

Article Effect of the Supporting Electrolyte on Chloroform Reduction at a Silver Electrode in Aqueous Solutions

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Abstract: Herein, we report, for the first time, a comparative study on the electrocatalytic reduction of chloroform on silver in different aqueous supporting electrolytes. Cyclic voltammetry measurements were performed at a wide range of scan rates and concentrations of CHCl₃ using 0.05 M NaClO₄, NaH₂PO4, and Na₂HPO₄ as supporting electrolytes. We observed that a type of supporting electrolyte anion strongly influences both the potential as well as the current density of the chloroform reduction peak, mainly due to the presence of OH⁻ in an alkaline Na₂HPO₄ solution, which is a specifically interacting anion. Moreover, the highest sensitivity of the Ag electrode toward CHCl₃ reduction was observed in a neutral NaClO₄ aqueous solution. It was found that the electroreduction of chloroform at the silver surface occurs via a concerted mechanism regardless of the type of the studied supporting electrolyte.

Keywords: silver electrode; supporting electrolyte; electrocatalytic reduction; trichloromethane

1. Introduction

Chloroform and other trihalomethanes (THMs) are known to be environmental pollutants [1], toxic, and possibly carcinogenic agents to humans [2]. The use of chloroform in consumer products (surgical anesthetic, cough syrups, toothpaste, etc.) was banned in the USA in 1976, as it was found to be carcinogenic in animal tests [3]. The issue of water contamination with THMs is reflected by the EU [4] and USA [5] regulations with the total trihalomethanes level in drinking water at maximum allowable averages of 100 and $80 \ \mu g \ L^{-1}$, respectively. The known sources of chloroform pollution are either environmental (volcanic emission and marine algae) or anthropogenic [6]. The CHCl₃ can be formed during waste incineration or as a water chlorination by-product. Moreover, chloroform is still widely used as a solvent in the pharmaceutical industry, dyes and pesticides synthesis, fire extinguishers, rubber industry, etc. and, therefore, it can be emitted to the environment in a form of exhaust gas and wastewater.

Catalytic cathodic reduction at silver electrodes, aiming to convert organic halides to the corresponding de-halogenated hydrocarbons, has been reported as a promising approach toward the reduced abatement of such environmental pollutants [7]. At this point, most of the reported studies on the electrocatalytic reduction of organic halides were oriented to explore the mechanism of bond-braking reactions in nonaqueous solutions [8–10]. The aqueous solutions were also studied, regardless of the fact that the electrochemical reduction of many organic halides takes place at very negative potentials, at which a simultaneous hydrogen evolution reaction (HER) occurs, significantly increasing the energy consumption of the process [11–19]. Nonetheless, further studies on principal aspects of the chloroform electrocatalytic reduction at silver electrodes in aqueous solutions is a crucial step toward environmental applications.



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To date, Rondinini et al. [20] studied the effect of the supporting electrolyte on the electrocatalytic activity of silver in nonaqueous solutions. They explored a large set of alkylammonium salts, and showed that the reduction potentials were spread widely in a range of about 300 mV. Authors reported that the major factor determining the reduction potential is the nature of the anion, while the cations exert a finer modulating effect. Moreover, the current density of the peak was also dependent on the supporting electrolyte. Fiori et al. [8] and Rondinini et al. [21] later confirmed the significant influence of the supporting electrolyte on the reduction potential of organic halides at Ag electrodes in a nonaqueous environment. For the chloroform reduction at silver electrodes in aqueous solutions, tetraethylammonium benzoate [22], KClO₄ [14,17,23,24], NaClO₄ [12], and NaOH [16] were reported as supporting electrolytes. NaClO₄ and KClO₄ are reportedly used due to negligible specific interactions of silver with perchlorate anions and sodium/potassium cations [25]. Taking into consideration the effect of the supporting electrolyte clearly demonstrated for nonaqueous solutions, it is surprising that no comparative study was reported to date for aqueous solutions. Therefore, in this work, aqueous NaClO₄, NaH₂PO₄, and Na₂HPO₄ solutions were selected and studied as supporting electrolytes for the electrochemical reduction of chloroform at Ag.

