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CO₂ Emission Performance, Mitigation Potential, and Marginal Abatement Cost of Industries Covered in China's Nationwide Emission Trading Scheme: A Meta-Frontier Analysis

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Academic Editor: Vincenzo Torretta Received: 24 April 2017; Accepted: 1 June 2017; Published: 2 June 2017

Abstract: China's nationwide emission trading scheme (CN-ETS) is scheduled to be launched in 2017. It is of great urgency and necessity to obtain a good understanding of the participating sectors of CN-ETS in terms of energy utilization and CO₂ emissions. In this regard, it should be noted that the findings may be biased without taking industry heterogeneity into consideration. To this end, a meta-frontier framework with the directional distance function is employed to estimate the CO₂ emission performance (CEP), mitigation potential (MP), and marginal abatement cost (MAC) at sector levels under the meta-frontier and the group-frontier. The results indicate that significant disparities in the CEP, MP, and MAC exist under both frontiers among various sectors, and the sectoral distributions of CEP, MP, and MAC are found to be different between the two frontiers. Additionally, the differences between the two frontiers in terms of CEP, MP, and MAC are considerable, and exhibit unequal distributions among these sectors. Notably, MAC under both frontiers and the difference between the mare found to be significantly correlated with the carbon intensity. Finally, policy implications are provided for the government and participating enterprises, respectively.

Keywords: China's nationwide emission trading scheme; directional distance function; meta-frontier analysis; CO₂ emission performance; mitigation potential; marginal abatement cost

1. Introduction

With climate change becoming an increasingly serious issue, the reduction of carbon dioxide (CO₂) emissions has attracted extensive attention worldwide. As the greatest CO₂ emitter in the world [1,2], China has shown its determination for developing a low-carbon economy and promised to abate its CO₂ emissions per unit of gross domestic product (GDP) (i.e., carbon intensity) by 40–45% by 2020 compared with that in 2005 [3]. Further, China set the latest target of abating carbon intensity by 18% by 2020, with 2015 as the reference year [4]. In order to achieve the above international commitments for mitigating CO₂ emissions, China's National Development and Reform Commission (NDRC) has launched seven pilot emission trading schemes (ETS) since 2013 [5], which are specifically located in Shenzhen, Guangdong, Shanghai, Beijing, Tianjin, Chongqing, and Hubei. These regional carbon markets are considered as experimental explorations for the establishment of China's nationwide emission trading scheme (CN-ETS), which is scheduled to be launched in 2017. It is reported the CN-ETS will cover seven emission-intensive industries, including paper making, electricity generation, metallurgy, non-ferrous metals, building materials, the chemical industry, and the aviation service industry [6].

In the context of achieving the construction and operation of CN-ETS, it is of great urgency and necessity to obtain a good understanding of the participating sectors in terms of energy utilization and CO₂ emissions [7]. In this regard, estimating the CO₂ emission performance (CEP), mitigation potential (MP), and marginal abatement costs (MAC) for these sectors can provide valuable information for the governments and participating enterprises. From the perspective of the government, a good knowledge of CEP, MP, and MAC could help design appropriate market mechanisms for the CN-ETS, e.g., the estimated MAC may be used as a reference for carbon pricing [8,9]. On the other hand, a comprehensive acquaintance of MAC among the participating sectors could help the participating enterprises to determine the best mitigation strategies [10]. Moreover, to the best knowledge of the authors, there have been few studies on the CEP, MP, and MAC of the sectors covered in the CN-ETS, and this paper aims to fill this research gap.

Additionally, it is accepted that there exists significant heterogeneity in terms of the production technology among various sectors [11], which is regarded as an obstacle to the objective evaluation of CEP, MP, and MAC [12]. Therefore, taking the technology heterogeneity into consideration, we employ a joint framework consisting of the directional distance function (DDF) and meta-frontier analysis to estimate CEP, MP, and MAC under the meta-frontier and the group-frontier, respectively. Following this, we investigate the sectoral distributions of CEP, MP, and MAC under both frontiers, and analyze the differences between the two frontiers at sector levels. Finally, several policy recommendations for the CN-ETS are made based on the conclusions. The rest of the paper proceeds as follows: Section 2 provides a literature review. Section 3 introduces the methods and materials. Section 4 presents the empirical results and discussion. Section 5 draws conclusions and policy implications.

2. Literature Review

DDF, theorized and developed by Chung et al. [13] and Chambers et al. [14,15], has been widely employed to study energy and environmental issues [16–20]. The main advantage of the DDF is that it can achieve the expansion of desirable outputs and reduction of undesirable outputs (e.g., CO₂), simultaneously [18]. Generally, two estimation techniques are often employed to estimate the DDF, namely the parametric and non-parametric methods. Compared with the former, the non-parametric technique does not need to pre-determine any functional and parametric forms, thereby avoiding the impacts of subjective factors on the results [17]. In this regard, the data envelopment analysis (DEA), a well-developed nonparametric frontier tool, is often combined with DDF to evaluate CEP [21–23], MP [9,24], and MAC [25–27]. For instance, Watanabe and Tanaka [21] employed the DEA-DDF to estimate the environmental performance for China's industrial sector at province levels from 1994 to 2002. Liu et al. [26] evaluated the carbon emission performance and marginal abatement cost for provinces in China by using a non-parametric DDF. Additionally, the same method was applied by Wei et al. [9] to measure the reduction potential of CO₂ emissions for the thermal power plants in China's Zhejiang province.

