

Brief Report

Modelisation of the Biomethane Accumulation in Anaerobic Co-Digestion of Whey and Sugarcane Molasse Mixtures

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Abstract: The biomethane accumulation of several combinations of whey and sugarcane molasses, inoculated with sludge from a treatment facility of one of the dairy enterprises of the Imbabura province in Ecuador, was assessed in the current experiment at a constant COD_0/VS_{in} ratio of 0.5. The whey/molasses (W:M) ratios for each treatment were (in % (m/m)) 0:100, 25:75, 50:50, 75:25, and 100:0, with a constant temperature of 37 °C and an initial pH adjustment of 7.5. Half a litre of total mixes was used for each treatment in duplicate. Six kinetic models were evaluated to account biomethane accumulation in anaerobic co-digestion processes in batch of whey and sugarcane molasses. Five of these have been tested by other researchers, and one was developed by modifying a first-order model to consider changes in the biomethane accumulation profile. This proposed model, along with the modified two-phase Gompertz model, resulted in the ones that were best able to adjust the experimental data, obtaining in all cases an $R^2 \geq 0.949$, indicating the accuracy of both models. In addition, the proposed here model has five parameters, one less than the modified two-phase Gompertz model, making it more straightforward and robust.

Keywords: anaerobic co-digestion; kinetics of biomethane accumulation; kinetic modelling; sugar-cane molasses; whey



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1. Introduction

The world's population has grown steadily over the past centuries, reaching 7.9 billion inhabitants today (Population Growth—Our World in Data). Alongside this, the demand for food and energy is also growing, and pressure on arable land and ecosystems is increasing [1]. With the increase in food production, there is also, logically, an increase in the waste generated. In contrast, the incessant increase in energy demand makes it essential to explore other renewable sources of energy to provide a future response to the depletion of traditional non-renewable sources, which will inexorably occur in the not-too-distant future.

In this sense, Ecuador, and specifically Zone 1 (formed by the provinces of Imbabura, Carchi, Esmeraldas, and Sucumbíos), is characterised by an active agricultural economy, which includes the daily production of more than 50% of Ecuador's milk production [2]. An essential part of this production is destined for cheese production, which generates significant quantities of cow-whey. In 2017, it was estimated that in the provinces of Imbabura and Carchi alone, more than 120 m³ of whey was generated daily [3]. About 70% of this whey is used for pig feed, but the rest must be treated in treatment plants due to its high polluting power [3].

On the other hand, one of the central sugar mills in the country is in the province of Imbabura, which generates sugar cane molasses as waste [4]. In this sense, agro-industrial waste could be studied as a possible source of raw material for biogas generation [5], an alternative to the circular economy for local industries.

Finally, modelling the complex fermentative processes that take place within anaerobic co-digestion [6–8], mediated by complex consortia of bacteria and yeast, including acetogenic and methanogenic bacteria and diverse sources of carbon and nitrogen, is of the utmost importance for the design of treatment processes for bioremediation and as a source of renewable biomethane from these agriculture or agro-industrial wastes [9–11].

The present work aims to evaluate the anaerobic digestion of cow's whey and sugarcane molasse, alone or formed by different mixtures, and to fit different kinetics models for the biomethane accumulation reported by other authors. A modified first-order model in two stages, not reported before, has also been evaluated.

2. Materials and Methods

2.1. Raw Materials Used

The whey used in this study came from the Ibarra branch of the company Floralp S.A. (Princesa Paccha 5-163, Caranqui, Ibarra, Imbabura, Ecuador, <https://floralp-sa.com> (accessed on 30 July 2023)). The company's waste treatment plant supplied the sludge. The sugar cane molasses was purchased on the local market from the *Ingenio Azucarero del Norte* (Panamericana Norte, km 25 vía Tulcán, Imbabura, Ecuador, <http://www.tababuela.com> (accessed on 30 July 2023)).

2.2. Physico-Chemical Characterisation

The total and volatile solids were determined according to the methods described in APHA 2540 B and APHA 2540 E, respectively [12]. For the determination of COD, the method described in APHA 5520 D was used [12].

A known volume was weighed to determine the density and pH, and the pH was measured in a conventional pH meter, previously adjusted between pH 4 and pH 10.

