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Energy Balances and Greenhouse Gas Emissions of Agriculture in the Shihezi Oasis of China

Zhengang Yan 1,2,3, Fuqin Hou 4 and Fujiang Hou 1,2,*

- State Key Laboratory of Grassland Agro-Ecosystems, Key Laboratory of Grassland Livestock Industry Innovation, Ministry of Agriculture, Lanzhou 730020, Gansu, China; yanzhg@gsau.edu.cn
- College of Pastoral Agriculture Science and Technology, Lanzhou University, Lanzhou 730000, Gansu, China
- College of Information and Science Technology, Gansu Agricultural University, Lanzhou 730070, Gansu, China
- ⁴ Xinjiang Western Animal Husbandry Limited Company, Shihezi 832000, Xinjiang, China; houfqxj@163.com
- * Correspondence: cyhoufj@lzu.edu.cn

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Abstract: The objective of this study was to evaluate the difference of crop and livestock products regarding energy balances, greenhouse gas (GHG) emissions, carbon economic efficiency, and water use efficiency using a life cycle assessment (LCA) methodology on farms in three sub-oases within the Shihezi Oasis of China. The three sub-oases were selected within the Gobi Desert, at Shizongchang (SZC), Xiayedi (XYD), and Mosuowan (MSW), to represent the various local oasis types: i. Oasis; ii. overlapping oasis-desert; and iii. Gobi oasis. The results indicated that crop production in XYD Oasis had higher energy balances (221.47 GJ/ha), and a net energy ratio (5.39), than in the other two oases (p < 0.01). The production of 1 kg CW of sheep in XYD Oasis resulted in significantly higher energy balances (18.31 MJ/kg CW), and an energy ratio (2.21), than in the other two oases (p < 0.01). The water use efficiency of crop production in the SZC Oasis was lower than that of the XYD and MSW oases (p < 0.05). Alfalfa production generated the lowest CO₂-eq emissions (8.09 Mg CO₂-eq/ha. year) and had the highest water use efficiency (45.82 MJ/m³). Alfalfa (1.18 ¥/kg CO₂-eq) and maize $(1.14 \text{ Y/kg CO}_2\text{-eq})$ had a higher carbon economic efficiency than other crops (p < 0.01). The main sources of GHG emissions for crop production were fertilizer and irrigation. The structural equation modelling (SEM) of agricultural systems in the Shihezi Oasis showed that the livestock category significantly influenced the economic income, energy, and carbon balances.

Keywords: Shihezi Oasis; agricultural production systems; life cycle assessment; energy balances; greenhouse gas emissions

1. Introduction

The typical Shihezi meta-ecosystem of mountains, oases, and desert in northwest China consists of Tianshan Mountain, Shihezi Oasis, and the Gurbantunggut Desert. The areas of the mountain, oasis, and desert in this ecosystem are 2541 km², 7681 km², and 10,996 km², respectively [1]. The Manasi River is the lifeblood of the Shihezi Oasis. It originates from Tianshan Mountain and runs dry in the Gurbantunggut Desert. Agricultural production in the Shihezi Oasis is located in three sub-oases: Shizongchang (SZC), Xiayedi (XYD), and Mosuowan (MSW). More than 95% of people, farm produce, and energy production are concentrated in these sub-oases. At present, many problems such as secondary salinization of cropland, desertification, chemical pollution, and climate change have seriously threatened the sustainability of oasis agriculture through a lack of knowledge and understanding of oasis agricultural management [2]. There exists a close relationship between agricultural production and energy use [3]. Agricultural production requires high human-applied energy inputs, of which a large proportion are imported (i.e., fertilizer, diesel, electricity, pesticide,

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etc.), the remaining being produced on farms as bio-energy (i.e., seed, manure, and animate energy provided by living plants in nature). Crop yields can be improved (i.e., at economic maximum) by increasing the net energy (outputs/inputs) of crops [4].

Since 1950, the average rate of increase in global temperatures has been 0.17 °C per decade due to agricultural practices [5]. Agriculture has been considered as a major contributor to GHG emissions [6]. Crop yields in developing countries are projected to decrease due to global climate change [7]. As elsewhere, there are enormous challenges of alleviating greenhouse gas (GHG) emissions from agricultural production due to the heterogeneous nature and biophysical complexity of farming systems in China [8].

Characterizing the energy balances and GHG emissions of agricultural practices in the Shihezi Oasis offer key information to mitigate carbon emissions and ensure food security [6]. Energy consumption and GHG emissions involve on-farm and off-farm inputs, which are carbon-based operations and products [8]. In this study, the system boundary and scope of calculating energy balances and GHG emissions only included farm agricultural production practices using the life cycle assessment (LCA) methodology. The objectives of this study were to calculate the differences of energy balances and GHG emissions from various oasis ecotypes (SZC vs. XYD vs. MSW) in the Shihezi Oasis associated with producing per unit of crop and livestock products using the LCA technique. Moreover, finding the difference in energy balance and GHG emissions between six crops and water use efficiency of crop production in the Shihezi Oasis was the other aim of this study.

2. Materials and Methods

2.1. Study Site

Shihezi Oasis, which is divided into the three sub-oases, the SZC, XYD, and MSW oases, is located in the center of the northern foothills of Tianshan Mountain in XinJiang, China (84°58′–86°64′ E, 43°26′–45°20′ N, Figures 1 and S1). Agricultural production in the Shihezi Oasis is divided geographically into three contrasting systems (Figures 1 and S1):

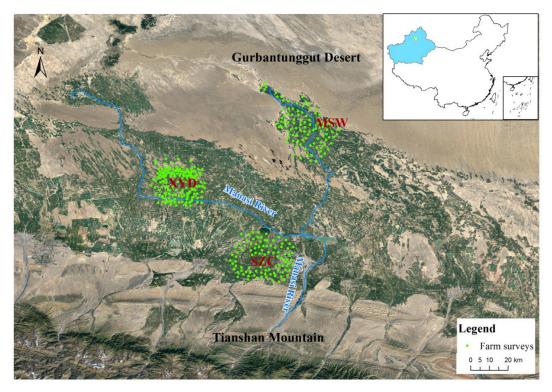


Figure 1. Satellite map of the study site in Xinjiang, China.

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(a) SZC sub-oasis, with a total area of 37.58×10^3 ha, of which 20.5×10^3 ha is cultivated, sits on an inclined piedmont plain. The Manasi River winds its way through the northeast of the sub-oasis; other surface and groundwater sources are abundant, and in a rural zone of Shihezi City, a freshwater spring overflow is used for flood (border dyke) irrigation;

- (b) XYD sub-oasis, with a total area of 300×10^3 ha, of which 61×10^3 ha is cultivated, lies where the Manasi River downstream oasis system and Gurban Tunggut Desert overlap;
- (c) MSW sub-oasis, with a total area of 137×10^3 ha, of which 44×10^3 ha is cultivated, lies where the Manasi River allows the sub-oasis to cut through the Gurban Tunggut Desert for a distance of 60 km.

