



Article

Atmospheric Monitoring of Methane in Beijing Using a Mobile Observatory

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Abstract: Cities have multiple fugitive emission sources of methane (CH₄) and policies adopted by China on replacing coal with natural gas in recent years can cause fine spatial heterogeneities at the range of kilometers within a city and also contribute to the CH₄ inventory. In this study, a mobile observatory was used to monitor the real-time CH₄ concentrations at fine spatial and temporal resolutions in Beijing, the most important pilot city of energy transition. Results showed that: several point sources, such as a liquefied natural gas (LNG) power plant which has not been included in the Chinese national greenhouse gas inventory yet, can be identified; the ratio "fingerprints" (CH₄:CO₂) for an LNG carrier, LNG filling station, and LNG power plant show a shape of "L"; for city observations, the distribution of CH₄ concentration, in the range of 1940–2370 ppbv, had small variations while that in the rural area had a much higher concentration gradient; significant correlations between CO2 and CH4 concentrations were found in the rural area but in the urban area there were no such significant correlations; a shape of "L" of CH₄:CO₂ ratios is obtained in the urban area in wintertime and it is assigned to fugitive emissions from LNG sources. This mobile measurement methodology is capable of monitoring point and non-point CH₄ sources in Beijing and the observation results could improve the CH₄ inventory and inform relevant policy-making on emission reduction in China.

Keywords: mobile measurement; methane; atmospheric concentration; source apportionment

1. Introduction

Methane (CH₄) contributes ~17% of the radiative forcing by the long-lived greenhouse gases (GHG) [1], with global warming potential (GWP) more than 25 times greater than that of CO_2 at a 100-yr time horizon [2]. Its mean annual increase rate witnessed no growth during 1999–2006 but has been increasing again since 2007 [1,3]. Increased emission from anthropogenic sources at mid-latitudes of the northern hemisphere has been considered as one possible main contributor of the recent increase [1,3], while some estimates indicated that China is the world's largest anthropogenic emitter of CH₄ [4]. Therefore, it is essential to quantify the anthropogenic sources of CH₄ emissions in China and further evaluate their contributions. Cities, especially in Northern China, are generally

important anthropogenic sources of CH_4 [5]. In megacities of China, the total natural gas (NG) consumption increased about seven times over the last 16 years (2000–2016) [6] with policies adopting the replacement of coal with gas launched by the Chinese government [5,7,8]. The increasing proportion of NG utilization together with increasing energy demands might lead to complex changes in the spatial and temporal pattern of atmospheric CH_4 concentrations, as well as the emission inventory [5,9,10]. To better understand the quantitative information, field monitoring at fine spatial and temporal scales in cities is necessary.

Presently, there are several monitoring tools for CH₄ emissions, such as satellite remote sensing [11-14], ground station observation [15,16], mobile observation by aircraft [14,17-19] or by car [20-24]. Satellite remote sensing technology is applied for observations mainly at the global or continental scales. Ground-based observation using stationary measurement towers is generally adopted by World Meteorological Organization (WMO) Global Atmosphere Watch (GAW) surface-based global or regional background stations [15,16]. It boasts higher precision than satellite remote sensing, but the high costs of establishment and maintenance have limited the development of an extensive network. Thus, it would be quite difficult to implement such a technology to resolve spatial heterogeneities in a city scale and to track moving sources [25]. In addition, aerial surveys can also offer intensive measurements for the distribution and quantification of sources and sinks of CH₄ at city, regional or terrestrial scales [19,26–30], but that kind of measurement is often restricted due to the strict air traffic control in China. A ground-based mobile observatory has more flexibility of use than aircraft, and it has been adopted by several CH₄ research projects [9,20-22,24,25,31,32]. The vehicles used include public transit, for example, light-rail platform [25], and ordinary cars [22,31]. Moreover, surveyed fugitive emission covered an extensive range of sources, including combustions [31,33], pipeline [21] and gas infrastructure leaks [22,23], NG developments [20], landfills [22], and oil producing regions [33]. Meanwhile, observed areas are mainly the urban areas [22,23,31,34]. Thus, the mobile ground-based observatory has been tested sufficiently with a wide range of applications to measure point and non-point sources of CH₄. However, although a qualitative measurement in Beijing for the purpose to verify the performances of their instrument was conducted [34], comprehensive CH₄ concentration of the entire city and the characteristic of different types of sources in Northern China by a mobile observatory has not been reported.

