

Article

# **Estimation of Black Carbon Emissions from Dry Dipterocarp Forest Fires in Thailand**

Ubonwan Chaiyo<sup>1,2</sup> and Savitri Garivait<sup>1,2,\*</sup>

- <sup>1</sup> The Joint Graduate School of Energy and Environment, King Mongkut's University of Technology Thonburi, Bangkok 10140, Thailand; E-Mail: uchaiyo@hotmail.com
- <sup>2</sup> Center of Excellence on Energy Technology and Environment, S&T Postgraduate Education and Research Development Office (PERDO), Commission on Higher Education (CHE), Ministry of Education, Bangkok 10140, Thailand
- \* Author to whom correspondence should be addressed; E-Mail: savitri\_g@jgsee.kmutt.ac.th; Tel.: +662-872-9014-5; Fax: +662-872-9805.

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**Abstract:** This study focused on the estimation of black carbon emissions from dry dipterocarp forest fires in Thailand. Field experiments were set up at the natural forest, Mae Nam Phachi wildlife sanctuary, Ratchaburi Province, Thailand. The dead leaves were the main component consumed of the surface biomass with coverage higher than 90% in volume and mass. The dead leaves load was  $342 \pm 190 \text{ g} \cdot \text{m}^{-2}$  and followed by a little mass load of twig, 100 g·m<sup>-2</sup>. The chemical analysis of the dead leaves showed that the carbon content in the experimental biomass fuel was  $45.81 \pm 0.04\%$ . From the field experiments, it was found that  $88.38 \pm 2.02\%$  of the carbon input was converted to carbon released to the atmosphere, while less than 10% were left in the form of residues, and returned to soil. The quantity of dead leaves consumed to produce each gram of carbon released was  $2.40 \pm 0.02 \text{ g}_{dry biomass burned}$ . From the study, the emissions factor of carbon dioxide, carbon monoxide, particulate matter (PM<sub>2.5</sub>) and black carbon amounted 1329, 90, 26.19 and 2.83 g·kg<sup>-1</sup> dry biomass burned, respectively. In Thailand, the amount of black carbon emissions from dry dipterocarp forest fires amounted 17.43 tonnes·y<sup>-1</sup>.

Keywords: Carbon mass balance; carbon released; emissions; tropical deciduous forest

## 1. Introduction

Forest fires in dry dipterocarp forest (DDF) occur in Thailand during the dry season, from December to April, when trees shed their leaves, and consequently, the leaf litter can accumulate quickly, and then serve as perfect fuel for burnings to be set by local communities. Fires of tropical deciduous forests, including dry dipterocarp forest and mixed deciduous forest in Thailand are generally man-made. The reasons frequently recorded by forest fire control stations are (1) to facilitate the gathering of non-timber forest product, (2) to ease the access to forest for hunting, and (3) to help land clearing for crop cultivation [1]. The type of fires occurred in DDF is characterized as surface fire, since the fuel is composed of biomass present on the ground surface of the forest, including leaf litter, twig, grass, undergrowth, shrub, climber, and seedling. The leaf litter constitutes the major component of biomass fuel. Studies on fuel consumption indicated that about 95% of biomass fuels in DDF are consumed by surface fires [2].

Regarding impacts of fires on the ecosystem, it was shown that the frequency of fire can drastically modify the structure and composition of aboveground biomass and affect the carbon cycle. Actually, the more frequent is the burning the lower is the growth rate. It was observed that in the annually burned area, the increase of tree diameter and basal area was lower than that of biennially, which was lower than that of triennially burned, with only 0.237 cm·y<sup>-1</sup> and 0.0007 m·y<sup>-1</sup>, respectively. Concerning the survival to fire which depends on the base diameter, it was found that the seedlings with base diameter less than 1 cm were completely dead after the fire passage [3]. It was revealed that fires directly affect biomass fuel characteristics, as well as underneath soil properties and nutrients dynamics [4]. Indeed, the heat from fires distresses insects on the ground or under the bark wood, and activates the natural regeneration and undergrowth development [3]. Although their adverse effects, it should be noted that fires play an important role in tropical deciduous forest management, because they enable to maintain its plant structure and ecosystem.

