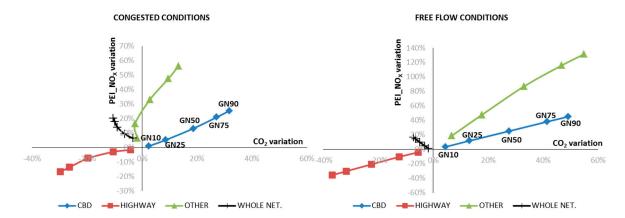
	VKM (veh × km)		VEH (ve	EH (veh × h) TEI_CO <sub>2</sub> (kg)		O <sub>2</sub> (kg)	TEI_NO <sub>x</sub> (kg)		PEI_NO <sub>x</sub> (kg × hab/km <sup>2</sup> )	
CBD NET.	Cong.	Free	Cong.	Free	Cong.	Free	Cong.	Free	Cong.	Free
ВС	870,955	314,411	29,662	7016	141,633	44,528	405	125	6,553,843	2,064,349
GN10	0.1%	0.3%	10.6%	0.6%	2.5%	4.4%	2.0%	4.0%	1.0%	3.3%
GN25	2.8%	7.8%	25.9%	10.6%	8.7%	13.2%	7.8%	13.1%	5.3%	11.7%
GN50	8.2%	20.2%	47.1%	28.1%	18.7%	27.7%	17.2%	28.2%	13.2%	25.2%
GN75	13.5%	31.6%	62.2%	44.9%	27.2%	41.5%	25.3%	42.4%	21.0%	38.1%
GN90	16.4%	37.8%	71.0%	54.5%	31.8%	49.2%	29.6%	50.4%	25.4%	45.2%
HIGHWAY.	Cong.	Free	Cong.	Free	Cong.	Free	Cong.	Free	Cong.	Free
ВС	3,545,963	1,412,048	45,751	13,622	541,775	222,941	1891	819	3,192,085	1,552,678
GN10	-5.0%	-9.3%	2.9%	-9.1%	-4.1%	-5.5%	-4.4%	-5.6%	-1.6%	-4.2%
GN25	-11.8%	-15.7%	3.4%	-14.9%	-10.5%	-12.3%	-11.5%	-12.7%	-3.0%	-10.4%
GN50	-20.9%	-25.0%	-3.5%	-22.3%	-19.6%	-22.5%	-20.8%	-23.3%	-7.4%	-20.8%
GN75	-27.4%	-33.4%	-11.1%	-29.6%	-26.4%	-31.6%	-27.4%	-32.8%	-13.6%	-30.3%
GN90	-30.4%	-38.1%	-14.5%	-34.0%	-29.7%	-36.7%	-30.9%	-38.2%	-16.7%	-35.7%
OTHER R.	Cong.	Free	Cong.	Free	Cong.	Free	Cong.	Free	Cong.	Free
ВС	1,467,996	444,370	21,127	4985	221,394	66,321	751	230	531,349	162,782
GN10	-2.6%	0.7%	9.8%	5.5%	-1.7%	6.8%	-3.1%	5.4%	6.4%	18.9%
GN25	-4.1%	11.4%	20.6%	26.9%	-2.5%	17.7%	-5.4%	14.6%	16.4%	47.6%
GN50	0.3%	26.7%	38.6%	58.3%	2.9%	33.1%	-1.8%	27.2%	33.2%	87.1%
GN75	6.2%	40.1%	53.9%	85.3%	9.6%	46.7%	3.7%	38.1%	47.6%	115.8%
GN90	9.7%	48.1%	62.9%	102.4%	13.3%	54.9%	6.3%	44.7%	56.4%	131.5%

Notes: CBD NET. stands for Central Business District Network, HIGHWAY. stands for Highways, and OTHER R. stands for Other Roads. BC for Base case scenario, VKM stands for vehicles-km travelled, i.e., traffic volume, VEH stands for vehicles-hour travelled, i.e., travel times, TEI\_CO<sub>2</sub> stands for the CO<sub>2</sub> traffic emission indicator, TEI\_NO<sub>x</sub> stands for the NO<sub>x</sub> traffic emission indicator.