2. Results and Discussion

During the reduction of chloroform at silver electrodes in nonaqueous solvents, 2 or 3 reduction peaks are observed [8–10,21,26], which can be ascribed to a gradual dehalogenation of the molecule, i.e., successive removal of halogen atoms [9]. In aqueous solutions, typically one wide peak of chloroform reduction, shifted toward less negative potential values as compared to nonaqueous solvents (up to 0.44 V), was observed [27]. The shift of the chloroform reduction peak toward less negative values, as compared to nonaqueous solvents (up to 0.44 V), was observed [27]. The shift of the chloroform reduction peak toward less negative values, as compared to nonaqueous solvents, was attributed to the ability of water to enhance the turnover frequency of catalytic sites, as it favors the desorption processes [13]. It is also assumed that the electrocatalytic reduction of CHCl₃ in an aqueous solvent is a complex phenomenon. Consequently, a superimposition of different peaks, corresponding to subsequent steps in the reaction mechanism, occurs [23], hindering the correct analysis of the process. Table 1 shows chloroform reduction peak potentials (E_p), scan rates (ν), and concentrations of CHCl₃ (c) reported in the literature.

Electrode	E _p , V vs. SCE	$v_{s}^{\nu_{r}}$	C CHCl3, mM	Supporting Electrolyte	Ref.
Ag	-1.05	0.1	5.0	0.010 M TEAB	[22]
Ag	-1.0	0.1	5	0.035 M NaClO ₄	[28]
Ag	-1.32	0.1	8.26	0.05 M KClO ₄	[17]
Ag NWs	-1.15	0.1	8.26	0.05 M KClO ₄	[17]
Ag NPs @ modified carbon powder	-0.85	0.2	10	0.1 M KClO ₄	[15]
Ag NPs in C-ME	-0.61	0.3	10	0.1 M KClO ₄	[14]
Ag	-1.127	0.1	10.63	0.1 M NaOH	[16]
Ag NCs	-1.067	0.1	10.63	0.1 M NaOH	[16]
Ag NWs	-0.957	0.1	10.63	0.1 M NaOH	[16]

 Table 1. Overview of experimental data for the chloroform reduction in aqueous solutions.

TEAB—tetraethylammonium benzoate, NWs—nanowires, NPs—nanoparticles, C-ME—microcavity electrode, NCs—nanocones.

From the data gathered in Table 1, one can conclude that the direct comparison of the influence of different electrode materials from the literature reports is not possible, as the scan rate, concentration of chloroform, as well as type and concentration of the supporting electrolyte, strongly differ. It was previously shown in cyclic voltammetry (CV) measurements that both the potential and current density of the chloroform reduction peak strongly depend on the concentration of CHCl₃ [17]. The CV measurements were performed at

the silver electrode (Figure 1) in order to assess the effect of the supporting electrolyte on the potential and current density of the chloroform reduction peak. As expected, the EDS spectrum of the Ag electrode (Figure 1c) confirmed its chemical composition and, consequently, the planes (111), (200), (220), (1311), and (222) indexed to face-centered-cubic silver (JCPDS No. 04-0783) were detected in the XRD pattern.



Figure 1. FE-SEM micrographs of the silver electrode (**a**) and a higher magnification of the selected area (**b**). The EDS analysis (**c**) and XRD diffraction pattern (**d**) of the Ag electrode.