On the other hand, emissions from forest fires strongly affect air quality. They include air pollutants and climate forcers such as carbon monoxide (CO), methane (CH<sub>4</sub>), nitrous oxide (N<sub>2</sub>O), fine particulate matter (PM<sub>10</sub> and PM<sub>2.5</sub>), carbonaceous aerosols composed of black carbon (BC) and organic carbon (OC), *etc.* While these compounds adversely impact the local air quality and long-term climate change, documentations on qualitative and quantitative characteristics of the emissions from dry dipterocarp forest fires in Thailand are still very scarce or inexistent.

In this study, we investigated the emission factor of black carbon from dry dipterocarp forest fires, by burning the leaf litter, which constitutes the major biomass fuel type. The estimation of the carbon dioxide, carbon monoxide, aerosols and carbon content in the biomass fuel and burned residues are presented. The amount of carbon released to the atmosphere and emission factor estimation of black carbon based on the carbon mass balance method of the combustion reaction was then presented and discussed.

## 2. Materials and Methods

#### 2.1. Study Site Description

Mae Nam Phachi wildlife sanctuary is located in Ratchaburi province, about 215 km far from Bangkok, in the western region of Thailand. It is lied between 13°8'N to 13°27'N latitude and 99°10'E

to 99°25'E longitude, on a high mountain covering the Suan Pueng district, Khing Amphoe Ban Kha and Park Tho district, close to the border of Thailand and Myanmar, and the altitude for this area ranges from 200 to 1400 m above the sea level. This wildlife sanctuary has a total area of 489 km<sup>2</sup>, and the forest types included therein are composed of dry evergreen forest, dry dipterocarp forest, mixed deciduous forest (Figure 1) and pure stand bamboo forest.

In DDF, the research study of Chaiyo *et al.* [5] reported that the dominant species are *Shorea obtusa, S. siamensis, Lannea coromandelica* and *Dipterocarpus obtusifolius*. The tree density in steep slope and terrain slope is  $2350 \pm 354$  and  $1225 \pm 71$  individuals ha<sup>-1</sup>, respectively. The aboveground biomass in steep slope and terrain slope is 91.96 and 30.95 tonnes ha<sup>-1</sup>, respectively. The surface biomass is mainly composed of leaf litter, *i.e.*, with coverage higher than 90% in volume and mass, that can be consumed to release the pollutants into the atmosphere.



Figure 1. Location of Mae Nam Phachi wildlife sanctuary, Ratchaburi Province, Thailand.

In Ratchaburi Province, the average total rainfall and mean temperature during 2005 to 2009, were 959 to 1285 mm and 28 °C, respectively, as reported by the Ratchaburi meteorological station. The peak of temperature during dry seasons, especially in April, lied between 30.3 to 31.3 °C. On the other hand, December was the coolest month with temperature ranges from 24.5 to 26.9 °C. January corresponded to the driest month with a total rainfall lesser than 6.5 mm rainfall, while October constituted the wettest month with 117.6 to 441.5 mm of rainfall (Figure 2).

The experimental plots were set during dry season, *i.e.*, in February to March. Three experimental plots of 40-m  $\times$  40-m plots were set in DDF. DDF1 and DDF2 were located in steep slope (*i.e.*, slope ranging from 20 to 40%) and DDF3 in terrain slope, respectively. The prescribed fire was set in DDF1, and the overall burning occurred in steep slope. The surface biomass fuel was sampled from the three plots.



#### 2.2. Surface Biomass Sampling

The surface biomass in this study was composed of leaf litter, twig, seedling, and understory (*i.e.*, grass, climber and herb). The systematic random sampling method was used to sample surface biomass from field experiment. Four samples of surface biomass were collected from four sub-plot of  $1 \text{-m} \times 1 \text{-m}$ , as shown in Figure 3. For each one square meter sub-plot, the surface coverage of surface biomass was observed and estimated in percentage of coverage, the surface biomass bed height was measured, and the biomass fuel was cut and sampled (Figure 4). For each sampling plot, surface biomass samples were separated into fuel type including leaf litter, twig, seedling, and understory. Each category was weighed in the field to determine the wet weight. All samples were then carried to laboratory to be oven-dried at 80 °C for 24 hours or until constant weight. All samples were then weighed to determine the dry weight and moisture content.