2.2. Data Collection

Data were collected from official records, published literature, and farm survey data. The farm surveys were carried out from 2015 to 2016 with a selection of a random 354 farmers in this study (Tables S1 and S2). Inputs of crop production comprised of labor, seed, fertilizer, pesticide, fuel and electricity consumption, plastic film, and the life and working hours of farm machines. Outputs of crop production accounted as crop products including grain, straw, and roots. Inputs of livestock production were comprised of labor, feed (intake, source, and processing), and fuel and electricity consumption. Outputs of livestock production only included livestock products (carcass weight, milk, and wool). To quantify the GHG emissions from livestock production, data from livestock production included categories, livestock numbers, age, feed sources, and feed usage. Structural equation modelling (SEM) can evaluate direct and indirect effects between variables by expected statistical relationships. An SEM analysis program named AMOS 19.0 (IBM Corporation, New York, USA) was used to evaluate the direct and indirect effects of predictor variables on net income, energy balances, and carbon balances. A good fit for that model is $0 \le \chi^2/\mathrm{df} \le 2$ and 0.05 [9].

2.3. Goal and Scope Definition

Based on the ISO standard, the overall goal and scope of this study was to evaluate energy balances and GHG emissions on farms [10]. The specific aim was to quantify the sustainability impacts associated with energy balances and GHG emissions per ha of cropland and kg of livestock products between three sub-oases in Shihezi of China. The target audience included the farm and its stakeholders, consumers, and the public.

2.4. Functional Unit

The functional unit was defined for the purposes of this study, with GJ/ha for carbon balances of crop production and MJ/kg of carcass weight (CW) and milk for carbon balances of livestock. Based on information from the global warming potential for a 100-year period, the data of CH₄ and N₂O emissions were converted into CO₂ equivalents, with 34 for CH₄ and 298 for N₂O. The units of GHG emissions were kg CO₂-eq/ha for cropland, kg CO₂-eq/kg for crop products, and kg CO₂-eq/kg CW and milk for livestock products.

2.5. Calculation of Energy Balances

The human-applied energy inputs of crop production are calculated as Equation (1).

$$EI_{crop} = CI_l \times EC_l + CI_s \times EC_s + CI_f \times EC_f + CI_p \times EC_p + CI_{ie} \times EC_{ie} + CI_{pm} \times EC_{pm} + CI_{dc} \times EC_{dc} + CI_{md} \times EC_{md}$$

$$(1)$$

where EI_{crop} is the unit area of total energy inputs for crop production (MJ/ha), CI is the unit area of energy input for crop production, and EC is the energy coefficient (Table 1). In regards to indexing, l, s, f, p, le, pm, de, and md are inputs of human labor (h), seed (kg), fertilizers (kg), pesticides (kg), electricity

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consumption for irrigation (kwh), and plastic films (kg), respectively. The energy outputs for each category of crop including the grain, straw, and roots are calculated as Equation (2).

$$EO_{crop} = CY_{grain} \times EC_{grain} + CY_{straw} \times EC_{straw} + CY_{root} \times EC_{root}$$
 (2)

where EO_{crop} is the unit area of total energy outputs from crop products (MJ/ha) and CY is the unit area of yield. In regards to indexing, grain, straw, and root are grain (kg), straw (kg), and roots (kg) of crop products, respectively.

Table 1. Factors used for the calculation of energy inputs and energy outputs.

Energy Factors of Ag	ricultural Production In	puts	
	Wheat (spring)	17.9	[11]
	Maize	104.65	[12]
Seed	Cotton	22.024	[13]
(MJ/kg seed)	Alfalfa	108.82	[11]
	Grape	15.16	[13]
	Tomato	16.33	[14]
Fertilizer	N	78.1	[12]
(MJ/kg fertilizer)	P	17.4	[12]
(WIJ/Kg Teltilizel)	K	13.7	[12]
Farmyard manure (MJ/kg manure)	Animal manure	14.63	[12]
Pesticide	Herbicides	278	[12]
(MJ/kg pesticide)	Insecticides	233	[12]
(MI)/NE pesticide)	Fungicides	121	[12]
Mulch (MJ/kg mulch)	Plastic mulch	51.9	[6]
Fuel (MJ/kg fuel)	Diesel	47.78	[6]
Electricity (MJ/kwh electricity)	Electricity	12	
Electricity (WIJ/KWII electricity)	for irrigation	12	
Transportation (MJ/kg truck)	Truck	8.8	[11]
Maintenance of machinery (MJ/kg tractor)	Tractor	5.21	[15]
Human labor (MJ/h)	Male	0.68	[12]
	Female	0.52	[12]
Forage feed	Wheat hay	15.05	[16]
(MJ/kg feed)	Maize hay	15.22	[16]
(WIJ/Kg ICCU)	Alfalfa hay	18.8	[16]
Concentrate feed	Maize	18.26	[16]
(MJ/kg feed)	Soybean	18.83	[16]
	Wheat husk	13.72	[16]
Energy Factors o	f Agricultural Products		
	Wheat (spring)	12.56	[16]
	Maize	18.26	[16]
Grain (MJ/kg grain)	Cotton	22.024	[16]
	Grape	2.341	[16]
	Tomato	1.258	[13]
	Wheat (spring)	15.05	[16]
Hay (MJ/kg hay)	Maize	15.22	[16]
110y (111)/NG 110y)	Alfalfa	18.8	[16]
	Cotton	17.37	[16]
	Lamb	12.877	[13]
Livestock products (MJ/kg product)	Beef	13.88	[13]
Livestock products (1913/kg product)	Milk	2.889	[13]
	Wool	23.41	[11]

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The energy inputs of livestock production are calculated as Equation (3).

$$EI_{livestock} = \left(\sum_{j=1}^{m} \left(LI_{feed,j} \times EC_{feed,j}\right) + LI_{drug} \times EC_{drug} + LI_{labor} \times EC_{labor} + LI_{elec} \times EC_{elec} + LI_{coal} \times EC_{coal}\right) / CW_{livestock}$$
(3)

where $EI_{livestock}$ is the unit carcass (milk) weight of total energy inputs for livestock production (MJ/kg CW and milk), LI is the unit livestock of energy input for livestock production, and $CW_{livestock}$ is the unit livestock of carcass weight (i.e., the weight of meat with bone except for fur, viscera, head, hooves, and blood). In regards to indexing, f_{led} , f_{labor} , f_{lec} , f_{labor} , f_{lec} , f_{load} , f_{labor} , f_{load} , f_{labor} , f_{load} , f_{labor} , f_{load} , f_{load

The energy outputs of livestock products are calculated as Equation (4).

$$EO_{livestock} = (LY_{carcass} \times EC_{carcass} + LY_{milk} \times EC_{milk} + LY_{wool} \times EC_{wool}) / CW_{livestock}$$
(4)

where $EO_{livestock}$ is the unit carcass (milk) weight of total energy outputs from livestock products (MJ/kg CW and milk) and LY is the unit livestock of yield. In regards to indexing, carcass, milk, and wool are carcass weight (kg), milk (kg), and wool of livestock (kg), respectively.

The energy indices of balances and ratios are calculated as follows:

$$EB_{crop\&livestock} = EO_{crop\&livestock} - EI_{crop\&livestock}$$
 (5)

$$NER_{crop\&livestock} = \frac{EO_{crop\&livestock}}{EI_{crop\&livestock}}$$
(6)

where $EB_{crop\&livestock}$ represents the energy balances of crops (MJ/ha) or livestock (MJ/CW and milk). $NER_{crop\&livestock}$ represents the net energy ratio (output/input) of crop or livestock production. $EO_{crop\&livestock}$ and $EI_{crop\&livestock}$ represent the same parameters of the above equation, respectively.

