Article

Major Ion Toxicity to Glochidia of Common and Imperiled Freshwater Mussel Species

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Abstract: Freshwater mussel taxa are severely imperiled and among the most sensitive to several contaminants, including chloride. Relatively little is known of the toxicity of major ions to glochidia (larvae), particularly for rare species, or the effects of hardness on major ion toxicity to glochidia. Therefore, the primary objectives of this work were to: (1) determine the acute toxicity of major ion salts to glochidia, (2) compare chloride sensitivity of glochidia from common and rare species, and (3) evaluate the relationship between water hardness and chloride toxicity to glochidia. We assessed 24 h EC50s for fatmucket (*Lampsilis siliquoides*) glochidia exposed to NaCl, KCl, MgCl₂, CaCl₂, Na₂SO₄, MgSO₄, CaSO₄, and NaHCO₃ in moderately hard water. We determined NaCl EC50s for four species and KCl EC50s for glochidia of three species. Toxicity of chloride was generally consistent among the various chloride salts except for KCl, which was more toxic than all others by 1–2 orders of magnitude. Our results suggest that commonly tested species may be generally representative of rare species. Water hardness ameliorated the toxicity of chloride for all species to varying degrees. Results of this work indicate that some existing water quality criteria that do not include mussel toxicity data may not be protective of freshwater mussels.

Keywords: chloride; hardness; Unionidae; sensitivity

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1. Introduction

In North America and elsewhere, water pollution from major ions (Na⁺, K⁺, Ca²⁺, Mg²⁺, Cl[−], SO₄[−], HCO₃[−]) stems largely from anthropogenic activities such as mineral mining, oil and gas extraction, irrigation, industrial effluents, and road deicing [1,2]. Major ion pollution can cause osmotic stress and increased metabolic costs in freshwater organisms [3,4], posing a threat to fish and aquatic invertebrates such as freshwater mussels [5–10]. Freshwater mussels (family Unionidae) are well documented among the most imperiled of aquatic species [11] and toxicity studies with early life stages of freshwater mussels exposed to major ions have demonstrated that glochidia (larvae) and juveniles are among the most sensitive of all species tested [5,10,12–16].

Glochidia exposed to high concentrations of major ions close their valves or suffer ionoregulatory disturbance rendering them unable to attach to the host fish required to complete their parasitic lifecycle [7]. Field studies corroborate lab toxicity testing; in the Clinch and Powell River systems in Tennessee, USA, declining mussel richness and density has been connected with rising concentrations of major ions from coal mining operations [15]. Similarly, in Ontario, Canada, stream salinization from road salt application during the winter resulted in elevated glochidia mortality [5,9].

The high sensitivity of mussels to some contaminants (e.g., ammonia, chloride, copper, cadmium) has resulted in revisions of some state and federal water quality guidelines for the protection of aquatic life. For example, in 2013 the US Environmental Protection Agency (USEPA) Water Quality Criteria for ammonia was updated based on mussel (and

snail) toxicity data, and in 2011, the Canadian Water Quality Guidelines for chloride were updated to be protective of mussels. Although existing data indicate that at least some types of major ion pollution pose a significant threat to freshwater mussels, to date only a few of the most common species have been evaluated and with the most common salts (i.e., NaCl); little is known about the sensitivity of rare mussel species, or the effects of various pairings of cations and anions.

The goal of the present study was to evaluate the relative risks of major ions to common and imperiled freshwater mussel species to better inform derivation of water quality guidelines for protection of aquatic species. Therefore, we conducted standardized acute toxicity tests with major ion salts and glochidia from four species: *Lampsilis siliquoidea*, *L. fasciola*, *Hamiota altilis*, and *Anodonta suborbiculata*. For *L. siliquoidea* and *L. fasciola*, we obtained glochidia from two geographically distinct populations. *L. siliquoidea*, commonly known as the fatmucket, is currently one of the most commonly used freshwater mussel species for toxicity tests and has a wide distribution in North America with secure populations throughout its range (<http://explorer.natureserve.org/servlet/NatureServe?searchName=Lampsilis+siliquoidea>; last accessed 15 October 2021).

