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### 2.3 Sample extraction and instrument analysis

Other instrumental parameters of UPLC/MS/MS were showed in Table S1 and Table S2.

Table S1

Time (min)	Composition of the mobile phase (%)	
Negative mode	Eluent A2 (0.01%Ammonium hydroxide)	Eluent B1 (Acetonitrile)
0	90	10
0.25	90	10
3.00	10	90
4.00	10	90
4.01	90	10
5.00	90	10

Table S2

Compound	Retention time/min	Parent ion (m/z)	Daughter ions(m/z)	Dewll time(s)	Conevol tage (v)	Collision energy (v)
Bisohenol F	2.74	199.10	93.10	0.042	30	20

			105.10*			20
			198.00*			22
Bisphenol E	2.89	213.10	/	0.042	40	/
			133.00			25
Bisphenol A	3.03	227.10	212.00*	0.042	31	17
			92.05			30
Bisphenol S	0.66	249.13	108.07*	0.161	42	24
			93.02			32
Bisphenol Z	3.45	267.22	173.17*	0.042	56	30
			197.10			25
Bisphenol AF	3.25	335.23	265.16*	0.042	32	22
			419.92*			40

\* represents quantification ion.

## 2.4 Quality assurance and quality control

The strict QA/QC protocol was used to detect the extraction efficiency of BPs in the water sample. The performance of the analytical method was evaluated in terms of linearity, limit of detection (LOD) and limit of quantitative (LOQ) and recovery rate. When analyzing each group of samples, running solvent, standar, and process blanks in turn to check background BPs, peak identification, and quantification. Respectively, LOD and LOQ were determined to be the minimum detectable amounts of the analyte with signal-to-noise (S/N) of 3 and 10. Through injecting different concentrations of the standard solutions (seven-point calibration curve) into the 1.0-200 ng/mL range ( $R^2 > 0.998$ ), the linearity of the target compounds was studied. In order to evaluate the recovery, the spiked samples were prepared using ordinary surface water samples. Prior to use, surface water samples used for recovery test were analyzed to detect the presence of BPs. The recovery data for BPs had been corrected to take into account the fact that the average blank peak area of BPs was subtracted from the average peak area of the other recovery points. Six separate chromatographic runs were performed on each of the two concentration levels. The analysis of the reagent blanks (n=3) showed that the analysis system and glassware did not contain BPs.

Table S3

Compound	Recovery [%] (n = 6)				Quantitation limits (n = 3)	
	Water sample		SPM samples		Water	SPM
	10 ng/L	100 ng/L	10 ng/g	100 ng/g	(ng/L)	(ng/g)
BPF	83.0±12.4	97.8±10.2	82.4±13.1	90.2±10.0	11.10	1.50
BPE	88.4±7.2	92.3±10.7	75.7±6.3	82.0±10.4	1.83	0.50
BPA	91.0±8.1	108±9.9	102±10.8	110±11.5	2.60	0.25
BPS	80.3±4.0	91.4±16.3	79.3±9.3	84.7±6.5	0.53	0.30
BPZ	70.3±5.6	85.6±10.0	73.5±7.4	80.2±7.1	7.07	1.00
BPAF	93.1±16.3	96.2±8.0	82.2±11.4	92.3±10.1	1.70	0.30

## 2.5 Parameter measurement and statistical analysis

Using the risk quotient (RQ) to evaluate the risk assessment of the target compounds in the urban water. The  $RQ_s$  was calculated as Eq (1).

$$RQ = \frac{MEC}{PNEC} = \frac{MEC}{EC_{50} \text{ or } LC_{50}/f} \quad (1)$$

MEC and PNEC are measured environmental concentrations and predicted no-effect concentrations. According to the REACH guidance document, in order to estimate PNEC based on toxicity data when only short-term/acute toxicity data  $EC_{50}$  or  $LC_{50}$  is available, PNEC is calculated by the  $EC_{50}$  or  $LC_{50}$  that divides the safety factor ( $f$ ) 1000. Once a long-term/chronic NOEC value of one, two or three nutritional levels is available, using the  $f$  of the 100, 50, or 10 (ECHA, 2008). PNEC is derived from chronic and acute toxicity data in the literature and is 100 or 1000  $f$  in our study. Table 3 provides PNEC calculations for algae, daphnia and fish.

Calculation of the oestrogen equivalent concentration (EEQ) of a chemically determined mixture is based on all measured xenoestrogens with a known oestrogen equivalency factor (EEF; Table 3), as shown in the following equation (Eq. (2)). When  $EEQ_{Total} > 1.0 \text{ ng E}_2/\text{L}$ , the contaminants are thought to affect the endocrine systems of organisms in the water bodies. So EEQ is also used to assess the risk of BPs to human health.

$$EEQ_{Total} = \sum EEQ_i = \sum (C_i \times EEF_i) \quad (2)$$

The  $C_i$  refers to the compound  $i$  with a concentration of  $C$  in the traditional dissolved phase.