Considering the technology heterogeneity across decision making units (DMUs), Battese et al. [28] and O'Donnell et al. [29] incorporated the meta-frontier approach into DDF to formulate a joint framework. In the framework, DMUs with different production technologies are classified into several groups in which DMUs are deemed to be homogeneous, and then evaluated under the meta-frontier and the group-frontier, respectively. Recently, this combined methodology has been widely applied in the energy and environmental field [30–39]. For instance, using the combined method of DDF and the meta-frontier approach, Lin et al. [33] evaluated the environmental performance of 63 countries during the period from 1981 to 2005. Further, Zhang et al. [34] proposed a meta-frontier non-radial DDF by combining the meta-frontier approach with the non-radial DDF, and used it to assess the energy and CO₂ emission performance of electricity generation in Korea. Furthermore, the model was applied by Yao et al. [36] to estimate China's energy efficiency, carbon emission performance, and mitigation potential at regional levels. Additionally, based on the same model, Li and Song [38] constructed a green development growth index to assess China's green development at province levels.

The literature on energy and environmental issues is abundant at industry levels [7,10,27,40,41]. For instance, Lee and Zhang [27] measured the reduction potential and marginal abatement cost of CO₂ emissions for 30 of China's manufacturing industries. Yuan et al. [40] estimated the shadow prices of CO₂ emissions for China's industrial sectors with the use of non-parametric DDF. Teng et al. [41] employed multiple methods to derive the marginal abatement cost curves for China's energy-intensive industries. Zhou et al. [10] applied multiple distance function approaches to approximate the shadow prices of CO₂ emissions for CO₂ emissions for Shanghai's industrial sectors. Xiao et al. [7] estimated the marginal abatement costs of CO₂ emissions for China's industrial sectors, while a comprehensive investigation on the participating sectors of CN-ETS has not been conducted. Furthermore, to the best of our knowledge, few studies take the industry heterogeneity into account, in addition to Xie et al. [11] and Chung and Heshmati [12]. In this context, we attempt to perform an empirical study on the participating sectors of CN-ETS in terms of CEP, MP, and MAC, taking into consideration the industry heterogeneity.

3. Methods and Materials

3.1. Environmental Production Technology

Consider a productive process in which various inputs of energy and non-energy resources are utilized to jointly produce desirable outputs and undesirable outputs. Mathematically, the joint production can be presented as Equation (1), which is the so-called environmental production technology.

$$P(x) = \{(y,b) : x \text{ can produce } (y,b)\}$$
(1)

where $x \in \Re^{I}_{+}$ denotes the input vector, $y \in \Re^{J}_{+}$ denotes the desirable output vector, and $b \in \Re^{K}_{+}$ denotes the undesirable output vector. Notably, in the joint-production process, the inputs and the desirable outputs are usually presumed to be strongly disposable, while the undesirable outputs are weakly disposable.

3.2. Directional Distance Function

The directional distance function (DDF) has been widely utilized to characterize the environmental production technology. Compared with traditional distance functions, DDF can achieve the simultaneous extension of desirable outputs and shrinkage of undesirable outputs, which can be mathematically defined in Equation (2).

$$\overrightarrow{D}(x,y,b;g_y,g_b) = max\{\beta: (y+\beta g_y,b+\beta g_b) \in P(x)\}$$
(2)

where (g_y, g_b) is the direction vector indicating the scale directions for the desirable outputs and undesirable outputs, which is generally specified as (y, -b). β is the outcome of DDF that estimates the greatest extent to which y and b can be respectively expanded and reduced given *x*.

Following Boyd et al. [42], the non-parametric DEA-DDF with constant returns to scale (CRS) is presented as follows:

$$\dot{D}(x, y, b; y, -b) = \max \beta_{0}$$

$$\sum_{j=1}^{N} \lambda_{j} x_{j} \leq x_{0}$$

$$\sum_{j=1}^{N} \lambda_{j} y_{j} \geq (1 + \beta_{0}) y_{0}$$

$$\sum_{j=1}^{N} \lambda_{j} b_{j} = (1 - \beta_{0}) b_{0}$$

$$\beta_{0} \geq 0, \ \lambda_{j} \geq 0$$
(3)

where x_0 , y_0 , b_0 , respectively, denote the inputs, desirable outputs, and undesirable outputs for the observed production unit, and λ_j is the intensity variable. Notably, the equal sign of the third constraint in Equation (3) reflects the weak disposability assumption of the undesirable outputs.

3.3. Meta-Frontier and Group-Frontier

Considering the existence of heterogeneity among various sectors in terms of energy utilization and carbon emissions, we thereby incorporate a meta-frontier analysis into the above DDF. Suppose N DMUs (industries in this paper) can be classified into I ($I \ge 1$) groups. The number of DMUs in the *i*th group is N_i (i = 1, 2, ..., I), and $\sum_{i=1}^{I} N_i = 1$. Thus, the environmental production technologies of the meta-frontier and group-frontier can be described as follows:

$$P^{M}(x) = \{(y,b) : x \ can \ produce \ (y,b)\}$$
(4)

$$P^{G-i}(x) = \{(y,b) : x \text{ can produce } (y,b)\}, i = 1, 2, \dots, I$$
(5)

where both $P^{M}(x)$ and $P^{G-i}(x)$ satisfy the assumptions of disposability mentioned earlier, and $P^{M}(x)$ consists of all the $P^{G-i}(x)$ (i = 1, 2, ..., I) : $P^{M}(x) = \{P^{G-1}(x) \cup P^{G-i}(x) \cup ... \cup P^{G-I}(x)\}$, namely, $P^{M}(x)$ includes all DMUs and envelops all group frontiers.