2.6. Calculation of GHG Emissions

The GHG emissions for each category of crop production input are calculated as Equation (7).

$$CE_{crop} = CI_s \times EF_s + CI_f \times EF_f + CI_p \times EF_p + CI_{ie} \times EF_{ie} + CI_{pm} \times EF_{pm} + AI_{dc} \times EF_{dc} + AI_{md} \times EF_{md} + SOIL_{res}$$

$$(7)$$

where CE_{crop} is the unit area of GHG emissions (kg CO₂-eq/ha) and EF is the emission factor (Table 2). $SOIL_{res}$ is the unit area of GHG emissions from soil respiration (kg CO₂-eq/ha) using the following, Equation (8) [17].

$$SOIL_{res} = 1.55 \times e^{0.031 \times T} \times \frac{P \times SOC}{(P + 0.68) \times (P + 2.23)}$$
 (8)

where T, P, and SOC are the annual mean temperature (°C), the annual rainfall (m), and soil organic carbon between a 0 and 20 cm depth (kg C.m⁻²), respectively [10].

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Table 2.	Factors	used for	the ca	alculation	of	GHG emissions.

Item	Sub-Item	Factors	References				
Emission Factors of GHG for Agricultural Production							
	Wheat (spring)	0.477	[18]				
	Maize	3.85	[19]				
Seed ¹	Cotton	2.383	[18]				
$(kg CO_2-eq/kg seed)$	Alfalfa	9.643	[18]				
	Grape	2.35	[2]				
	Tomato	1.63	[20]				
	N	6.38	[21]				
Fertilizer	P	0.733	[22]				
(kg CO ₂ -eq/kg fertilizer)	K	0.55	[22]				
(Rg CO2 cq/Rg Tertifizer)	Soil emissions CO ₂ after N application	0.633	[11]				
	Soil emissions N ₂ O after N application	6.205	[23]				
Pesticide	Herbicides	23.1	[5]				
(kg CO ₂ -eq/kg pesticide)	Insecticides	18.7	[5]				
	Fungicides	13.933	[24]				
Mulch (kg CO ₂ -eq/kg mulch)	Plastic mulch	18.993	[6]				
Electricity (tCO ₂ -eq/kwh electricity)	Electricity for irrigation	0.917	[25]				
Fuel (kg CO ₂ -eq/L fuel)	Diesel	2.629	[6]				
	Tractor 7810	14.07	[26]				
	Tractor 55/60	0.49	[26]				
Tractor depreciation	Tractor 1002/1202	1.32	[26]				
(kg CO ₂ -eq/year)	Tractor 250	0.16	[26]				
	Harvester 1200	0.66	[26]				
	Harvester 154	1.34	[26]				
Food myo cossing	Maize	0.0102	[27]				
Feed processing (kg CO ₂ -eq/kg feed)	Soybean	0.1013	[27]				
(kg CO ₂ -eq/kg leed)	Wheat	0.0319	[27]				
CH ₄ emissions from enteric	Sheep	125	[11]				
fermentation	Beef cattle	1175	[11]				
(kg CO_2 -eq/head/year)	Dairy cattle	1525	[11]				
CH ₄ emissions from	Sheep	2.75	[11]				
manure management	Beef cattle	25	[11]				
(kg CO_2 -eq/head/year)	Dairy cattle	250	[11]				
N ₂ O emissions from	Sheep	62.3	[11]				
manure management	Beef cattle	120.4	[11]				
kg CO ₂ -eq/head/year)	Dairy cattle	106.7	[11]				

 $^{^{1}}$ Including seed production, cleaning, and packaging; 2 Including exhaled carbon dioxide of adult labor.

Carbon stocks of crop production are calculated using Equation (9) [28].

$$CS_{crop} = CS_{grain} + CS_{stem} + CS_{root} (9)$$

where CS_{crop} , CS_{gain} , CS_{stem} , and CS_{root} represent the unit area of carbon values accumulated in crops (kg CO_2 -eq/ha), grain (kg CO_2 -eq/ha), stems (kg CO_2 -eq/ha), and roots (kg CO_2 -eq/ha), respectively. Carbon balances of crop production are calculated as Equation (10).

$$CB_{crop} = CS_{crop} - CE_{crop} \tag{10}$$

where CB_{crop} represents the unit area of carbon balances (kg CO_2 -eq/ha).

The yearly GHG emissions from each livestock species are calculated using Equation (11).

$$CE_{livestock} = (TCO_{2feed} + TCO_{2drug} + TCO_{2elec} + TCO_{2coal} + TCH_{4enteric} + TCH_{4manure} + TN_2O_{manure})/CW_{livestock}$$

$$(11)$$

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where $CE_{livestock}$ is the unit carcass (milk) weight of GHG emissions from livestock (kg CO₂-eq/kg CW and milk), TCO_2 is the emissions of carbon dioxide, TCH_4 is the emissions of methane, and TN_2O is the emissions of nitrous oxide. In regards to indexing, enteric and enteric are ruminant enteric fermentation (kg CO₂-eq/kg CW and milk) and manure management (kg CO₂-eq/kg CW and milk), respectively.

The carbon stock and carbon balances of livestock production are calculated as the following, Equations (12) and (13) [29].

$$CS_{livestock} = (LW \times 0.2) / CW_{livestock}$$
 (12)

$$CB_{livestock} = CS_{livestock} - CE_{livestock}$$
 (13)

where $CS_{livestock}$ is the carbon stock of livestock (kg CO₂-eq/kg CW and milk) and LW is the live weight of livestock. $CB_{livestock}$ is the carbon balance (kg CO₂-eq/kg CW and milk).

2.7. Calculation of Carbon Economic Efficiency

Based on emissions per one kilogram of carbon dioxide equivalence from agricultural products, the total carbon economic efficiency associated with the mean market price of these products in 2015 and 2016 (Table S3) is calculated using Equation (14) [28].

$$CEE = \frac{YP \times PRICE}{CE} \tag{14}$$

where *CEE* is the carbon economic efficiency ($\frac{4}{kg}$ CO₂-eq), *YP* is the yield of agricultural products ($\frac{4}{kg}$), *PRICE* is the price of agricultural products ($\frac{4}{kg}$), and *CE* is the GHG emissions from agricultural production ($\frac{4}{kg}$ CO₂-eq).

2.8. Calculation of Water Use Efficiency

The analysis of energy balance per 1 cubic meter of water can find the relationship between irrigation and the net energy of crop production in the Shihezi Oasis. The water use efficiency of crop production is calculated using Equation (15):

$$WUE_{crop} = \frac{EB_{crop}}{WU_{crop}} \tag{15}$$

where WUE_{crop} , EB_{crop} , and WU_{crop} represent the water use efficiency of crops grown (MJ/m³), energy balances (MJ/ha), and the water use of crops grown (m³/ha), respectively.

2.9. Statistical Analyses

We used the statistical program named Genstat17.0 (VSNI Corporation, Hemel Hempstead, UK) to analyze differences in energy indices (energy balances and net energy ratio), carbon indices (GHG emissions, carbon stock, carbon balances, and carbon economic efficiency), water use, and water use efficiency using inline linear models.