The goal of this work was to set up a ground-based mobile observatory and to measure the concentration of atmospheric CH_4 in Beijing with high precision, which has a population of ~19.6 million and has the largest demand of winter-heating services among cities in China [35]. Beijing is also suffering heavy air pollution [36,37] and taking efforts to replace coal by NG to improve air quality. Nearly all of the coal-fired boilers, including power generating boilers, have been changed into gas ones or air conditioners up to 2017. Studying the concentration of atmospheric CH_4 in Beijing could tremendously inform the CH_4 inventory and relevant policy-making on GHG emission reduction in China.

2. Experimental Design

2.1. Mobile Observatory for CH₄

The structure of the mobile observatory (MO) used in this work is illustrated in Figure 1. A multi-purpose vehicle (MPV, A800, Chang' an Automobile Group, Chongqing, China) with a sunroof was selected as the mobile platform. The components of this MO included electrical power, a high precision CH₄ analyzer, a global positioning system (GPS) receiver and a data acquisition device. The electric power was provided as 220V AC via an inverter-charger which was a 12 V, 3000 W, pure sine-wave inverter that served a number of different functions (HS-3KRH, Foshan Baykee New Energy Technology Incorporated Company, Foshan, China). The inverter-charger provided 220V AC from a 12V DC supply associated with a battery bank. The battery bank consisted of two lithium ion batteries (300 Ah each), wired in parallel and connected to the inverter-charger. The high precision

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CH₄ analyzed was based on a cavity ring-down spectrometer (CRDS) technique (G1301 and 2401, Picarro, Santa Clara, CA, USA). The measurement precision of G1301 was better than 0.7 ppbv for CH₄ (5 min average) and that of G2401 was 0.5 ppbv. Ambient air was collected through tubing (1/4" OD, ~2 m long, Synflex-1300, Aurora, CO, USA) from the inlet ~0.35 m over the sunroof, or ~2.2 m above the road surface. Particles were removed by a filter (0.2 µm, Ploycap 75, 0.2 PTFE, Buckinghamshire, UK) before it was assessed by the CRDS analyzer, with a measurement interval of approximately 3 s (G1302) and 1.2 s (G2401). Considering that the speed of the car was 40 to 80 kilometers per hour, the spatial resolution of this observation method was approximately 10-20 m. Additional technical information for the CRDS analyzer can be found in previous studies [38]. Before and after each observation mission, the CRDS analyzer was calibrated by standard gases which traced to WMO X2004A scale for CH₄, X2007 scale for CO₂, and WMO-2014A scale for CO to correct the drift of the analyzer. The geographic location of the vehicle platform was obtained by a GPS receiver (BS-70DU, Shenzhen Beitian Communication Co., Ltd., Shenzhen, China), mounted inside the car with an accuracy of approximately 2 m. Data from the CRDS analyzer and GPS receiver were recorded each 1.2 s by a laptop computer (Tsinghua Tongfang Co., Ltd., Beiing, China, T45PRO-Gcc-21053). The data system time was synchronized with the GPS system time so that the data file and GPS files could be merged in real time.

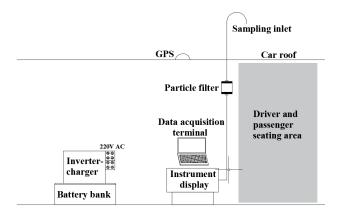


Figure 1. Schematic of the mobile observatory. The viewing angle is from the passenger side of the car.