Applying the DDF depicted in Equation (2) to specify Equations (4) and (5), we can obtain the meta-frontier DDF and the *i*th group-frontier DDF:

$$\overrightarrow{D}^{M}(x,y,b;g_{y},g_{b}) = max\left\{\beta^{M}:\left(y+\beta^{M}g_{y},b+\beta^{M}g_{b}\right)\in P^{M}(x)\right\}$$
(6)

$$\overset{\to G^{-i}}{D}(x, y, b; g_y, g_b) = max \Big\{ \beta^{G^{-i}} : \Big(y + \beta^{G^{-i}} g_y, b + \beta^{G^{-i}} g_b \Big) \in P^{G^{-i}}(x) \Big\}, \ i = 1, 2, \dots, I$$
(7)

Finally, we employ Equation (3) to estimate the DDF values under the meta-frontier (β^{M}) and group-frontier (β^{G-i}), respectively.

3.4. CO₂ Emission Performance and Mitigation Potential

According to Hu and Wang [43], Zhou and Ang [44], and Choi et al. [45], the CEP can be measured by the ratio of target CO₂ emissions to the actual CO₂ emissions:

$$CEP_j = \frac{C_j - \beta_j C_j}{C_j} = 1 - \beta_j$$
(8)

where CEP_j , C_j , and β_j denote the CEP, actual CO₂ emissions, and DDF value of DMU_j, respectively. In addition, $\beta_j C_j$ denotes the excessive CO₂ emissions of DMU_j compared with the frontier, which is considered as achievable emission reductions. Thus, the MP of DMU_j can be measured by Equation (9):

$$MP_j = \frac{\beta_j C_j}{C_j} \times 100\% = \beta_j \times 100\%$$
(9)

where MP_j denotes the MP of DMU_j , and the sum of CEP_j plus MP_j equals 1, which implies that a greater reduction potential signifies a worse environmental performance of the observation.

3.5. Shadow Price

The shadow price is often utilized to approximate the marginal abatement cost of undesirable outputs, which is generally read as the opportunity cost of eliminating one extra unit of undesirable output in terms of lost desirable outputs in the production process. In this context, Färe and Grosskopf [16] developed the duality relationship between the distance function and revenue function,

which is of great significance for the derivation of the shadow price of undesirable outputs. With the use of the Lagrangian technique and the envelope theorem, the shadow price of undesirable outputs is derived as follows:

$$p^{b} = -p^{y} \times \frac{\partial \vec{D}(x, y, b; g_{y}, g_{b}) / \partial b}{\partial \vec{D}(x, y, b; g_{y}, g_{b}) / \partial y}$$
(10)

Noticing the deficiency that different observations sharing the same frontier point enjoy at the same shadow price, Lee et al. [46] thereby incorporated an inefficiency factor into Equation (8) and reformulated the shadow-pricing equation as:

$$p^{b} = -p^{y} \times \frac{\partial \overrightarrow{D}(x, y, b; g_{y}, g_{b}) / \partial(\sigma_{b}b)}{\partial \overrightarrow{D}(x, y, b; g_{y}, g_{b}) / \partial(\sigma_{y}y)} \times \frac{\sigma_{b}}{\sigma_{y}}$$
(11)

where p^y and p^b , respectively denote the shadow prices of the desirable and undesirable outputs. According to Färe et al. [47], p^y is usually assumed to equal its market price (i.e., 1). σ_y, σ_b are the inefficiency factors and σ_b/σ_y equals $(1 - \beta)/(1 + \beta)$.

3.6. Materials

3.6.1. Sectors and Variables

It is reported that the CN-ETS will cover the emission-intensive industries of paper making, electricity generation, metallurgy, non-ferrous metals, building materials, the chemical industry, and the aviation service industry [6]. According to the Standard Industrial Classification formulated by the National Bureau of Statistics of China (NBSC), the seven industries covered in the CN-ETS can be further divided into 39 sectors. Furthermore, according to the carbon intensity, the 39 sectors can be classified into three groups: high-carbon, medium-carbon, and low-carbon groups, and each group consists of 13 sectors. The specific sectors of the three groups are listed in Table 1.

Table 1. Specific sectors and codes for high-, medium-, and low-carbon groups.