## 2.3. Fire Experiment Set Up

In order to measure the fire characteristics (including fire spread rate and flame height), which led to emissions of compounds of interest, wood sticks of 1.5 meters tall with 1 meter height scale mark were placed every 5 meters on the left, central, and right lines of the plot as shown in Figure 3. The rate of fire spread was calculated using the measurement of the time taken by the flame to move between two sticks in each line. The flame height at each wood stick was recorded manually using a meter tape.



Figure 3. Fire experiment set-up.

Legends:  $\Box$  1-m × 1-m plot for pre-burnt fuel loads, • wooden stick for determining the rate of fire spread and flame height, the burning direction and FEB (Field Equipment Box) to move on the central line.

Figure 4. Surface biomass (a) and residues after burning (b) in a  $1-m \times 1-m$  sub-plot sampling.



The fire was set by igniting in-line the located upwind side of the main plot. The in-line ignition was used to minimize the edge effect that may affect the fire propagation. The measurements were recorded from the three rows of wood sticks, *i.e.*, left, central and right lines of the main plot, and not from the left and right line only to avoid perturbation of the measurement that may result from the edge effect.

## 2.4. After-Fire Residues Sampling

The residues after burning composed of ash and charred leaves were collected after the fire experiment from four sub-plots of  $1-m \times 1-m$  using simple random method (Figure 4). The same sample handling and treatment as for surface biomass were applied to the residue samples for dry matter, and moisture and carbon content determination.

## 2.5. Analysis of Carbon Content in Biomass and Residues after Burning

After oven-dried at 80 °C for at least 24 hours or until reaching a constant weight, the leaf litter samples were cut using the cutting mill (SM 2000 model, Retsch, Germany) and ground with the ball mill (PM<sub>10</sub> model, Retsch, Germany) until obtaining a powder with diameter lesser than 100  $\mu$ m. The ground samples were then analyzed for the carbon content by Organic Elemental Analyzer (Flash EA1112 NC-Soil model, ThermoFinnigan, Italy). The same analytical procedure was applied to residues after burning composed of a mixture of ash and charred leaves.

#### 2.6. Real-Time Measurement and Sampling of Emissions

The background concentration measurement was conducted by real-time measurement set-up at the center of the plot for 30 min before ignition started. The fire was ignited at noon and the emission measurement in the plume was performed until no visible smoke was observed. The weather conditions including relative humidity, temperature, wind speed and wind directions were monitored during the whole duration of the background and fire emission measurement using a portable weather station (La Cross Technology, France). The weather station was set at about 10 meters downwind from the prescribed plot.

## 2.6.1. Measurement of CO<sub>2</sub> and CO

Carbon dioxide (CO<sub>2</sub>) was analyzed using a non-dipersive infared (NDIR) detector. Carbon monoxide (CO) analysis was performed using chemical cell detection technique with a full scale resolution of 1000 ppm, repeatability of  $3\% \pm 2$  LSD (Least significant digits), and display resolution of 1 ppm. For CO<sub>2</sub>, a full scale resolution of 20,000 ppm with display resolution of  $\pm 300$  ppm was selected. The outflow connecting tubing was connected with the micro-aethalometer at a flow rate of 50 mL·min<sup>-1</sup>.

## 2.6.2. Measurement of PM<sub>2.5</sub>

DustTrak aerosol monitor (TSI Inc., USA), a real-time optical scattering measurement equipment, was used to measure the concentration of  $PM_{2.5}$  emitted from the burning. The principle of the

(1)

measurement consists in pumping the air or smoke through the chamber to be scattered by the infra-red diode laser beam. The aerosols flow through the inlet with selected size by pump. The light scattering laser of the photo-detector determines the mass concentration of aerosol drawn to the unit in a constant stream. The mass concentration measurement is based on a calibration using the ISO 12103-1, A1 dust (Arizona Road Dust) standard.

### 2.6.3. Sampling of Total Particulate Matter

Quartz filters of 47 mm diameter (Pall Life Science, USA) were preheated at 800 °C for 5 hours [6] then cooled down in desiccator for 24 hours, weighed and placed in desiccator before use to collect particulate matter of the emission plume. The quartz filter was placed in an open-faced stainless steel filter holder (Pall Life Sciences, USA) connected with an individual pump operating at a flow rate of  $5 \text{ L}\cdot\text{min}^{-1}$ .