L. fasciola, the wavy-rayed lampmussel, was designated as a Species of Special Concern in Canada [17] but otherwise has a wide distribution and secure status in the United States [18]. *H. altilis*, commonly known as the finelined pocketbook, is endemic to the eastern reaches of the Mobile basin in Alabama, Georgia, and Tennessee, and historically in Mississippi. Decreasing range and declining populations resulted in *H. altilis* being listed as Threatened 1993 by the US Fish & Wildlife Service and as Endangered on the IUCN Red List [19]. *Anodonta suborbiculata*, the flat floater, is broadly distributed across the US and is characterized as having stable or expanding populations (<http://explorer.natureserve.org/servlet/NatureServe?searchName=Anodonta+suborbiculata>; last accessed 15 October 2021).

2. Materials and Methods

2.1. Test Organisms

Gravid adult female *L. siliquoidea* were obtained from captive populations maintained by the Kentucky Department of Fish and Wildlife Resources Center for Mollusk Conservation (Frankfort, KY, USA) (heretofore referred to as *L. siliquoidea* KY) and Missouri State University at the Kansas City Zoo (heretofore referred to as *L. siliquoidea* MO). Gravid adult female *L. fasciola* glochidia were obtained from captive populations maintained at the Alabama Aquatic Biodiversity Center (Marion, AL, USA) operated by the Alabama Department of Conservation and Natural Resources (heretofore referred to as *L. fasciola* AL) and from the Conservation Aquaculture Center (Marion, NC, USA) operated by the North Carolina Wildlife Resources Commission (heretofore referred to as *L. fasciola* NC). Gravid female *H. altilis* were also obtained from the Alabama Aquatic Biodiversity Center (Marion, AL, USA). Gravid female *A. suborbiculata* were obtained from culture ponds at Auburn University (Auburn, AL, USA).

The female mussels were transported in coolers with 10 L of their source water to the Aquatic Biology and Ecotoxicology Lab at the University of Georgia, where they were acclimated to dechlorinated tap water over 24–48 h and then placed in a Living Stream (Frigid Units, Inc., Toledo, OH, USA) with dechlorinated tap water at 17 °C. Mussels were fed every other day with a prepared algal diet [20]. All mussels were maintained for 1–2 weeks in the Living Stream prior to obtaining glochidia for testing.

Consistent with ASTM [20] guidelines for toxicity testing with mussel early life stages, viability was determined for glochidia from each female mussel prior to pooling the glochidia for distribution into the test beakers. Briefly, the glochidia were obtained from individual females ($n = 3$ to 6) by using a syringe filled with culture water to gently rupture the marsupial gills and flush the glochidia from a single female into a 1 L beaker with approximately 300 mL of culture water. Glochidia were gently suspended with a 100 mL pipette while 10 subsamples of 200 μ L each were collected and placed on a petri dish. Each

200 μ L drop was observed under a stereomicroscope (Leica MZ12, Leica Microsystems, Wetzlar, Germany) to determine the number of open and closed glochidia prior to addition of a drop of saturated NaCl solution, and again 60 s after the NaCl was added. Viability was calculated using the following equation: percent viability = (number closed after NaCl – number closed before NaCl)/(number closed after NaCl + number closed before NaCl) * 100. Roughly equal numbers (within 2-fold) of glochidia were pooled from the various females with viability >80%.

2.2. Test Chemicals and Target Concentrations

All test chemicals (salts) were certified ACS grade purchased from Fisher Scientific unless otherwise specified. Chloride salts included sodium chloride (NaCl, >99.0% purity, 7647-14-5), magnesium chloride ($\text{MgCl}_2 \cdot 6\text{H}_2\text{O}$, 99% purity, CAS 7791-18-6), potassium chloride (KCl, >99.0% purity, CAS 7447-40-7), and calcium chloride ($\text{CaCl}_2 \cdot 2\text{H}_2\text{O}$, 99% purity, CAS 10035-04-8). Sulfate salts included sodium sulfate (Na_2SO_4 , >99.0% purity, CAS 7757-82-6), magnesium sulfate (MgSO_4 , certified grade, CAS 7487-88-9), and calcium sulfate ($\text{CaSO}_4 \cdot 2\text{H}_2\text{O}$, >98% purity, CAS 10101-41-4).