2.2. Measurement Routes and Strategies

Table 1 shows the measurement strategies in Beijing. Firstly, five kinds of point sources, including a liquefied natural gas (LNG) carrier, an LNG station, a residential area, an LNG power plant and a refuse landfill, were surveyed to elucidate their concentration characteristics and ratio "fingerprints" (CH₄:CO₂). Secondly, the city was surveyed to determine their overall concentration characteristics and to identify the major CH₄ emitters. Outward from the city center of Beijing, there are five ring roads that circle the city, the 2nd (32.7 km), the 3rd (48 km), the 4th (64 km), the 5th (99 km), and the 6th (192 km) ring road. The 2nd-4th, 5th, and 6th ring roads are located in the urban, suburban and rural area of the city, respectively. In this study, observations at the 2nd-4th ring roads represented the urban area and the 6th ring road represented the rural area. Climatically, there are four distinctive seasons (spring, summer, autumn, and winter) in Beijing. Heating is necessary over the winter months (annually often from mid-November to mid-March), but not at other times of the year. In this study, "wintertime" and "non-wintertime" were used to characterize the application of heating services. When conducting the monitoring, since the battery used in the measurement can only supply power for ~5 h to the G1302 in the non-wintertime, the 2nd-4th and 6th ring roads were surveyed on different days and each for three times. For G2401 used in the wintertime, the battery was sufficient for surveys of all the ring roads (2nd-6th) on a single day. All the measurements were conducted during the daytime. The weather conditions and average wind speed, which were directly associated with the diffusion conditions, were also captured from observation, as listed in Table 1.

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| Point Source/Ring Road | Measurement Species | Date | Time | Wintertime (Y/N) | Location | Weather | Mean Wind Speed (m/s) |
|---------------------------|--------------------------------------|----------------------|-------------|---------------------|----------|--------------|--------------------------|
| LNG carrier | CH ₄ /CO ₂ | 18th October, 2018 | 10:40-10:55 | N | Rural | Clear/cloudy | 2.5 |
| LNG station | CH ₄ /CO ₂ | 19th October, 2018 | 11:58-12:04 | N | Suburban | Cloudy | 1.4 |
| Residential area | CH ₄ /CO ₂ | 24th October, 2018 | 17:12-17:30 | N | Rural | Cloudy | 1.5 |
| Power plant | CO ₂ /CH ₄ /CO | 21st February, 2019 | 16:18-16:50 | Y | Suburban | Clear/cloudy | 2.0 |
| Landfill | CH ₄ /CO ₂ /CO | 21st February, 2019 | 17:00-17:35 | Y | Suburban | Clear/cloudy | 2.0 |
| 2nd-4th | CH ₄ /CO ₂ | 26th September, 2018 | 12:30-17:30 | N | Urban | Cloudy | 2.3 |
| 2nd-4th | CH ₄ /CO ₂ | 23rd October, 2018 | 14:03-19:12 | N | Urban | Clear | 2.4 |
| 2nd-4th | CH ₄ /CO ₂ | 30th October, 2018 | 12:45-18:19 | N | Urban | Clear | 2.3 |
| 6th | CH ₄ /CO ₂ | 28th September,2018 | 10:24-13:39 | N | Rural | Cloudy | 2.0 |
| 6th | CH ₄ /CO ₂ | 18th October, 2018 | 14:29-16:17 | N | Rural | Clear | 2.5 |
| 6th | CH ₄ /CO ₂ | 24th October, 2018 | 14:39-18:39 | N | Rural | Cloudy | 1.5 |
| 2nd-6th | CH ₄ /CO ₂ /CO | 20th February, 2019 | 09:22-18:18 | Y | All | Clear | 1.5 |

Table 1. Measurement routes and strategies.

2.3. Data Analysis

According to the observation log, invalid data were identified and flagged. The concentration data of CH_4 and CO_2 were calibrated by standard gases, as was the time data for the delayed response of the sampling system for 55 s. Finally, the CH_4 concentration data, along with the meteorological data obtained from the meteorological stations adjacent to the observed areas, were combined with GPS data to generate spatial information data using the Arc GIS software.

To understand the correlation of CH_4 and CO_2 for fugitive emissions and city-scale emissions of Beijing, the minimum concentration value was subtracted from each survey as background to create a dataset of differences, or "enhanced" (e) concentrations which means the observed concentration minus background values and represent the enhancement for a gas species (eCO₂, eCH₄). The data processing method refers to what Hurry et al. used in the observation activity in oil producing regions in North America [33].