#### 2.6.4. Measurement of Black Carbon

The micro-aethalometer (Magee Scientific, AE51 model, USA) uses the optical technique principle to measure the black carbon concentration at 880 nm wavelength. The equipment was set up for real-time measurement of the black carbon concentration at the flow rate of 50 mL $\cdot$ min<sup>-1</sup>.

## 2.7. Field Equipment Box and Tower

The field equipments for emission concentration measurement, including real-time monitoring of carbon dioxide, carbon monoxide (Quest Technologies, IAQ model AQ500Pro, USA), black carbon (Micro-aethalometer, Magee Scientific, USA), and  $PM_{2.5}$  (DustTrak, 8520 model, TSI, USA), and the batch sampling of total particulate matter, were set up in the field equipment box (FEB) at the top of a tower of 3 meter-height.

## 2.8. Estimation of Emission Factors of Carbonaceous Compound

### 2.8.1. Carbon Mass Balance Method

The carbon mass balance (CMB) method [7,8] was used to calculate the carbon oxidized by the combustion of biomass fuel as follows.

$$C_{input} = C_{output} \Leftrightarrow C_{biomass} = C_{ash} + C_{released} \tag{1}$$

where:  $C_{biomass}$  is the amount of carbon in the biomass consumed (gC·m<sup>-2</sup>),  $C_{ash}$  is the amount of carbon in residues after burned (gC·m<sup>-2</sup>),  $C_{released}$  is the amount of carbon released to the atmosphere (gC·m<sup>-2</sup>).

The carbon content of biomass fuel was used to estimate the amount of carbon in the ash and released to the atmosphere. The CMB method [7,8] was used to calculate the carbon oxidized related to the carbon in biomass before and after burning. The general equation to calculate the CMB in biomass combustion is illustrated in Equation (1).

In case of a complete combustion, the carbon released to the atmosphere would be equal to the carbon in the form of carbon dioxide  $(CO_2)$ . As forest fire is always an incomplete combustion, the

carbon released is composed of carbon dioxide (CO<sub>2</sub>) representing more than 90% of the total, carbon monoxide (CO) corresponding to minor component (*i.e.*, lesser than 5% of the total) [5,6] and trace carbonaceous compounds including gaseous species such as methane (CH<sub>4</sub>), non-methane volatile organic compounds (NMVOCs), *etc.*, and carbon-contained particulate matter, composed of black carbon (BC) and organic carbon (OC). In this study, the carbon released to the atmosphere was estimated from the carbon in residues after burned, which was then subtracted from carbon in biomass fuel consumed.

## 2.8.2. Carbon Emitted in the Form of CO<sub>2</sub> and CO, and Modified Combustion Efficiency (MCE)

The CMB method was used to estimate the emission factor of  $CO_2$  because carbon dioxide is the main form of carbon released to the atmosphere with a share higher than 90% of the total carbon released [9,10]. Emission information is obtained from carbon dioxide (CO<sub>2</sub>) and carbon monoxide (CO) emission ratio and emission factor, which relate the emission of a compound of interest and the total emission to the amount of fuel burned, respectively. The carbon mass in the form of CO<sub>2</sub> and CO were calculated by Equation (2) adapted from the equation developed by Neto *et al.* [11].

$$MC_i (gC \cdot m^{-3}) = \frac{[i] \times 10^6 \times M_i \times 10^3 \times C}{V_i \times 10^{-3} m^3 \times MW_i}$$
(2)

where: *i* is the concentration of gas *i* (ppmv),  $M_i$  is the mass of carbon emitted in the form of gas *i* in the smoke plume (g),  $V_i$  is the volume of gas *i* (24.45 L for 1 mole of gas at 25 °C and 1 atm),  $MW_i$  is the molecular mass of compound of gas *i* (g·mole<sup>-1</sup>), *C* is the molar mass of carbon (g·mole<sup>-1</sup>).