Target test concentrations were based on literature for acute toxicity tests with other taxa (e.g., fish, cladocerans) and salt solubility limits. Target treatment levels for all toxicity tests (all mussel species and hardness levels) with NaCl, MgCl_2 , CaCl_2 , Na_2SO_4 , and MgSO_4 were control, 0.625, 1.25, 2.5, 5.0, and 10.0 g of salt/L added to test water. Target test concentrations for CaSO_4 were lower due to solubility limits and included: control, 0.125, 0.25, 0.5, 1.0, 2.0 g salt/L. Target test concentrations for KCl were control, 0.0063, 0.0125, 0.025, 0.05, and 0.1 g salt/L. Controls consisted of reconstituted water only, as described for each test below.

2.3. Glochidia Toxicity Test Methods

All experiments followed established methodology for conducting toxicity tests with early life stages of freshwater mussels [20]. Briefly, all toxicity tests consisted of six treatments (see above) with three replicates per treatment for a total of 18 experimental units, each consisting of 500–1000 glochidia in 100 mL of test solution in a 150 mL glass beaker. All replicates were randomized in a water bath maintained at 20 °C under ambient laboratory light at 14:10 h light:dark. Test solutions were not renewed during the test and glochidia viability was assessed as described previously in a sample of 100–200 glochidia at 24 h. Water samples (10 mL) were collected from the highest test concentration at the start of each test and acidified ($\text{pH} < 2$) and stored at 4 °C until analysis of appropriate target ions at the University of Georgia Center for Applied Isotope Studies where target cations (Na^+ , K^+ , Mg^{2+} , Ca^{2+}) were measured by inductively coupled plasma mass spectrometry (ICP-MS) and anions (Cl^- , SO_4^{2-}) were measured by ion chromatography. Water quality parameters (temperature, dissolved oxygen, pH) were measured at the beginning of each test in one randomly selected replicate from each treatment. Hardness and alkalinity of the dilution water for each test were measured with standard titration methods [21].

Survival in each test concentration was corrected for mortality in the controls. Data analysis was performed to determine the EC50 (median effect concentration) with the Toxicity Relationship Analysis Program (TRAP; https://archive.epa.gov/med/med_archive_03/web/html/trap.html; last accessed 20 July 2021).

2.4. Toxicity of Chloride and Sulfate Salts to *Lampsilis siliquoidea*

We evaluated the toxicity of chloride salts (sodium chloride, magnesium chloride, calcium chloride, and potassium chloride) and sulfate salts (sodium sulfate, magnesium sulfate, and calcium sulfate) to *L. siliquoidea* (MO) glochidia. Toxicity of sodium chloride was also assessed with *L. siliquoidea* (KY) for comparison among geographically distinct populations. Pooled initial viability for the MO population ($n = 4$) was 84.0% and the KY population ($n = 4$) was 93.1%. All tests were conducted in moderately hard reconstituted water [22].

2.5. Species Sensitivity Comparison for Chloride Salts

We compared toxicity of chloride salts (NaCl and KCl) to glochidia of common and rare mussel species, and in some cases, the same species obtained from distinct populations (i.e., different states). NaCl toxicity was assessed with *L. siliquoidea* (KY and MO), *L. fasciola* (NC and AL), *A. suborbiculata*, and *H. altilis*. Toxicity of KCl was evaluated with *L. siliquoidea* (MO), *L. fasciola* (AL), and *H. altilis*. All tests were conducted in moderately hard reconstituted water [22].

2.6. Water Hardness Effects on Chloride Toxicity

We assessed the effects of water hardness on chloride (NaCl) toxicity in glochidia of three species: *L. siliquoidea* (KY), *L. fasciola* (NC and AL), and *A. suborbiculata*. Test water hardness categories included soft (0–60 mg CaCO₃/L), moderately hard (61–120 mg CaCO₃/L), hard (121–180 mg CaCO₃/L), and very hard (181–300 mg CaCO₃/L) as characterized by USGS (<https://water.usgs.gov/owq/hardness-alkalinity.html>; last accessed 26 May 2020), with an additional category for extremely hard (≥ 301 mg CaCO₃/L). Test water hardness was established by proportionally adjusting the amounts of the five salts used to create moderately hard reconstituted water, as described by Smith et al. [22]. Tests with *L. siliquoidea* and *A. suborbiculata* were performed at all five hardness levels; tests with *L. fasciola* (NC and AL) were conducted at all levels except extremely hard due to insufficient numbers of glochidia.