3. Results

3.1. Yield and Water Use of Crop Production

The results of the yield and water use of crop production in three sub-oases of the Shihezi Oasis are presented in Table 3. The yields (kg DM/ha) of maize and cotton in SZC Oasis were significantly higher than in the other two oases (p < 0.05). The water use for wheat (spring) and maize production in SZC Oasis was higher than in the other two oases (p < 0.05).

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	SZC ¹	XYD ²	MSW ³	SED ⁴	<i>p</i> -Value
	Cro	op products (Mg DM/ha)		
Wheat (spring)	15.65 a	16.44 a	10.67 b	0.379	< 0.05
Maize	60.00 a	45.45 ^b	48.89 ^b	0.218	< 0.05
Cotton	13.42 a	15.68 ^b	15.53 ^c	0.157	< 0.05
Alfalfa	-	14.81	13.62	-	-
Grape	-	30.34	30.67	-	-
Tomato	8175	-	-	-	-
	V	Vater use (10	00 m ³ /ha)		
Wheat (spring)	6.77 ^a	5.59 ^c	6.39 ^b	0.176	< 0.001
Maize	9.06 a	7.59 ^b	7.623 ^b	0.244	< 0.05
Cotton	8.32 ^a	8.23 ^a	6.30 ^b	0.33	< 0.05
Alfalfa	-	4.17	5.06	-	-
Grape	-	7.39	5.96	-	-
Tomato	6.56	-	-	-	-
Farmland	7.67 ^a	7.53 a	6.501 ^b	0.202	< 0.05

Table 3. Yield and water use of crop production in the Shihezi sub-oases.

3.2. Energy Balances, Net Energy Ratio, and Water Use Efficiency of Agricultural Production

The results of energy balances, the net energy ratio, and water use efficiency in three sub-oases of Shihezi Oasis are presented in Table 4 and Table S4. For crop production, the output energy (308.69 GJ/ha), energy balances (221.47 GJ/ha), and net energy ratio (5.39) in the XYD Oasis were the highest among the three production systems; the corresponding values of MSW Oasis were higher than in the SZC Oasis, respectively (Table 4). However, the input energy (42.58 GJ/ha) in the MSW Oasis was lower than in the other two oases (p < 0.01). The net energy ratio (6.45) and energy balances (309.46 MJ/ha) of maize in the XYD Oasis were significantly higher than that of the SZC and MSW oases (Table 4). Alfalfa (12.71) and maize (6.27) had higher net energy ratios than other crops (p < 0.01) (Table S4). For livestock production, the production of 1 kg CW of sheep in XYD Oasis resulted in significantly higher energy balances than in the other two oases. Similarly, the net energy ratio (2.21) of sheep in the XYD Oasis was significantly higher than that of the SZC and MSW oases (Table 4).

Table 4. Energy balances, net energy ratio, and water use efficiency of agricultural production in the Shihezi sub-oases.

	SZC ³	XYD ⁴	MSW ⁵	SED ⁶	<i>p</i> -Value		
Energy input							
	C	rop production					
Wheat (spring)	63.19 a	60.01 a	55.39 b	3.239	< 0.05		
Maize	56.93 a	42.17 ^c	50.03 b	1.913	< 0.001		
Cotton	76.91 a	58.37 ^b	41.12 ^c	5.166	< 0.001		
Alfalfa	-	19.09	23.61	-	-		
Grape	-	35.58	37.54	-	-		
Tomato	49.55	-	-	-	-		
Farmland	69.33 ^a	58.47 ^b	42.58 ^c	4.416	< 0.001		
Livestock production (MJ/kg CW and milk)							
Sheep	14.96 a	10.96 ^b	11.97 ^b	1.225	< 0.05		
Beef cattle	38.84 a	30.32 ^b	30.01 ^b	5.230	< 0.05		
Dairy cattle	1.44 ^a	1.30 ^b	1.32 ^b	0.040	< 0.05		
Energy output (GJ/ha)							
Crop production (GJ/ha)							
Wheat (spring)	255.82 ^b	268.81 a	174.36 ^c	14.779	< 0.001		
Maize	313.63	360.01	350.24	7.048	0.317		
Cotton	99.72 ^b	116.49 a	115.41 a	2.709	<0.05		

 $^{^1}$ SZC: Shizongchang; 2 SZC: Xiayedi; 3 SZC: Mosuowan; 4 SED: Standard error of differences; a , b , and c represent means, with different letters in a row differing significantly (p < 0.05).

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Table 4. Cont.

	SZC ³	XYD ⁴	MSW ⁵	SED ⁶	<i>p</i> -Value
	SZC			SED '	<i>p</i> -value
Alfalfa	-	278.48	255.97	-	
Grape	· · -	101.28	101.9	-	-
Tomato	204.15	-	-	- <u>-</u>	-
Farmland	230.69 °	308.69 a	247.64 ^b	16.517	< 0.001
			IJ/kg CW and		
Sheep	29.23	29.26	29.29	0.010	0.35
Beef cattle	87.04	87.01	87.02	0.002	0.561
Dairy cattle	2.56	2.54	2.54	0.001	0.570
		nergy balance			
		rop production			
Wheat (spring)	195.63 b	205.8 a	118.97 ^c	13.699	< 0.001
Maize	257.07 ^b	309.46 a	269.97 ^b	8.519	< 0.05
Cotton	22.82 ^c	58.12 ^b	74.29 ^a	7.598	< 0.001
Alfalfa	-	255.39	236.36	-	-
Grape	-	265.7	264.36	-	-
Tomato	154.6	-	-	-	-
Farmland	161.12 ^c	221.47 ^a	207.95 ^b	11.452	< 0.001
		roduction (M	IJ/kg CW and	Milk)	
Sheep	17.11 ^b	18.31 ^a	17.18 ^b	1.035	< 0.05
Beef cattle	48.13	49.01	48.22	0.924	0.125
Dairy cattle	1.10	1.12	1.08	0.051	0.214
		Net energy	ratio ¹		
		Crop produ			
Wheat	4.05 ^a	4.27 ^a	3.15 ^b	0.185	< 0.05
Maize	6.09 ^c	6.45 ^a	6.20 ^b	0.225	< 0.001
Cotton	1.30 ^c	2.03 ^b	2.81 ^a	0.219	< 0.001
Alfalfa	-	12.06	13.05	-	-
Grape	-	8.47	8.04	-	-
Tomato	4.12	-	-	-	-
Farmland	3.32 ^c	5.39 a	4.63 b	0.625	< 0.001
		Livestock pro	duction		
Sheep	1.95 ^b	2.21 ^a	1.97 ^b	0.167	< 0.05
Beef cattle	2.29 ^b	2.43 a	2.31 ^b	0.256	< 0.05
Dairy cattle	1.78	1.80	1.77	0.004	0.524
•	,	Water use eff	iciency ²		
	Cı	op productio	on (MJ/m ³)		
Wheat (spring)	28.92 ^{a,b}	36.84 ^a	18.61 b	3.196	< 0.05
Maize	7.57 ^c	14.42 a	11.80 ^b	1.102	< 0.001
Cotton	2.73 ^c	6.99 ^b	11.79 a	1.321	< 0.001
Alfalfa	_	27.85	23.50	-	-
Grape	_	25.57	24.38	-	-
Tomato	23.59	-	_	-	-
Farmland	9.90 b	13.99 a	14.15 ^a	0.812	< 0.05
		2			

 $^{^{1}}$ net energy ratio = output energy/input energy; 2 water use efficiency: Water use based on energy balances; 3 SZC: Shizongchang; 4 SZC: Xiayedi; 5 SZC: Mosuowan; 6 SED: Standard error of differences; a , b , and c represent means, with different letters in a row differing significantly (p < 0.05).