The emission factor of  $CO_2$  or CO can be estimated using Equation (3) adapted from Urbanski *et al.* [12] and Neto *et al.* [11].

$$EF_{i} (\mathbf{g} \cdot \mathbf{kg}^{-1}) = \frac{MC_{i} \times 1000 \times MW_{i}}{C \times C_{t}} \times C_{m}$$
(3)

where:  $C_m$  is the mass fraction of carbon released to dry biomass consumed, 1000 (g·kg<sup>-1</sup>) and  $C_t$  is the total mass of carbon released in the form of CO<sub>2</sub>, CO (mgC·m<sup>-3</sup>).

The modified combustion efficiency (MCE) is defined as the ratio of  $CO_2$  released to the total release of CO and  $CO_2$  [13]. In the open burning process, the MCE never reaches 100%. The MCE can be used to describe the relative contribution of flaming and smoldering phase of biomass burning.

$$MCE = \frac{\Delta CO_2}{\Delta CO_2 + \Delta CO}$$
(4)

## 2.8.3. Estimation of Emission Factor of Emitted Carbonaceous Aerosols

Black carbon (BC) and organic carbon (OC) are the solid carbon contained found generally in the fine fraction of atmospheric particles (*i.e.*,  $PM_{2.5}$ ) The OC and total carbon (TC) content of atmospheric particulate matter can be estimated by Equations (5) and (6). From Ward *et al.* [14] and Neto *et al.* [11], it was found that the mass of TC represents about 60% of the total mass of PM<sub>2.5</sub> emitted from biomass burning. The emission factor of particle species can be calculated as Equation (7).

$$\mathrm{EF}_{Ai} \left( \mathbf{g} \cdot \mathbf{k} \mathbf{g}^{-1} \right) = \frac{M_{Ai} \times 1000}{C_{i}} \times C_{m} \tag{7}$$

where:  $M_{Ai}$  is the emission concentrations of carbonaceous aerosols or particulate matter of species *i* (mgC·m<sup>-3</sup>),  $C_m$  is the mass fraction of carbon released to dry biomass consumed, 1000 (g·kg<sup>-1</sup>),  $C_t$  is the total mass of carbon released in form of CO<sub>2</sub>, CO and PM<sub>2.5</sub> (mgC·m<sup>-3</sup>).

## 2.9. Morphology of Carbonaceous Aerosols Using Scanning Electron Microscope

Particulates collected on quartz filter were conditioned in desiccator for 24 hours before analysis of the morphology by scanning electron microscopy with energy-dispersive X-ray spectroscopy (SEM-EDX) (JEOL, Model: ISM-S410LV, Japan). Each sampled filter was punched to obtain 10 mm diameter samples which were then gold coated before the SEM-EDX analysis. Micrographs were taken at 20 Kv for identification and characterization of aerosols.

#### 3. Results and Discussion

## 3.1. Surface Biomass Fuel and After-Fire Residues Characteristics

The depth of surface biomass varied among sites, depending on the vegetation structure and the repartitioning between living and dead fraction. The dead leaves and twig depths were about 5–10 cm and 2–6 cm, respectively. Overall 90% coverage of surface fuel was dead leaves, followed by twig. The dead leaves load in DDF1, DDF2 and DDF3 were 561, 212 and 253 g·m<sup>-2</sup>, respectively. The moisture content of dead leaves was 4.69, 6.55 and 7.11% in DDF plot number 1 to 3, respectively. The average of dead leaves load in four plots was  $342 \pm 190$  g·m<sup>-2</sup> and the twig load was lesser than 100 g·m<sup>-2</sup>. It was confirmed that the dead leaves constituted the main component of the surface biomass.

After burning, the amount of surface biomass consumed was  $271 \pm 138 \text{ g} \cdot \text{m}^{-2}$ . It was observed that twig kept the same shape as before burning. Therefore, dead leaves were the major fuel consumed. The residues after burning or fire included the charred leaves and ash. The mass of residues after burning was  $39 \pm 13 \text{ g} \cdot \text{m}^{-2}$ . It should be noted that it was difficult to separate charred leaves from ash for quantification of the mass load.