High-Carbon Group	Medium-Carbon Group	Low-Carbon Group
Electricity Generation and Supply (S1)	Manufacture of Glass Products (S14)	Manufacture of Paper (S27)
Pressing of Steel (S2)	Manufacture of Ceramics Products (S15)	Manufacture of Paper Products (S28)
Smelting of Steel (S3)	Manufacture of Glass Fiber Products (S16)	Manufacture of Paper Pulp (S29)
Casting of Ferrous Metals (S4)	Manufacture of Glass (S17)	Manufacture of Synthetic Fibers (S30)
Smelting of Ferroalloy (S5)	Manufacture of Raw Chemical Materials (S18)	Manufacture of Fiber Materials (S31)
Smelting of Iron (S6)	Manufacture of Special Chemical Products (S19)	Pressing of Non-ferrous Metals (S32)
Processing of Petroleum (S7)	Manufacture of Synthetic Materials (S20)	Smelting of Common Non-ferrous Metals (S33)
Processing of Coking (S8)	Manufacture of Fertilizers (S21)	Smelting of Non-ferrous metal alloy (S34)
Manufacture of Tile and Construction stone (S9)	Manufacture of Paints and Pigments (S22)	Smelting of Precious Metals (S35)
Manufacture of Cement and Limestone Products (S10)	Manufacture of Common Chemical Products (S23)	Smelting of Rare Earth Metals (S36)
Manufacture of Cement and Limestone (S11)	Manufacture of Pesticides (S24)	Casting Pressing of Non-ferrous Metals (S37)
Manufacture of other Non-metallic Mineral Products (S12)	Manufacture of Explosives and Fireworks (S25)	Manufacture of Plastics Products (S38)
Manufacture of Fireproof Materials (S13)	Civil Aviation Services (S26)	Manufacture of Rubber Products (S39)

Note: sector's carbon intensity gradually decreases from S1 to S39.

Following the common practice in the literature, this paper selects five variables as the inputoutput indicators. Specifically, three variables of capital stock, labor employment, and energy consumption serve as the inputs, while the industrial value added (IVA) and CO_2 emissions are viewed as the desirable output and the undesirable output, respectively. The input-output variables are summarized in Table 2.

Category	Variable	Unit
	Capital	100 million Yuan
Input	Labor	10 thousand persons
	Energy	10 thousand tons of coal equivalent (10,000 tce)
Desirable output	IVA	100 million Yuan
Undesirable output	CO ₂ emission	10 thousand tons

Table 2. Variables of inputs and output	s.
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Note: Yuan denotes Chinese currency.

3.6.2. Data

A statistical description of the input-output data is shown in Table 3, and a detailed description of the data sources of each variable is provided below.

Voor	Itoms		Inputs		Desirable Output	Undesirable Output
iear	items	Capital	Labor	Energy	IVA	CO ₂
	Mean	4394.54	66.93	7393.14	2284.41	25,858.95
0014	Std. dev.	10678.39	64.43	22,921.26	2857.27	87,583.80
2014	Max	66,386.48	259.51	138,403.59	13,761.05	536,897.12
	Min	55.01	1.59	14.95	37.31	55.93
	Mean	6108.10	69.16	8418.21	2839.78	28,994.29
0017	Std. dev.	14,842.20	66.58	26,099.32	3551.90	98,203.14
2017	Max	92,272.49	268.17	157,593.40	17,106.51	601,994.73
	Min	76.46	1.64	17.02	46.38	62.71

Table 3. Descriptive statistics of input-output data (2014 and 2017).

Note: due to limitations of statistical data, the data in this paper only contain industrial enterprises above designated size.

Capital stock is generally estimated by the Perpetual Inventory Method (PIM), which requires the data of the initial capital stock and industrial capital depreciation rate. However, due to the limitations of statistical data, these data also need to be estimated. Thus, to decrease the deviation from data estimation, we adopt an alternative approach of taking the net value of fixed assets as the capital stock [48], which are collected from the "China Industrial Economic Statistical Yearbook 2015 [49]". Moreover, the data of the "Civil Aviation Services (S26)" are collected from "From the Statistical View of Civil Aviation 2015 [50]".

Labor employment is directly collected from the "China Industrial Economic Statistical Yearbook 2015 [49]" and "From the Statistical View of Civil Aviation 2015 [50]".

Energy consumption denotes the comprehensive consumption of end-use energy, which includes 17 fuels of raw coal, cleaned coal, other washed coal, coke, coke oven gas, other gas, other coking products, crude oil, gasoline, kerosene, diesel oil, fuel oil, liquefied petroleum gas (LPG), refinery gas, other petroleum products, natural gas, and liquefied natural gas (LNG). The related data are obtained from the "China Energy Statistical Yearbook [51]" and "From the Statistical View of Civil Aviation 2015 [50]". All kinds of energy usages are converted into standard coal equivalents with the use of conversion factors which are provided in Appendix IV of the China Energy Statistical Yearbook.

Considering that IVA cannot be directly gathered from relevant statistical yearbooks since 2008, we calculate the data in 2014 by multiplying those in 2007 with the growth rates of the industrial added

value, which are acquired from the National Data Website [52]. Notably, the industrial added value of the Civil Aviation Service industry is converted from the main business income, by multiplying the ratio of the industrial added value of the entire industrial sector to the corresponding main business income. The main business income of the Civil Aviation Service industry is gathered from "From the Statistical View of Civil Aviation 2015 [52]".

 CO_2 emissions are generally estimated with the reference approach of the Intergovernmental Panel on Climate Change (IPCC), due to the lack of official statistical data on CO_2 emissions in China. Therefore, this paper also adopts the approach to estimate CO_2 emissions with the use of energy usage data:

$$CO_2 \ emissions = \sum_i E_i \times NCV_i \times CEF_i \times COF_i \times (44/12)$$
(12)

where E_i denotes the final usage of energy source *i*; NCV_i (net caloric value), CEF_i (CO₂ emission factor), and COF_i (CO₂ oxidation factor) denote the heat equivalent, the CO₂ emission factor, and the CO₂ oxidation factor, respectively. 44/12 is the ratio of the CO₂ molecular weight (44) to the C atomic weight (12), which is named the CO₂ gasification coefficient.