The voltammograms obtained at the Ag electrode in the absence and presence of 8.26 mM chloroform for selected scan rates are shown in Figure 2 for 0.05 M Na₂HPO₄, NaClO₄, and NaH₂PO₄ solutions. The potentials are referenced to the saturated calomel electrode (SCE). In all studied electrolytes in the potential range of -0.547 to -1.447 V vs. SCE, a competitive reaction (i.e., decomposition of water or supporting electrolyte) was not observed. Only small capacitive currents were observed in the absence of chloroform. Conversely, independently of the studied electrolyte, after the addition of chloroform, one irreversible peak was observed similarly to previous reports (Table 1). Both the potential as well as the current density of the chloroform reduction peak shifted with the scan rate. Such an effect was previously reported for the chloroform reduction in aqueous solutions at Ag [17] and during benzyl chloride reduction in acetonitrile at Ag [29] and AgPd [30] electrodes. It is important to note that the magnitude of the shift strongly depends on the type of studied supporting electrolyte—especially when the results are compared with previously reported data for the chloroform reduction in 0.05 M KClO₄ in the same setup [17]. The most significant shift in the reduction potential and highest current density were observed for Na₂HPO₄. This fact will be discussed later.



Figure 2. Cyclic voltammograms obtained at the silver electrode in the absence and presence of 8.26 mM chloroform in 0.05 M (**a**) Na₂HPO₄, (**b**) NaClO₄, and (**c**) NaH₂PO₄ solutions for selected scan rates.

The reduction of organic halides occurs according to a stepwise or concerted pathway (Figure 3).



Figure 3. Schematic representation of the mechanism of chloroform reduction at the surface of Ag electrode in aqueous solutions.

In the stepwise mechanism, an intermediate radical anion is formed before the C–X bond cleavage. In the concerted mechanism, the electron transfer and the C–X bond cleavage occur simultaneously. Those reactions are considered the slowest step of the reduction. The second electron transfer to the formed radical leads to the formation of organic anions, and is recognized as a very fast step occurring at less negative potentials. The created carbanion is protonated rapidly by a proton donor [27]. The cleavage mechanism of organic halide R–X bonds depends mainly on the type of reduced molecule and electrode material [31]. A kinetic indicator κ can be a useful tool to discriminate between the mechanisms involved in the R–X bond cleavage [32], and can be determined by voltammetric analysis using Equations (1) and (2) [26]:

$$\kappa_{1} = -\frac{1.151 \text{RT}}{\text{F}\left(\frac{\partial E_{p}}{\partial \log \nu}\right)} \tag{1}$$

$$\varsigma_2 = -\frac{1.857RT}{F(E_{p/2} - E_p)}$$
 (2)

where: κ —electron transfer coefficient; R—universal gas constant, R = 8.3145 J mol⁻¹ K⁻¹; T—temperature, K; v—scan rate, V s⁻¹; E_p—reduction potential of chloroform, V; E_{p/2}—potential corresponding to half of the peak current, V; F—Faraday constant, F = 9.6485 · 10⁴ C mol⁻¹.

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If κ is in the range between 0.35 and 0.5, the process occurs via the stepwise mechanism with an electrochemical rate-determining step. For $\kappa < 0.35$, the reaction follows the concerted mechanism and κ values can be considered as a transfer coefficient [17]. Moreover, Equation (1) clearly shows that the observed shift in the reduction potential with scan rate is correlated with the electron transfer coefficient. As can be seen in Figure 4, the peak current density (j_p) is linearly dependent on the square root of the scan rate ($\nu^{0.5}$) for all used supporting electrolytes, which indicates that the reaction is diffusion-controlled [31]. Since E_p vs. log ν dependencies (insets in Figure 4) are linear with negative slopes, it was possible to estimate κ values (Table 2) from Equations (1) and (2).



Figure 4. The dependencies of the cathodic current density of the CHCl₃ reduction peak (j_p) vs. the square root of the scan rate ($v^{0.5}$) measured in 0.05 M (**a**) Na₂HPO₄, (**b**) NaClO₄, and (**c**) NaH₂PO₄. Insets: The corresponding dependencies of the peak potential (E_p) vs. log v.