3. Results

3.1. Glochidia Viability, Water Quality Parameters, and Ion Concentrations

Pooled initial glochidia viability before testing was above 80% for all samples and ranged from 84 to 93.2%. Water quality parameters were within acceptable ranges for all toxicity tests: temperature ranged from 19.7–20.2 °C, dissolved oxygen ranged from 7.68–9.45 mg/L, and pH ranged from 7.68–8.32. Measured anion concentrations of test salts averaged 96.0 ($\pm 7.3\%$, $n = 22$) of target concentrations. Consistent with test guidelines [20], control survival was $\geq 90\%$ for all tests.

3.2. Toxicity of Chloride and Sulfate Salts to *L. siliquoidea*

Median effective concentrations (24-h EC₅₀) of chloride salts to *L. siliquoidea* ranged from 2.47 g/L to 3.6 g/L for all chloride salts except KCl, which had an EC₅₀ of 0.056 g/L (Figure 1). Toxicity of sulfate salts was generally similar to chloride salts (other than KCl) (Figure 1). The EC₅₀ for Na₂SO₄ was 5.32 g salt/L (4.93–5.78) and MgSO₄ was 3.52 g salt/L (3.20–3.85). An EC₅₀ for CaSO₄ could not be calculated because of low mortality in the highest treatment concentrations, so it was reported as >2.0 g salt/L, which was the highest concentration tested and it approached solubility limits. The difference in NaCl EC₅₀s between the two *L. siliquoidea* populations (MO, 2.47 g/L and KY, 3.60 g/L) exceeded the variability in EC₅₀s for to NaCl, MgCl₂, and CaCl₂ within a single (MO) population.

3.3. Species Sensitivity Comparison for Chloride Salts

NaCl 24-hr EC₅₀s varied among the four species by a factor of 2.4 (Figure 2); the lowest EC₅₀ was 1.56 g/L (1.33–1.80) for *L. fasciola* (AL), and the highest was 3.60 g/L for *L. siliquoidea* (KY). The Federally Threatened species, *H. altilis*, had an EC₅₀ of 2.42 g/L (2.11–2.77), near the geometric mean ($n = 6$) of 2.36 g/L. Within species, NaCl EC₅₀s for the two *L. siliquoidea* populations (MO, 2.47 g/L and KY, 3.60 g/L) varied more than the two *L. fasciola* populations (AL, 1.56 g/L and NC, 1.95 g/L), but neither species varied by more than 1.5-fold.

Figure 1. Twenty-four-hour EC50s for *Lampsilis siliquoidea* glochidia exposed to chloride (gray bars) or sulfate (black bars) salts in moderately hard water (hardness 80–90 mg/L) at 20 °C. Origin of glochidia was Missouri, USA, except for one batch noted from Kentucky (KY). Error bars represent 95% confidence limits.

Figure 2. Twenty-four-hour NaCl EC50s for glochidia from *Lampsilis siliquoidea*, *L. fasciola*, *Anodonta suborbiculata*, and *Hamiota altilis* (state of origin in parenthesis). All tests were conducted in moderately hard reconstituted water (hardness 80–90 mg/L) at 20 °C. The dashed line indicates the geometric mean of the EC50s for the six tests. Error bars represent 95% confidence limits.

Toxicity of KCl was generally similar among the three species tested, varying by a factor of 1.3 (Figure 3). The 24 h EC50s ranged from a low of 0.040 g/L (0.037–0.043) in *H. altilis* to 0.053 g/L (0.048–0.057) in *L. siliquoidea*, with a geometric mean ($n = 3$) of 0.046 g/L.