Based on the economic development strategy and mitigation policy objectives which are included in "China's 12th Five Year Plan [53]", "China's 13th Five Year Plan [4]", and the "Industrial Action Plan on Climate Change (2012–2020) [54]", we calculate the annual growth rates of capital stock, labor employment, energy consumption, IVA, and CO_2 emissions. For instance, the annual average growth rates of energy consumption and CO_2 emissions are calculated based on the projected goals set for the energy consumption reduction per unit of GDP and CO_2 emissions reduction per unit of GDP. Additionally, the annual growth rates of the capital stock and labor employment are assumed to be unchanged during the two periods [55]. Table 4 reports the annual growth rates of the variables. By multiplying the respective annual growth rate, the input-output data in 2014 are used to forecast those in 2017, and the results are listed in Table 3.

Annual Growth Rate	2011-2015	2016-2020
φ^k	11.59%	11.59%
φ^l	1.11%	1.11%
φ^e	6.53%	3.43%
μ	9.60%	6.50%
η	6.05%	2.75%

Table 4. Annual growth rate of input and output variables.

Note: φ^k , φ^l , and φ^e denote the annual growth rate of capital stock, labor employment, and energy consumption, while μ and η denote that of IVA and CO₂ emission.

4. Results and Discussions

In this section, we first report the estimates of DDF values (β), CO₂ emission performance (CEP), mitigation potential (MP), and marginal abatement cost (MAC) under the meta-frontier and group-frontier at group levels. Then, we investigate the sectoral distributions of CEP, MP, and MAC under both frontiers, and analyze the differences between the two frontiers in terms of the CEP, MP, and MAC at sector levels.

4.1. Statistical Summary of Estimates

4.1.1. DDF Values

DDF values under the meta-frontier (β^M) and the group-frontier (β^G) are estimated by combining Equation (3) with Equations (6) and (7), respectively, and the results are listed in Table 5. From the table, it can be observed that β^M and β^G vary from 0 to 0.888 and from 0 to 0.687, respectively. Moreover, the mean values of β^M are generally larger than those of β^G . From the perspective of total sectors, the average β^M is found to be 0.544, which is more than twice that of β^G (0.225), indicating that there is a

significant difference. Additionally, the differences in the average DDF value between the two frontiers are found to be considerable at group levels, especially for the high-carbon and medium-carbon groups. Specifically, the average β^M (0.688) of the high-carbon group is found to be almost three times that of β^G (0.234). Further, the average DDF value of the medium-carbon group varies between the meta-frontier and group-frontier at greater levels, which is found to jump from 0.169 to 0.672. As for the low-carbon group, the difference between the two frontiers could be neglected. Notably, there are significant differences among the DDF values of the three groups when the meta-frontier is used as the basis of evaluation, while relatively moderate distinctions are observed when using the group frontiers.

	Groups	Ν	Min	Max	Mean	Std. Dev.
	High-carbon	13	0.000	0.888	0.688	0.214
Mate for all an	Medium-carbon	13	0.000	0.840	0.672	0.211
Meta-frontier	Low-carbon	13	0.000	0.635	0.271	0.216
	Total	39	0.000	0.888	0.544	0.285
	High-carbon	13	0.000	0.687	0.234	0.217
Crown frontior	Medium-carbon	13	0.000	0.532	0.169	0.209
Gloup-Holluei	Low-carbon	13	0.000	0.635	0.271	0.216
	Total	39	0.000	0.687	0.225	0.213

Table 5. Estimates of DDF values.

4.1.2. CO₂ Emission Performance

Table 6 reports the estimates of CO_2 emission performance under the meta-frontier (MCEP) and the group-frontier (GCEP). From the table, it is found that MCEP and GCEP vary from 0.112 to 1 and from 0.313 to 1, respectively. Moreover, the mean values of GCEP are generally higher than those of MCEP. From the perspective of total sectors, the average GCEP is found to be 0.775, while that under the meta-frontier is only 0.456, which indicates that there is a significant difference. Additionally, the differences in the average CEP scores between the two frontiers are found to be considerable at group levels, especially for the high-carbon and medium-carbon groups. Specifically, the average GCEP (0.766) of the high-carbon group is found to be more than twice that of MCEP (0.312). Similarly, the average CEP of the medium-carbon group is found to jump from 0.328 to 0.831, which is an increase of 153%. As for the low-carbon group, the difference between the two frontiers could be neglected. Notably, there are significant differences among the CEP scores of the three groups when the meta-frontier is used as the basis of evaluation, while relatively moderate distinctions are found when using the group frontiers.