$EEQ_{Total}$  is the total estradiol equivalent and  $EEF_i$  is the estradiol equivalent.

**Table S4**

Sampling locations	Time	Concentrations (ng/L) in traditionally soluble phase				Reference
		BPA	BPS	BPF	BPAF	
Study area	2018 <sup>a</sup>	290 (244)	43.6(42.1)	2.53 (2.23)	6.68 (4.51)	This study
	(07)	133-576	5.87-83.5	ND-5.44	1.62-17.8	
Wujin district	2018	217 (157)	60.5(32.1)	7.13 (4.61)	2.70 (2.05)	[1]
	(08)	73.5-678	7.80-319	1.14-40.1	0.30-17.7	
Taihu Lake	2013	8.5 (7.9)	6.0 (2.0)	0.83 (0.5)	0.28 (0.2)	[2]
	(09)	4.2-14	0.28-67	ND-5.6	0.13-1.1	
Taihu Lake	2015	9.7 (7.3)	2.6 (0.94)	1.24 (1.1)	0.27 (0.1)	[3]
	(05)	3.9-33.2	0.32-27.3	0.5-3.28	0.06-2	
Taihu Lake	2015	92.6 (53.2)				[4]
	(11)	28-565				
Taihu Lake	2016	97	120	140	8.2	[5]
	(04)	28-560	4.5-1600	ND-1600	0.7-23	
Taihu Lake	2016	25.7 (23.8)	15.9 (6.6)	78 (30)	114 (111)	[6]
	(11)	19.4-68.5	41.-157	25.6-723	110-140	
Luoma Lake	2016	86	21	6.8	17	[5]
	(04)	49-110	ND-94	3.5-14	12-84	
Liaohe River	2013	47 (29)	14 (8.9)		1.9 (1.0)	[2]
	(09)	5.9-141	0.22-52	ND <sup>b</sup>	0.5-9.6	
Hunhe River	2013	40 (42)	11 (8.4)		2.4 (0.94)	[2]
	(09)	4.4-107	0.61-46	ND	0.61-11	
Pearl River	2013	73 (73)	135 (135)	773 (757)		[7]
	(07)	ND-98	ND-135	448-1110	ND	
West River	-2014	43 (43)		64 (64)		[7]
	(03)	ND-43	ND	ND-105	ND	
20 source water China	2017	12.8 (10.5)	1.1 (0.4)	2.2 (ND)	3.0 (0.1)	[8]
	(11)	ND-34.9	ND-5.2	ND-12.6	ND-10.8	
Several Rivers, Bay (Japan)		104	5.3	638		[7]
	2013	ND-431	ND-15	ND-2850	ND	
Several Rivers (Korea)	(07)	105.7	41	633		[7]
	-2014	1.0-272	ND-42	ND-1300	ND	
Several Rivers, Lake (India)	(03)	551	2174	91.5		[7]
		ND-1950	ND-7200	ND-289	ND	

<sup>a</sup> Year (Month).<sup>b</sup> ND: not detected.

**Table S5**

Compound	Non-target organism	Test Endpoint	Toxicity data (mg/L)	PNEC (ng/L)	Reference	EEF Ref. [17]
BPA	Algae	72h-EC50	2.2 (Growth)	2200	[9]	$1.07 \times 10^{-4}$
	Daphnia spp.	48h-EC50	3.9 (Immobility)	3900	[10]	
	Fish	48h-EC50	3.6 (Pigmentation)	3600	[11]	
BPS	Algae	96h-EC50	6.9	6900	[13] <sup>a</sup>	$1.06 \times 10^{-6}$
	Daphnia spp.	48h-EC50	55 (Immobility)	55000	[14]	
	Fish	72 hpF-EC50	155 (Mortality)	155000	[15]	
BPF	Algae	72h-IC50	22.1 (Growth)	22100	[11]	$1.08 \times 10^{-4}$
	Daphnia spp.	21d-NOEC	0.84 (Reproduction)	8400	[11]	
	Fish	48h-EC50	1.1 (Pigmentation)	1100	[11]	
BPAF	Algae	72h-IC50	3.0 (Growth)	3000	[11]	$7.23 \times 10^{-4}$
	Daphnia spp.	21d-NOEC	0.23 (Reproduction)	2300	[11]	
	Fish	72hpF-EC50	0.92 (Mortality)	920	[15]	
BPE	Daphnia spp.	48h-EC50	18	18000	[14]	$5.92 \times 10^{-5}$
	Fish	EC50	0.0579	57.9	[16]	

<sup>a</sup> The toxicity data was calculated from the ecological structure activity relationships (ECOSAR) model.

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