2.3. Experimental Procedure

The experimental units were prepared with a constant $COD_0/V_{S_{in}}$ ratio of 0.5, where COD_0 is the amount (in grams) of COD at the start of the anaerobic co-digestion (AcD) process and contains the COD inputs from both the inoculum (activated sludge) and the substrates (milk whey and sugarcane molasse) and $V_{S_{in}}$ represents the quantity (in grams) of volatile solids present in the inoculum (activate sludge). The working volumes of activated sludge, whey, and sugarcane molasse mixtures were between 66–452 mL. For treatments that did not reach 452 mL, a volume of sterile deionised water was added until all the volume (including sterile deionised water) reached 452 mL. After that, the bottle caps were worn, and neoprene caps were placed to connect the pipes for the exit of the gases produced by anaerobic digestion.

Five of the six vials were inoculated with 10 mL of activated sludge, while to the sixth, with a "50:50" mixture, was added 10 mL sterile water and served as a "negative control" of the process. This last bottle did not produce gases practically.

The flasks were placed in a thermostatically controlled bath, maintaining the temperature at 37 ± 1 °C. The experimental setup consists of six 500 mL flasks, where anaerobic digestion occurs discontinuously, which are connected to six 250 mL flasks, which act as a trap to capture the CO_2 produced. Each trap flask was connected to 250 mL test tubes, inverted, and filled with the same solution as the traps (0.375 M NaOH + phenolphthalein), allowing the measurement of methane gas by liquid displacement, as described by other authors [7,13,14]. All test tubes were placed in a cuvette, partially filled with the same alkaline solution (Figure 1).



Figure 1. The experimental facility used in the research. (A) Thermostated water-bath with recirculation. (B) Anaerobic digester flasks. (C) Bubbling traps for CO₂ capture. (D) Immersion cuvette. (E) Inverted test tubes for methane measurement.

Before this, the sludge was adapted for 15 days, with similar amounts as in the whey and cane molasses mixture evaluation being supplied every 2–3 days, and when an appreciable decrease in gas bubbling was observed. The sludge was inoculated into the reactors once the fizzing had ceased after the last addition of the substrate.

Experimental blocks with six variants in each (five treatments + “negative control”) maintained a constant ratio of COD₀/VS_{in} equal to 0.5 and were performed twice.

2.4. Kinetic Model for the Anaerobic Co-Digestion Mixes of Whey and Molasses

To kinetically characterise the process and model the generation of the primary metabolite, methane, the modified first-order in two-stage model (Equation (2)) was used together with other traditional models described by other authors [8], like the modified two-phase Gompertz model (Equation (3)), the multi-stage first-order model (Equation (4)), all conceived to describe the accumulative biomethane production obtained from complex substrates in which the diauxic growth has been observed.

Additionally, the three simplest models with three parameters each were also evaluated. The Fitzhugh model (Equation (5)), the transference-function model (Equation (6)), and Cone’s model (Equation (7)), despite their simplicity, in most cases, as will demonstrate further, adjust the experimental values accurately.

The model used here is based on the first-order model and was conceived for anaerobic digestions of substrate mixtures and where the phenomenon of diauxic is observed. For this, we should estimate t_{d_i} when a change in the methane accumulation profile is observed. Therefore, it is a **modified first-order model** for mixtures of many substrates and multi-stages are available.

$$G = \begin{cases} G_{m1} [1 - e^{-k_{01}t}] & \text{for } 0 \leq t < t_{d1} \\ G_{m2} [1 - e^{-k_{02}(t-t_{d1})}] & \text{for } t_{d1} \leq t < t_{d2} \\ G_{m3} [1 - e^{-k_{03}(t-t_{d2})}] & \text{for } t_{d2} \leq t \leq t_{d3} \\ \vdots & \\ G_{mn} [1 - e^{-k_{0n}(t-t_{dn-1})}] & \text{for } t_{dn-1} \leq t \leq t_f \end{cases} \quad (1)$$

For two stages, the above model will transform into

$$G = \begin{cases} G_{m1} [1 - e^{-k_{01}t}] & \text{for } 0 \leq t < t_{d1} \\ G_{m2} [1 - e^{-k_{02}(t-t_{d1})}] & \text{for } t_{d1} \leq t < t_f \end{cases} \quad (2)$$

where G_{m1} and G_{m2} are the maximum accumulated value of methane in each stage, in Nml CH₄; k_{01} and k_{02} are the first-order constants of the kinetics of biomethane accumulation, in d⁻¹; t_d and t_f are the times where diauxic phenomenon and end of the AcD process are observed in days.