CBD areas are the most populated areas in the region of Madrid. Although for the entire region traffic emissions (both  $CO_2$  and  $NO_x$ ) decreased, for CBD areas they increased. Although gross  $NO_x$  emissions are reduced at a regional level, population exposure levels increased (Table 2). Furthermore, population exposure to  $NO_x$  increased more quickly on CBD roads than it did for the entire region. For low GN penetration rates (GN10 to GN50), they increased in similar proportion, yet for GN10, the increase for the entire region is 6.4%, while for CBD roads, it was only 2%. However, for higher GN penetration rates (GN75 and GN90), CBD increased by almost double over the increases in the entire region. Population exposure to air pollution decreased in highways under free-flow traffic and congested conditions for all penetration rates considered.

To better understand the impact of this measure on the spatial redistribution of the emissions, Figure 6 plots TEI\_CO<sub>2</sub> reductions against PEI\_NO<sub>x</sub> reductions by road type. The penetration rate of green drivers is represented by dots, triangles, squares, or addition signs depending on the type of road being represented. The types of roads are also assigned different colors. Analyzing the results for the entire region (black line), we can consider GN systems as successful for reducing CO<sub>2</sub> emissions, but ineffective at decreasing local NO<sub>x</sub> exposure. This is because while the relative increases in CO<sub>2</sub> emissions in the CBD were offset by the greater relative decreases in the highways, the relocation of NO<sub>x</sub> emissions towards the densely populated CBD were not compensated by a reduction in the highways.

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**Figure 6.** Impacts of the different levels of GN drivers on the spatial allocation of  $CO_2$  emissions and population exposure to  $NO_x$  by the road type and traffic situation considered. Percentages on each axis indicate the percent variation in each measure with positive percentages indicating an increase and negative percentages indicating a decrease.

## 5. Conclusions

A proper environmental assessment of transport policies should take climate change and air pollution co-benefits and conflicts into account. Although this research does not take all the drivers to evaluate population exposure to air pollution into account, weighting emissions by population density reveals important insights. PEI makes emissions produced in dense areas more significant than those produced in less dense areas.

According to the results of this study, GN systems are an effective tool for fighting climate change, but are ineffective at reducing local air pollution, since it increases in highly-populated areas. These results are consistent under both the congested and free-flow conditions. On the one hand,  $CO_2$  emissions reductions increase with the level of penetration of GN drivers. Furthermore, the greatest  $CO_2$  reductions occur in highly-congested scenarios. On the other hand, considering  $NO_x$  emissions at the regional level, the results showed a reduction for the Madrid region, but with a spatial reallocation at the CBD areas. Besides, the results showed an upsurge in the population's exposure to  $NO_x$  air pollution for the whole region. This increase is greater in congested situations than in free-flow situations, and it rises along with the percentage of green drivers.

Traffic performance worsens considerably as eco-routing drivers increase due to the reallocation of traffic to shorter routes. On the one hand, considering the entire network, travel times increase for all GN penetration levels in highly-congested scenarios, and for penetration levels higher than 25% for free-flow situations. On the other hand, although global traffic volumes diminish, congestion increases on CBD and OTHER peripheral roads. Finally, other important urban sustainability factors, such as noise and accidents, create a vulnerable situation for travelers such as pedestrians and cyclists, who are expected to be negatively affected by the increase in traffic volume on CBD roads.

On-board navigation systems need be enhanced to take both traffic performance and traffic emissions into account. There are some advances in this regard, e.g., Jimenez et al. [67] have developed an objective function that optimizes the energy used on the trip, while observing the time schedule. However, the different spatial impact of transport emissions should also be taken into account, in order to efficiently achieve not only greener, but also cleaner cities. Therefore, there is a need to develop navigation algorithms that balance traffic performance, air pollution and climate change impacts to obtain the optimal route, taking the impacts on the entire network into account, instead of focusing on individual fuel consumption reductions.

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