3. Result and Discussion

3.1. Point Source Observation in Beijing

Figure 2 shows map transections of CH₄ and ratio "fingerprints" (CH₄:CO₂) of five point sources in Beijing. Figure 2A shows the screening result of an LNG carrier. A hotspot was captured with peak concentration up to 2524 ppbv. Its CH₄:CO₂ ratios show a shape of "L" of high CH₄ values versus low CO₂ values (slope = 0.02, r = 0.82, p < 0.001), and high CO₂ concentrations versus low CH₄ concentrations (slope = 0.0002, r = 0.43, p < 0.001). This is similar to the survey result of gas regulator stations [22]. The slope 0.0002 could be used to deduce that it mainly originated from the vehicle exhaust [33] and the slope 0.02 was associated with the release of CH₄ from storage facilities in the carrier [22].

Figure 2B presents the CH_4 map transection of an LNG filling station in the suburban area. One hotspot with maximum concentration 4777 ppbv appeared at the site adjacent to the LNG storage facility in the station while another with 2799 ppbv on the side of a working gas gun. Its $CH_4:CO_2$ ratios also shows a shape of "L" ($CH_4:CO_2 = 0.18$ and 0.001). The slope 0.18 is higher than 0.02 shown in Figure 2A by one order of magnitude which describes a different type of source profile, although it may be specific to the time and location.

Figure 2C shows the CH₄ map transection of a residential area at cooking time using gas for cooking. One hotspot with peak concentration 3436 ppbv emerged in the residential area while another one with maximum concentration 3448 ppbv on the side of it. The CH₄:CO₂ ratios of the two hotspots are 0.014 (r = 0.9, p < 0.001) and 0.007 (r = 0.82, p < 0.001). They may associate with the leakage of CH₄ pipelines or combustion devices [22,33].

Figure 2D shows the survey result of an LNG power plant. The MO circled the plant for three times and the concentration distributions obtained were similar, namely, concentrations being low in the upwind side and high in the downwind side. The $CH_4:CO_2$ ratios also reflected a characteristic "L" pattern (-0.02, r = -0.55, p < 0.001 and 3.8, r = 0.5, p < 0.001), but the "L" was different from Figure 2A,B

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and one of the slopes was negative. This may have been associated with the relationship between the CH_4 supply and the combustion efficiency of the boiler.

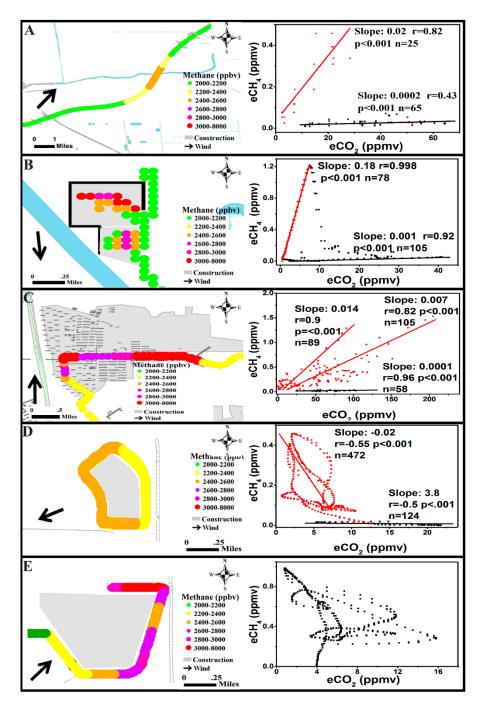


Figure 2. CH₄ concentration map and ratio "fingerprints" (CH₄:CO₂) of five point sources as a function of location: (**A**) an liquified natural gas (LNG) carrier; (**B**) an LNG filling station; (**C**) a residential area; (**D**) an LNG power station; (**E**) a refuse landfill. In A–D, the red dots represent the measurements of which high levels of CH₄ were associated with low CO₂ concentrations and black dots represent that the measurements where high CO₂ concentrations were associated with low CH₄ concentrations. The red and black lines are the linear regression results corresponding to the two groups of dots, respectively. The linear regressions are as the basis of ordinary least squares.