The **two-phase modified Gompertz model** was suggested to represent the accumulation of biomethane in AcD processes, where the phenomenon of diauxic growth is observed [14]. This model is based on six parameters (G_{m1} , G_{m2} , R_{m1} , R_{m2} , λ_1 and λ_2) ($f = 6$).

$$G = G_{m1} e^{\{-e^{(R_{m1} \cdot e^{(\lambda_1 - t)/G_{m1}} + 1)}\}} + G_{m2} e^{\{-e^{(R_{m2} \cdot e^{(\lambda_2 - t)/G_{m2}} + 1)}\}} \quad (3)$$

The G_{m1} , G_{m2} , R_{m1} , R_{m2} , λ_1 and λ_2 parameters that can be obtained, like that of the rest of the models, experimentally from having experimental data relating to G vs. t , and employing a non-linear regression analysis, represent the maximum values of biomethane accumulation (G_{m1} and G_{m2} , in Nml CH₄), biomethane generation rate (R_{m1} and R_{m2} , in Nml CH₄/d) and the duration of the lag phase (λ_1 and λ_2 , in days), for each of the two phases of diauxic growth.

The **multi-stage first-order model** was conceived to model the production of biomethane in the presence of complex substrates formed by various sources of carbon, and their interactions, which lead to anaerobic digestion passing through different stages [15].

$$G = G_{m1} [1 - e^{-k_{01}t}] + G_{m2} [1 - e^{-k_{02}t}] + G_{m12} \left[1 - \frac{k_{02} \cdot e^{-k_{01}t}}{k_{02} - k_{01}} - \frac{k_{01} \cdot e^{-k_{02}t}}{k_{01} - k_{02}} \right] \quad (4)$$

It is a five-factor ($f = 5$) model (G_{m1} , G_{m2} , G_{m12} , k_{01} and k_{02}), where G_{m1} , G_{m2} and G_{m12} represent the maximum accumulation of biomethane (Nml CH₄) in the stages "1", "2" and during the interaction of both substrates ("12"), whereas k_{01} and k_{02} , represent the first-order kinetic constants in the states "1" and "2", respectively.

The last three models to be analysed are simple models formed by only three factors ($f = 3$).

The **Fitzhugh model**, initially developed to monitor the production of biomethane by the action of microorganisms present in livestock rumen [16,17], has also been successfully used by other researchers to co-digest food waste with activated sludge [18]. It is a simple three-factor model (G_m , k_0 and n , $f = 3$), where n represents the presence (if $n \geq 1$) or the absence (if $n < 1$) of a lag phase in the anaerobic process.

$$G = G_m [1 - e^{(-k_0 t)^n}] \quad (5)$$

G_m , k_0 and n ($f = 3$), represent the maximum accumulation of biomethane (in Nml CH₄), the first order kinetic constant (in d⁻¹), and a dimensional constant, related to the existence or not of a lag phase in the AcD process, respectively.

Additionally, the **transference function model** was also assessed (Equation (6)). In some cases, this model has been used to describe anaerobic digestion [19].

$$G = G_m [1 - e^{-(R_m/G_m) \cdot (t-\lambda)}] \quad (6)$$

Cone's empirical model, like others here, was initially developed to quantify methane production by the rumen microorganisms by metabolizing the grass [20].

$$G = \frac{G_m}{1 + (kt)^{-n}} \quad (7)$$

The values that need to be adjusted are G_m , k and n , representing the maximum cumulative amount of methane (in Nml CH₄), the first-order kinetic constant (d⁻¹), and a nondimensional number, respectively.

The experimental data ($N = 19$) for each mix were fitted by the least squares method and using the generalized reduced gradient (GRG) method [21], a nonlinear numerical optimization algorithm provided by the MS Office-365 Excel *Solver* tool.