Figure 3. Twenty-four-hour KCl EC50s for glochidia from *Lampsilis siliquioidea*, *L. fasciola*, and *Hamiota altilis*. All tests were conducted in moderately hard reconstituted water (hardness 80–90 mg/L) at 20 °C. The dashed line indicates the geometric mean of the EC50s for the three species. Error bars represent 95% confidence limits.

3.4. Water Hardness Effects on Sodium Chloride Toxicity

For glochidia from all three species (*L. siliquioidea*, *L. fasciola*, and *A. suborbiculata*), the lowest NaCl EC50s occurred in the softest water (Figure 4). Toxicity generally decreased with increasing water hardness, at least to a point. For example, the 24 h NaCl EC50 for *L. siliquioidea* was lowest at 2.70 g/L in water with hardness 42 mg/L as CaCO_3 and the EC50 increased by 2.24-fold to 6.06 g salt/L at a hardness of 168 mg/L as CaCO_3 , though at a hardness of 464 mg/L as CaCO_3 the EC50 was 3.57 g salt/L (Figure 4). A similar trend generally occurred with *A. suborbiculata* but the highest EC50s (Approx. 5.8 g salt/L) were not reached until the highest hardness levels of 310 and 548 mg CaCO_3 /L, with no decline of EC50 at the highest hardness level. The trend for a protective effect of hardness on NaCl toxicity was similar but less pronounced for the *L. fasciola* (NC) population with NaCl EC50s increasing 2.5-fold from a low of 1.31 g/L at a hardness of 43 mg/L as CaCO_3 to a high of 3.21 g/L at a hardness of 302 mg/L as CaCO_3 , which was the highest hardness level tested for this species.

Figure 4. Effect of water hardness on 24 h NaCl EC50s for glochidia from *Lampsilis siliquioidea*, *L. Fasciola*, and *Anodonta suborbiculata*. Error bars represent the upper and lower 95% confidence limits. Dashed lines indicate general trends and are not fitted regression lines.

4. Discussion

4.1. Toxicity of Chloride and Sulfate Salts to Glochidia

Acute toxicity tests with salts of chloride and sulfate revealed that freshwater mussel glochidia are sensitive to these major ions, consistent with results from previous glochidia ion toxicity research [5,9,12] as well as ion testing with juvenile mussels [16]. Among chloride toxicity tests in the present study, glochidia were substantially more sensitive to KCl than other chloride salts, consistent with juvenile mussel KCl toxicity test results [16], suggesting that potassium may be highly regulated in mussel physiology. While chloride has been the focus in previous studies on freshwater mussel major ion sensitivity [5,14], this differential toxicity among chloride salts also suggests that the cations (e.g., calcium, sodium, magnesium, and potassium) of the salts studied should be taken into consideration when evaluating acute toxicity of major anions like chloride and sulfate. Similar trends have been reported for other freshwater organisms such as fish and cladocerans [23], and the approximate order of chloride salt toxicity is $\text{KCl} > \text{MgCl}_2 > \text{CaCl}_2 > \text{NaCl}$.

The similarity in sensitivity among the four species tested with NaCl and among the three species tested with KCl suggests that glochidia of different unionid species may be similar in their tolerance to at least some major ions. *Hamiota altilis*, the federally threatened species, was the most sensitive of the three species tested with KCl, but *H. altilis* sensitivity to NaCl was near the mean of the four species (six tests). Additional testing is warranted to determine if other rare species are more sensitive to major ions. For NaCl, the variability in sensitivity between different populations of the same species was more limited than the overall variability among species. Unfortunately, insufficient numbers of glochidia were available to perform similar comparisons with different populations and other salts.

Not only was sensitivity to NaCl similar among the four species of glochidia we tested, but it was also similar to the range of sensitivity reported for juvenile mussels of a variety of species ([16] Figure 5). Like juveniles, glochidia are among the most sensitive organisms tested to date with NaCl ([16] Figure 5). Similarly, our tests with glochidia from three species demonstrated that glochidia are the most sensitive organisms tested to date with KCl, and glochidia appear to be about an order of magnitude more sensitive to KCl than juvenile mussels ([16] Figure 6). These findings suggest that glochidia would drive any potential water quality criteria derivation for potassium or KCl.