Table 2. Data for the chloroform reduction at the silver electrode in selected aqueous supporting electrolytes.

Electrolyte	рН	E _p , V vs. SCE at 100 mV s ⁻¹	j _p , mA cm ⁻² at 100 mV s ⁻¹	$\frac{\partial E_p}{\partial logv}$	к ₁	к ₂
Na ₂ HPO ₄	9.4	-1.215	-4.89	-0.171	0.17	0.17
NaClO ₄	7.0	-1.344	-4.61	-0.284	0.10	0.15
NaH ₂ PO ₄	4.6	-1.389	-4.77	-0.287	0.10	0.10

For all studied supporting electrolytes, the observed current density of the chloroform reduction peak is similar; however, the peak potential strongly differs. Moreover, the calculated κ values are significantly lower than 0.35 and are in good agreement for both estimation methods. It indicates that the chloroform reduction in studied supporting electrolytes takes place via the concerted mechanism at the surface of the polycrystalline silver electrode (Figure 5). Such a mechanism is often observed for cases when a proton donor is present in the solution [31]. For instance, the transfer coefficient reported for the chloroform reduction at the Ag electrode was in the range of 0.11–0.13 for 8.26 mM CHCl₃ in a 0.05 M aqueous solution of KClO₄ [17]. The charge transfer coefficient depends on the peak potential of analyte reduction [33,34]; therefore, observed discrepancies are expected.



Figure 5. Kinetic indicator κ for identification of the mechanism involved in the R-X bond cleavage together with experimental κ values obtained for the reduction of chloroform in different aqueous supporting electrolytes.

Similarly, as in the case of the electrocatalytic reduction of organic halides in nonaqueous solutions, the selected aqueous supporting electrolytes influence the reduction potential of the studied compound [20]. Importantly, the nucleophilic character of the hydrogen phosphate ion (HPO₄^{2–}) has to be considered, which is about twice as great as for the hydroxyl ion, leading to hydrolysis-like reactions and providing easier removal of Cl[–] ions [35]. This could lead to (*i*) lowering of the concentration of chloroform in the electrolyte and, therefore, a less negative potential of the chloroform reduction, and (*ii*) a slightly higher cathodic current density of the reduction peak due to a higher turnover frequency in the case of Na₂HPO₄. However, further extensive studies are necessary for the better understanding of this phenomenon. Moreover, in the aqueous solutions, additional hydrolysis of chloroform can occur, according to a neutral mechanism (Equation (3)) or an alkaline-catalyzed mechanism (Equation (4)) [35]:

$$CHCl_3 + 5H_2O \rightarrow HCOOH + 3H_3O^+ + 3Cl^-$$
(3)

$$CHCl_3 + 4OH^- \rightarrow HCOO^- + 2H_2O + 3Cl^-$$
(4)

The neutral hydrolysis is the dominant pathway at pH < 4. Between pH 4 and 8, none of the hydrolysis mechanisms are predominant, and at pH > 8, the alkaline-catalyzed hydrolysis pathway becomes important due to the strong neutrophilic character of the hydroxyl ion [35]. The presence of OH^- , specifically interacting ion in a solution, is therefore another reason for the less negative potential of chloroform reduction observed in an alkaline solution due to a lower concentration of chloroform. The hydrolysis half-life of chloroform at room temperature is 1000 years at pH = 7, and even longer for more acidic solutions; therefore, the influence of hydrolysis could be neglected.