2.5. Statistical Comparison of Models

Three known formulas will be used to judge whether the models represent the observed experimental data sufficiently well: the square regression coefficient (R^2 , Equation (8)), the normalized root mean square error (*NRMSE*, Equation (9)) and the corrected Akaike information criterion (AIC_C , Equation (10)) [15,18,22].

$$R^2 = 1 - \frac{\sum_{i=1}^{19} (G_{exp} - G_{model})_i^2}{\sum_{i=1}^{19} (G_{exp} - \bar{G}_{exp})_i^2} \quad (8)$$

And

$$NRMSE = \left(\frac{\sqrt{\frac{\sum_{i=1}^{19} (G_{exp} - G_{model})_i^2}{N}}}{G_{exp_{max}} - G_{exp_{min}}} \right) \times 100 \quad (9)$$

The correction that is introduced in the nondimensional Akaike Information Criterion (the last term on right in Equation (10)) [23] is recommended when the values obtained from AIC are small, and the number N of experimental data is not too large, as is the present case [24].

$$AIC_C = N \cdot \ln \left(\frac{\sum_{i=1}^{19} (G_{exp} - G_{model})_i^2}{N} \right) + 2f + \left(\frac{2f(f+1)}{N-f-1} \right) \quad (10)$$

where N represents the number of experimental points used to construct each model ($N = 19$), and f represents the number of factors the model possesses.

In this case, models with R^2 values closer to one and with lower *NRMSE* and AIC_C values are considered the most appropriate models to represent the observed experimental data.

3. Results and Discussion

For whey, sugarcane molasse and activated sludge, the values of volatile solids were 164.24, 726.94, and 935.4 g VS/L, respectively. The total solids were 237.70, 824.70, and 12.96 g TS/L, respectively, while the COD reached values of 0.64, 8.14, and 1.56 g COD/L in the same order. Additionally, the density was 0.98, 1.20, and 0.98 g/mL, while the initial pH that was had was of 6.90, 5.60, and 3.90, respectively.

According to the characterisation of the substrates in terms of volatile solids, total solids, and COD, it can be concluded that molasses has 4.4, 3.4 and 12.7 times more, respectively, than whey, suggesting a priori that molasses have a higher potential than whey for methane production.

The methane yield values are low, so it is suggested in further studies to raise the COD₀/VS_{in} ratio to values ≥ 1 .

The values of each two-mixture treatment were used to represent the models for accumulative methane production. The models were charted alongside the observed experimental data, separating the five- and six-parameter models (Figure 2 (a1–a5)) from the simpler three-factor models (Figure 2(b1–b5)).



Figure 2. Biomethane accumulation kinetics for different mixtures of whey and sugarcane molasse. On the left side (a1–a5) are represented the three models that have between 5 and 6 factors, being from (a1–a5), the ratio of whey and molasse (W:M, in % (m/m)): 0:100; 25:75; 50:50; 75:25; and 100:0, respectively. On the right side (b1–b5) the three most straightforward, three-factor models were charted, being of (b1–b5), the ratios (W:M, in % (m/m)): 0:100; 25:75; 50:50; 75:25, and 100:0, respectively.

It is somewhat disconcerting to note the observation of the 50:50 mixture, where a decrease in the accumulation of biomethane is observed (Figure 2(a3,b3)). The causes of the biomethane reabsorption in the liquid phase, which leads to decreased accumulated volume, should be investigated. This phenomenon may be related, although additional experiments would be needed to prove it, to the temperature fluctuations in the lab between day and night, which can reach ≥ 15 °C.

The R^2 , $NRMSE$, and AIC_C values of the five- and six-factor models exhibit better results, especially in those cases where changes in the methane accumulation profile are observed, and within these the modified first-order model and two-phase modified Gompertz model have shown higher performance than multi-stage first-order model (Table 1).

Table 1. Parameters of the kinetic models analysed in the present study and their respective statistical values of adjustment goodness.