Groups	N	Min	Max	Mean	Std. Dev.
h-carbon	13	0.112	1.000	0.312	0.214
um-carbon	13	0.160	1.000	0.328	0.211
w-carbon	13	0.365	1.000	0.729	0.216
Total	39	0.112	1.000	0.456	0.285
h-carbon	13	0.313	1.000	0.766	0.217
um-carbon	13	0.468	1.000	0.831	0.209
w-carbon	13	0.365	1.000	0.729	0.216
Total	39	0.313	1.000	0.775	0.213
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4.1.3. Mitigation Potential

Table 7 reports the estimates of Mitigation Potential under the meta-frontier (MMP) and the group-frontier (GMP). From the table, it can be observed that MMP and GMP vary from 0% to 88.8% and from 0% to 68.7%, respectively. Moreover, the mean values of MMP are generally higher than

those of GMP. From the perspective of total sectors, the average MMP is found to be 54.4%, which is more than twice the mean value of GMP (22.5%), indicating that there is a significant difference. Additionally, the differences in the average MP values between the two frontiers are found to be considerable at group levels, especially for the high-carbon and medium-carbon groups. Specifically, the average MMP of the high-carbon group (68.8%) is found to be almost three times that of GMP (23.4%). Further, the average MP value of the medium-carbon group varies between the meta-frontier and group-frontier at greater levels, which is found to jump from 16.9% to 67.2%. As for the low-carbon group, the difference between the two frontiers could be neglected. Notably, there are significant differences among the MP values of the three groups when the meta-frontier is used as the basis of evaluation, while relatively moderate distinctions are observed when using the group frontiers.

	Groups	Ν	Min	Max	Mean	Std. Dev.
	High-carbon	13	0.0%	88.8%	68.8%	0.214
	Medium-carbon	13	0.0%	84.0%	67.2%	0.211
Meta-frontier	Low-carbon	13	0.0%	63.5%	27.1%	0.216
	Total	39	0.0%	88.8%	54.4%	0.285
	High-carbon	13	0.0%	68.7%	23.4%	0.217
Crown frontion	Medium-carbon	13	0.0%	53.2%	16.9%	0.209
Gloup-Holluel	Low-carbon	13	0.0%	63.5%	27.1%	0.216
	Total	39	0.0%	68.7%	22.5%	0.213

Table 7. Estimates of mitigation potential.

4.1.4. Marginal Abatement Cost

Table 8 reports the estimates of the marginal abatement cost under the meta-frontier (MMAC) and the group-frontier (GMAC). From the table, it is found that MMAC and GMAC vary from 240 to 12,970 Yuan/ton and from 160 to 16,160 Yuan/ton, respectively. From the perspective of the meta-frontier, the low-carbon group is found to have the highest weighted average MMAC of 4970 Yuan/ton, followed by the medium group with a value of 1310 Yuan/ton, and the high-carbon group is at the bottom, with a value of 500 Yuan/ton. On the other hand, the weighted average GMAC values of the three groups also follow the same order, which are found to be 10,320, 2080, and 400 Yuan/ton. The differences between the mean values of MMAC and GMAC at group levels are considerable, especially for the low-carbon group, which is found to have a twofold change in the values. Additionally, the average weighted MAC values of the total sectors under the group-frontier and the meta-frontier are found as 670 and 770 Yuan/ton, respectively. As a matter of fact, the sectors have to sacrifice relatively more IVA for reducing an additional unit of CO_2 emissions, if they are producing at the more efficient overall production frontier with relatively less mitigation potential. So, in general, marginal abatement costs are negatively correlated with the mitigation potential.

Table 8. Estimates of marginal abatement cost (unit: 10,000 Yuan per ton).

	Groups	Ν	Min	Max	Weighted Mean ¹	Std. Dev.
	High-carbon	13	0.024	0.246	0.050	0.054
	Medium-carbon	13	0.083	0.506	0.131	0.110
Meta-frontier	Low-carbon	13	0.242	1.297	0.497	0.329
	Total	39	0.024	1.297	0.067	0.343
	High-carbon	13	0.016	0.147	0.040	0.033
Crown frontion	Medium-carbon	13	0.079	0.258	0.208	0.074
Gloup-Holluel	Low-carbon	13	0.361	1.616	1.032	0.447
	Total	39	0.016	1.616	0.077	0.485

¹ Weight is the share of each sector's emissions in the all.

It is interesting to compare our MAC estimates with those of previous studies. A summary of MAC estimates is reported in Table 9. From the table, it can be observed that the estimates of the previous studies lie in different ranges. The different samples and approaches in these studies are the main reasons for the differences in the results. The results of this study are found to be apparently lower than those of Xiao et al. [7] in terms of the mean and weighted mean of MAC. This is reasonable because our study only covers the emission-intensive industries of the industrial sector, while Xiao et al. [7] consider all industrial sectors as research objects, which is more expensive to reduce emissions. Additionally, the significant disparity of the average MAC between Yuan et al. [40] and Xie et al. [11] can also be attributed to the difference in research objects.

Studies	Methodology	Sample	Mean	Weighted Mean
Lee and Zhang (2012) [27]	IDF/P/T	30 Manufacturing industries, 2009	19.7	/
Yuan et al. (2012) [40]	DDF/N/DEA	24 Industrial sectors, 2004 and 2008	16,360	/
Zhou et al. (2015) [10]	MDF/N/T/Q	Shanghai's industrial sectors	/	394.5–1906.1
Xie et al. (2016) [11]	DDF/N/Three-stage DEA	9 key industries, 2005–2014.	1345	/
Xiao et al. (2017) [7]	DDF/P/Q	39 industrial sectors, 2005–2011	13,131	3517
This paper	Meta-DDF/N/DEA	39 participating sectors of CN-ETS	3180 and 4240	670 and 770

Table 9. Summary of MAC estimates (unit: Yuan per ton).