#### 3.2. Carbon Content in Surface Biomass and After-Fire Residues

The carbon content of dead leaves and small twig before burning was  $45.81 \pm 0.04\%$  and  $43.49 \pm 0.09\%$ , respectively. The carbon content in the residues was  $35.0 \pm 0.06\%$ . The carbon content in the dead leaves and residues after burning are reported in Table 1.

From Table 1, it was found that the carbon released to the atmosphere from the DDF fires represented  $88.38 \pm 2.02\%$  of the initial dead leaves carbon content. The results showed that each gram of carbon released to the atmosphere corresponded to  $2.40 \pm 0.02$  g of dead leaves consumed by fires in DDF.

	Carbon (gC·m <sup>-2</sup> )	
Dead leaves consumed	After-fire Residues	Released to atmosphere
$124.83 \pm 63.36$	$13.75 \pm 4.57$	$111.07 \pm 58.88$

**Table 1.** Carbon content in the dead leaves consumed, after-fire residues and released to the atmosphere.

Carbon in after-fire residues would be left on the ground surface and returned to forest soil to enhance its carbon stock in the long term. The percentage of carbon in the residues was  $11.25 \pm 2.11\%$  of the total carbon in the surface biomass fuel consumed. It was shown that the amount of carbon returned to the ecosystem in the form of a mixture of ash and charred leaves in this forest type is  $53.85 \pm 9.67 \text{ gC} \cdot \text{kg}^{-1}_{\text{dry biomass.}}$ 

## 3.3. Carbon Released to the Atmosphere

In this study, the carbon released to the atmosphere resulted from dead leaves burning was  $88.38 \pm 2.02\%$  the initial dead leaves fuel carbon content. This indicates that each gram of carbon released was resulted from about  $2.40 \pm 0.02$  g of dry dead leaves, which is in good agreement with the result obtained by Nelson from laboratory fire experiments [13]. It is also of the same order of magnitude as the laboratory experimental result found by Chaiyo *et al.* [15] which stated that each gram of carbon released was produced from the combustion of DDF dead leaves was  $2.27 \pm 0.01$  g.

The assessment of the amount of carbon released to the atmosphere from fires in Mae Nam Phachi wildlife sanctuary during 2007 to 2010 indicated that the average of carbon to the atmosphere was  $6.04 \pm 3.88$  tonnes C·y<sup>-1</sup>.

## 3.4. Modified Combustion Efficiency (MCE) and Emission of Carbonaceous Compound

The amount of biomass burned in DDF1 was 430.41 g. Overall, the combustion in DDF plot was flaming dominant, indicating that the combustion completeness should be higher than 0.90 [16], and MCE should range between 0.70–0.99. The MCE found in this study was  $0.93 \pm 0.05$ .

#### 3.5. Estimated Emission Factors of CO<sub>2</sub>, CO and Carbonaceous Related Compounds

The total amount of carbon emitted from DDF was calculated from the biomass consumed for each gram of carbon emitted. The fraction of  $PM_{2.5}$  which includes BC and OC was estimated from the total carbon associated to  $PM_{2.5}$ . The total carbon (TC) was assumed to represent 60% of the total mass of  $PM_{2.5}$  based on Ward *et al.* [17]. The percentage of carbon associated to  $CO_2$  never exceeded 90% of the total mass of carbon released (Table 2).

**Table 2.** Mass of carbon associated with measured pollutants, mass of biomass consumed, and % carbon associated to  $CO_2$  released from leaf litter burning.

			Mass of Biomass					
Plot	m <sub>C-CO2</sub>	m <sub>C-CO</sub>	CO <b>m</b> <sub>C-BC</sub> <b>m</b> <sub>C-PM<sub>2.5</sub> <b>m</b><sub>C-OC</sub></sub>		m <sub>C-total</sub>	Consumed (g·gC <sup>-1</sup> )	%C <sub>CO2</sub>	
DDF1	196.75	20.92	0.42	8.53	8.11	226.20	2.40	86.98

The emission factors reported in Table 3 for CO<sub>2</sub>, CO, PM<sub>2.5</sub> and BC were calculated from the real-time measurement data from field experiment. Emission concentrations of OC was deduced from total organic carbon (TC) associated in PM<sub>2.5</sub>. In the fire experimental plot, the emission factors of BC and OC were 0.77 and 14.94 g·kg<sup>-1</sup><sub>dry biomass burned</sub>, respectively.