Subsequently, the influence of chloroform concentration on the peak potential and current density for all studied electrolytes was explored. The CV measurements were performed in a 0.05 M solution of supporting electrolyte at the scan rate of 100 mV s⁻¹. The registered voltammograms are shown in Figure 6. For all studied supporting electrolytes, a linear correlation between the peak current density and CHCl₃ concentration was observed for the concentration range of 1.03–6.20 mM (see insets in Figure 6). Similarly to previous reports [10,17], the potential of chloroform reduction shifts toward more negative values with its concentration. Therefore, it is important to consider the concentration of organic halides, when the catalytic activity of different electrodes is compared. The calculated sensitivity, limit of detection (LOD, calculated as $3\sigma_b/a$), and limit of quantification (LOQ, calculated as $10\sigma_b/a$) are gathered in Table 3. Clearly, the lowest LOD and LOQ values, as well as the highest sensitivity, were obtained for NaClO₄, which implies an inert character of the supporting electrolyte and justifies the claim that this supporting electrolyte is the best choice among those studied here.



Figure 6. Cyclic voltammograms obtained at 100 mV s⁻¹ for the silver electrode in a 0.05 M solution of (**a**) Na₂HPO₄, (**b**) NaClO₄, and (**c**) NaH₂PO₄ containing trichloromethane in the concentration range of 0–6.20 mM. Insets: The corresponding dependencies of the peak cathodic current density of CHCl₃ reduction (j_p) vs. chloroform concentration.

Electrolyte	LOD, mM	LOQ, mM	Sensitivity, mA cm ⁻² M ⁻¹
Na ₂ HPO ₄	0.348	1.16	1.04
NaClO ₄	0.204	0.68	1.10
NaH ₂ PO ₄	0.354	1.18	1.02

Table 3. The analytical parameters of the chloroform reduction at silver electrode in different aqueous solutions.

3. Materials and Methods

Pure disodium phosphate (Na₂HPO₄·2H₂O), sodium phosphate (NaH₂PO₄·H₂O), and sodium perchlorate (NaClO₄·H₂O) were purchased from Sigma-Aldrich (St. Louis, MO, USA), and chloroform (CHCl₃) was acquired from Chempur[®] (Piekary Śląskie, Poland). The polishing Al₂O₃ fine abrasive powder was purchased from Alfa Aesar[®] (Thermo Fisher Scientific, Lancashire, United Kingdom). Highly deionized water (18.2 MΩcm) was obtained using the MilliQ[®] Millipore System (Merck KGaA, Darmstadt, Germany).

All electrochemical measurements were performed in a custom-made, conventional three-electrode setup with the Gamry Instrument Reference 3000 potentiostat/galvanostat (Warminster, PA, USA). For cyclic voltammetry measurements, Teflon-coated polycrystalline silver (Ag, 7 mm diameter, 99.999% Specpure, Johnson Matthey Chemicals Ltd., England) served as a working electrode, the Ag | AgCl | Cl⁻ (3 M NaCl) electrode in a Luggin capillary (G0100 Vycor, BASi, West Lafayette, IN, USA) filled with a 0.1 M solution of the selected supporting electrolyte was used as a reference electrode, and a Pt wire coiled on a Teflon[®] frame was used as a counter electrode [17]. The CV measurements were recorded at the potential range from -0.547 to -1.447 V at selected scan rates $(25-200 \text{ mV s}^{-1})$ and chloroform concentrations (0–6.02 mM). All used electrolytes, i.e., NaClO₄, NaH₂PO₄, Na₂HPO₄, were prepared using deionized water. Before each series of experiments, the surface of the silver electrode was polished with the Al_2O_3 powder, and electrochemically cleaned between each measurement in the 0.05 M supporting electrolyte by CV cycling in abovementioned potential range. Due to the fact that in the literature, the reduction potentials for organic halides are consistently provided versus the SCE, all potentials measured in this study were also recalculated to this reference.

The surface morphology of Ag electrode and its composition was studied using a field emission scanning electron microscope (FE-SEM, Hitachi S-4700, Hitachi High-Tech Corporation, Tokyo, Japan) equipped with an energy-dispersive X-ray spectrometer (EDS, Noran System 7, Thermo Electron Scientific Instruments LLC, Madison, WI, USA). The crystallinity of the Ag electrode was examined using a Rigaku Mini Flex II diffractometer with Cu K α radiation (1.54060 Å) at 30–90° 2 θ with a step size of 0.005° and at a rate of 5° min.⁻¹.