Note: P = Parametric; N = Nonparametric; T = Translog functional form; Q = Quadratic functional form; IDF = Input distance function; DDF = Directional distance function; MDF = Multiple distance function; CN-ETS = China's nationwide emission trading scheme.

4.2. Sectoral Analysis

4.2.1. Distribution Analysis

Figure 1 indicates the estimates of CEP, MP, and MAC under the meta-frontier and the group-frontier at the sector level. From the figure, we can observe that there exist significant disparities in the CEP, MP, and MAC under both frontiers among the 39 sectors. Specifically, from the meta-frontier perspective, the sectoral distribution of MCEP presents an increasing trend from S1 to S39 (i.e., the decreasing order of carbon intensity), while that of MMP presents the opposite tendency. Similarly, the sectoral distributions of GCEP and GMP show the upward and downward trends along with the decrease of carbon intensity in their respective groups, respectively. For instance, in the high-carbon group, the sectoral distribution of GCEP presents an increasing tendency from S1 to S13, while a decreasing trend is observed for GMP. On the other hand, it is found that both of the sectoral distributions of MMAC and GMAC present an increasing tendency from S1 to S39.

In light of this interesting phenomenon, we conducted a correlation analysis between MMAC, GMAC, and carbon intensity, respectively, and the results are listed in Table 10. According to the correlation results, there exists a significantly negative correlation (p = 0.000 < 0.01, Pearson's r = -0.537) between carbon intensity and MMAC, which has been previously reported in the literature, such as in Zhou et al. [10]. Additionally, GMAC is also found to be significantly negatively correlated with carbon intensity (p = 0.000 < 0.01, Pearson's r = -0.566).

Estimates	Items	Carbon Intensity
	Pearson correlation	-0.537 **
MMAC	Sig. (2-tailed)	0.000
	N	39
	Pearson correlation	-0.566 **
GMAC	Sig. (2-tailed)	0.000
	N	39
	Pearson correlation	-0.320 *
D_MAC	Sig. (2-tailed)	0.047
	N	39

Table 10. Correlation analysis.

** Correlation is significant at the 0.01 level (2-tailed); * Correlation is significant at the 0.05 level (2-tailed).



Figure 1. Estimates of CEP, MP, and MAC by sectors.

In order to provide an intuitive understanding of the distribution patterns of CEP, MP, and MAC, and make a comparison between the meta-frontier and the group-frontier, the kernel density curves of CEP, MP, and MAC under both frontiers are plotted based on the R-program. From Figure 2, it can be found that there exist some differences in the distribution patterns of CEP, MP, and MAC between the two frontiers. Specifically, the kernel curve of CEP moves rightward and the dispersion range of points become narrower when the evaluation basis switches from the meta-frontier to the group-frontier, which means an increase in the mean value and a decrease in the variance of CEP,

respectively. In the meanwhile, the kernel curve of MP moves leftward and the dispersion range of points become narrower, which means that the mean value and the variance of MP both decrease. On the contrary, the kernel curve of MAC holds the position unchanged, but exhibits a significant decrease in the peak value and an expansion in the dispersion range of points when the evaluation basis changes, which jointly means a considerable increase in the variance of MAC.



Figure 2. Kernel density of CEP, MP, and MAC.

4.2.2. Difference Analysis

Figure 3 reports the differences in CEP, MP, and MAC between the meta-frontier and the group-frontier at sector levels. From the figure, we can observe that there exist significant disparities between the two frontiers in terms of CEP, MP, and MAC, and the differences exhibit unequal distributions among the 39 sectors. The difference in CEP (D_CEP) is found to be significant in the high-carbon and medium-carbon sectors. For instance, the D_CEP of Pressing of Steel (S2) is found to be 0.691. Similar results can be observed for sectors such as the Manufacture of Fireproof Materials (S13), Manufacture of Special Chemical Products (S19), and Manufacture of Explosives and Fireworks (S25). This means that there is significant technology heterogeneity between the meta-frontier and group-frontier for the two groups. Conversely, it is found that there is no difference in the low-carbon sectors, which may be attributed to the fact that low-carbon sectors play an important role in benchmarking efficiency levels under meta-frontier technologies, and thereby there are no technology gaps for the low-carbon sectors. Similarly, the difference in MP (D_MP) between the two frontiers is also found to be significant in the high-carbon and medium-carbon sectors, while the opposite is true for the low-carbon sectors.

On the other hand, the difference in MAC (D_MAC) exhibits a completely different situation, in which the D_MAC is found to be larger, equal to, and smaller than 0, depending on the observations. For instance, the Pressing of Non-ferrous Metals (S32) and Casting Pressing of Non-ferrous Metals (S37) are found to have the lowest and the highest D_MAC, with values of -5610 Yuan/ton and 8680 Yuan/ton, respectively. There is no definite relationship between the MMAC and GMAC of sectors; the MMAC of sectors can be larger, equal to, or smaller than the GMAC, depending on the relative slopes of the meta- and the group-frontiers, respectively [19]. However, as mentioned earlier, both MMAC and GMAC are significantly correlated with carbon intensity. Therefore, it is interesting to explore the relationship between D_MAC and carbon intensity. In this regard, the correlation analysis between D_MAC and carbon intensity was conducted, and the results are listed in Table 10. According to the correlation results, there is also a significantly negative correlation (p = 0.047 < 0.05, Pearson's r = -0.320) between D_MAC and carbon intensity.