**Table 3.** Emission factors of measured compound emitted from dry dipterocarp forest determined using.

DI /	Emission Factor (g·kg <sup>-1</sup> dry biomass burned)					
Plot	EF <sub>CO</sub> ,	EF <sub>CO</sub>	EF <sub>BC</sub>	EF <sub>PM25</sub>	EFoc	
DDF1	1329	90	0.77	26.19	14.94	

The comparison of emission factors from this study with other research works is displayed in Table 4. It was shown that the overall emission of  $CO_2$  from others was higher compared to this study. However, the ratios of CO and  $CO_2$  are of the same range. The BC emission factor from this study was found higher than from the same forest zone such as tropical and extra-tropical forest fires, while the ratio of CO and  $CO_2$  was similar, indicating a quite equivalent fire behavior.

F	EF(g·kg <sup>-1</sup> drv biomass burned)							
Forest Type	CO <sub>2</sub>	CO	BC	PM <sub>2.5</sub>	TPM	OC	ĸei.	
Dry dipterocarp forest	1329	90	0.77	26.19	-	14.94	This study	
Savanna and grassland	$1613\pm95$	$65 \pm 20$	$0.48\pm0.18$	$5.4 \pm 1.5$	$8.3 \pm 3.2$	$3.4 \pm 1.4$	[10]	
Tropical forest	$1580\pm90$	$104 \pm 20$	$0.66\pm0.31$	$9.1 \pm 1.5$	6.5-10.5	$5.2 \pm 1.5$	[10]	
Extra tropical forest	$1569 \pm 131$	$107 \pm 37$	$0.56\pm0.19$	$13 \pm 7.0$	$17.6 \pm 6.4$	8.6–9.7	[10]	
Amazonian forest					-		[11]	
Canisters	1599	111.3		4.84	-	-		
RT FASS *	1617	137.6				-		
Primary tropical forest	1601	107.8		14.8			[18]	
Pasture maintenance fire	1506	152.4		18.7			[18]	
Tropical forest	1643	93	0.52	9.1		4.71	[19]	
Extra tropical forest	1509	122	0.56	15.0			[19]	

**Table 4.** Comparison of emission factors determined in this study with other research works on emissions from tropical forest fires.

Notes: \* RT-FASS means Real-Time Fire-Atmosphere Sampling System.

Table 5 reported the emission fluxes of BC and OC estimated for DDF fires in Thailand as documented in other studies using the emission factors of BC and OC determined in this study. The estimation may be affected by some uncertainty because the amount of residues after burning and their carbon content were not available, and so the combustion completeness. The BC and OC emission fluxes, expressed in  $g \cdot m^{-2}$ , found for DDF located in Kanchanaburi, HKK and Phu Kradueng are similar to that in MNP, *i.e.*, our study site.

The emission factors of BC and OC obtained in this study were used to estimate the emissions of BC and OC from DDF fires in Thailand during 2002 to 2006. It was found that the emission of BC and OC from DDF amounted 17.43 and 336.83 tonnes  $y^{-1}$ , respectively.

Location Study	Fuel Load	Fuel Consumed	<b>Emission Fl</b>	uxes (g·m <sup>-2</sup> )	D. (
	$(g \cdot m^{-2})$	$(g \cdot m^{-2})$	BC	OC	Reference
Kanchanaburi	440	300	0.23	4.48	[20]
HKK *	500	450	0.35	6.72	[21]
НКК	810	320	0.25	4.78	[21]
НКК	110	530	0.41	7.92	[22]
Phu Kradueng	470	290	0.22	4.33	[23]
MNP **	342	271	0.21	4.05	This study

**Table 5.** Comparison of emission fluxes of BC and OC from dry dipterocarp forest fires in this site with the other research studies.

Notes: \* HKK means Huay Kha Khaeng Wildlife Sanctuary, Uthai Thani Province; \*\* MNP means Mae Nam Phachi Wildlife Sanctuary, Ratchaburi Province.