4. Conclusions

In this work, for the first time, the influence of the supporting electrolyte on the potential of chloroform reduction in aqueous environment at the surface of the silver electrode was reported. Similarly to previously reported nonaqueous supporting electrolytes, we observed that both the peak current density and peak potential of the chloroform reduction depend on the type of the used supporting electrolyte. On the other hand, it was found that the electroreduction of chloroform at the silver surface occurs via a concerted mechanism regardless of the type of studied supporting electrolyte. The highest sensitivity of the Ag electrode toward CHCl₃ reduction was observed in a neutral NaClO₄ aqueous solution. NaClO₄ is widely reported in the literature as an inert aqueous supporting electrolyte for the electrocatalytic reduction of organic halides at Ag. However, the influence of other supporting electrolytes and noninert additives affecting the solution pH is often ignored. This is particularly important for the electrocatalytic reduction of organic halides in real water samples, where multiple anions and cations can be present. **Author Contributions:** Conceptualization, A.M.B. and A.B.; methodology, A.M.B. and A.B.; formal analysis, A.M.B.; investigation, A.M.B. and A.B.; writing—original draft preparation, A.M.B.; writing—review and editing, A.B. and G.D.S.; visualization, A.M.B. and G.D.S.; supervision, G.D.S.; project administration, A.M.B.; funding acquisition, A.M.B. and G.D.S. All authors have read and agreed to the published version of the manuscript.

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References

- 1. Trihalomethanes in Drinking-Water. In *Background Document for Development of WHO Guidelines for Drinking-Water Quality,* WHO/SDE/WSH/03.04/64; World Health Organization: Geneva, Switzerland, 2004; pp. 1–18.
- 2. Chloroform. IARC Monographs on the Evaluation of Carcinogenic Risks to Humans; IARC: Lyon, France, 1999; Volume 73, pp. 131–182.
- 3. Huang, B.; Lei, C.; Wei, C.; Zeng, G. Chlorinated volatile organic compounds (Cl-VOCs) in environment—sources, potential human health impacts, and current remediation technologies. *Environ. Int.* **2014**, *71*, 118–138. [CrossRef]
- 4. Council Directive 98/83/EC of 3 November 1998 on the Quality of Water Intended for Human Consumption. *Off. J. L* **1998**, 330, 32–54.
- 5. *Stage 2 Disinfectants and Disinfection Byproducts Rule (Stage 2 DBPR) 71 FR 388;* United States Environmental Protection: Washington, DC, USA, 2006; Volume 71, pp. 1–2.