Models	Parameters	Mix (W:M), % (m/m)				
		0:100	25:75	50:50	75:25	100:0
Modif. first-order ($f = 5$) Equation (2)	Gm_1 , Nml CH ₄	127.00	18.21	32.90	92.86	16.47
	k_{01} , d ⁻¹	0.95	1.55	1.37	1.25	0.86
	Gm_2 , Nml CH ₄	182.00	19.83	24.90	100.00	20.00
	k_{02} , d ⁻¹	3.92	32.18	4.72	12.53	12.65
	t_d , d	6.17	4.80	4.17	6.51	6.21
	R^2 , -	0.949	0.983	0.979	0.986	0.966
	$NRMSE$, %	5.59%	2.84%	3.17%	2.62%	4.48%
	AIC_C , -	102.75	-5.95	15.12	51.24	10.44
Modif. two-phase Gompertz ($f = 6$) Equation (3)	Gm_1 , Nml CH ₄	125.71	1.90	32.30	66.89	15.90
	Rm_1 , Nml CH ₄ ·d ⁻¹	59.39	35.66	28.95	159.11	10.80
	λ_1 , d	0.13	4.88	0.16	0.39	0.20
	Gm_2 , Nml CH ₄	56.31	17.94	-8.02	33.45	4.10
	Rm_2 , Nml CH ₄ ·d ⁻¹	251.23	20.22	-13.43	6.90	18.10
	λ_2 , d	6.16	0.18	4.20	0.00	6.14
	R^2 , -	0.990	0.983	0.993	0.995	0.975
	$NRMSE$, %	2.66%	2.81%	1.86%	1.48%	3.85%
	AIC_C , -	78.94	-2.02	-0.68	33.90	9.11
Multi-stage first-order ($f = 5$) Equation (4)	Gm_1 , Nml CH ₄	193.22	3.99	36.66	47.91	49.88
	k_{01} , d ⁻¹	0.16	0.15	1.20	0.42	0.02
	Gm_2 , Nml CH ₄	1.10	10.22	1013.16	25.60	6.38
	k_{02} , d ⁻¹	3.50	1.67	0.00	6.37	1.13
	Gm_{12} , Nml CH ₄	32.5	6.70	1144.77	27.5	6.4
	R^2 , -	0.926	0.968	0.921	0.996	0.932
		$NRMSE$, %	7.15%	3.84%	5.88%	1.46%
	AIC_C , -	112.11	5.56	38.63	28.88	22.91
Fitzhugh ($f = 3$) Equation (5)	Gm , Nml CH ₄	194.38	19.29	27.14	96.39	18.89
	k_0 , d ⁻¹	0.16	0.99	1.79	1.09	0.48
	n , -	1.65	1.28	1.19	0.99	1.21
	R^2 , -	0.915	0.960	0.667	0.971	0.898
		$NRMSE$, %	8.09%	4.33%	10.93%	3.83%
	AIC_C , -	109.80	3.05	55.18	58.63	24.25
Transference function ($f = 3$) Equation (6)	Gm , Nml CH ₄	194.40	19.30	27.14	96.39	18.89
	Rm , Nml CH ₄ ·d ⁻¹	52.63	24.47	58.03	104.96	10.99
	λ , d	0.00	0.00	0.00	0.00	0.00
	R^2 , -	0.915	0.960	0.672	0.971	0.898
		$NRMSE$, %	8.09%	4.33%	10.93%	3.83%
	AIC_C , -	109.80	3.05	55.18	58.63	24.25
Cone ($f = 3$) Equation (7)	Gm , Nml CH ₄	746.80	20.24	27.18	116.36	26.06
	k , d ⁻¹	0.02	2.00	1.40	1.76	0.48
	n , -	0.61	1.26	5.78	0.67	0.77
	R^2 , -	0.933	0.968	0.688	0.996	0.919
		$NRMSE$, %	6.91%	3.85%	10.72%	1.42%
	AIC_C , -	103.79	-1.40	54.44	20.93	19.32

Except for mixing (W:M) 50:50 in three-parameter models ($f = 3$), for the rest of the cases, all models showed good performance, with $R^2 > 0.89$.

In the present study, only the modified first-order ($f = 5$) and the two-phase modified Gompertz ($f = 6$) models were able to represent all the experimental values of the mixtures adequately and in which notable results were consistently obtained as demonstrated by the $R^2 \geq 0.949$, and values of $NRMSE \leq 5.59\%$, and $AIC_C \leq 102.75$, for all the mixes.

It should be noted that the modified two-phase Gompertz model has been used successfully to represent the accumulation of methane and its yield, in numerous studies of anaerobic digestion [25–28], both in the single substrate and in mixtures, where the phenomenon of diauxic growth has often been observed [29].

Both models quite accurately represent the experimental data obtained. The modified first order model, however, does so with one factor less, which means that, for equal values of R^2 , as is the case for mixing (W:M) 25:75 (see Figure 2(a2) and Table 1), the AIC_C value of the modified second order model is lower, and therefore better, than the one obtained for the two-phase modified Gompertz model.

4. Conclusions

In the study presented here, we managed to model the cumulative production of biomethane from various mixtures of whey and sugarcane molasse, using various models reported elsewhere. In addition, a new model was suggested that accurately predicts the observed experimental behaviour and is less complex than the best of the models used here. Further studies are in progress to evaluate and validate this model for other anaerobic co-digestion processes.

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