Figure 2E shows the survey result of a refuse landfill. High concentration was in the downwind east and northeast of site. The wind distribution and the existence of the CH₄ disposing plant located

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in the east of the landfill have contributed to such concentration distribution. The characteristic of CH_4 : CO_2 ratios were complex, which were different to the other four sources mentioned above. The complexity may have generated from the multiple CH_4 sources in the landfill, such as the landfill itself with the biological processes, the landfill leachate, the collecting pipelines of CH_4 , the burning of the collected CH_4 in the disposing plant, and the garbage trucks.

In short, the MO had enough temporal and spatial resolution to identify the typical point sources of CH_4 in both urban and rural areas in Beijing. The ratio "fingerprints" for the point sources using LNG all showed a characteristic "L" pattern, which was useful to distinguish different CH_4 sources in the city.

3.2. City Observation in the Non-Wintertime Period

Figure 3 shows results of the urban (the 2nd–4th ring roads) and rural area (the 6th ring road) of Beijing in non-wintertime. The boxplot method was used to show the distribution characteristics of the observed CH_4 concentration data, whose statistical results are shown in Table 2. It includes the values of minimum, maximum, median, upper quartile (Q1), lower quartile (Q2), inter-quartile range (IQR), and maximum outlier and number of valid data for each measurement. Note that the outliers are the values outside of the box body (less than Q1 – $1.5 \times IQR$ or higher than Q3 + $1.5 \times IQR$).

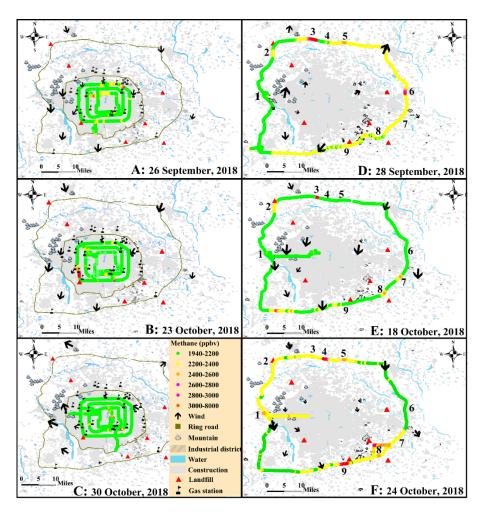


Figure 3. CH4 concentration maps of the 2nd–4th and the 6th ring roads as a function of location in non-wintertime. Note: The directions and magnitudes of arrows show the wind directions and speeds when the meteorological station was at these locations.

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| | Figure | Minimum (ppbv) | Maximum (ppbv) | Median (ppbv) | Q1 ¹⁾ (ppbv) | Q3 ¹⁾ (ppbv) | IQR ²⁾ (ppbv) | Maximum Outlier (ppbv) | N^3 |
|-------|--------|-------------------|-------------------|------------------|----------------------------|----------------------------|-----------------------------|---------------------------|--------|
| Urban | 3A | 2015 | 2370 | 2127 | 2082 | 2198 | 116 | 3047 | 15,760 |
| | 3B | 1956 | 2244 | 2056 | 2011 | 2104 | 93 | 3182 | 3243 |
| | 3C | 1940 | 2050 | 1986 | 1973 | 2004 | 31 | 2864 | 3016 |
| | 4 | 2059 | 2248 | 2145 | 2169 | 2116 | 53 | 2436 | 13,793 |
| Rural | 3D | 2049 | 2515 | 2221 | 2138 | 2290 | 152 | 7657 | 12,339 |
| | 3E | 2016 | 2238 | 2126 | 2103 | 2158 | 55 | 4441 | 2641 |
| | 3F | 2074 | 2467 | 2255 | 2197 | 2307 | 110 | 4216 | 2502 |
| | 4 | 1953 | 2276 | 2009 | 1979 | 2098 | 119 | 2789 | 8143 |

Table 2. Statistics analysis for CH₄ concentration data of the urban and rural area.

Q: quartile; IQR: inter-quartile range; N: number of valid data.