3.3. GHG Emissions and Carbon Economic Efficiency of Agriculture Production

Carbon indices to evaluate agriculture production are presented in Table 5 and Table S4. Carbon balances ($-1.12 \text{ Mg CO}_2\text{-eq/ha}$) of crop production in SZC Oasis were lower than in the other two oases (p < 0.05) (Table 5). GHG emissions (17.72 Mg CO₂-eq/ha) from cotton production were significantly higher than for the other five crops (p < 0.01) (Table S4). Alfalfa had higher carbon stocks (23.76 Mg CO₂-eq/ha) and carbon balances (15.97 Mg CO₂-eq/ha) than other crops (p < 0.01) (Table S4). According to the price of agricultural products (Table S3) and carbon input for farm production, SZC Oasis had a lower carbon economic efficiency (0.34 \frac{1}{2}/kg CO₂-eq) of crop production than the other

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two oases (p < 0.05) (Table 5). The carbon economic efficiency of maize (1.14 \pm /kg CO₂-eq) and alfalfa (1.18 \pm /kg CO₂-eq) were significantly higher than other crops (p < 0.01) (Table S4).

Table 5. GHG emissions, carbon stocks, carbon balances, and the carbon economic efficiency of agricultural production in the Shihezi sub-oases.

	SZC ²	XYD ³	MSW ⁴	SED ⁵	<i>p</i> -Value				
		GHG emis							
			CO ₂ -eq/ha. ye	ear)					
Wheat (spring)	8.58 ^b	8.60 ^b	8.99 a	0.065	< 0.05				
Maize	12.45 ^a	12.12 ^b	12.09 ^b	0.057	< 0.05				
Cotton	17.72	17.75	17.69	0.137	0.981				
Alfalfa	-	8.09	8.10	-	-				
Grape	-	12.26	12.19	-	-				
Tomato	17.02	-	-	-	-				
Farmland	13.22	12.94	12.27	0.377	0.534				
Livestock production (kg CO ₂ -eq/kg CW and milk)									
Sheep	9.23 ^a	8.35 ^{a,b}	7.62 b	0.305	0.072				
Beef cattle	22.95 ^b	24.19 a	22.87 ^b	0.237	< 0.05				
Dairy cattle	0.67	0.70	0.71	0.013	0.639				
,		Carbon st	ocks						
	Crop pro	duction (Mg	CO ₂ -eq/ha ye	ar)					
Wheat (spring)	10.44 b	10.40 b	10.86 a	0.075	0.069				
Maize	23.52 a	22.84 b	22.83 b	0.107	< 0.05				
Cotton	13.13	13.10	13.11	0.101	0.995				
Alfalfa	-	23.83	23.74	-	-				
Grape	_	10.98	11.00	_	_				
Tomato	10.35	-	-	_	_				
Farmland	12.10 b	15.94 ^a	17.33 a	0.582	< 0.001				
			ارد. 0 ₂ -eq/kg CW و		\0.001				
Sheep	1.06 ^a	1.76 ^{a,b}	1.93 b	0.061	0.610				
Beef cattle	3.57	3.22	3.54	0.001	0.287				
Dairy cattle	0.10	0.11	0.11	0.003	0.593				
Dairy Cattic	0.10	Carbon bal		0.003	0.575				
	Crop pro		CO ₂ -eq/ha ye	ar)					
Wheat (spring)	1.85	1.80	1.87	0.012	0.160				
Maize	11.06 ^a	10.72 b	10.75 b	0.050	< 0.05				
Cotton	-4.59	-4.64	-4.56	0.037	0.601				
Alfalfa	-	15.74	15.63	-	-				
Grape	_	-1.28	-1.18	_	_				
Tomato	-6.68	-	-	_	_				
Farmland	-1.12 b	2.99 a	5.06 a	0.798	< 0.05				
			ان.00 2-eq/kg CW و		<0.03				
	–7.63 ^b	-6.60 ^{a,b}	–5.69 ^a	0.330	< 0.05				
Sheep Beef cattle	−7.03 −19.39 ^a	-20.97 b	−3.07 −19.32 ^a	0.298	< 0.05				
	-0.53	-20.97	-0.57	0.298	0.653				
Dairy cattle					0.633				
	Carbon eco		ency (¥/kg CO	₂ -eq)					
TATIL and	0.17	Crop produ		0.001	0.105				
Wheat	0.17	0.18	0.17	0.001	0.185				
Maize	1.16	1.17	1.14	0.016	0.802				
Cotton	0.70	0.72	0.70	0.015	0.872				
Alfalfa	-	1.18	1.17	-	-				
Grape	- 0.22	0.41	0.42	-	-				
Tomato	0.32	0.75.3	0.02.3	0.042	-0.001				
Farmland	0.34 ^b	0.75 ^a	0.82 a	0.042	< 0.001				
C1		Livestock pro		0.000	0.015				
Sheep	0.24 a	0.22 ^{a,b}	0.20 b	0.008	0.063				
Beef cattle	0.38 b	0.40 a	0.38 b	0.004	< 0.05				
Dairy cattle	0.17	0.18	0.18	0.609	0.524				

¹ net energy ratio = output energy/input energy; ² SZC: Shizongchang; ³ SZC: Xiayedi; ⁴ SZC: Mosuowan; ⁵ SED: Standard error of differences; ^a, ^b, and ^c represent means, with different letters in a row differing significantly (p < 0.05).

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3.4. Water Use Efficiency Based on Energy

Water use and the water use efficiency of crop production in the Shihezi Oasis are presented in Tables 3 and 4. Among the three sub-oases in the Shihezi Oasis, the water use of crop production in the MSW Oasis is lower than the corresponding value of the SZC and XYD oases (p < 0.05). However, the water use efficiency of crop production in the SZC Oasis is lower than that of the XYD and MSW oases (p < 0.05). The water use (4590 m³/ha) for alfalfa production in the Shihezi Oasis is the lowest, with the highest water use efficiency (45.82 MJ/m³) (Table S4).

3.5. Contribution of Carbon Emissions

To further illustrate the relationship between GHG emissions and other factors, Figure 2 shows the contribution of GHG emissions from inputs of crop production. The major GHG emission sources within the crop production system are fertilizer and irrigation.

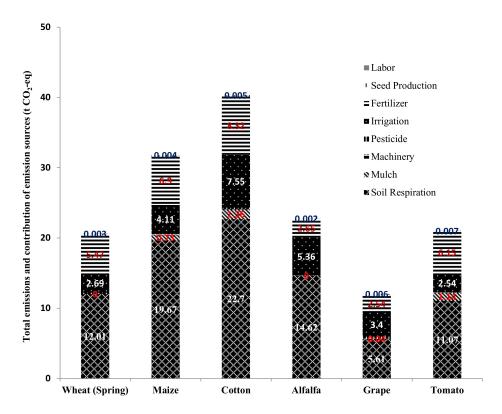


Figure 2. Total GHG emissions from crop production and the contribution of emission sources.