Figure 3. Differences in CEP, MP, and MAC by sectors. (Note: D_CEP, D_MP, and D_MAC denote the differences between the two frontiers in terms of CEP, MP, and MAC, respectively; D_CEP = GCEP-MCEP, D_MP = GMP-MMP, D_MAC = GMAC-MMAC).

4.3. Discussions

In general, whether from the perspective of group levels or from the perspective of sector levels, it is found that there exist significant disparities in the CEP, MP, and MAC under both frontiers among various sectors. Additionally, the differences between the two frontiers in terms of CEP, MP, and MAC are considerable, and exhibit unequal distributions among the 39 sectors. This can be attributed to the significant heterogeneity of production technology among various sectors. Considering that carbon intensity is widely considered as the measurement of carbon emission efficiency, the industries with a relatively low carbon intensity are more efficient than those with a high carbon intensity in terms of energy utilization and carbon emissions. Thus, low-carbon sectors are found to have a higher CEP, less MP, and larger MAC than medium-carbon and high-carbon sectors.

As the most efficient DMUs, low-carbon sectors are found to play a more important role in benchmarking efficiency levels under the meta-frontier. In this case, no difference between the meta-frontier and the group-frontier for the low-carbon group in terms of CEP and MP exists, while the opposite is true for the medium-carbon and high-carbon sectors. As a result, considerable differences in the CEP and MP between the two frontiers are observed for the medium-carbon and high-carbon sectors. As for MAC, theoretically, the MMAC can be larger, equal to, or smaller than the GMAC, depending on the relative slopes of the meta-frontier and the group-frontier. Nevertheless, it is found that D_MAC has a significantly negative correlation with carbon intensity.

5. Conclusions and Policy Implications

CN-ETS that covers seven emission-intensive industries is scheduled to be launched in 2017. In this context, it is of great urgency and necessity to obtain a good understanding of participating sectors in terms of energy utilization and carbon emissions. In this regard, estimating the CEP, MP, and MAC for these sectors can provide valuable information for the governments and participating enterprises. Therefore, taking the industry heterogeneity into consideration, we employed a joint framework consisting of the DDF and meta-frontier analysis to estimate CEP, MP, and MAC under the meta-frontier and the group-frontier, respectively. Following this, we investigated the sectoral distributions of CEP, MP, and MAC under both frontiers, and analyzed the differences between the two frontiers in terms of CEP, MP, and MAC at sector levels.

Based on the detailed analysis, the main conclusions are drawn as follows: First, there exist significant disparities in the CEP, MP, and MAC under both frontiers among various sectors. Specifically, high-carbon and medium-carbon sectors are found to display a low CEP, large MP, and high MAC, while the opposite situation is observed for low-carbon sectors which have a high CEP, small MP, and high MAC. Furthermore, the sectoral distributions of CEP, MP, and MAC are found to be different between the two frontiers. Additionally, the differences between the two frontiers in terms of CEP, MP, and MAC are considerable, and exhibit unequal distributions among the 39 sectors. Notably, the MAC values under both frontiers and the difference between them are all found to be significantly correlated with carbon intensity.

Based on the above conclusions, possible policy implications are provided for the government and participating enterprises, respectively. First of all, based on the estimates of CEP, MP, and MAC for sectors, our study allows policy makers to pinpoint where the greatest emissions cuts—at the least expense—can be made in China's emission-intensive sectors. From the perspective of the government, different policies should be implemented for the critical emission reduction sectors based on their CEP, MP, and MAC. In this regard, high-carbon sectors with a low CEP, large MP, and low MAC should shoulder more responsibility for the reduction of emissions. In particular, necessary policy measures such as introducing carbon-emission-reduction technologies, eliminating backward production facilities, raising the industry entry threshold, and increasing the industry concentration should be implemented for high- and medium-carbon sectors. Furthermore, the CEP, MP, and MAC (especially GCEP, GMP, and GMAC) of participating sectors should be taken into consideration when formulating the criteria for the initial allocation of carbon allowances before transactions. Additionally, the weighted average MAC could offer a reference for the carbon price of CN-ETS, since the industries covered in CN-ETS are taken as the research objects. On the other hand, from the perspective of enterprises, by comparing the MAC of participating sectors with the carbon price of CN-ETS, participating enterprises could identify the least-costly emission reduction strategy from a list of policy alternatives such as abating carbon emissions, buying carbon allowances, and selling carbon allowances, while low-carbon enterprises may choose to buy carbon allowances.

Despite the contributions, this paper has a limitation in assuming the same growth rate for various sectors in light of their heterogeneity. Moreover, our study cannot consider dynamic effects since we are limited to using a one-year data cross-section. Additionally, due to the limitation of statistical data, this study selects 39 sectors as research objects rather than the specific enterprises covered in CN-ETS. However, this study could be easily extended to enterprises, so future research requires the collection of data at enterprise levels in order to obtain more accurate estimates.

Acknowledgments: The authors would like to thank the financial support provided by the National Natural Science Foundation of China (Grant No. 41471457), and the Fundamental Research Funds for the Central Universities (Grant No. 2016B09414).

Author Contributions: Zhencheng Xing designed the research and drafted the manuscript; Jigan Wang collected the data and conducted the model simulation; Jie Zhang reviewed, commented on, and revised the manuscript. All authors have read and approved the final manuscript.

Conflicts of Interest: The authors declare no conflict of interest.

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