Figure 3A–C present the CH_4 map transections of the urban area on 26 September, 23 October, and 30 October, 2018, respectively. Generally, the concentrations for the three survey days mainly ranged from 1940 to 2370 ppbv, which are obviously higher than that reported in 2000 (1870–2000 ppbv) [39] in the urban area and similar to that measured during 2009–2014 by WMO/GAW Shangdianzi regional background station (1850–2400 ppbv) which is 100 km away from the urban area 16 . The concentration gradient was relatively small, implying that the urban area might be treated as a typical non-point source of CH_4 . However, some hotspots, whose concentration were higher than 2800 ppbv, were observed during all surveys with different locations, indicating that some intermittent or mobile fugitive leaks may exist, for example, underground transmission, distribution pipelines or usages by the residents [31].

Figure 3D-F show the CH₄ map transections of the rural area on 28 September, 18 October, and 24 October, 2018, respectively. The concentrations for the three surveys mainly distributed between 2016–2515 ppby, which were notably higher than those of the urban area mentioned above. The dispersion of these data was also higher than those of the urban area through comparing the inter-quartile range (IQRs). There were obviously concentration gradients up to several hundred ppbv, in Figure 3D,F, but not in Figure 3E. Higher concentrations were observed both in areas upwind and downwind of urban Beijing, which revealed that the local emission might be the main impact factor rather than the CH₄ emission from urban area. Moreover, eight notable hotspots were captured. The concentrations of hotspot 1 (2400–2600 ppbv) were just slightly higher than its surroundings (2000–2400 ppbv). It was identified to be a LNG power plant, which was a source not included by the methane inventory of China [5], through the method mentioned in Section 3.1 and the field investigation. The hotspot 2 (2200–2400 ppbv) could be recognized in all three surveys, mainly due to the nearby landfill. Hotspot 3 emerged with the maximum concentration 7657 ppbv in Figure 3D,E. It might have been a stationary fugitive source, such as NG pipeline leak, which existed before 28 September, 2018 and was presumably remediated between 18 October, and 24 October, 2018. Hotspots 4-9 emerged randomly in the three figures. Hotspot 7 was identified as the LNG carrier (Figure 2A) and hotspot 8 was a residential area (Figure 2C). But the sources for hotspots 4–6 and 9 were not identified. According to their characteristic, they were speculated to be associated with mobile gas carriers or intermittent sources, such as leakages of underground transmission and distribution pipelines.

3.3. City Observation in the Wintertime Period

Figure 4 shows the result of survey on 20 February, 2019 which was in the wintertime. The statistical results of the urban and rural area are presented in Table 2. The CH_4 concentrations for the urban area (the 2nd–4th ring roads) were mainly in the range of 2059–2248 ppbv and it was similar to that obtained in the non-wintertime. The CH_4 concentrations for the rural area (the 6th ring road) were mainly between 1953–2276 ppbv, lower than that obtained in the non-wintertime. There was a remarkable concentration gradient. Higher concentrations were found in the southwest area and lower in the northeast part. Meanwhile, concentrations were much higher in the downwind area than the upwind area, similar to the observation result in urban area of eastern Germany [23]. Several hotspots were observed and their identifications were similar to that discussed in Section 3.2.

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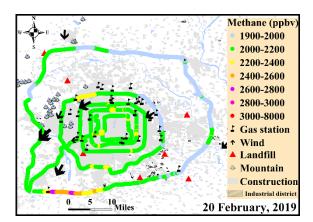


Figure 4. CH₄ concentration map as a function of location in wintertime.