## 3.6. Morphology of Carbonaceous Aerosols

The SEM micrographs of TPM showed various morphological types (Figure 5). The micrographs from the burning of DDF indicated 8 shapes of carbonaceous aerosols. The elemental composition analysis of each micrograph found the presence of carbon (C), oxygen (O), aluminum (Al), chlorine (Cl), potassium (K), iron (Fe) and calcium (Ca). DA micrograph shows the dominant shape, followed by DB, DH and DG micrographs, respectively. The DB and DH micrographs display similar shapes of branched clusters of soot particles as obtained by Pachauri *et al.* [24].

# 3.7. Contribution to Global Warming of Emissions from DDF Fires

Table 6 presents the emissions from annual DDF fires in Thailand in CO<sub>2</sub> equivalent, using the Global Warming Potential (GWP) of CO<sub>2</sub>, CO, BC and OC, the main carbon components measured in this study.

**Figure 5.** SEM micrographs of particulate matter emitted from dry dipterocarp forest fires in Thailand.





Figure 5. Cont.



Figure 5. Cont.



Figure 5. Cont.



**Table 6.** Contribution to global warming of emissions from dry dipterocarp forest fires in Thailand.

Species -	G	WP	Emissions	$\mathbf{Tonnes} \cdot \mathbf{CO}_{2eq} \cdot \mathbf{y}^{-1})$		
	20 y *	100 y *	(tonnes·y <sup>−1</sup> )	20 у	100 y	
$CO_2$	1	1	29,964	+29,964	+29,964	
CO	6	2	2027	+12,165	+4055	
BC	1600	460	17	+27,886	+8017	
OC	-240	-69	337	-80,840	-23,242	

Notes: \* [25], (+) means warming and (-) means cooling.

The GWP metric is used to determine the relative radiation effect of a given emission mass of a species compared to  $CO_2$  integrated over a chosen time horizon. Generally, 20 year and 100 year time horizons are used to calculate the GWP of a given compound. In addition, BC and OC are usually estimated based on 20 year time horizon, because of their shorter lifetime compared to traditional greenhouse gas. It was found that on the basis of 20 year time horizon in tonnes  $CO_{2eq}$  per year unit, the emission of BC by dry dipterocarp forest fires can induce similar warming effect as the co-emitted  $CO_2$ : 27,886 tonnes  $CO_{2eq} \cdot y^{-1} vs$ . 29,964 tonnes  $CO_{2eq} \cdot y^{-1}$  (Table 6).

#### 4. Conclusions

The surface biomass is composed of leaf litter and small twig. Dead leaves represent more than 90% of the total surface biomass whether weight or ground surface coverage basis, and hence constitute the main fuel for combustion. Another fraction that may influence the biomass combustion process investigated in this study are small twigs, but at a level much smaller than dead leaves. Based on the field experiment measurement data, it was estimated that the amount of carbon released by DDF fires in Thailand during 2002 to 2006 was 9241.46 tonnes  $\cdot y^{-1}$ , and about 3.37 tonnes  $\cdot y^{-1}$  of carbon was returned to soil. Also, it was shown that Chiangmai represents the province where more than 50% of

the total carbon released from DDF fires in 2009 followed by Lampoon, Khon Kaen, and Karlasin, respectively.

The morphology of carbon contained particulate matter emitted from DDF fires investigated in the study included 8 shapes. Moreover, it was observed that the color of the particles collected from DDF fires were grey to black, indicating high presence of BC. It was found that BC represented about 8.25% of the total mass of PM<sub>2.5</sub> emitted from DDF fires. Using the emission fluxes of BC and carbon released obtained in this study, it was found that BC constitutes about  $1.89 \times 10^{-3}$  (~ 0.19%) of the carbon released to the atmosphere from DDF fires.

Using the 20 year and 100 year time horizon GWP of BC, this study showed that for 20 year time horizon, BC emitted from DDF in Thailand may induce the same global warming effect as co-emitted CO<sub>2</sub>.

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#### **Author Contributions**

Both authors contributed to the research and field experiment design of this study, and to the elaboration of this manuscript. Ubonwan Chaiyo performed the data collection and analysis, and prepared the first draft of this paper. Savitri Garivait supervised the overall research work, provided guidelines for the write up of the manuscript, and contributed to its editing and finalization.

#### **Conflicts of Interest**

The authors declare no conflict of interest.

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