3.6. Effect Analysis of Structural Equation Modelling (SEM)

The effect analysis of SEM is presented in Table S5. The path modes show that the direct and total effects of the livestock breeding structure on predicted variables (farm net income, energy, and carbon balances) are much stronger than other dependent variables (Figure 3a, total effects = 0.647; Figure 3b, total effects = 0.898; Figure 3c, total effects = 0.991; Figure 3d, total effects = 0.980; Figure 3e, total effects = 0.898; and Figure 3f, total effects = 0.1.091). Similarly, the path modes show that the indirect effects of water use efficiency on economy income, energy, and carbon balances are much stronger than those of other variables (Figure 3a, indirect effects = 0.196; Figure 3b, indirect effects = 0.297; and Figure 3c, indirect effects = 0.430).

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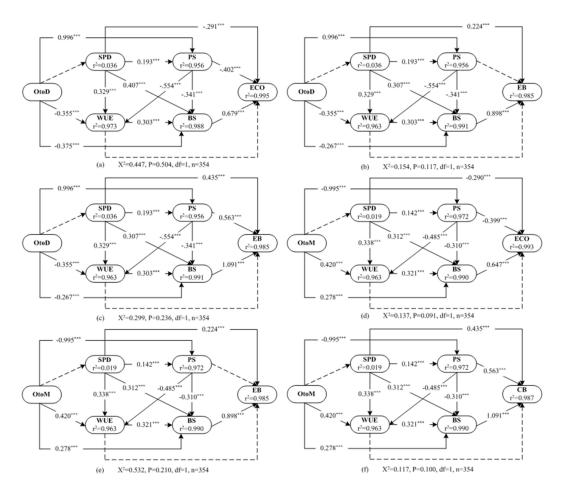


Figure 3. Path models showing the direct and indirect effects of predictor variables on farm net income, energy balances, and carbon balances. The path models with significant correlations are presented as solid lines. The values on solid lines represent standardized regression weights. Interrupted lines indicate no significant correlation between two variables. Black arrows indicate positive effects. For each endogenous variable, the relative amount of explained variance is given. Shading indicates the greatest positive direct effect, indirect effect, and total effect between dependent and independent variables. OtoD: The distance from the oasis to the desert (km); SPD: Soil particle diameter (μ m); PS: Planting structure (planting crop type); WUE: Water use efficiency (MJ/m³); BS: Livestock breeding structure (breeding livestock category); ECO: Net income (1000 ¥/ha); EB: Energy balance (GJ/ha); CB: The difference value of carbon stock minus GHG emissions from agricultural production inputs (Mg CO₂-eq/ha); and OtoM: The distance from the oasis to the desert (km). χ^2 : Chi-square. p: Probability level. df: Degrees of freedom. n: Sample size.

4. Discussion

4.1. Energy Balances and Net Energy Ratio

The present energy balances of agricultural production in the Shihezi Oasis are comparable to those published elsewhere in the world using the similar methodology of calculating the input and output energy values on farms. For example, the input energy of cotton production (58.80 GJ/ha) in the Shihezi Oasis is much higher than that (31.237 GJ/ha) in Iran [30]. This difference refers to the low energy input of weed controlling and harvesting operations using human labor instead of applying machinery in Iran. Nevertheless, Shihezi's input energy for maize production (49.71 GJ/ha) is close to that reported in Iran (50.458 GJ/ha) due to similar intensive and high input crop production systems [31]. However, our output energy including grain, straw, and roots of maize production (341.21 GJ/ha) is much higher than that only related to grain estimated in Iran (134.946 GJ/ha) [30].

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The net energy ratios of wheat and maize production in the research area are much higher than that (3.95 vs. 2.13, 6.27 vs. 2.67, respectively) in Iran [31,32]. The reason for the lower output energy and net energy ratio of crop production estimated in Iran is that the outputs included only grain. However, the net energy ratio of wheat production in this study is higher than that (3.95 vs. 1.59) in Pakistan due to less wheat yield as output energy [33]. The corresponding value of maize production in the research area is also higher than that (6.87 vs. 5.52) in Turkey due to less crop yields [34].

The flow of energy and matter is the driving force of agricultural development [35]. Agricultural production in the Shihezi Oasis is a high inputs and outputs system. With the rapid development of agriculture, high inorganic energy inputs can satisfy people's expectations of living standards, but at the cost of environmental sustainability. From data in this research study (cf. Table S1), a more sustainable approach for Shihezi Oasis farmers would be to increase both alfalfa acreage and sheep numbers, in tandem.

4.2. GHG Emissions from Agricultural Systems

Similarly, GHG emissions from agricultural systems in the Shihezi Oasis are comparable to those evaluated in other places. For example, GHG emissions for maize production (12.1 MgCO₂-eq/ha) in this study are similar to those reported in Iran (12.9 Mg CO₂-eq/ha) due to similar energy inputs of maize production [31]. As the effective value of economic output produced by the unit carbon input, the signals of carbon economic efficiency can explain the benefits of carbon cost to society. The present carbon economic efficiency for wheat production ($0.027/kg CO_2$ -eq) is higher than that ($0.023/kg CO_2$ -eq) in the USA [36]. The most significant reason for this difference is GHG emissions from wheat production, ranging from $0.14-0.38 CO_2$ -eq per kg of wheat produced on farms in the USA. Apart from this paper, there is no other related research on carbon balances in China which is significant enough to suggest adjusting the balance of crops with livestock production. As already stated, in the Shihezi Oasis, high inputs such as fertilizer, mulch, and machinery resulted in high outputs in crop production, but also in high GHG emissions. In addition, if GHG emissions are to be related to climate policy framework in the future, similar to legislation introduced in some other countries, it will be essential to know the impact of those polices on crop and livestock production costs.

4.3. Balance between Livestock and Forage Crops

There exists an imbalance between livestock and forage crop energy inputs vs. outputs in the Shihezi Oasis (Figure 4). Livestock numbers in the Shihezi Oasis are far greater than can be supported by the forage crop supply. For example, the current forage crop supply in XYD Oasis could not meet the feed demand of livestock, which had the largest number of livestock among the three sub-oases in the Shihezi Oasis. Livestock feedstuff prices in the Shihezi Oasis, whether bought or sold, are the same (Table S2), indicating that the cost of growing feedstuff "on farm" is less than the cost of buying it from outside. This is good reason for oasis forage maize and alfalfa acreage to be greatly increased, and thus bringing the demand and supply into balance.

Similarly, the revenue (13,976 \(\frac{1}{2}\)ha) of alfalfa production in 2014 was higher than that (10,275 \(\frac{1}{2}\)ha) of cotton production in the oasis of the Manasi River. The differences of net income per 100 sheep in 2014 were \(\frac{1}{2}\)29,492 between proposed crop-livestock production compared to the current practices of Nileke County of the Xinjiang Autonomous Region in China [37]. The revenue per hectare of maize production in Dingxi City of Gansu Province in China was much higher than that (\(\frac{1}{2}\)14,070 vs. \(\frac{1}{2}\)3315) of wheat production in 2018. The net income of integrated crop-livestock production in 2018 was 3.19 times that of intensive crop production [38].