Figure 5A-D shows the relationship between the CH₄ and CO₂ concentrations measured in non-winter and wintertime. Regression analysis was used to determine the CH₄/CO₂ ratio. The regression coefficients (r) in Figure 5A,C were both quite small, demonstrating that the correlation of CH₄ and CO₂ enhanced concentrations in the urban of Beijing was quite weak, in contrast to the significant correlation in research conducted in Nanjing [40]. Our study provided high spatial resolution data of CH₄ and CO₂, while a fixed sensor is used by the study in Nanjing on a higher floor of a building, which might explain that stronger correlations between CH₄ and CO₂ were observed since they were sensitive to more mixed air within the footprint of their sensor. In Figure 5C, the red outliers are with high CH₄ concentrations but small or even negative CO₂ elevation makes the whole figure a shape of "L". This "L" pattern was similar to that of the LNG point sources, such as the passing-by LNG carrier, the LNG filling station, or the LNG power plant (see Figure 2A,B,D, respectively), which indicated that LNG sources might have contributed to CH₄ enhancement in this survey. The enhanced CH_4 concentrations had significant correlation (p < 0.001) with CO_2 enhancements in the rural area with regression coefficient (r) of 0.5 and 0.79 in Figure 5B,D, respectively. The regression slope (CH₄:CO₂) for the rural area was 0.003 in non-wintertime and 0.004 in wintertime. The slope 0.004 was comparable to that obtained in Boulder [41] and Indianapolis [30], USA. It should be noted that different methodologies were applied in these studies and the differences introduced by methodologies were not considered.

CO concentrations were also observed in wintertime. Figure 5E,F shows the relationship between the CH₄ and CO concentrations in wintertime. In Figure 5E, similar to that of CH₄/CO₂ in Figure 5A,C, the regression coefficients (r) was quite small, revealing that the correlation of CH₄ and CO enhanced concentrations in the urban of Beijing was quite weak. Meanwhile, obvious CH₄ fugitive emissions could be seen (red dots). This revealed that the CH₄ sources were complicated in the urban area. Shown in Figure 5F, the enhanced CH₄ concentrations had significant correlation (p < 0.001) with CO enhancements in the rural area with a regression coefficient (r) of 0.87, which was similar to the significant correlation of CH₄ and CO₂ shown in Figure 5B,D. This indicated the high CO₂/CH₄/CO concentrations in winter might have been influenced by combustion or by well mixed anthropogenic emissions from other sources in Beijing city.

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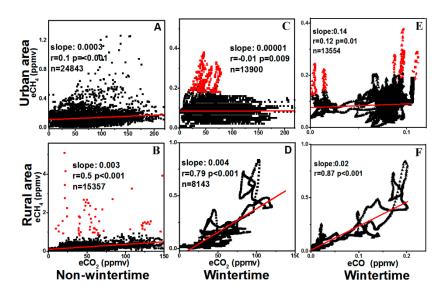


Figure 5. Correlation analysis of CH₄ and CO₂ enhanced concentrations in the non-wintertime in the urban area (**A**) and in the rural area (**B**) and in the wintertime in the urban area (**C**) and in the rural area (**D**) and of CH₄ and CO enhanced concentrations in the wintertime in the urban area (**E**) and in the rural area (**F**). For the rural area measurement, three measurement data were merged for each figure. The slope, the regression coefficient (r) and the confidence level (p), the number of observations (n), are also shown. Note that the scales are different for urban and rural areas. The red lines are the linear regression results. In B, C, and E, the red dots represent the measurements of obvious fugitive CH₄ and they were not used in the regression analysis.

4. Conclusions

A mobile observatory was applied to real-time monitor the CH₄ concentration of point and non-point (city) sources in Beijing. The results indicated that this mobile measurement methodology has enough temporal and spatial resolution to identify typical CH₄ sources in Beijing. The ratio "fingerprints" (CH4:CO2) for the LNG carrier, LNG filling station, and LNG power plant showed a characteristic "L" pattern which may reveal typical point sources with LNG leakage. For city observations in non-wintertime, the CH₄ concentrations were in the range of 1940–2370 ppbv and those in the rural area were 2016–2515 ppbv. Several fugitive point sources, including LNG power plant and landfill, were successfully identified in the rural measurement. As for city observation in wintertime, the CH₄ concentrations were mainly in the range of 2059–2248 ppby, which was comparable to that in the non-wintertime observations, while concentrations in the rural area (1953–2276 ppbv) also had an obvious concentration gradient. According to the correlation analysis, the CH₄ and CO₂ enhanced concentration showed significant correlation in the rural area, but not in the urban area. A shape of "L" was obtained and it suggested fugitive emissions from the NG sources in wintertime. Further attempts of CH₄ emission study by inverse modeling on a city scale would help understanding the observed signals. Similar studies report the quantification of CH₄ emission by mobile measurement platform such as airplanes [42,43].

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