4.2. Effect of Water Hardness on Chloride Sensitivity

Chloride (NaCl) toxicity tests at a range of water hardness in the present study suggested that mussels residing in naturally hard waters may be afforded some protection against major ion toxicity; however, further testing is needed to confirm this finding. These results are consistent with hardness testing results from major ion tests with other commonly tested taxa [24,25]. However, the protective effect of hardness to glochidia may be limited; *L. siliquoidea* glochidia in the present study demonstrated greater sensitivity to NaCl as hardness exceeded the very hard (170 mg/L as CaCO_3) range. This toxicity should not be surprising because hardness is comprised primarily of Ca and Mg, which caused toxicity in the g/L range to glochidia in our other tests when paired with chloride and sulfate. Greater buffering of toxicity by hardness may be possible with longer acclimation periods with high hardness water, but unfortunately glochidia remain viable outside of the adult female for only a few days [26], so long acclimation periods (days) prior to testing are not feasible with 24 h toxicity tests. One possibility is to acclimate brooding females in harder waters for long periods prior to obtaining glochidia for testing.

4.3. Implications

The current USEPA criterion maximum (short-term) concentration for chloride is 860 mg Cl^- /L and the criterion continuous (long-term) concentration for chloride is 230 mg Cl^- /L [27]. Derivation of these criteria did not include freshwater mussel toxicity data, largely because standardized freshwater mussel toxicity testing guidelines were not available at the time of the derivation in 1988. Current Canadian Water Quality Guidelines

for Chloride were published in 2011 and included glochidia and juvenile mussel toxicity data [28]. As expected, these criteria are lower than USEPA criteria, with a short-term exposure limit of 640 mg Cl[−]/L and a long-term exposure limit of 120 mg Cl[−]/L. In the present study, chloride (NaCl) EC50 for *L. fasciola* was 590 mg Cl[−]/L in soft water and 938 mg Cl[−]/L in moderately hard water, suggesting that current EPA chloride criteria may not adequately protect this species in softer waters. These EC50 values are similar to previous reports for toxicity of NaCl to a closely related species, *L. siliquoidea* [5,12].

Additional research to further characterize freshwater mussel sensitivity to major ions is needed to advance understanding of freshwater mussel declines in North America and elsewhere. Among other needs, testing with more species to better represent the approximately 300 extant species in North America is a high priority. Furthermore, comparisons of sensitivity of populations from different regions throughout a species' range are rare and more of these types of tests would likely help explain some of the variability in sensitivity among tests with the same species. Based on our findings, some of this variability may also be attributed to water quality, particularly hardness, of the sites where mussels are collected for testing. Additionally, little is known about the variability in sensitivity of glochidia that are collected at different times of the year. For example, some species, such as those of the genus *Lampsilis*, brood glochidia nearly year-round, whereas other species brood for only a few weeks out of the year. Few studies have compared the sensitivity of glochidia collected early in their brooding period to those collected near the end of the period, but Fritts et al. [29] reported that *L. cardium* glochidia collected early in their brooding period were significantly less sensitive to NaCl than those collected late in the brooding period. Tests that utilize the most sensitive life stages would be most protective of natural populations.

Beyond additional laboratory toxicity testing, ambient water quality should be monitored, especially at sites of great mussel diversity, to determine if major ion concentrations are at or above levels that freshwater mussels can tolerate. For example, Gillis [5] found that chloride concentrations measured at mussel beds in Canada periodically reached acutely toxic levels for the mussel species tested in that study. In addition, Prosser et al. [9] found that several rivers in Ontario showed a risk of reaching chronically toxic levels of major ions to freshwater mussels. Monitoring of ambient chloride, potassium, and other major ions is especially needed during times of the year when freshwater mussels are releasing glochidia since glochidia sensitivity to major ions appears to be similar to juveniles.

5. Conclusions

In summary, freshwater mussels appear to be among the most sensitive organisms to at least some major ions, such as chloride and potassium. Hardness may provide some amelioration of toxic effects of chloride (and possibly other ions). Finally, NaCl toxicity tests indicate that commonly tested freshwater mussel species may be generally protective of rare species; however, additional testing is needed to confirm this finding. Further testing with variation in other water quality parameters would add to the environmental realism and application of standardized toxicity tests.

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