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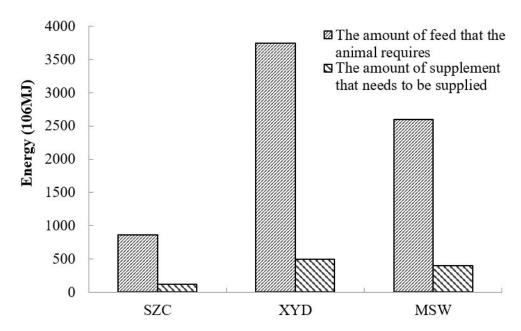


Figure 4. Feed demand compared to supplemental feed in the research site.

4.4. Uncertainty of Energy Balances and GHG Emissions Assessment

As a basic assessment of energy balances and GHG emissions from agricultural systems in the Shihezi Oasis, there still existed many factors of uncertainty. Firstly, there were differences in data of arable land area, livestock numbers, and inputs of agricultural production within each sub-oasis. Secondly, the statistical method of partial data in China's literature is different from that in developed countries. Thirdly, the coefficient for calculation of carbon balances and GHG emissions in the Shihezi Oasis were estimated using overseas data reported by Zeng et al. [39], and Cheng et al. [6]. In addition, the other factors such as energy consumption and GHG emissions from plastic mulch were only associated with plastic production. In this study, we only used the Tier 1 method of IPCC 2013 to evaluate GHG emissions from livestock production [11]. Nevertheless, the basic estimation of energy balances and GHG emissions in the Shihezi Oasis may offer fundamental information for the Chinese government to develop long-term agricultural policies on food security, energy conservation, and GHG emissions reduction in northwest China.

5. Conclusions

The models of energy balances, carbon balances, carbon economic efficiency, and water use efficiency developed in this study were used to calculate the differences of energy balances and GHG emissions from the three various local oasis ecotypes in the Shihezi Oasis. The evaluation indicated that crop production in the XYD Oasis had higher energy balances (221.47 GJ/ha) and an energy ratio (5.39) than in the SZC and MSW oases. The sheep production per kg of CW in the XYD Oasis had higher energy balances (18.31 MJ/kg CW) and an energy ratio (2.21) than in the other two oases. The water use efficiency of crop production in the SZC Oasis was lower than that of the other two oases (p < 0.05). These evaluation models in the present study were also used to calculate the differences of energy balances, GHG emissions, and carbon economic efficiency from local main crops in the Shihezi Oasis. We found one hectare of forage crops (i.e., alfalfa and maize) generated fewer emissions than any of the other four crops (wheat, cotton, grapes, and tomatoes). Alfalfa production in the Shihezi Oasis had the lowest water use (4594.0 m³/ha) and highest water use efficiency (45.82 MJ/m³) compared with all other five crops (wheat, maize, cotton, grapes, and tomatoes). The carbon economic efficiency of maize, relative to its market value, was significantly higher than that for each of the other five crops. Fertilizer and irrigation were the two main GHG emission sources from crop production in all

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three sub-oases together. Analysis of SEM showed that the livestock breeding structure and water use efficiency significantly influenced farm income, and energy and carbon balances in the Shihezi Oasis. The agro-system of Shihezi Oasis was dominated by crop production derived by high energy inputs (e.g., irrigation, fertilizer, and pesticide). The present production model of Shihezi Oasis will bring high risk to the environment and market. However, integrated crop-livestock production will increase energy use efficiency and decrease GHG emissions from the agriculture system in the Shihezi Oasis. A more sustainable approach for the agricultural development of Shihezi Oasis would be to increase both forage crop (i.e., alfalfa and maize) acreage and sheep numbers, in tandem.

Supplementary Materials: The following are available online at http://www.mdpi.com/2073-4433/11/8/781/s1, Figure S1: Longitudinal section of Shihezi study site to show Mountain-Oasis-Desert coupling ecological system (84°58′–86°64′ E), Table S1: Average data of climate, crop, and livestock of the three sub-oases in the Shihezi Oasis (2015–2016), Table S2: The structured questionnaire of farm survey, Table S3: Average market price of inputs and outputs for agricultural production (2015–2016), Table S4: Energy balances, carbon balances, carbon economic efficiency, water use, and water use efficiency of crop grown in the Shihezi Oasis, Table S5: The standardized direct, indirect and total effects between dependent variables and predict variables.

Author Contributions: F.H. (Fujiang Hou) and Z.Y. conceived and designed the experiments; F.H. (Fuqin Hou) and Z.Y. performed the experiments and farm survey; Z.Y. and F.H. (Fuqin Hou) analyzed the data; Z.Y. wrote the main manuscript text; and F.H. (Fujiang Hou) reviewed the manuscript. All authors have read and agreed to the published version of the manuscript.

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Conflicts of Interest: The authors declare no conflict of interest.

References

- 1. Ren, J.Z.; He, D.H.; Wang, N.; Zhu, X.Y.; Li, Z.Q. Models of coupling agro-grassland systems in desert-oasis region. *Acta Prataculturae Sin.* **1995**, *2*, 11–19. (In Chinese)
- 2. Kramer, K.J.; Moll, H.C.; Nonhebel, S. Total greenhouse gas emissions related to the Dutch crop production system. *Agric. Ecosyst. Environ.* **1999**, 72, 9–16. [CrossRef]
- 3. Ozkan, B.; Akcaoz, H.; Fert, C. Energy input–output analysis in Turkish agriculture. *Renew. Energy* **2004**, 29, 39–51. [CrossRef]
- 4. Khoshnevisan, B.; Rafiee, S.; Omid, M.; Mousazadeh, H.; Rajaeifar, M.A. Application of artificial neural networks for prediction of output energy and GHG emissions in potato production in Iran. *Agric. Syst.* **2014**, 123, 120–127. [CrossRef]
- 5. Lal, R. Carbon emission from farm operations. Environ. Int. 2004, 30, 981–990. [CrossRef] [PubMed]
- 6. Cheng, K.; Pan, G.; Smith, P.; Luo, T.; Li, L.; Zheng, J.; Zhang, X.; Han, X.; Yan, M. Carbon footprint of China's crop production—An estimation using agro-statistics data over 1993–2007. *Agric. Ecosyst. Environ.* **2011**, 142, 231–237. [CrossRef]
- 7. Broad, K.; Agrawala, S. The ethiopia food crisis-uses and limits of climate forecasts. *Science* **2000**, 289, 1693–1694. [CrossRef]
- 8. Marland, G.; West, T.; Schlamadinger, B.; Canella, L. Managing soil organic carbon in agriculture: The net effect on greenhouse gas emissions. *Tellus B* **2003**, 613–621. [CrossRef]
- 9. Yan, Z.G.; Li, W.; Yan, T.H.; Chang, S.H.; Hou, F.J. Evaluation of energy balances and greenhouse gas emissions from different agricultural production systems in Minqin Oasis, China. *PeerJ* **2019**, 7, e6890. [CrossRef]
- 10. International Organization for Standardization. *ISO14044: Environmental Management—Life Cycle Assessment—Requirements and Guidelines; ISO: Geneva, Switzerland, 2006; pp. 1–16.*
- 11. Intergovernmental Panel on Climate Change (IPCC). Climate Change 2014: Mitigation of Climate Change. Available online: http://www.ipcc.ch/report/ar5/wg3/ (accessed on 21 July 2020).
- 12. Wen, D.; Pimentel, D. Energy flow through an organic agroecosystem in China. *Agric. Ecosyst. Environ.* **1984**, 11, 145–160. [CrossRef]

Atmosphere **2020**, 11, 781 16 of 17

13. Pimentel, D. *Handbook of Energy Utilization in Agriculture*; Bender, M., Ed.; CRC Press: Boca Raton, FL, USA, 1980; pp. 155–161.