- Ivahnenko, T.; Zogorski, J.S. Sources and Occurrence of Chloroform and Other Trihalomethanes in Drinking-Water Supply Wells in the United States. 1986–2001; Scientific Investigations Report 2006–5015; U.S. Geological Survey: Storrs, CT, USA, 2006.
- Sonoyama, N.; Seike, S.; Sueoka, T.; Sakata, T. Electrochemical decomposition of ppb level trihalomethane in tap water. J. Appl. Electrochem. 2003, 33, 1049–1055. [CrossRef]
- 8. Fiori, G.; Rondinini, S.; Sello, G.; Vertova, A.; Cirja, M.; Conti, L. Electroreduction of volatile organic halides on activated silver cathodes. *J. Appl. Electrochem.* 2005, *35*, 363–368. [CrossRef]
- 9. Isse, A.A.; Sandonà, G.; Durante, C.; Gennaro, A. Voltammetric investigation of the dissociative electron transfer to polychloromethanes at catalytic and noncatalytic electrodes. *Electrochim. Acta* 2009, *54*, 3235–3243. [CrossRef]
- Vertova, A.; Barhdadi, R.; Cachet-Vivier, C.; Locateli, C.; Minguzzi, A.; Nedelec, J.-Y.; Rondinini, S. Cavity microelectrodes for the voltammetric investigation of electrocatalysts: The electroreduction of volatile organic halides on micro- sized silver powders. *J. Appl. Electrochem.* 2008, *38*, 965–971. [CrossRef]
- 11. Xu, Y.; Zhu, Y.; Zhao, F.; Ma, F.C. Electrocatalytic reductive dehalogenation of polyhalogenated phenols in aqueous solution on Ag electrodes. *Appl. Catal. A* 2007, 324, 83–86. [CrossRef]
- 12. Scialdone, O.; Guarisco, C.; Galia, A.; Filardo, G.; Silvestri, G.; Amatore, C.; Sella, C.; Thouin, L. Anodic abatement of organic pollutants in water in micro reactors. *J. Electroanal. Chem.* **2010**, *638*, 293–296. [CrossRef]
- 13. Scialdone, O.; Guarisco, C.; Galia, A.; Herbois, R. Electroreduction of aliphatic chlorides at silver cathodes in water. *J. Electroanal. Chem.* **2010**, *641*, 14–22. [CrossRef]
- 14. Rondinini, S.; Mussini, P.R.; Muttini, P.; Sello, G. Silver as a powerful electrocatalyst for organic halide reduction: The critical role of molecular structure. *Electrochim. Acta* 2001, *46*, 3245–3258. [CrossRef]
- 15. Sonoyama, N.; Sakata, T. Electrochemical continuous decomposition of chloroform and other volatile chlorinated hydrocarbons in water using a column type metal impregnated carbon fiber electrode. *Environ. Sci. Technol.* **1999**, *33*, 3438–3442. [CrossRef]
- 16. Brzózka, A.; Jeleń, A.; Brudzisz, A.M.; Marzec, M.M.; Sulka, G.D. Electrocatalytic reduction of chloroform at nanostructured silver electrodes. *Electrochim. Acta* 2017, 225, 574–583. [CrossRef]
- 17. Brudzisz, A.; Sulka, G.D.; Brzózka, A. A facile approach to silver nanowire array electrode preparation and its application for chloroform reduction. *Electrochim. Acta* 2020, *362*, 137110. [CrossRef]