- 14. Huang, Z.W.; Yang, D.G.; Li, X.P.; Refu, K.T. Analysis on the energy flow of the farmer households and the characteristics of the eco-economic fractals in the middle and lower reaches of the Tarim river. *Arid Zone Res.* **2004**, *21*, 308–312. (In Chinese)
- 15. Lu, F.B. Energy flow in the agroecosystem of farming-livestock-fruit. *Ecol. Agric. Res.* **1994**, 2, 40–46. (In Chinese)
- 16. Steve, M. How much carbon dioxide does the human breathe out every day in the world? *China Sci. Tech. Panor. Mag.* **2005**, *6*, 176–177. (In Chinese)
- 17. Chen, S.T.; Huang, Y.; Zou, J.; Shen, Q.; Hu, Z.; Qin, Y.; Chen, H.; Pan, G. Modeling interannual variability of global soil respiration from climate and soil properties. *Agric. For. Meteorol.* **2010**, *150*, 590–605. [CrossRef]
- 18. West, T.O.; Marland, G.A. Synthesis of carbon sequestration, carbon emissions, and net carbon flux in agriculture: Comparing tillage practices in the United States. *Agric. Ecosyst. Environ.* **2002**, *91*, 217–232. [CrossRef]
- 19. Shi, L.G.; Chen, F.; Kong, F.L.; Fan, S.C. The carbon footprint of winter wheat-summer maize cropping pattern on north China plain. *China Popul. Resour. Environ.* **2011**, *21*, 93–98. (In Chinese)
- 20. Blook, H.; Kool, A.; Luske, B.; Ponsioen, T.; Scholten, J. Methodology for Assessing Carbon Footprints of Horticultural Products. Available online: https://www.researchgate.net/publication/285690888 (accessed on 21 July 2020).
- 21. Lu, F.; Wang, X.K.; Han, B. Assessment on the availability of nitrogen fertilization in improving carbon sequestration potential of China's cropland soil. *Chin. J. App. Ecol.* **2008**, *19*, 2239–2250. (In Chinese)
- 22. Dubey, A.; Lal, R. Carbon footprint and sustainability of agricultural production systems in Punjab, India, and Ohio, USA. *J. Crop Improv.* **2009**, *23*, 332–350. [CrossRef]
- 23. Adom, F.; Maes, A.; Workman, C.; Clayton-Nierderman, Z.; Thoma, G.; Shonnard, D. Regional carbon footprint analysis of dairy feeds for milk production in the USA. *Int. J. Life Cycle Assess.* **2012**, *17*, 520–534. [CrossRef]
- 24. Lal, R. Enhancing crop yields in the developing countries through restoration of the soil organic carbon pool in agricultural lands. *Land Degrad. Dev.* **2010**, *17*, 197–209. [CrossRef]
- 25. Wang, J.Q. *Agricultural Standards—Feeding Standard of Meat Producting Sheep and Goats (NY/T 816-2004)*; Ministry of Agriculture of China, Ed.; China Agriculture Press: Beijing, China, 2004; pp. 17–20.
- 26. Dyer, J.A.; Desjardins, R.L. Carbon dioxide emissions associated with the manufacturing of tractors and farm machinery in Canada. *Biosyst. Eng.* **2006**, *93*, 107–118. [CrossRef]
- 27. Meng, X.H.; Cheng, G.Q.; Zhang, J.B.; Wang, Y.B.; Zhou, H.C. Analyze on the spatialtemporal characteristics of GHG estimation of livestock's by life cycle assessment in China. *China Environ. Sci.* **2014**, *34*, 2167–2176. (In Chinese)
- 28. Shi, L.G.; Fan, S.G.; Kong, F.L.; Chen, F. Preliminary study on the carbon efficiency of main crops production in north China plain. *Acta Agron. Sin.* **2011**, *37*, 1485–1490. (In Chinese) [CrossRef]
- 29. Wu, C.C.; Gao, X.Y.; Hou, F.J. Carbon balance of household production system in the transition zone from the loess plateau to the Qinghai-Tibet Plateau, China. *Chin. J. Appl. Ecol.* **2017**, *28*, 3341–3350. (In Chinese)
- 30. Pishgarkomleh, S.H.; Sefeedpari, P.; Ghahderijani, M. Exploring energy consumption and CO₂ emission of cotton production in Iran. *J. Renew. Sustain. Energy* **2012**, *4*, 427–438.
- 31. Yousefi, M.; Damghani, A.M.; Khoramivafa, M. Energy consumption, greenhouse gas emissions and assessment of sustainability index in corn agroecosystems of Iran. *Sci. Total Environ.* **2014**, 493, 330–335. [CrossRef] [PubMed]
- 32. Khoshroo, A. Energy use pattern and greenhouse gas emission of wheat production: A case study in Iran. *Agric. Commun.* **2014**, *2*, 9–14. Available online: https://www.researchgate.net/publication/262105641 (accessed on 21 July 2020).
- 33. Abbas, A.; Yang, M.L.; Ahmad, R.; Yousaf, K.; Iqbal, T. Energy use efficiency in wheat production, a case study of Paunjab Pakistan. *Fresenius Environ. Bull.* **2017**, 26, 6773–6779. Available online: https://www.prt-parlar.de/download_feb_2017/ (accessed on 21 July 2020).
- 34. Baran, M.F.; Gokdogan, O. Comparison of energy use efficiency of different tillage methods on the secondary crop corn silage production. *Fresenius Environ. Bull.* **2016**, 25, 3808–3814. Available online: https://www.prt-parlar.de/download_afs_2016/ (accessed on 21 July 2020).

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35. Sere, C.; Steinfeld, H.; Groenewold, J. Food and Agricultural Organization of the United Nations (FAOUN). Available online: http://www.fao.org/WAIRDOCS/LEAD/X6101E/X6101E00.HTM (accessed on 21 July 2020).

- 36. Soussana, J.F.; Tallec, T.; Blanfort, V. Mitigating the greenhouse gas balance of ruminant production systems through carbon sequestration in grasslands. *Animal* **2010**, *4*, 334–350. [CrossRef] [PubMed]
- 37. Zhang, Y.H. The Research of Grassland Agricultural Development Model in Xinjiang. Ph.D. Thesis, Shihezi University, Xinjiang, China, 2015. (In Chinese)
- 38. Zhang, P.P. Comparison of Pastoral Agriculture and Traditional Agricultural Production Pattern in Dingxi City. Master's Thesis, Lanzhou University, Gansu, China, 2019. (In Chinese)
- 39. Zeng, X.F.; Zhao, S.W.; Li, X.X.; Li, T.; Liu, J. Main crops carbon footprint in pingluo county of the Ningxia Hui autonomous region. *Bull. Soil Water Conserv.* **2012**, *32*, 61–65. (In Chinese)



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