- 18. Petro, A.G.C.; Thapa, B.; Karty, J.A.; Raghavachari, K.; Baker, L.A.; Peters, D.G. Direct electrochemical reduction of acetochlor at carbon and silver cathodes in dimethylformamide. *J. Electrochem. Soc.* **2020**, *167*, 155517. [CrossRef]
- 19. Huang, B.; Li, J.; Cao, X.; Zhu, Y.; Chen, W.; Lei, C. Electrochemical reduction of p-chloronitrobenzene (p-CNB) at silver cathode in dimethylformamide. *Electrochim. Acta* 2019, 296, 980–988. [CrossRef]
- 20. Rondinini, S.B.; Mussini, P.R.; Crippa, F.; Sello, G. Electrocatalytic potentialities of silver as a cathode for organic halide reductions. *Electrochem. Comm.* 2000, *2*, 491–496. [CrossRef]
- 21. Rondinini, S.; Vertova, A. Electrocatalysis on silver and silver alloys for dichloromethane and trichloromethane dehalogenation. *Electrochim. Acta* **2004**, *49*, 4035–4046. [CrossRef]
- 22. Peverly, A.A.; Peters, D.G. Electrochemical determination of trihalomethanes in water by means of stripping analysis. *Anal. Chem.* **2012**, *84*, 6110–6115. [CrossRef]
- Lugaresi, O.; Preales-Rondon, J.V.; Minguzzi, A.; Solla-Gullon, J.; Rondinini, S.; Feliu, J.M.; Sanchez-Sanchez, C.M. Rapid screening of silver nanoparticles for the catalytic degradation of chlorinated pollutants in water. *Appl. Catal. B.* 2015, 163, 554–563. [CrossRef]
- 24. Minguzzi, A.; Lugaresi, O.; Aricci, G.; Rondinini, S.; Vertova, A. Silver nanoparticles for hydrodehalogenation reduction: Evidence of a synergistic effect between catalyst and support. *Electrochem. Commun.* **2012**, *22*, 25–28. [CrossRef]
- 25. Ardizzone, S.; Cappelletti, G.; Mussini, P.R.; Rondinini, S.; Doubova, L.M. Adsorption competition effects in the electrocatalytic reduction of organic halides on silver. *J. Electroanal. Chem.* **2002**, *532*, 285–293. [CrossRef]
- 26. Durante, C.; Perazzolo, V.; Perini, L.; Favaro, M.; Granozzi, G.; Gennaro, A. Electrochemical activation of carbon–halogen bonds: Electrocatalysis at silver/copper nanoparticles. *Appl. Catal. B* 2014, 158–159, 286–295. [CrossRef]
- 27. Durante, C.; Isse, A.A.; Sandona, G.; Gennaro, A. Electrochemical hydrodehalogenation of polychloromethanes at silver and carbon electrodes. *Appl. Catal. B* 2009, *88*, 479–489. [CrossRef]
- 28. Rondinini, S.; Aricci, G.; Krpetic, Z.; Locatelli, C.; Minguzzi, A.; Porta, F.; Vertova, A. Electroreductions on silver-based electrocatalysts: The use of Ag nanoparticles for CHCl₃ to CH₄ conversion. *Fuel Cells* **2009**, *9*, 253–263. [CrossRef]
- Wang, A.; Huang, Y.-F.; Sur, U.K.; Wu, D.-Y.; Ren, B.; Rondinini, S.; Amatore, C.; Tian, Z.-Q. In situ identification of intermediates of benzyl chloride reduction at a silver electrode by SERS coupled with DFT calculations. *J. Am. Chem. Soc.* 2010, 132, 9534–9536. [CrossRef] [PubMed]
- 30. An, C.; Kuang, Y.; Fu, C.; Zeng, F.; Wang, W.; Zhu, H. Study on Ag–Pd bimetallic nanoparticles for electrocatalytic reduction of benzyl chloride. *Electrochem. Commun.* **2011**, *13*, 1413–1416. [CrossRef]
- Costentin, C.; Robert, M.; Savéant, J.M. Stepwise and concerted electron- transfer/bond breaking reactions. Solvent control of the existence of unstable πion radicals and of the activation barriers of their heterolytic cleavage. *J. Am. Chem. Soc.* 2004, 126, 16834–16840. [CrossRef] [PubMed]
- 32. Bard, A.J.; Faulkner, L.R. Electrochemical Methods: Fundamentals and Applications, 2nd ed.; Wiley: Hoboken, NJ, USA, 2001.
- Corrigan, D.A.; Evans, D.H. Cyclic voltammetric study of tertnitrobutane reduction in acetonitrile at mercury and platinum electrodes: Observation of a potential dependent electrochemical transfer coefficient and the influence of the electrolyte cation on the rate constant. J. Electroanal. Chem. Interf. Electrochem. 1980, 106, 287–304. [CrossRef]
- 34. Garreau, D.; Saveant, J.M.; Tessier, D. Potential dependence of the electrochemical transfer coefficient. An impedance study of the reduction of aromatic compounds. *J. Phys. Chem.* **1979**, *83*, 3003–3007. [CrossRef]
- 35. Amonette, J.E.; Jeffers, P.M.; Qafoku, O.; Russell, C.K.; Humphrys, D.R.; Wietsma, T.W.; Truex, M.J. Abiotic Degradation Rates for Carbon Tetrachloride and Chloroform: Final Report; U.S. Department of Energy: Germantown, MD, USA, 2012. [CrossRef]